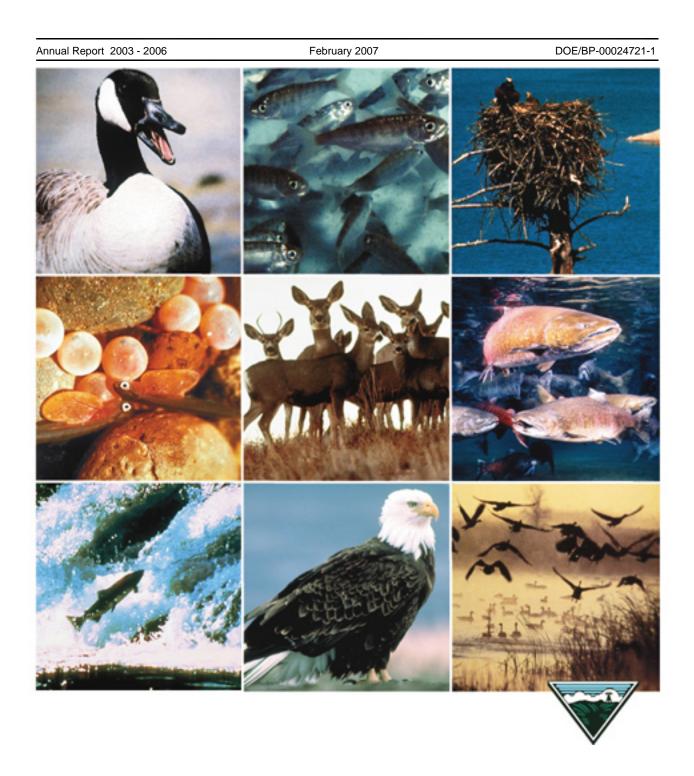
Evaluation of Juvenile Salmonid Outmigration and Survival in the Lower Umatilla River Basin



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EVALUATION OF JUVENILE SALMONID OUTMIGRATION AND SURVIVAL IN THE LOWER UMATILLA RIVER BASIN

ANNUAL REPORT 2004-2006

(1 OCTOBER 2003 - 30 SEPTEMBER 2006)

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EXECUTIVE SUMMARY

This report summarizes activities conducted by the Oregon Department of Fish and Wildlife's Juvenile Outmigration and Survival M&E project in the Umatilla River subbasin between 2004-2006. Information is used to make informed decisions on hatchery effectiveness, natural production success, passage improvement and flow enhancement strategies. Data collected includes annual estimates of smolt abundance, migration timing, and survival, life history characteristics and productivity status and trends for spring and fall Chinook salmon, coho salmon and summer steelhead. Productivity data provided is the key subbasin scale measure of the effectiveness of salmon and steelhead restoration actions in the Umatilla River. Information is also used for regional planning and recovery efforts of Mid-Columbia River (MCR) ESA-listed summer steelhead. Monitoring is conducted via smolt trapping and PIT-tag interrogation at Three Mile Falls Dam.

Project Objectives

The Umatilla Juvenile Outmigration and Survival Project was established in 1994 to evaluate the success of management actions and fisheries restoration efforts in the Umatilla River Basin. Project objectives for the 2004-2006 period were to: (1) operate the PIT tag detection system at Three Mile Falls Dam (TMFD), (2) enhance provisional PIT-tag interrogation equipment at the east bank adult fish ladder, (3) monitor the migration timing, abundance and survival of naturally-produced juvenile salmonids and trends in natural production, (4) determine migration parameters and survival of hatchery-produced fish representing various rearing, acclimation and release strategies, (5) evaluate the relative survival between transported and non-transported fish, (6) monitor juvenile life history characteristics and evaluate trends over time, (7) investigate the effects of river, canal, fishway operations and environmental conditions on smolt migration and survival, (8) document the temporal distribution and diversity of resident fish species, and (9) participate in planning and coordination activities within the basin and dissemination of results.

Significant Findings and Accomplishments

Outmigration Monitoring: Results indicate migration patterns for hatchery salmonids are similar to those of natural salmonids, although the magnitude of smolt emigration varied significantly. Median emigration of natural summer steelhead is roughly one week earlier then their hatchery counterparts. Median emigration of natural fall Chinook salmon is approximately one week later than hatchery conspecifics. Correlations between migration timing and environmental factors are species dependent and are variable from year to year.

Size and age at emigration for Umatilla salmon and summer steelhead is similar to surrounding basins. Spring Chinook salmon emigrate as yearlings, fall Chinook as subyearlings and summer steelhead at a broad distribution of ages.

PIT tag data suggests, passage route selection at TMFD is influenced by canal diversion. When West Extension Canal is operating, the majority of salmonids emigrate past TMFD using the juvenile bypass. When the canal is not operating, juvenile fish passage is more prevalent

through the east bank adult fish ladder. During high flow events most fish appear to pass over the crest of the dam.

Smolt Abundance: Abundance of summer steelhead has fluctuated since 1995, however trends reflect little or no change in productivity. The average annual smolt production of Umatilla summer steelhead is 50,217 (CI +/- 6,116). Smolt production of spring Chinook salmon has ranged between 8,458 and 14,159 for outmigration years 2004-2006. Production of fall Chinook salmon has fluctuated dramatically over the same period (12,781 to 107,858).

Juvenile Survival: Overall performance of hatchery smolts has been variable and less then satisfactory for most species. In-basin survival has averaged 31% for hatchery-reared coho salmon, 38% for summer steelhead, 59% for subyearling fall Chinook salmon, 62% for spring Chinook salmon, and 70% for yearling fall Chinook salmon. Enhanced detection capabilities at TMFD and application of the SURPH model in 2005 has provided more robust estimates of survival.

Survival from TMFD to John Day Dam (JDD) of naturally-reared spring Chinook salmon, fall Chinook salmon and summer steelhead, has averaged 65%, 16% and 59%, respectively, between 1999-2006. Egg-to-smolt survival (ESS) rates for naturally-produced summer steelhead averaged 0.77% for brood years (BY) 1993-2004. Umatilla summer steelhead continue to have similar smolt-to-adult return (SAR) rates compared to steelhead in surrounding basins. The mean SAR rate for naturally-produced Umatilla summer steelhead between outmigration years 1995-2002, was 3.53%.

Smolt-per-spawner (SPS) estimates for Umatilla salmon and steelhead are considerably lower than that of surrounding basins. Mean SPS for BY 1993-2004 were 27.1 for natural spring Chinook salmon, 75.8 for natural fall Chinook salmon and 26.6 for natural summer steelhead. The average number of summer steelhead smolts produced per female (SPF) between BY's 1993-2004 was 40.0.

PIT tag interrogation: PIT-tag interrogation has been successful and incrementally improved at TMFD since 1999. Over the past 3 years, equipment has been upgraded to improve system performance, reliability and remote system monitoring. Three detection arrays are currently in place at TMFD; two within the juvenile bypass and one in the adult fish ladder. Detection efficiency at the 3 sites ranges from 47.9% to 99.9%. Tag information from all 3 sites is coordinated to remotely upload at one central location.

Engineered design plans and specifications for a permanent adult ladder detection system were completed in 2005.

INTRODUCTION

The Umatilla River historically supported large runs of salmon (*Oncorhynchus spp.*), steelhead (*O. mykiss*), and Pacific lamprey (*Lampetra tridentata*) for productive Tribal and recreational fisheries. By the early 1920s, these runs were decimated. The extirpation and degradation of salmon and steelhead populations in the Umatilla River was a result of extensive agricultural development and associated water withdrawals, habitat destruction, water quality degradation, passage problems, over-harvest and habitat loss both inside and outside of the basin (Saul et al. 2001).

In 1980, the Northwest Power Act served as a springboard for focusing attention and effort on restoring salmon and steelhead populations throughout the Columbia River basin. A multitude of projects were initiated, to help address rehabilitation efforts and achieve fisheries restoration program goals (NPPC 1984, 1987, 1994 and 2000). Rehabilitation of Umatilla fish populations called for restoration of spring and fall races of Chinook salmon (*O. tshawytscha*), coho salmon (*O. kisutch*), and supplementation of summer steelhead (CTUIR and ODFW 1989). Proposed measures included passage improvements at irrigation diversion dams, habitat restoration, hatchery production, construction of acclimation facilities, flow enhancement, and fish transport during low flows (CTUIR and ODFW 1984; ODFW 1986, CTUIR and ODFW 1989).

Monitoring and research efforts to evaluate Umatilla restoration and enhancement projects were implemented in the late 80's and early 90's. The Umatilla Passage Evaluation study was established in 1989 to evaluate passage of juvenile and adult salmonids following construction of canal screening and bypass facilities (Cameron et al. 1997). The Umatilla Hatchery Monitoring and Evaluation (M&E) project was implemented in 1990, following construction of the Umatilla Hatchery (Grant et al. 2007). In 1991, the Natural Production M&E Project was initiated to monitor salmon reintroduction efforts and the effects of supplementation on natural summer steelhead populations. These efforts, however, did not include a comprehensive evaluation of the migration success, abundance and survival of hatchery-reared and naturally-produced juvenile salmonids in the Umatilla River.

Long-term trend monitoring of juvenile abundance and survival was considered necessary as habitat was improved, flow enhanced, natural production expanded, and hatchery practices adjusted. Monitoring and sampling at the basin level were considered crucial for gathering necessary information on life history characteristics, population abundance, and life-stage specific survival of natural salmonids. Specific questions still remained regarding in-basin survival and passage problems for juvenile fish, production potential for natural populations, and aquatic community health. Furthermore, as production strategies evolved, results needed to be monitored and merits evaluated to help guide management decisions.

In 1994, the Umatilla Juvenile Outmigration and Survival M&E project was established in attempt to address some of the above noted concerns. Outmigration monitoring was originally conducted via branding and color-marked fish (Knapp et al. 1996, 1998a, 1998b). The advent of PIT tag technology at the John Day Dam in 1998, however prompted the use of PIT tags (400 kHz) in the Umatilla Basin the same year. In 2000, detection equipment was upgraded to a 134 kHz system, to further enhance monitoring capabilities. Additional upgrades have recently been

implemented to improve performance, reliability and remote system monitoring. Eight years of PIT-tag interrogation at TMFD has provided increasingly accurate estimates of juvenile abundance, migration timing and survival, while minimizing the number of fish handled.

The Umatilla Outmigration and Survival Project is one of three intimately-linked Research, Monitoring and Evaluation (RM&E) projects designed to evaluate the success of adaptive management strategies and restoration efforts in the Umatilla River. Data collected is used to resolve critical uncertainties identified in the Umatilla Subbasin Plan (Debano et al. 2004) and Umatilla RM&E Plan (CTUIR and ODFW 2006). Outmigration and Survival activities complement and support those carried out by the Umatilla Hatchery M&E project (BPA Project 199000500) and the Umatilla Natural Production M&E project (BPA Project 199000501).

This report summarizes work performed by the ODFW's Umatilla Juvenile Outmigration and Survival M&E project between 2004 and 2006. Information gathered has helped clarify the status and trends of juvenile salmonids in the Umatilla River, provided insight into the effects of river, canal flow, water temperature, and other environmental variables on fish passage and survival, helped evaluate juvenile life history characteristics and trends in natural production and aided in monitoring the movement of juvenile Pacific lamprey and resident fish.

STUDY AREA

The Umatilla River basin is located within Umatilla and Morrow counties of northeast Oregon (Figure 1). Draining an area of 2,290 square miles, it flows in a northwesterly direction into RM 289 of the Columbia River. The uppermost reaches of the basin are situated along the steep timbered slopes of the Blue Mountains of the Umatilla National Forest (Saul et al. 2001). The remainder of the drainage lies within the broad upland plain of the Deschutes-Umatilla Plateau (Contor and Kissner 2000).

The Umatilla River contains a mainstem length of 115 miles (Contor et al 2000) and is fed by eight major tributaries. Elevation ranges from nearly 5,800 feet at the headwaters, to 260 feet at its confluence with the Columbia River (Saul et al. 2001). Identified by hydrologic unit number 17070103 (US Geological Survey 1989), it receives a mean annual precipitation of 10 to 50 in./yr within the lower and upper basin, respectively (Contor et al 2000; Saul et al. 2001). The Umatilla River basin lies within Oregon's North Central bio-geoclimatic zone (Zone 6).

The upper portion of the basin encompasses a section of the Umatilla National Forest as well as 172,000 acres of tribal land (Confederated Tribes of the Umatilla Indian Reservation, CTUIR). The majority of the land in the Umatilla basin is privately owned (82%), with the remainder being divided amongst the State of Oregon, Umatilla County, various cities and CTUIR (Saul et al. 2001).

Project activities are concentrated in the lower Umatilla River mainstem, between RM 1.2 and RM 3.7 (Figure 1). The juvenile screening facility (incline plane trap and PIT tag detection system) is located within West Extension Canal off TMFD (RM 3.7). The secondary trapping location (rotary-screw trap) is situated in a deep pool beneath the Interstate 82 Bridge (RM 1.2). Sampling is conducted year round, with operations focusing at RM 3.7 between February and July and at RM 1.2 from October to January. Sampling at RM 1.2 was discontinued in 2005.

Release sites for trap and haul operations and trap efficiency tests are located at various points along the Umatilla mainstem. More specifically, the sites are situated between the mouth and RM 27.2 (Westland Diversion Dam).

The average monthly discharge within the lower river varies from a low of 23 cfs in the summer (July) up to 1,095 cfs during spring runoff (April). Water temperatures have been known to peak at lethal levels of between 18°C and 27°C (Saul et al. 2001).

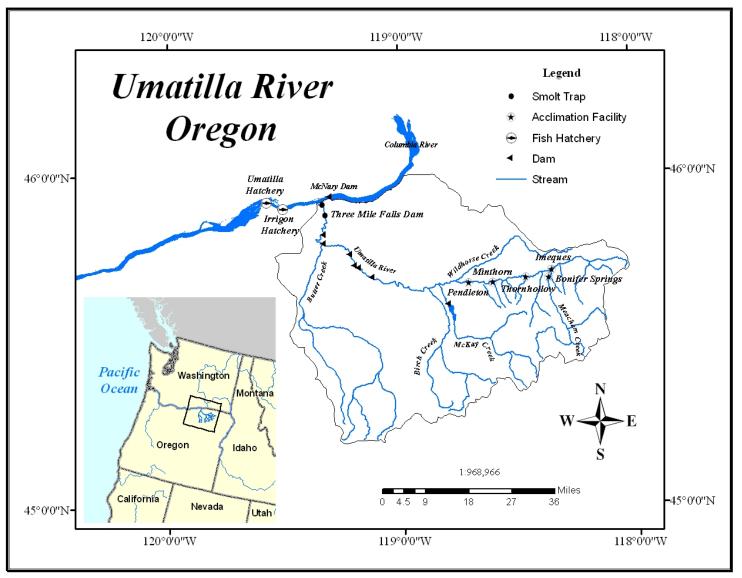


Figure 1. Map of general study area of Pacific Northwest, U.S.A. and detailed map of the trapping and tagging locations from the Umatilla River.

METHODOLOGY

Outmigration Monitoring

Fish Trapping and PIT Tag Operations

Outmigration monitoring was conducted via smolt trapping and PIT tag interrogation at TMFD (RM 3.7). An inclined plane trap set in the juvenile bypass facility at West Extension Canal was used to capture emigrating salmonids. Trapping was conducted between February and June to coincide with the primary smolt emigration period. This was also the primary period when water was diverted into West Extension Canal. In 2004, a rotary screw trap was also operated from October to January in attempt to capture early/late natural migrants. The rotary screw trap was located beneath the I-82 bridge on the lower Umatilla River (RM 1.2).

PIT-tagged fish were interrogated at TMFD via one of three antenna arrays; in the juvenile bypass trap, at the juvenile bypass outfall or the east bank adult fish ladder. Detection data was stored on a PC and automatically uploaded to the PIT Tag Information System (PTAGIS) database on a daily basis via the Minimon program and modem. A full description of PIT-tag interrogation, marking, and file management operations can be found in White et al. 2004.

Approximately 1,200 natural salmonids were PIT-tagged annually to monitor trends in natural production. As many as 14,679 hatchery-reared fish were also tagged for juvenile outmigration and survival monitoring.

Fish Condition and Life History Characteristics

Regardless of trap type or location, all salmonids captured were anesthetized with a stock solution of MS-222 (40 mg/l) prior to sampling. Fish were enumerated by species, race, and origin. Origin was categorized as "natural" or "hatchery" based on the presence/absence of a fin clip, wire tag, or the appearance of wear on the dorsal or ventral fins. Race of natural Chinook salmon was categorized as spring or fall using body morphology, length and age characteristics.

All natural smolts and a subsample of hatchery smolts were examined to assess age, size, condition, health and smolt status at emigration. Size at emigration was quantified from fish captured in traps for each species of salmonid. Fork length was recorded to the nearest mm and a t-test for two independent samples performed to determine if there was a significant difference among length distributions of hatchery and natural fish. Scale samples on a subsample of natural summer steelhead were collected for age analysis and applied to smolt abundance estimates to determine age class contributions to the emigrating population. Natural Chinook salmon scales were collected to assist in the identification of spring and fall Chinook salmon. Scales were mounted on mylar strips and examined under a microfiche at 24X or greater magnification to discern annuli patterns reflecting freshwater age. Developmental (smoltification) stage was ranked as parr, intermediate, or smolt based on brightness and the absence or presence of parr marks.

Condition was characterized as the proportion of cumulative scale loss evident on the fish at the time of emigration Fish condition was partitioned into one of three categories: good (< 3%), partially descaled (3-20%), or descaled (> 20%) following criteria used by the Umatilla Hatchery Monitoring and Evaluation project (Keefe et al. 1994). Juvenile fish health was monitored by daily examination of randomly selected emigrants per species/origin type for body injuries, external parasites, bird marks, obvious fungal infections of the body surface, and signs of potential disease. Fish mortalities were noted by species and identified as being from an unknown source or a direct result of sampling activities. All natural fish that died from an unknown cause and some diseased and dead hatchery fish were forwarded to the ODFW Fish Pathology Lab. Analysis of samples followed standard protocols defined in the latest edition of the American Fisheries Society "Fish Health Blue Book" (procedures for the detection and Identification of Certain Fish Pathogens).

Migration Parameters

Migration parameters were analyzed for hatchery juvenile salmonids using PIT tag detections at West Extension Canal. Parameters included emigration timing, pattern, duration, and travel speed for each species. Peak movement was determined by selecting the date when the maximum number of tagged emigrants were detected. Median emigration was the date when 50% of the tag detections were observed and diel movement was determined by the percentage of fish detected within hourly blocks of time. Travel speed was calculated for tagged fish detected at West Extension Canal using the following equation:

$$TS = (RM-3.7)/(D-R)$$

where TS = travel speed, RM = river mile of release, D = date and time of detection at West Extension Canal, and R = date and time of forced release. Median travel speed was calculated for each hatchery release group. Median rather than mean travel speeds were computed because detection distributions tended to be skewed. Negative travel speed estimates due to early movement from acclimation ponds during volitional release periods were omitted from the analysis. Tagged fish collected during fish sampling operations were also removed due to the inability to assign an accurate time of recapture.

The number of fish captured at the rotary-screw trap and West Extension Canal was expanded to compare the timing of natural and hatchery smolts emigrating from the Umatilla River using the following equation:

$$C_{w} = \sum_{\underline{(C_{\underline{d}}/R_{\underline{d}})}} \underline{(C_{\underline{d}}/R_{\underline{d}})}$$

where, C_w = estimated weekly number of fish captured at the rotary-screw trap or moving through the juvenile bypass, C_d = daily number fish captured, R_d = daily sample rate, P_w = ratio of time sampled to unsampled within the week. Weekly estimates were summed to derive and then plotted as a proportion of the season total. Cumulative frequencies were calculated and the week of the 10^{th} , 50^{th} , and 90^{th} percentiles were determined. The Kolmogorov-Smirnov test was

performed to assess statistical differences in the emigration timing of hatchery and natural summer steelhead.

Paired release tests were completed at the juvenile bypass facility in 2004 and 2005 to evaluate temporal and spatial variability in recapture rates and travel speeds of fish previously exposed to the trap (run-of-river) compared to fish with no prior trap exposure (naïve). In 2004, three paired releases of approximately 60 PIT tagged fish each were made at two separate locations upstream of the juvenile bypass facility; the waste water treatment plant (1.5 miles upstream of the trap) and the canal headgates (230 feet upstream of the trap). Run-of-river fish were collected at West Extension Canal, while naïve fish were collected from upriver acclimation sites. In 2005, similar tests were conducted however comparisons between run-of-river and naïve fishes were not made. A two-sample Proportion Test was used to test for differences ($\alpha = 0.05$) in recapture rates within and among releases. The travel times of individual fish were loge transformed to meet the assumption of normality and ANOVA was used to test for differences within and among releases. When significant differences ($\alpha = 0.05$) were found a pairwise comparison using the Fisher's test for least significance difference was conducted. Associations between specific abiotic variables and recapture rates were assessed using regression analysis.

Abundance and Survival

Trap Efficiencies

To calibrate the collection efficiency of the traps and estimate outmigrant abundance, groups of 25 to 60 fish per species were collected, PIT-tagged, and released 1.3 miles upstream of the trap for recapture. Fish were PIT-tagged according to standards outlined in the PIT Tag Marking Procedures Manual (CBFWA, PIT Tag Steering Committee, 1999). Tests were conducted 2 times a week for each species while sufficient numbers of fish were being captured. Tagged fish were held for 24 hours prior to release to assess tagging mortality, tag loss, and to determine the probability of survival of individual release groups. The probability of survival and estimated survival of tagged fish released was calculated as:

$$s = L/H$$
,
and
 $M = N(s)$

where s = probability of survival, L = number of live tagged fish after holding, H = initial number of tagged fish, M = estimated number of tagged fish released, and N = total number of tagged fish released. Tagged fish that died or dropped their tags prior to release were removed from the release group. Tag retention and fish survival for all factors other than tagging were assumed to be 100% after release. It was also assumed that all marked and unmarked smolts migrated downstream independently of one another and had equal catchability. Specific details regarding tagging, holding and fish transport operations can be found in White et al. 2004.

Recaptured fish were enumerated by species/origin and trap efficiency estimates computed using the following formula:

$$TE = R/M$$

where, TE = estimated trap efficiency, R = number of recaptured fish, and M = estimated number of tagged fish released. Separate trap efficiency estimates within a species were compared using chi-square analysis. If estimates were not significantly different ($\alpha = 0.05$) they were pooled. Pooling was continued until a significant difference was determined or no trend of change was evident

Smolt Abundance

Smolt emigrant abundance was defined as the number of smolts leaving the Umatilla River or reaching TMFD. Abundance was calculated for natural emigrants only and is the key component required to address critical uncertainties surrounding in-basin productivity and natural production capacity. Smolt abundance was derived based on the number of fish collected at lower river traps and the estimated trap efficiency. Abundance of fish passing TMFD was estimated as:

$$A = B/TE$$
and
$$B = (C/T)$$

where, A = estimated number of outmigrants, B = estimated number of fish that passed through the trap (total passage), TE = estimated trap efficiency, C = sample rate, and T = proportion of time sampled.

Emigrant abundance was estimated on a monthly basis and then summed to derive the total number of outmigrants for the season. For months where trap efficiencies of natural species were not available or were sparse, trap efficiency estimates from hatchery conspecifics were used. If hatchery conspecifics were not available for a particular month, efficiency estimates from the month before or month after were used. The Bootstrap method (Efron and Tibshirani 1986; Thedinga et al. 1994) with 1,000 iterations was used to derive a variance and 95% confidence intervals for abundance estimates.

The downstream movement of Chinook salmon fry (< 45 mm) and subyearling coho salmon (45 - 105 mm) was assumed to be a natural dispersal mechanism to help distribute fish among the suitable rearing habitat and thus were not classified as emigrating smolts. Freshwater age-0 summer steelhead (<75mm) were assumed not to be outmigrants based on the fact that a subyearling life history pattern has never been detected on scale samples collected from adult steelhead escaping to the Umatilla River.

Smolt-Yield-Per-Spawner

Smolt-yield-per-spawner was estimated to measure the productivity of naturally spawning hatchery and natural fish in the Umatilla River basin. Estimates were calculated using the following equation:

$$SYS = A/NS$$

where, SYS = smolt-yield-per-spawner, A = estimated smolt abundance at TMFD, and NS = number of potential spawners.

The number of spawners for spring Chinook salmon was estimated by multiplying the total number of redds counted during spawning ground surveys by two. Fall Chinook salmon spawner estimates were computed using the female spawning escapement at TMFD plus the number of female outplants times two. Escapement data from TMFD was used to determine the number of spawners for natural summer steelhead.

Smolts -Per-Female

Smolts-per-female were estimated for natural summer steelhead using the female spawning escapement and estimated smolt abundance at TMFD. Smolt yield was proportionately partitioned by brood year based on freshwater age composition. Age structure was determined from scale analysis of 100 natural adult returns per year to TMFD (1991-2000). Missing scale data for BY's 2001-2004 was reconstructed using the mean freshwater age composition from BY 1991-2000.

Smolt Survival

Survival was estimated for hatchery and natural salmonids to assess in and out-of-basin loss by species and life-stage. Survival estimates were also generated to support hatchery production monitoring and evaluation of optimal rearing and release strategies. Mark-recapture methodology utilizing PIT tags and subsequent detections at TMFD and Columbia River dams was used to calculate survival. Hatchery fish were PIT-tagged at the hatchery or acclimation facilities prior to release. Natural fish were captured and tagged at TMFD during smolt trapping operations.

Survival to TMFD and JDD was estimated using the SURPH 2 model (v 2.1). SampleSize 1.3 was used to determine tag sizes needed to estimate survival rates at desired levels of precision (Tables 1 and 2; Lady et al. 2001). A 20% CV provided the most realistic sample sizes under current operations and funding conditions.

Table 1. Mean survival and capture probabilities used to estimate tag sizes.

	Survival probability		Capture p	robability	
	Release	TMFD to	Release	TMFD to	Final period
Species	to TMFD	JDD	to TMFD	JDD	survival/capture
Hatchery spring Chinook salmon	0.62	0.56	0.12	0.35	0.18
Hatchery yearling fall Chinook salmon	0.69	0.67	0.24	0.30	0.15
Hatchery subyearling fall Chinook salmon	0.61	0.53	0.46	0.32	0.11
Hatchery summer steelhead	0.36	0.55	0.30	0.27	0.21
Natural spring Chinook salmon	0.34	0.68	0.21	0.39	0.18
Natural fall Chinook salmon		0.15		0.70	0.11
Natural summer steelhead	0.34	0.57	0.28	0.34	0.18

¹Derived from Umatilla fish PIT tagged and detected between 1999-2005; Survival and capture probabilities calculated through SURPH.

Table 2. Current and proposed tag sizes needed to obtain survival rates to TMFD and JDD with 5%, 10% or 20% coefficient of variation (CV) for hatchery-reared and naturally-produced juvenile salmonids.

			Proposed tag sizes				
	Current tag	Re	lease to TM	IFD	Т	MFD to JD	D
Species	sizes	5% CV	10% CV	20% CV	5% CV	10% CV	20% CV
Hatchery spring Chinook salmon	2,000	14,000	3,600	900	25,400	6,200	1,600
Hatchery yearling fall Chinook salmon	600	5,200	1,300	400	16,800	4,200	1,000
Hatchery subyearling fall Chinook salmon	1,200	3,200	800	200	24,500	6,200	1,500
Hatchery summer steelhead	900	9,500	2,400	600	30,700	7,500	1,800
Natural spring Chinook salmon		9,500	2,400	600	21,600	5,400	1,400
Natural fall Chinook salmon		6,700	1,700	400	30,200	6,600	1,700
Natural summer steelhead		9,700	2,300	600	27,400	6,700	1,700

Egg-to-Smolt Survival

Egg-to-smolt survival was estimated for natural summer steelhead as a measure of life-stage specific survival. Egg-to-smolt survival was computed as:

$$ESS = A/ED$$

and

$$ED = (3979*S1) + (6965*S2)$$

where, ESS = egg-to-smolt survival, A = estimated smolt abundance at TMFD, and ED = egg deposition of salt-1 [S1] and salt-2 [S2] females available to spawn. Smolt abundance was partitioned by brood year based on freshwater age composition. Age classification of female spawners was determined from brood fish collected between 1993-2004.; S1<600mm FL and S2 >600 mm FL (BY 1993-2004).

Smolt-to-Adult Return

Smolt-to-adult return for natural summer steelhead was estimated as:

$$SAR = AR/A$$

where SAR = smolt-adult return, AR = total number of adult returns to TMFD, A = estimated smolt abundance. Smolt abundance was derived from annual outmigrant estimates at TMFD and partitioned by age structure. The number of adult returns (salt-1, salt-2, and salt-3 fish) were aligned by smolt outmigration year using adult counts at TMFD and scale analysis.

Transport Evaluation

Transport evaluation tests were not conducted in 2004 due to logistical constraints and unfavorable environmental conditions. Transport evaluation tests were discontinued in 2005.

Transport evaluation tests were conducted to evaluate the relative survival of transported versus non-transported fish. In the past, transported (treatment) fish were hauled to the mouth of the Umatilla River and released. Non-transported (control) fish were released directly into the river beneath the Stanfield Bridge (RM 23).

Environmental Conditions and Fish Passage

Environmental variables including river flow, flow augmentation, water temperature, and water clarity were monitored and analyzed using associative and time-series analysis to characterize conditions in the Umatilla River and to assess their effects on emigration timing and fish passage. Daily river flow, discharge from McKay Reservoir, and water temperature data were obtained from the USBR Hydromet Archives: http://www.usbr.gov/pn/hydromet/. Weekly mean flow from the Umatilla gauging station (RM 2.1) and water temperature from the Yoakum gauging station (RM 37.6) was plotted against time. Missing flow or temperature records were estimated using associative models with upstream gauging stations or by using an average of the day before and after the missing measurement. Water clarity was measured daily to the nearest 0.1 m using a 7-in-diameter secchi disk at RM 1.2 or 3.7 and weekly mean secchi depth was plotted against time. Weekly mean discharge for flow augmentation from McKay Reservoir and water exchange as a proportion of total flow at the Umatilla gauge was plotted against time and overlaid with cumulative smolt emigration.

The relationship between specific abiotic variables and the daily proportion of emigrants passing a trap site was tested using the Spearman rank correlation test. Abiotic variables included river flow, water temperature, and water clarity. Abiotic variables were the average of the mean of the day before and the day of passage. The time period used for the analysis was between the days when the first and last emigrant was observed. Flow and temperature data from the Yoakum gauging station (RM 37.6) and water clarity measured at RM 1.2 or 3.7 were utilized for the analysis. Any missing flow or temperature records were estimated by taking the average of the mean daily flow or temperature three days prior and three days after the missing record.

Multiple regression techniques were used to analyze recapture rates of PIT tagged fish in an attempt to determine factors that influence route selection at Three Mile Falls Dam. Least squares multiple regression was used to test for association between specific abiotic variables and the efficiency of the juvenile bypass/trapping facility. Abiotic variables included water temperature, river flow, water clarity, diversion rate, volume of water diverted into the facility and facility entrance velocity. Arcsine transformed trap efficiency estimates (see **Trap Efficiencies**) were used as the dependent variable. Mean values of abiotic variables based on the day of and the day after the trap efficiency releases were calculated and used as the independent variables. Flow and temperature variables from the Umatilla gauging station were utilized for the analysis and canal diversion data was obtained from a gauging station located immediately downstream (2001 - 2006) of the canal check gates. Canal diversion rate was calculated by dividing the daily canal flow by the daily river flow. Daily river flow was calculated by adding the Umatilla gauge reading and the daily canal flow.

Resident Species and Lamprey

Resident fish were enumerated by species, and fork length was recorded to the nearest mm. Total length was recorded and developmental stage was classified for juvenile lamprey. Length data was used to create length-frequency distributions by life stage for juvenile lamprey and summary statistics; including sample size, mean fork length, and minimum and maximum fork lengths for resident fishes. Lamprey data was submitted to CTUIR's lamprey restoration project for additional analyses and dissemination.

RESULTS AND DISCUSSION

Outmigration Monitoring

Little change was observed in migration patterns of hatchery or natural salmonids compared with previous years. Migration patterns for hatchery salmonids are similar to those of natural salmonids, although the magnitude of smolt emigration can vary significantly (Figures 2-4). Median emigration for natural summer steelhead is roughly one week earlier than their hatchery counterparts. Median emigration of natural fall Chinook salmon is approximately one week later than hatchery conspecifics (Appendix Tables A1 to A3). Peaks in emigration for most hatchery migrants are observed shortly after release; however, coho salmon tend to have a protracted outmigration period.

Natural and hatchery summer steelhead exhibit similar trends in outmigration timing, but the cumulative distributions are statistically different (Figure 5; D = 0.2567, P = 0.0085). As would be expected, the temporal range of natural smolts is broader than that of hatchery; however, distributions are still different when adjusted for the volitional release date of hatchery smolts (D = 0.2212, P = 0.0087). This is mostly due to the inherent sensitivity of the Kolmogorov-Smirnov Test, differences in means and variances and partially due to the natural smolt emigration already being in progress. In addition, a low number of hatchery steelhead were typically observed leaving the acclimation facilities prior to forced release (Grant et al. 2007). It took about 6 weeks for 90% of the hatchery smolts to emigrate from the system compared to 17 weeks for natural smolts (5 and 15 weeks for 50%; respectively). The broad temporal range in the emigration timing of natural smolts may help buffer against the highly variable flow regime of the Umatilla River subbasin and provide a survival advantage when compared to hatchery smolts. In addition, patterns in the emigration timing of natural smolts may be linked to genetic variation in the out-migrating juveniles (Grant et al. 2007).

Annual correlations between migration timing and environmental factors have been species dependant and variable from year to year. We have observed associations between flow and temperature and migration timing of Chinook and coho salmon, but summer steelhead are not influenced by abiotic factors. The majority of anadromous fish movement is observed on the falling limb of the hydrograph (Figure 6). In addition, fish movement is affected by changes in daylight; diel patterns of movement are different between fish in-river and at canal facilities; and movement is influenced by canal operations (Figure 7).

Passage route selection at TMFD is influenced by canal diversion. Detection data suggests when West Extension Canal is diverting water, the majority of fish migrate past TMFD using the juvenile bypass. However, when the canal is not operating, juvenile fish passage is more prevalent through the east bank adult fish ladder (Figure 7).

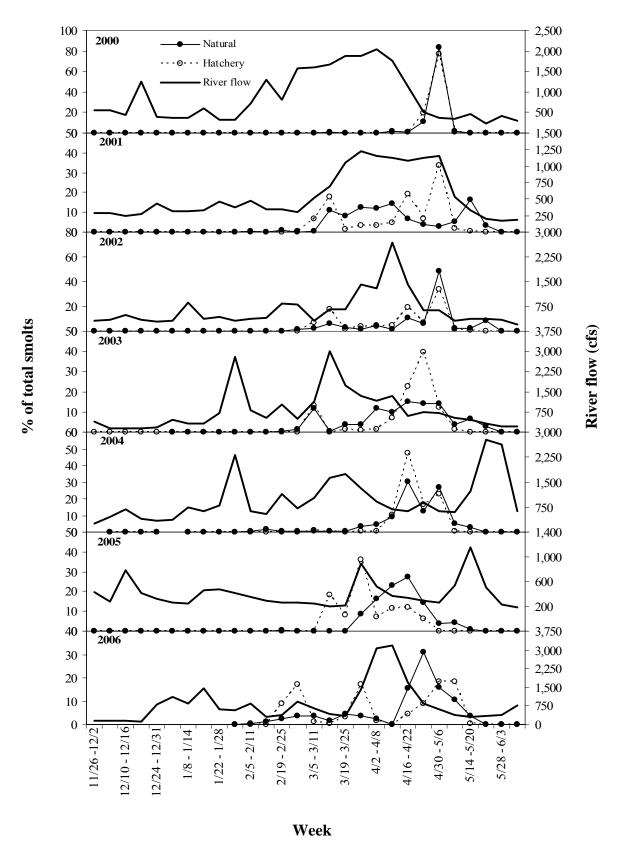


Figure 2. Emigration timing for natural and hatchery spring Chinook salmon smolts in the Umatilla River, 2000-2006.

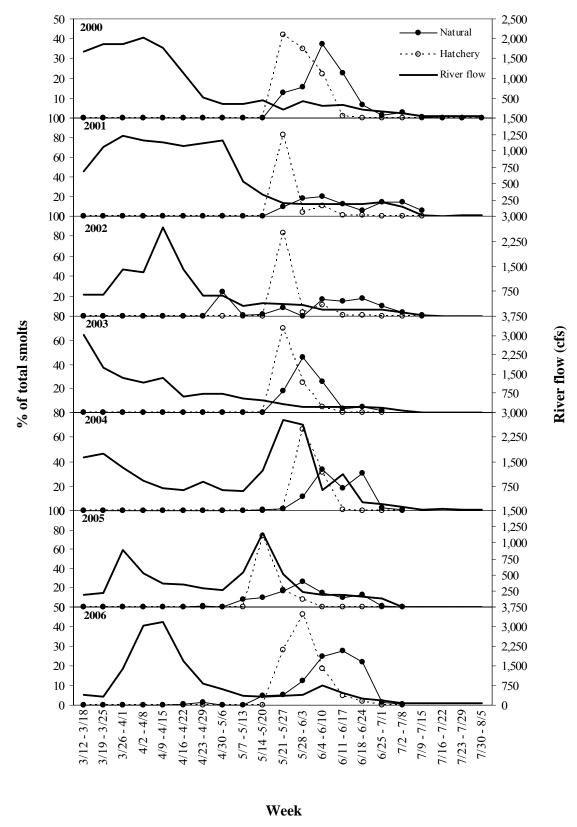


Figure 3. Emigration timing for natural and hatchery subyearling fall Chinook salmon smolts in the Umatilla River, 2000-2006.

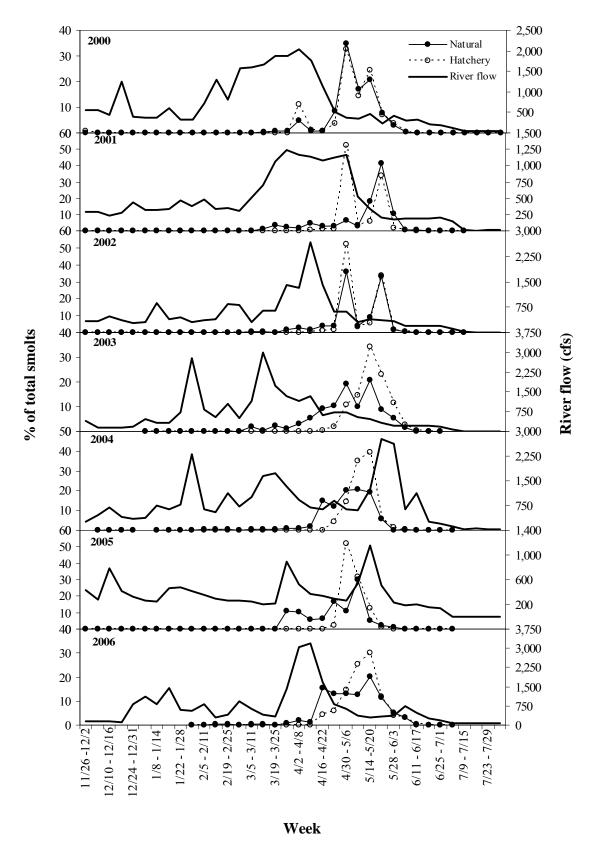


Figure 4. Emigration timing for natural and hatchery summer steelhead smolts in the Umatilla River, 2000-2006.

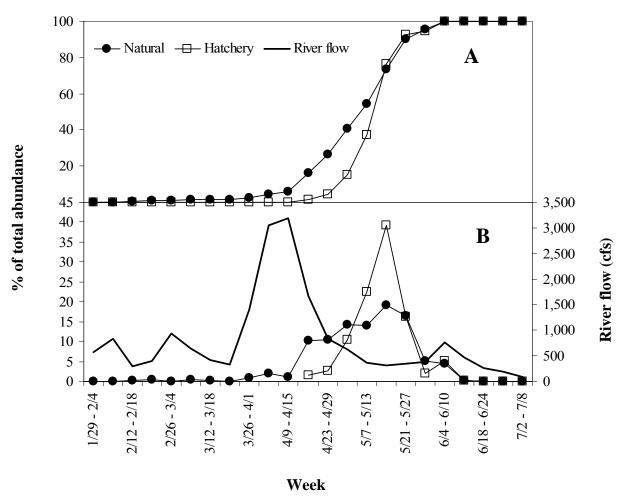


Figure 5. Cumulative (A) and percent abundance (B) of summer steelhead emigrating from the Umatilla River in 2006.

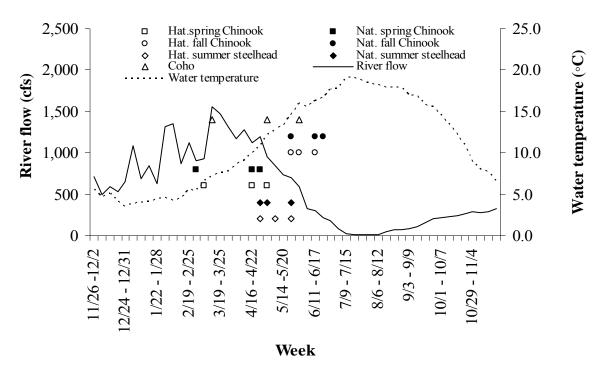


Figure 6. Mean weekly flows and temperatures (1995-2006) measured in the Umatilla River compared to 10, 50, and 90% emigration dates (left to right respectively) of Chinook salmon, coho salmon, and summer steelhead. Emigration timing distributions are the mean values calculated from outmigration years 1995-2006.

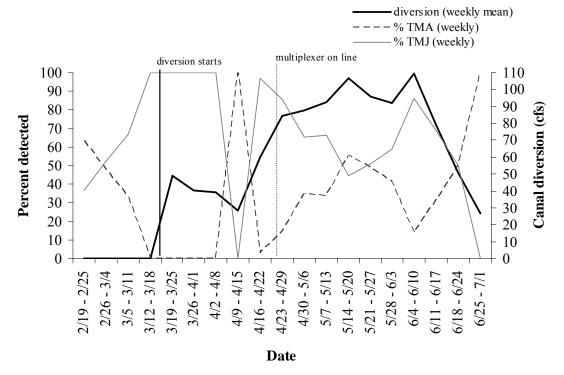


Figure 7. Passage route selection at TMFD by PIT tagged production released salmonids associated with canal diversion at West Extension Canal. TMA is the east bank adult fish ladder and TMJ is the juvenile bypass.

Size and age at emigration for Chinook salmon and summer steelhead in the Umatilla River is similar to that of surrounding subbasins (Contor editor 2003; Fast et al. 1991; Gallinat et al. 2003; Lindsay et al. 1986; Mayer and Schuck 2004; Peven et al. 1994). Spring Chinook salmon are emigrating as yearlings, fall Chinook salmon as subyearlings, and summer steelhead at a broad distribution of ages. One, two and three-year freshwater rearing accounts for 9.1%, 82.2% and 8.4% respectively of the life history of natural summer steelhead emigrating past TMFD (Tables 3 and 4). Both yearling and subyearling coho salmon have been observed emigrating from the Umatilla River.

Fish condition tends to deteriorate throughout the monitoring season as flows decrease and water temperature increases. In addition, hatchery fish typically exhibit higher rates of descaling compared to natural fish, which is likely due to fish acclimation activities, including transportation, netting, tagging, and seining during forced releases. Bird marks are observed at a higher rate for hatchery summer steelhead compared to other species/origin type (Tables 5). This is possibly a result of size-dependant selection by avian predators (Collis et al. 2001). *Neascus sp.* (black spot) infestation is prevalent in natural spring Chinook salmon and summer steelhead migrants (Table 5). Mortality rates tended to be higher for natural compared to hatchery-reared fish (Table 6).

Efficiency of the juvenile bypass is influenced by water temperature, river flow and canal diversion rate; relationships are species-specific and variable from year to year. Natural spring Chinook salmon exhibit the strongest relationship with canal diversion rate (Figure 8), while summer steelhead exhibit the weakest. Trap efficiencies are slightly lower for hatchery summer steelhead, coho salmon, and natural fall Chinook salmon compared to all other species (Table 7; Appendix Table A-4). In addition, efficiency estimates are more variable for Chinook salmon compared to coho salmon and summer steelhead. Termination of canal diversion with the onset of Phase I pump exchange briefly diminishes bypass efficiency.

Paired release tests in 2004 and 2005 indicated temporal and spatial variability in the recapture rate and travel speed of hatchery Chinook salmon (Figure 9 and 10). Recapture rates were considerably lower with greater variability and travel speed faster for fish released upstream of the juvenile facility compared to those released within the facility. Prior trap exposure had little or no effect on recapture rate or travel speed.

Patterns in fish movement were primarily a result of seasonal patterns in river conditions, canal operations, the level of smoltification and behavioral responses to hydraulic heterogeneity associated with TMFD and its fish passage routes. Similar trends in entrainment rate have been observed at Chandler Fish Facility (Sandford and Ruehle 1996) and corresponding patterns between migration and smoltification were observed by Beckman et. al. (2000) for spring Chinook salmon in the Yakima River. In addition, laboratory tests by Kemp et. al. (2005) determined that Pacific salmonid smolts more readily pass through an open channel compared to a constricted channel and believed this to be a result of rapidly accelerating flow. They suggest that fine-scale behavior of smolts as a result of local hydraulic conditions may help explain lower-than expected entrainment rates for fish passage facilities.

Table 3. Natural summer steelhead smolt size and age upon emigration from the Umatilla River, 1995-2006.

Outmigration	n Mean fork lengt	h	Fork leng	th (mm) by	freshwate	r age class	Perce	nt of freshw	ater age cla	ass ^b
year	(mm) (SD)	Sample size ^a	1 (SD)	2 (SD)	3 (SD)	4 (SD)	1	2	3	4
1995	175 (28)	1,612								
1996	176 (24)	2,970								
1997	157 (23)	183								
1998	186 (33)	2,547								
1999	181 (22)	1,704								
2000	180 (26)	619								
2001	178 (28)	844								
2002	166 (30)	571								
2003	176 (30)	959	102 (6)	170 (27)	211 (42)	270 (53)	3.4	79.6	16.0	1.0
2004	167 (30)	655	104 (16)	165 (25)	292 (32)	202 ()	5.9	82.6	11.3	0.2
2005	179 (25)	1,511	160 (28)	185 (27)	210 (44)		7.5	88.1	4.4	0.0
2006	179 (26)	1,005	164 (31)	184 (28)	191 (23)		17.6	77.4	5.0	0.0
Average	178 (28)	15,180	150 (37)	177 (28)	202 (37)	256 (55)	9.1	82.2	8.4	0.2

^a Sample sizes for age/length analysis from 2003 to 2006 were 382, 477, 589 and 563; respectively. ^b Derived from scale analysis of smolts trapped at TMFD.

Table 4. Natural Chinook salmon smolt size upon emigration from the Umatilla River, 1995-2006.

	Spring Ch	inook	Fall Chi	inook
Outmigration	Mean fork length		Mean fork length	
year	(mm)(SD)	Sample size	(mm) (SD)	Sample size
1995	109 (19)	561	73 (12)	395
1996	105 (14)	160	61 (9)	9
1997	99 (10)	14	64 (8)	21
1998	108 (14)	1,033	73 (11)	5,050
1999	103 (9)	616	87 (12)	24
2000	106 (15)	82	77 (14)	2,385
2001	103 (7)	125	74 (12)	1,336
2002	101 (14)	861	74 (11)	723
2003	100 (10)	545	78 (9)	787
2004	103 (9)	350	72 (11)	1,054
2005	112 (10)	596	75 (13)	2,723
2006	105 (9)	443	91 (10)	618
Average	105 (13)	5,379	75 (13)	15,130

Table 5. Condition summary for hatchery and natural smolts sampled at TMFD, 2004-2006.

		Hat	chery			Natı	ıral	
Outmigration	No.		·					
year	examined	% bird	% injury	% parasite	No. examined	% bird	% injury	% parasite
				Spring Chinook				
2004	2,355	1.2	2.9	0.8	349	0.9	2.6	44.7
2005	1,313	1.9	1.1	0.0	588	0.5	0.7	27.9
2006	1,383	2.5	1.7	0.1	443	0.5	0.7	27.1
			Fal	l Chinook yearlin	igs			
2004	3,188	1.7	2.9	0.8				
2005	1,696	2.4	1.7	0.2				
2006	1,341	1.4	0.7	0.6				
			Fall (Chinook subyearl	lings			
2004	1,604	0.3	1.5	0.1	1,035	0.1	2.5	0.0
2005	1,065	0.4	1.2	0.0	2,668	0.2	1.9	0.9
2006	738	0.3	0.7	0.1	617	0.0	0.6	0.2
				$Coho^{a}$				
2004	149	2.0	9.4	0.7	2,474	1.5	4.4	1.1
2005	366	1.6	0.5	0.3	2,184	1.4	1.6	0.3
2006	103	1.0	1.9	1.0	907	2.0	1.5	1.7
			S	ummer steelhead	!			
2004	510	3.1	3.9	0.2	657	0.9	4.3	19.3
2005	482	6.2	3.9	0.0	1,498	2.5	2.1	10.0
2006	610	3.9	1.5	0.8	1,003	1.1	2.4	17.4

a Includes natural and unmarked hatchery fish.

Table 6. Unknown source and sampling mortality for hatchery and natural smolts sampled at TMFD, 2004-2006.

Outmigration _	Hate	hery	Natural			
year	No. sampled % mortality		No. sampled	% mortality		
		Service - Chie	1-			
2004	(270	Spring Chine		0.2		
2004	6,278	0.2	351	0.3		
2005	16,481	0.1	606	1.0		
2006	12,317	0.2	445	2.0		
		Fall Chinook ye	arlings			
2004	10,532	0.7				
2005	9,958	0.3				
2006	9,654	0.2				
		Fall Chinook suby	vearlings			
2004	5,297	0.9	2,206	2.3		
2005	6,825	0.3	9,488	0.9		
2006	9,864	0.3	1,159	1.7		
		Coho ^a				
2004	266	0.4	5,563	0.4		
2005	682	0.1	13,371	1.6		
2006	446	0.2	7,799	0.2		
		Summer steell	head			
2004	1071	0.3	660	0.9		
2005	2197	4.7	1,992	0.5		
2006	1,720	0.5	1,020	1.8		

^a Includes natural and unmarked hatchery fish.

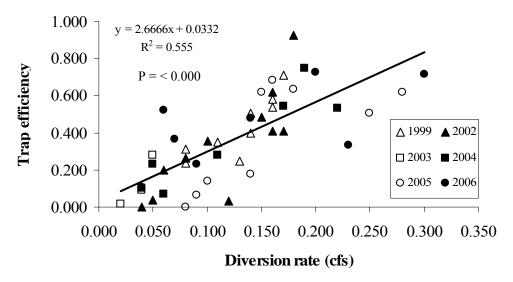


Figure 8. Relationship between canal diversion rate and efficiency of the juvenile bypass/trapping facility. Efficiency estimates used are from natural spring Chinook salmon.

Table 7. Species-specific mean efficiency estimates for the juvenile bypass/trapping facility at West Extension Canal, 1999-2006.

Species	No. of releases	Trap efficiency (SD)
Hatchery Chinook (1+)	129	0.30 (0.21)
Hatchery fall Chinook (0+)	70	0.34 (0.22)
Hatchery summer steelhead	77	0.22 (0.12)
Coho	62	0.26 (0.14)
Natural spring Chinook	47	0.37 (0.24)
Natural fall Chinook	27	0.21 (0.18)
Natural summer steelhead	73	0.32 (0.12)
Total/average	485	0.29 (0.19)

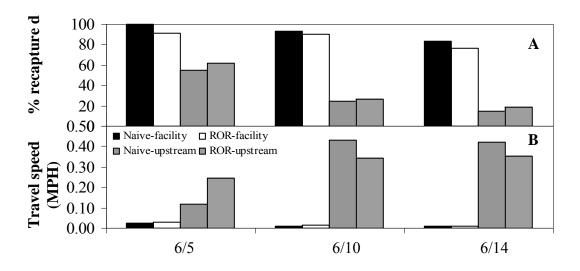


Figure 9. Recapture rate (A) and travel speed (B) for hatchery subyearling fall Chinook salmon released in 2004.

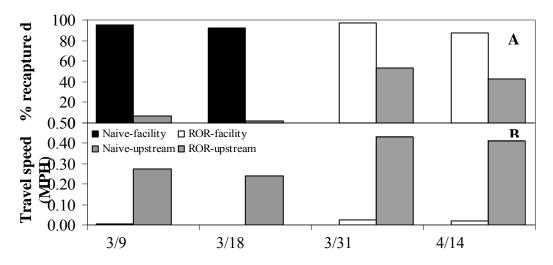


Figure 10. Recapture rate (A) and travel speed (B) for hatchery spring Chinook salmon released in 2005.

Abundance and Survival:

Abundance estimates for naturally-produced Chinook salmon and summer steelhead smolts are presented in Table 8. Abundance of summer steelhead has fluctuated since 1995, however trends reflect little or no change in productivity. The average annual smolt production is 50,217 (+/-6,116). Smolt yield for spring Chinook salmon has ranged between 603 and 52,645 over the past 11 years. Estimates for the 2004-2006 period are below the 11 year mean (Table 8). Production of fall Chinook smolts has fluctuated dramatically over the same time frame; smolt estimates have ranged between 169 to 242,100 from 1995-2006. A suite of factors including spawning escapement, quality and quantity of rearing habitat, river flow and overwinter survival are likely contributing to the highly variable smolt production. In 2005, a stable flow regime combined with a late season freshet likely played a key role in the strong overwinter survival and outmigration success of fall Chinook in the lower river (Figure 11). In contrast, dramatic flow fluctuations combined with a high magnitude spring freshet may have contributed to the lower abundance of Chinook emigrants in 2006.

Table 8. Abundance estimates of natural Chinook salmon and steelhead smolts emigrating from the Umatilla River, 1995-2006.

		Species	
Outmigration	1		
year	Summer steelhead	Spring Chinook	Fall Chinook
1995	54,361	19,900	6,912
1996	73,361	1,885	169
1997	22,221	603	830
1998	59,182	13,045	242,100
1999	46,530	23,809	1,292
2000	81,759	9,051	32,542
2001	33,844	4,746	28,597
2002	77,016	35,033	9,812
2003	24,773	52,645	30,998
2004	35,640	14,159	27,789
2005	59,807	13,137	107,858
2006	34,110	8,458	12,781
Average	50,217	16,373	41,807

Smolt-per-spawner ratios for Umatilla salmon and summer steelhead are considerably lower than that of surrounding subbasins (Berg and Fast 2002; Gallinat and Ross 2006). The average number of smolts produced per spawner for brood years 1993-2004 was 27.1 natural spring Chinook salmon, 75.8 fall Chinook salmon and 26.6 summer steelhead (Figure 12).

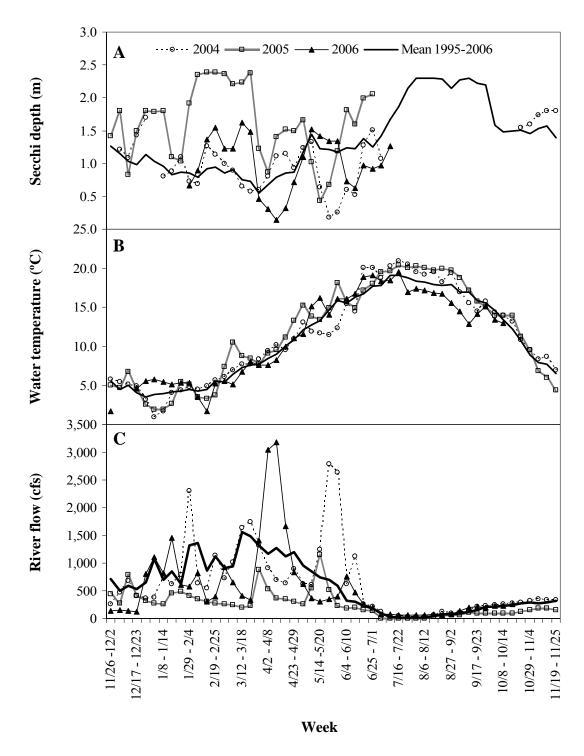


Figure 11. Weekly secchi depth (A), water temperature (B), and river flow (C), Umatilla River, 2004 - 2006.

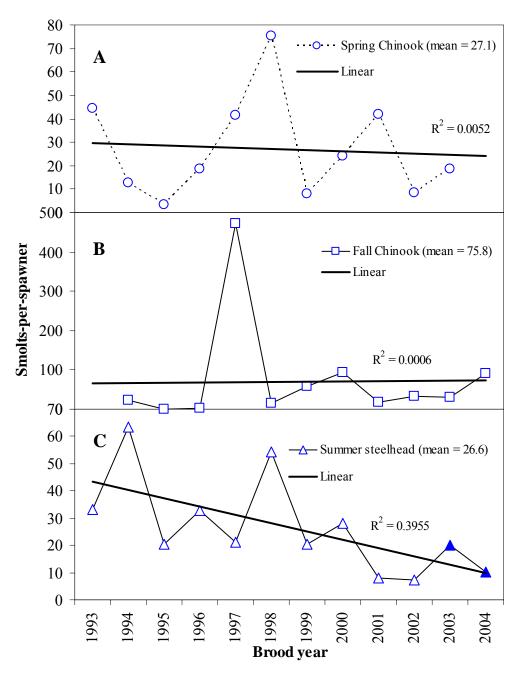


Figure 12. Smolt-per-spawner estimates of natural (A) spring Chinook salmon, (B) fall Chinook salmon and (C) summer steelhead in the Umatilla River, BY1993-2004. Adjusted for freshwater age structure. \triangle indicates dataset incomplete for age 3+ and 4+ smolts.

Summer steelhead smolts-per-spawner plotted against total spawning escapement indicated a declining smolt output with increasing escapement (Figure 13). Smolt estimates for natural summer steelhead between smolt outmigration year (SOY) 1995-2006 were reasonably strong, suggesting other in-basin rearing and migration conditions may be limiting the smolt production. Contor (2003) suggested flows in summer rearing habitat may be a limiting factor for Umatilla summer steelhead. Additional reports by Contor et al (1997) linked annual winter/spring flows with adult returns two years later. This suggests flow impacts freshwater survival across multiple life stages and age classes.

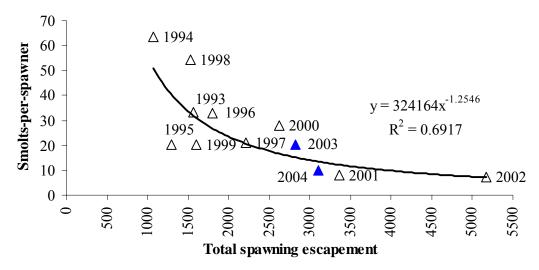


Figure 13. Relationship between total spawning escapement and smolts produced per spawner for natural summer steelhead between brood years 1993-2004. ▲ indicates return data incomplete.

The average number of summer steelhead smolts-produced-per-female between BY's 1993-2004 was 40.0 (Figure 14). Egg-to-smolt survival rates for natural summer steelhead ranged between 0.21-1.58% for BY 1993-2004, with a mean of 0.77% (Figure 15; Appendix Table A-5). Smolt-to-adult return (SAR) rates for naturally-reared summer steelhead were within the range of nearby subbasins. The mean SAR rate for Umatilla summer steelhead between SOY 1995-2002 was 3.53% (Figure 16; Appendix Table A-6). Preliminary estimates for the Yakima River (smolts at Prosser Dam to adults at McNary Dam) averaged 5.12% for SOY 2002-2003 (D. Lind. pers. comm.). SAR estimates for the John Day River subbasin averaged 1.65% between 2001-2004 (W. Wilson. pers. comm.).

Survival estimates for natural smolts are limited, however data suggests out-of-basin survival from TMFD to JDD is higher than in-basin survival to TMFD for most species. In-basin survival to TMFD averaged 44%, 46% and 34% for naturally-produced spring Chinook salmon, coho salmon and summer steelhead, respectively, from 1999-2001 (Table 9). Out-of-basin estimates have ranged from 16% in fall Chinook salmon to 65% in naturally-reared spring Chinook salmon over the 1999-2006 period.

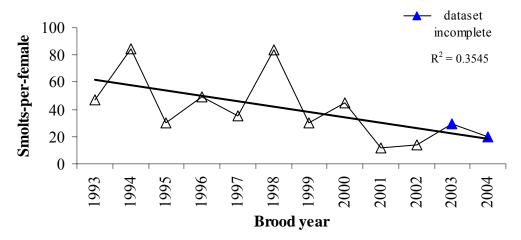


Figure 14. Natural summer steelhead smolts-per-female by brood year, 1993-2004.

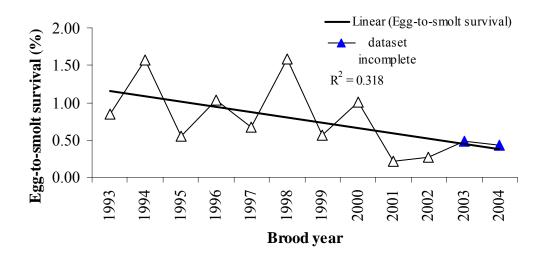


Figure 15. Natural summer steelhead egg-to-smolt survival by brood year, 1993-2004.

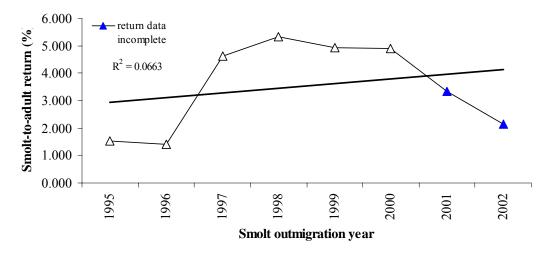


Figure 16. Natural summer steelhead smolt-to-adult return (%) by smolt outmigration year, 1995-2002.

Table 9. In- and out-of-basin survival estimates for juvenile natural salmon and steelhead emigrating from the Umatilla River, 1999-2006. Standard errors are in parentheses.

	Species										
	Spring Chinook salmon		Fall Chino	ook salmon	Coho sa	almon	Summer steelhead				
Outmigration	Release to	TMFD to	Release to	TMFD to	Release to	TMFD to	Release to	TMFD to			
year ^a	TMFD	JDD	TMFD	JDD	TMFD	JDD	TMFD	JDD			
1999	0.41 (0.07)	0.74 (0.12)	(b)	(b)	(b)	(b)	0.26 (0.02)	0.69 (0.07)			
2000	(b)	(b)	(b)	(b)	0.46 (0.22)	(b)	0.37 (0.04)	0.57 (0.14)			
2001	0.47 (0.08)	0.62 (0.12)	(b)	0.11 (0.10)	(b)	(b)	0.40 (0.05)	0.53 (0.20)			
2002	(b)	0.70 (0.12)	(b)	(b)	(b)	(b)	(b)	0.61 (0.17)			
2003	(b)	0.83 (0.26)	(b)	(b)	(b)	(b)	(b)	0.64 (0.12)			
2004	(b)	0.31 (0.13)	(b)	(b)	(b)	(b)	(b)	0.44 (0.16)			
2005	(b)	0.90 (0.22)	(b)	0.19 (0.11)	(b)	(b)	(b)	0.54 (0.12)			
2006	(b)	0.46 (0.11)	(b)	0.18 (0.08)	(b)	(b)	(b)	0.68 (0.14)			
Average	0.44	0.65		0.16	0.46		0.34	0.59			

Average 0.44 0.65 -- 0.16 0.46 -- 0.34 0.59

Tagged fish used to estimate survival from TMFD to JDD were released approximately 1.3 miles above TMFD for trap calibration tests.

Overall performance of hatchery smolts has been variable and less than satisfactory for most species. In-basin survival estimates have averaged 31% for coho salmon, 38% for summer steelhead, 59% for subyearling fall Chinook salmon, 62% for spring Chinook salmon, and 70% for yearling fall Chinook salmon (Table 10). Improved detection capabilities at TMFD and application of the SURPH 2 Model has provided more robust in-basin estimates over the past 3 years. Out-of-basin estimates from TMFD to JDD have ranged from 52% in hatchery subyearling fall Chinook salmon to 70% in hatchery coho salmon (Table 10).

Monitoring of hatchery release groups has shown no significant difference in survival between standard-transferred and fall-transferred (overwintered) hatchery spring Chinook salmon (Table 11). Yearling fall Chinook and coho salmon released in April survived at a higher rate than those released in March (Tables 12 and 13). Subyearling fall Chinook salmon direct-released at RM 48.5 generally survive at a higher rate than those acclimated and released at RM 73.5 (Table 14). Summer steelhead released low in the basin consistently outperform those released higher up (Table 15).

Inter-annual variation in the survival of salmon and steelhead is a result of numerous factors, including spatial and temporal distribution of release, type of release, size at tagging and/or release, rearing strategies, rate of residualism, water flow and temperature, turbidity, and the number of fish tagged. The error associated with most estimates in our time series data is too great to directly link survival to any one variant with a high degree of confidence. As the number of fish tagged is increased and errors reduced, we will be able to better evaluate the key variables influencing the survival of hatchery and natural smolts in the Umatilla River.

No marked fish released or insufficient data to calculate survival estimate and/or standard error

Table 10. In- and out-of-basin survival estimates for juvenile hatchery salmon and steelhead released in the Umatilla River, 1995-2006. Standard errors are in parentheses.

					Spe	cies				
	Spring Chin	ook salmon	Fall Chinook	salmon (+1)	Fall Chinook	salmon (+0)	Coho salmon		Summer steelhead	
Outmigration	Release to	TMFD to	Release to	TMFD to	Release to	TMFD to	Release to	TMFD to	Release to	TMFD to
year ^a	TMFD	JDD	TMFD	JDD	TMFD	JDD	TMFD	JDD	TMFD	JDD
1995	0.67 (b)		(b)		0.18 (b)		(b)		(b)	
1996	0.34 (0.39)		0.40 (0.03)		1.41 (0.02)		0.38 (0.03)		0.94 (0.05)	
1997	(b)		(b)		0.35 (0.03)		0.34 (0.10)		(b)	
1998	0.73 (0.03)		0.70 (0.08)		1.52 (0.04)		1.29 (0.06)		0.50 (0.04)	
1999	0.45 (0.13)	(b)	0.76 (0.17)	(b)	0.51 (0.04)	(b)			0.36 (0.06)	(b)
2000	0.67 (0.10)	0.60 (0.11)	0.82 (0.50)	0.50 (0.17)	0.91 (0.04)	0.49 (0.06)			0.23 (0.03)	0.60 (0.16)
2001	0.65 (0.06)	0.57 (0.06)	0.65 (0.07)	0.71 (0.11)	0.54 (0.03)	0.58 (0.11)	0.47 (0.11)	0.42 (0.16)	0.30 (0.04)	0.30 (0.06)
2002	0.59 (0.05)	0.55 (0.07)	0.81 (0.17)	0.52 (0.16)	0.49 (0.03)	0.53 (0.14)	0.39 (0.04)	0.74 (0.18)	0.36 (0.03)	0.75 (0.20)
2003	0.81 (0.13)	0.47 (0.10)	0.70 (0.09)	0.69 (0.17)	0.47 (0.05)	0.59 (0.10)	0.27 (0.03)	0.66 (0.10)	0.34 (0.08)	0.50 (0.17)
2004	0.65 (0.10)	0.61 (0.14)	0.76 (0.10)	0.67 (0.31)	0.62 (0.06)	0.69 (0.19)	0.24 (0.05)	2.24 (2.18)	0.44 (0.07)	1.09 (0.58)
2005	0.61 (0.07)	0.57 (0.10)	0.48 (0.11)	0.90 (0.48)	0.70 (0.11)	0.30 (0.15)	0.27 (0.05)	0.91 (0.49)	0.58 (0.08)	0.61 (0.21)
2006 ^c	0.53 (0.04)	0.54 (0.09)	0.65 (0.04)	0.70 (0.12)	0.45 (0.05)	0.49 (0.15)	0.23 (0.03)	0.78 (0.35)	0.41 (0.01)	0.64 (0.08)
Average ^d	0.62	0.56	0.70	0.67	0.59	0.52	0.31	0.70	0.38	0.57

^a Survival estimates for 1995-1998 were calculated using the Migrant Abundance methodology (Dauble et al. 1993). CJS survival estimates were calculated by the computer program SURPH (Lady et al. 2001) for 1999-2006.

b Insufficient data to calculate survival estimate and/or standard error.

c March releases of spring Chinook salmon, fall Chinook salmon yearlings and coho salmon were released on 2/18/06 due to freezing conditions at the acclimation site.

d Average is calculated using CJS survival estimates only, and does not include estimates that exceed 1.00

Table 11. In- and out-of-basin survival estimates for standard- and fall-transferred spring Chinook salmon reared at Umatilla Fish Hatchery, 1999-2006. Standard errors are in parentheses.

_	Treatment group								
_	Standard	d-transfer	Fall-transfer						
Outmigration	Release to	TMFD to	Release to	TMFD to					
year ^a	TMFD	JDD	TMFD	JDD					
1999	0.45 (0.13)	(b)	0.33 (0.30)	(b)					
2000	0.35 (0.13)	1.10 (0.43)	0.89 (0.55)	0.81 (0.57)					
2001	0.63 (0.10)	0.53 (0.10)	0.47 (0.07)	0.79 (0.16)					
2002	0.73 (0.16)	0.71 (0.25)	0.63 (0.13)	0.47 (0.13)					
2003	0.91 (0.33)	0.53 (0.25)	0.70 (0.27)	0.92 (0.54)					
2004	0.56 (0.33)	0.76 (0.48)	0.49 (0.30)	0.59 (0.39)					
2005	0.58 (0.10)	0.46 (0.11)	0.68 (0.14)	0.64 (0.20)					
2006 ^c	0.50 (0.07)	0.70 (0.29)	0.53 (0.10)	0.54 (0.18)					
Average ^d	0.59	0.62	0.59	0.68					

^a All standard- and fall-transfer fish reared in Oregon raceways for migration years 1999-2003 and Michigan raceways for migration years 2004-2006.

Table 12. In- and out-of-basin survival estimates for March and April released yearling fall Chinook salmon reared at Bonneville Fish Hatchery, 1999-2006. Standard errors are in parentheses.

_	Treatment group							
_	March	release	April release					
Outmigration	Release to	TMFD to	Release to	TMFD to				
year	TMFD	JDD	TMFD	JDD				
1999	0.50 (0.10)	(b)	0.76 (0.17)	(b)				
2000	0.82 (0.50)	0.68 (0.50)	0.76 (0.18)	0.39 (0.15)				
2001	0.46 (0.12)	0.67 (0.20)	0.80 (0.10)	0.75 (0.14)				
2002	0.70 (0.19)	0.34 (0.13)	0.73 (0.24)	0.93 (0.45)				
2003	0.47 (0.17)	0.69 (0.31)	0.85 (0.11)	0.74 (0.23)				
2004	0.79 (0.35)	0.38 (0.28)	0.89 (0.10)	0.88 (0.56)				
2005 ^a	0.41 (0.25)	(b)	0.63 (0.15)	0.75 (0.39)				
2006	$0.49 (0.10)^{c}$	0.59 (0.24) ^c	0.86 (0.05)	0.71 (0.14)				
Average	0.58	0.56	0.79	0.74				

^a Fish were direct stream released at RM 48.5 in February and March for migration year 2005.

^b Insufficient data to calculate a survival estimate.

^c Fish were released on 2/18/06.due to freezing conditions at the acclimation site.

^d Average does not include estimates that exceed 1.00

^b Insufficient data to calculate a survival estimate.

^c Fish were released on 2/18/06.due to freezing conditions at the acclimation site.

Table 13. In- and out-of-basin survival estimates for March and April released juvenile coho salmon acclimated and released at RM 56, 2001-2006. Standard errors are in parentheses.

		Treatment group										
		Herman Creek chery		Cascade Hatchery								
	March	release	March	release	April	release						
Outmigration	Release to	TMFD to	Release to	TMFD to	Release to	TMFD to						
year	TMFD	JDD	TMFD	JDD	TMFD	JDD						
2001	(b)	(b)	0.58 (0.53)	0.18 (0.22)	0.64 (0.15)	0.45 (0.17)						
2002	0.43 (0.16)	0.38 (0.20)	0.61 (0.31)	0.26 (0.16)	0.43 (0.05)	1.32 (0.53)						
2003	0.13 (0.04)	0.70 (0.29)	0.19 (0.04)	0.93 (0.32)	0.38 (0.05)	0.59 (0.11)						
2004	0.30 (0.27)	(b)	0.28 (0.17)	0.94 (1.01)	0.26 (0.06)	(b)						
2005 ^a	0.39 (0.24)	(b)	0.19 (0.06)	1.18 (1.07)	0.34 (0.10)	0.81 (0.53)						
2006			$0.04 (0.02)^{c}$	(b)	0.41 (0.06)	0.68 (0.30)						
Average ^d	0.31	0.54	0.32	0.58	0.41	0.63						

^a Fish were direct stream released at RM 48.5 in February and March for migration year 2005.

Table 14. In- and out-of-basin survival estimates for subyearling fall Chinook salmon reared at Umatilla Fish Hatchery and acclimated and released at RM 73.5 or direct stream released at RM 48.5, 1999-2006. Standard errors are in parentheses.

	Treatment group								
_	Acclimate	d (RM 73.5)	Direct released (RM 48.5)						
Outmigration	Release to	TMFD to	Release to	TMFD to					
year ^a	TMFD	JDD	TMFD	JDD					
1999	0.50 (0.04)	(b)							
2000	1.05 (0.18)	0.85 (0.47)	1.01 (0.15)	0.40 (0.15)					
2001	0.63 (0.05)	0.49 (0.11)	0.45 (0.03)	0.75 (0.27)					
2002	0.21 (0.03)	2.16 (2.02)	0.73 (0.05)	0.37 (0.09)					
2003	0.29 (0.06)	0.53 (0.16)	0.65 (0.08)	0.62 (0.12)					
2004	0.54 (0.09)	0.88 (0.36)	0.65 (0.06)	0.56 (0.21)					
2005			0.70 (0.11)	0.30 (0.15)					
2006	0.57 (0.11)	0.43 (0.20)	0.39 (0.05)	0.49 (0.21)					
Average ^c	0.46	0.64	0.60	0.50					

^a In 1999, fish were acclimated and released from RM 80 and in 2000 fish were direct stream released at RM 73.5.

b Insufficient data to calculate a survival estimate.

^c Fish were emergency released on 2/19/06.

^d Average does not include estimates that exceed 1.00.

b Insufficient data to calculate a survival estimate.

^c Average does not include estimates that exceed 1.00.

Table 15. In- and out-of-basin survival estimates for juvenile summer steelhead reared at Umatilla Fish Hatchery, 2002-2006. Acclimated and released at Bonifer Springs (RM 79), Minthorn Springs (RM 64.9), and Pendleton (RM 56).

	Treatment group										
	RM	79 ^a	RM	64.5	RN	1 56					
Outmigration year	Release to TMFD	TMFD to JDD	Release to TMFD	TMFD to JDD	Release to TMFD	TMFD to JDD					
2002 2003	1.35 (0.71) (b)	0.23 (0.15) (b)	0.31 (0.04) 0.24 (0.06)	1.60 (1.05) 0.60 (0.22)	0.37 (0.06) 0.47 (0.23)	0.46 (0.18) 0.43 (0.26)					
2004 ^c	0.37 (0.12)	(b)	0.38 (0.08)	1.33 (1.20)	0.59 (0.18)	0.43 (0.27)					
2005 ^d 2006	 0.31 (0.02)	 0.51 (0.11)	0.57 (0.10) 0.45 (0.02)	0.86 (0.52) 0.62 (0.14)	0.59 (0.14) 0.47 (0.02)	0.48 (0.20) 0.72 (0.14)					
Average ^e	0.34	0.37	0.39	0.69	0.50	0.50					

^a Rivermile 2 of Meacham Creek, which flows into the Umatilla River at rivermile 79.

PIT Tag Interrogation

PIT-tag marking and remote interrogation has been successful and incrementally improved at TMFD since 1999. Software and hardware has been improved to enhance performance, reliability and remote system monitoring. Equipment in the juvenile bypass trap has progressed from a 400 kHz detection system seven years ago, to a stationary antenna array (134 kHz) with remote data transmission (Figures 17 and 18). Detection efficiency at the juvenile bypass is 99.9%. A second antenna array (flatplate) was installed in the juvenile bypass outfall in 2005 to help monitor potential passage delays and interrogate fish during periods of non-trapping (Figures 17 and 19). Estimated efficiency of the outfall detector is 47.9%, when the bypass is operating at 5 cfs.

Efforts to upgrade provisional PIT-tag detection equipment at the east bank adult fish ladder were unsuccessful in 2004 due to budgetary and logistical constraints. In 2005, designs plans and specifications to install a permanent adult ladder detector were completed (Appendix B). Plans included installation of 2 stationary pass-through antennas molded into high impact housing and mounted within the vertical slots (weir walls) of the adult fish ladder. Initial attempts to secure funding for the construction phase of the project were unsuccessful. However, in the spring of 2006 ODFW collaborated with the Bureau of Reclamation to enhance detection efficiency at the adult fish ladder. Paddle style antennas fed by portable transceivers were replaced with a single antenna enclosed in PVC piping and mounted to the backlight chamber of the viewing window (Figure 20). The new antenna array was powered by a multiplexer unit, which increased detection at the ladder location 3-fold (Figure 21). Mean juvenile detection efficiency at the adult ladder detector is currently 89%.

Through the help of Pacific States Marine Fisheries Commission (PSMFC), additional upgrades were conducted in fall of 2006 to provide for remote system monitoring and coordination of data

^b Insufficient data to calculate survival estimate and/or standard error.

^c Fish were direct stream released downstream of acclimation pond in 2004.

^d Fish were acclimated and released at RM 73.5 instead of RM 64.5 in 2005.

^e Average does not include estimates that exceed 1.00

from all 3 detection arrays at TMFD (juvenile bypass trap, juvenile bypass outfall and adult fish ladder). Tag data from the two arrays along the west bank of the dam is now transmitted to the east bank of the dam via radio modem/antenna, so tag information from all 3 sites is currently uploaded at one central location (Figure 15). A phone line with modem was installed along the east bank of the dam to remotely upload PIT-tag data every 3 hours to the PTAGIS database. In addition, an electronics trailer equipped with climate control, was stationed along the east bank of the dam to house radio modems, transceivers and data logging equipment (Figure 17). The design scheme for the permanent adult ladder detector was incorporated into the upgrades. Therefore, should installation of the permanent interrogation system proceed, all necessary electronics and communication equipment is in place.

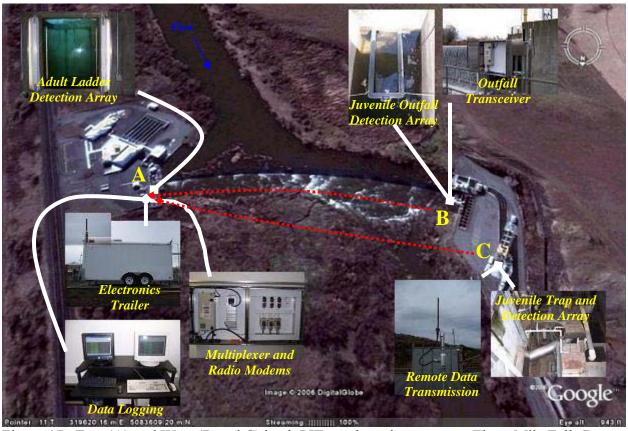


Figure 17. East (A) and West (B and C) bank PIT tag detection arrays at Three Mile Falls Dam.

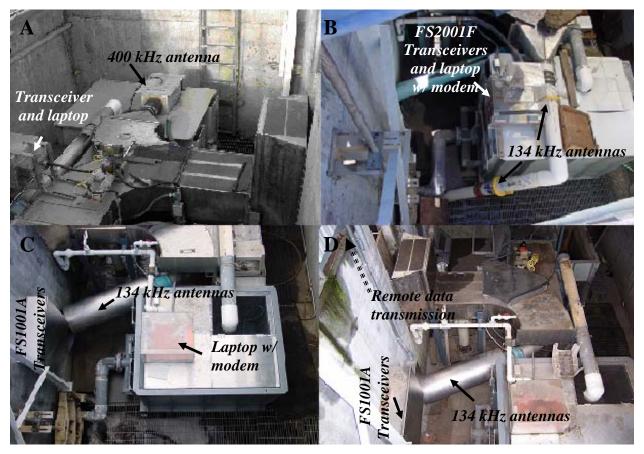


Figure 18. PIT tag detection array within the trap at the juvenile bypass facility in 1999 (A), 2000 (B), 2004 (C) and 2006 (D).

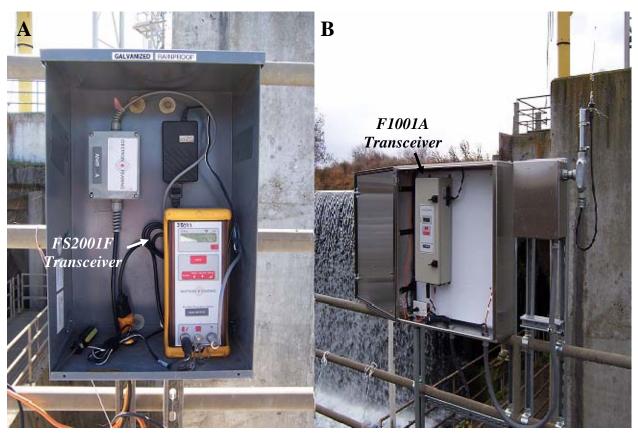


Figure 19. PIT tag detection array within the outfall at the juvenile bypass facility in 2005 (A) and 2006 (B).



Figure 20. PIT tag detection array within the viewing window at the adult fish ladder in 2000 (A) and 2006 (B).

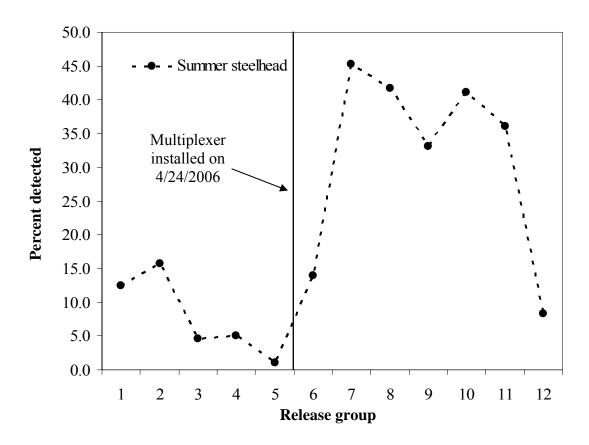


Figure 21. Improved detection efficiency of natural summer steelhead in the adult fish ladder at Three Mile Falls Dam, following installation of the multiplexer and upgraded antenna array, April 2006. Associative analysis suggests the low detection rate for the last release group (12) was a result of high flows.

Resident Species and Lamprey

Endemic species dominate the catch of resident species in the lower Umatilla River with several non-endemic species present in smaller numbers (Table 16. An increasing number of juvenile Pacific lamprey have been captured over the past three years despite a reduction in trapping efforts targeted at this species. The increase in juvenile numbers likely corresponds to the outplanting of adults in 2000 by CTUIR (Close 2002). A wide range in size and time of capture suggests that multiple life history strategies of both endemic and non-endemic species exist. In addition, temporal changes in the magnitude of resident fish and lamprey captured in lower river traps appear to be related to flow and temperature (Figure 22).

Table 16 Number of non-target species and Pacific lamprey captured and length summary statistics, Umatilla River, 2004 – 2006. Number captured is derived from both trapping sites, but not expanded for trap-efficiency or interpolated for unsampled periods. Standard deviation in parenthesis.

	O	Outmigration year 2004			O	Outmigration year 2005				Outmigration year 2006			
	No.		Fork lengt	h	No.		Fork lengt	:h	No.		Fork leng	th	
Species	captured	N ^a	Mean	Range	captured	N ^a	Mean	Range	captured	N ^a	Mean	Range	
Adult lamprey	3	3	449 (41)	470-545	1	1	470 ()	470					
Bluegill	115	110	53 (24)	25-156	15	15	100 (14)	70-118	30	30	104 (24)	43-149	
Bullhead sp.	304	269	71 (37)	28-223	20	15	86 (56)	30-186	212	81	99 (61)	38-282	
Chiselmouth	590	506	167 (61)	26-292	439	308	139 (69)	39-288	270	154	169 (26)	101-260	
Common carp	2	1	24 ()	24					6	5	134 (37)	88-170	
Crappie sp.	20	19	79 (42)	27-164	9	9	107 (28)	83-159	1				
Dace sp.	96	96	48 (7)	31-75	130	122	45 (5)	35-65	2	2	55 (0)	55	
Largemouth bass	816	80	42 (43)	23-340	7	7	71 (18)	56-107	7	6	46 (10)	33-60	
Larvae lamprey	164	154	137 (30)	48-192	120	74	126 (37)	56-177	404	82	146 (9)	130-169	
Metamorphosed lamprey	34	34	139 (14)	115-167	349	119	137 (12)	106-171	29	3	153 (11)	140-160	
Mosquitofish	1	1	26 ()	26					1				
Mountain whitefish					1	1	183 ()	183	44	33	186 (15)	152-213	
Northern pikeminnow	127	126	212 (46)	68-283	145	32	175 (41)	50-230	85	84	216 (40)	86-300	
O. mykiss kelt	2	2	586 (35)	561-610					19	19	644 (69)	540-750	
Redside shiner	31	27	65 (25)	25-163	13	11	43 (13)	26-59	7	3	63 (5)	59-68	
Sculpin sp.	5	5	90 (64)	19-147	1	1	121 ()	121	7	6	105 (5)	100-110	
Smallmouth bass	65	65	108 (38)	24-212	245	91	102 (35)	56-287	122	99	133 (41)	73-310	
Sucker sp.	1,249	982	151 (53)	24-445	468	297	153 (51)	15-254	708	447	187 (55)	62-470	
Sunfish sp.									1	1	129 ()	129	
Yellow perch	670	634	62 (14)	23-191	29	29	109 (45)	25-169	30	28	80 (32)	55-169	
<u>Total</u>	4,294	3,114	119 (68)	19-610	1,992	1,132	126 (60)	15-470	1,985	1,083	173 (86)	33-750	

^a N = number of observations.

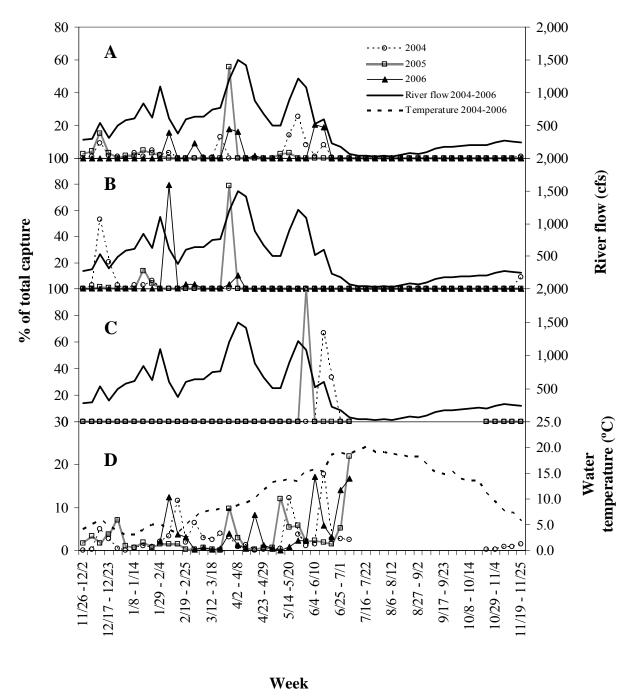


Figure 22. Capture timing of larvae (A), metamorphosed (B), and adult (C) Pacific lamprey and non-target species (D) in the Umatilla River, 2004-2006. Data includes both trapping sites and gaps in data series indicate periods when traps were not operating.

CONCLUSIONS AND RECOMMENDATIONS

Information gathered by the ODFW's Umatilla Juvenile Outmigration and Survival M&E Project between 2004-2006 has helped characterize the status and trends of juvenile salmonids in the Umatilla River subbasin, provided insight into the effects of river flow, canal flow, water temperature, and other environmental variables on fish passage and survival, helped evaluate juvenile life history characteristics and trends in natural production and aided in monitoring the movement of juvenile Pacific lamprey and resident fish.

The following recommendations are provided to assist in adaptive management of the Umatilla subbasin and answering critical uncertainties of the fisheries resource.

Juvenile outmigration and survival monitoring is important for the long-term tracking of
population responses over time. Information collected on smolt abundance, survival and life
history characteristics is essential for characterizing the status and trends of salmon and
steelhead populations in the Umatilla River subbasin and is needed to evaluate in-basin
capacity, productivity and ascertain loss by life-stage. Productivity metrics produced from
project data are the key subbasin scale measure of the effectiveness of salmon and steelhead
restoration actions in the Umatilla River subbasin.

<u>Recommendation:</u> Continued monitoring at TMFD will provide critical information on key performance measures including smolts-per-spawner, egg-to-smolt survival and smolt-to-adult return rates for naturally-produced and hatchery-reared salmon and steelhead in the Umatilla basin. These metrics enable us to look at productivity across multiple phases of the salmon/steelhead lifecycle and better ascertain limiting factors within and outside the basin.

2. Data collected by this project is critical to MCR Recovery Planning of ESA-listed summer steelhead and the 2004 Biological Opinion Remand Process. Although relatively small (10 yr lifespan) the Umatilla dataset is one of the only long-term datasets available for summer steelhead in the MCR ESU that allows us to look at life-stage specific survival at a tributary level. There is currently a large void in PIT-tag migration and survival data for MCR summer steelhead that is hampering efforts to identify recovery needs for these fish.

<u>Recommendation</u>: Additional PIT-tagging of natural summer steelhead is necessary to support SAR estimates and better understand the effects of lower Columbia River dams on listed populations.

3. The MCR recovery team is building a model to analyze the effects of tributary habitat actions on steelhead viability and the Umatilla dataset is being used to test the model. The smolt data is essential for modeling to understand how tributary, habitat, hatchery and hydrosystem actions influence productivity.

<u>Recommendation:</u> Continue to estimate smolt abundance, migration timing and survival of ESA-listed natural summer steelhead to assist in development of recovery goals for these fish.

4. PIT-tag interrogation has been successful and incrementally improved at TMFD since 1999. Over the past 3 years, equipment has been upgraded, detection efficiency enhanced and tag information coordinated to remotely upload at one central location. As a result, we have observed more robust estimates in smolt survival and comparisons of migration timing for hatchery and natural salmonids.

<u>Recommendation:</u> Continue to PIT-tag hatchery and natural smolts and monitor trends in natural abundance. Enhance detection efficiency and estimates of smolt survival at TMFD through installation (construction) of the permanent PIT-tag detection system in the east bank adult ladder.

5. Data suggests passage route selection at TMFD is influenced by canal diversion.

<u>Recommendation:</u> Installation of a permanent PIT-tag detection system in the adult fish ladder would further enhance the understanding of route selection past TMFD during canal diversion and various flow enhancement strategies.

6. An increased understanding of flow enhancement efforts and their effects on all life stages of salmonids is required to effectively manage water in the Umatilla River.

<u>Recommendation:</u> Additional research should be conducted to identify the potential affects of McKay Reservoir releases on natural and hatchery salmonids, particularly subyearling fall Chinook salmon.

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APPENDIX A Auxiliary Information from Outmigration Studies

Appendix Table A-1. Week of 10th, 50th, and 90th percentiles of summer steelhead emigrating from the Umatilla River, 1995-2006.

		Natural sumi	mer steelhead		Hatchery summer steelhead				
Outmigration	10%	50%	90%	Peak	10%	50%	90%	Peak	
year	emigration	emigration	emigration	emigration	emigration	emigration	emigration	emigration	
1995	2/19 - 2/25	4/23 - 4/29	5/14 - 5/20	4/23 - 4/29	4/19 - 4/15	4/23 - 4/29	5/21 - 5/27	4/23 - 4/29	
1996	4/23 - 4/29	5/14 - 5/20	5/28 - 6/3	5/14 - 5/20	5/7 - 5/13	5/14 - 5/20	5/28 - 6/3	5/14 - 5/20	
1997	2/19 - 2/25	4/23 - 4/29	5/7 - 5/13	4/23 - 4/29	4/9 - 4/15	4/30 - 5/6	5/14 - 5/20	4/30 - 5/6	
1998	4/16 - 4/22	4/30 - 5/6	5/21 - 5/27	4/23 - 4/29	4/23 - 4/29	4/30 - 5/6	5/21 - 5/27	4/30 - 5/6	
1999	3/26 - 4/1	5/21 - 5/27	5/28 - 6/3	5/21 - 5/27	4/30 - 5/6	5/21 - 5/27	5/28 - 6/3	5/21 - 5/27	
2000	4/23 - 4/29	4/30 - 5/6	5/21 - 5/27	4/30 - 5/6	4/2 - 4/8	4/30 - 5/6	5/21 - 5/27	4/30 - 5/6	
2001	4/9 - 4/15	5/21 - 5/27	5/28 - 6/3	5/21 - 5/27	4/9 - 4/15	4/30 - 5/6	5/21 - 5/27	4/30 - 5/6	
2002	4/16 - 4/22	4/30 - 5/6	5/21 - 5/27	4/30 - 5/6	4/30 - 5/6	4/30 - 5/6	5/21 - 5/27	4/30 - 5/6	
2003	4/9 - 4/15	4/30 - 5/6	5/21 - 5/27	5/14 - 5/20	4/30 - 5/6	5/14 - 5/20	5/28 - 6/3	5/14 - 5/20	
2004	4/16 - 4/22	4/30 - 5/6	5/14 - 5/20	5/7 - 5/13	4/30 - 5/6	5/7 - 5/13	5/14 - 5/20	5/14 - 5/20	
2005	3/26 - 4/1	4/23 - 4/29	5/7 - 5/13	5/7 - 5/13	4/30 - 5/6	4/30 - 5/6	5/14 - 5/20	4/30 - 5/6	
2006	4/16 - 4/22	5/7 - 5/13	5/21 - 5/27	5/14 - 5/20	4/23 - 4/29	5/7 - 5/13	5/21 - 5/27	5/14 - 5/20	
Mean	4/2 - 4/8	4/30 - 5/6	5/21 - 5/27	5/7 - 5/13	4/23 - 4/29	5/7 - 5/13	5/21 - 5/27	5/7 - 5/13	

Appendix Table A-2. Week of 10th, 50th, and 90th percentiles of spring Chinook salmon emigrating from the Umatilla River, 1995-2006.

		Natural spring Ch	ninook salmon		Hatchery spring Chinook salmon					
Outmigration	10%	50%	90%	Peak	10%	50%	90%	Peak		
year	emigration	emigration	emigration	emigration	emigration	emigration	emigration	emigration		
1995	12/3 - 12/9	2/26 - 3/4	4/16 - 4/22	12/3 - 12/9	12/3 - 12/9	4/9 - 4/15	4/16 - 4/22	4/16 - 4/22		
1996	3/19 - 3/25	4/2 - 4/8	4/23 - 4/29	4/2 - 4/8	3/12 - 3/18	3/12 - 3/18	4/2 - 4/8	3/12 - 3/18		
1997	12/24 - 12/31	1/15 - 1/21	4/9 - 4/15	1/15 - 1/21	3/26 - 4/1	3/26 - 4/1	4/16 - 4/22	3/26 - 4/1		
1998	4/9 - 4/15	4/23 - 4/29	4/30 - 5/6	4/23 - 4/29	4/16 - 4/22	4/16 - 4/22	4/30 - 5/6	4/16 - 4/22		
1999	3/5 - 3/11	4/2 - 4/8	4/30 - 5/6	3/12 - 3/18	3/5 - 3/11	3/5 - 3/11	3/5 - 3/11	3/5 - 3/11		
2000	4/23 - 4/29	4/30 - 5/6	4/30 - 5/6	4/30 - 5/6	4/23 - 4/29	4/30 - 5/6	4/30 - 5/6	4/30 - 5/6		
2001	3/12 - 3/18	4/9 - 4/15	5/14 - 5/20	5/14 - 5/20	3/12 - 3/18	4/16 - 4/22	5/7 - 5/13	4/23 - 4/29		
2002	3/12 - 3/18	4/30 - 5/6	5/14 - 5/20	4/30 - 5/6	3/12 - 3/18	4/16 - 4/22	4/30 - 5/6	4/30 - 5/6		
2003	3/5 - 3/11	4/16 - 4/22	5/7 - 5/13	4/16 - 4/22	3/5 - 3/11	4/23 - 4/29	4/30 - 5/6	4/23 - 4/29		
2004	4/2 - 4/8	4/16 - 4/22	4/30 - 5/6	4/16 - 4/22	4/9 - 4/15	4/16 - 4/22	4/30 - 5/6	4/16 - 4/22		
2005	3/26 - 4/1	4/9 - 4/15	4/23 - 4/29	4/16 - 4/22	3/12 - 3/18	3/26 - 4/1	4/16 - 4/22	3/26 - 4/1		
2006	3/5 - 3/11	4/23 - 4/29	5/7 - 5/13	4/23 - 4/29	2/26 - 3/4	4/16 - 4/22	5/7 - 5/13	5/7 - 5/13		
Mean	2/26 - 3/4	4/16 - 4/22	4/30 - 5/6	4/16 - 4/22	3/5 - 3/11	4/16 - 4/22	4/30 - 5/6	4/9 - 4/15		

Appendix Table A-3. Week of 10th, 50th, and 90th percentiles of fall Chinook salmon emigrating from the Umatilla River, 1995-2006.

		Natural fall C	hinook salmon		Hatchery fall Chinook salmon					
Outmigration	10%	50%	90%	Peak	10%	50%	90%	Peak		
year	emigration	emigration	emigration	emigration	emigration	emigration	emigration	emigration		
1995	6/18 - 6/24	6/18 - 6/24	6/25 - 7/1	6/25 - 7/1	5/21 - 5/27	5/21 - 5/27	6/4 - 6/10	5/21 - 5/27		
1996	5/14 - 5/20	5/28 - 6/3	6/11 - 6/17	5/14 - 5/20	6/4 - 6/10	6/4 - 6/10	6/4 - 6/10	6/4 - 6/10		
1997	5/28 - 6/3	6/11-6/17	6/18 - 6/24	6/11- 6/17	5/28 - 6/3	6/4 - 6/10	6/18 - 6/24	5/28 - 6/3		
1998	6/4 - 6/10	6/11-6/17	6/25 - 7/1	6/11- 6/17	5/28 - 6/3	6/4 - 6/10	6/11-6/17	6/4 - 6/10		
1999	4/30 - 5/6	6/11-6/17	6/25 - 7/1	6/11- 6/17	6/11-6/17	6/18 - 6/24	6/18 - 6/24	6/11- 6/17		
2000	5/21 - 5/27	6/4 - 6/10	6/18 - 6/24	6/4 - 6/10	5/21 - 5/27	5/28 - 6/3	6/4 - 6/10	5/21 - 5/27		
2001	5/21 - 5/27	6/11- 6/17	7/2 - 7/8	6/4 - 6/10	5/21 - 5/27	5/28 - 6/3	5/28 - 6/3	5/28 - 6/3		
2002	4/30 - 5/6	6/4 - 6/10	6/25 - 7/1	4/30 - 5/6	5/21 - 5/27	5/21 - 5/27	6/4 - 6/10	5/21 - 5/27		
2003	5/21 - 5/27	5/28 - 6/3	6/4 - 6/10	5/28 - 6/3	5/21 - 5/27	5/21 - 5/27	5/28 - 6/3	5/21 - 5/27		
2004	5/28 - 6/3	6/11-6/17	6/18 - 6/24	6/4 - 6/10	5/28 - 6/3	5/28 - 6/3	6/4 - 6/10	5/28 - 6/3		
2005	5/7 - 5/13	5/28 - 6/3	6/18 - 6/24	5/28 - 6/3	5/14 - 5/20	5/14 - 5/20	5/21 - 5/27	5/14 - 5/20		
2006	5/21 - 5/27	6/11- 6/17	6/18 - 6/24	6/4 - 6/10	5/21 - 5/27	5/28 - 6/3	6/4 - 6/10	5/28 - 6/3		
Mean	5/21 - 5/27	6/11- 6/17	6/18 - 6/24	6/4 - 6/10	5/21 - 5/27	5/28 - 6/3	6/11- 6/17	5/28 - 6/3		

Appendix Table A-4. Species-specific trap efficiency estimates for the juvenile bypass/trapping facility at TMFD, 2004-2006 (ND = Not determined).

]	Februa	ry		Marcl	ı	April			May			June			
Year	Min	Max	Mean	Min	Max	Mean	Min	Max	Mean	Min	Max	Mean	Min	Max	Mean	Mean
	Hatchery Yearling Spring Chinook Salmon															
2004	ND	ND	ND	1.7	43.3	15.6	8.3	81.7	53.0	23.1	76.7	55.1	ND	ND	ND	46.0
2005	ND	ND	ND	2.0	53.3	22.2	38.3	68.8	53.8	ND	ND	ND	ND	ND	ND	38.0
						Hatcher										
2004	ND	ND	ND	10.0	55.2	27.4	24.1	76.7	55.1	8.6	65.0	41.2	ND	ND	ND	46.7
2005	20.0	20.0	20.0	10.8	30.8	20.6	40.0	67.5	59.7	ND	ND	ND	ND	ND	ND	37.9
					т.	F 4 1	0.1	1' F	11.01.	1.0.1						
2004	NID	NID	ND	ND		latchery					1	ND	15.0	(5.0	20.0	20.2
2004 2005	ND ND	ND ND	ND ND	ND ND	ND ND	ND ND	ND ND	ND ND	ND ND	ND 22.0	ND 51.0	ND 36.5	15.0 23.4	65.0 56.0	29.0 39.7	39.2 38.1
2005	ND	ND	ND	ND	ND	ND	ND	ND	ND	38.0	38.0	38.0	ND	ND	ND	38.0
2000	ND	ND	ND	ND	ND	ND	ND	ND	ND	36.0	36.0	36.0	ND	ND	ND	36.0
							Hatche	rv Col	no Salmo	nn -			1			
2004	ND	ND	ND	5.0	6.7	6.1	24.0		32.0	11.7	62.7	42.5	ND	ND	ND	31.7
2005		11.0	11.0	8.0	28.0	18.0	24.1	26.7	25.2	15.8	46.2	31.0	ND	ND	ND	23.1
						Н	atchery	Summ	er Steel	head						
2004	ND	ND	ND	ND	ND	ND	50.0	50.0	50.0	21.9	53.3	31.5	ND	ND	ND	34.1
2005	ND	ND	ND	ND	ND	ND	ND	ND	ND	17.0	41.6	26.2	ND	ND	ND	26.2
2006	ND	ND	ND	ND	ND	ND	46.3	46.3	46.3	16.0	45.8	26.9	14.3	14.3	14.3	27.9
							-		g Salmo							
2004	ND	ND	ND	23.1	23.1	23.1	10.0	54.5	30.9	7.1	75.0	45.2	ND	ND	ND	35.9
2005	0.0	13.8	6.8	17.6	17.6	17.6	50.0	63.4	59.3	68.3	68.3	68.3	ND	ND	ND	38.2
2006	23.3	23.3	23.3	10.7	36.7	26.9	48.1	52.4	50.2	71.4	72.7	72.1	ND	ND	ND	43.6
							~ •									
2004	NID	ND	NID	NID		Natural						NID	2.1	<i>51.</i> 2	22.6	22.6
2004	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND 72.4	ND 56.2	3.1	51.2	22.6	22.6
2005	ND	ND	ND	ND	ND	ND	ND	ND	ND	40.0	72.4	56.2	22.2	37.0	29.4	40.1
2006	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	24.2	43.1	32.2	32.2
						N.	Intural 6	- 	r Staalh	and						
2004	ND	ND	ND	44.0	44.0	44.0	47.8	49.9	r Steelh 48.7	26.9	40.0	34.2	ND	ND	ND	40.1
2004	25.0	25.0	25.0	8.3	16.7	11.4	39.0	54.0	44.1	14.1	47.1	31.5	ND	ND	ND	28.6
2005	26.3	33.3	29.8	31.8		35.9	51.3	61.0	56.2	27.8	42.6	36.7	29.2	29.2	29.2	38.1
2000	20.3	22.2	27.0	51.0	10.0	33.7	31.3	01.0	20.2	27.0	12.0	50.7	27.2	27.2	27.2	20.1
																-

Appendix Table A-5. Egg-to-smolt survival of naturally spawning summer steelhead in the Umatilla River, BY 1993-2004.

Decad	F.	mala Essanama	a	Niah anb	NTI- ou ^C	Essts
Brood		male Escapeme		Number ^b	Number ^c	Egg to
Year	Total	S1	S2	of eggs	of smolts	smolt survival (%)
1993	1102	522	580	6116187	52,010	0.85
1994	806	432	374	4323435	68,162	1.58
1995	866	404	462	4824913	26,295	0.54
1996	1205	881	324	5761557	59,278	1.03
1997	1330	768	562	6969537	46,532	0.67
1998	996	559	437	5267468	83,144	1.58
1999	1082	578	504	5809681	32,573	0.56
2000	1614	1,327	287	7278281	73,177	1.01
2001	2293	1,118	1175	12631251	26,813	0.21
2002	2741	1,767	974	13813433	37,559	0.27
2003^{d}	1937	561	1376	11815091	56,855	0.48
2004 ^d	1589	1,290	299	7214651	31,423	0.44
Mean ^e	1463	851	613	7652124	49,485 ^d	0.77

^a Total no. females available to spawn (hatchery and natural origin).

Appendix Table A-6. Smolt-to-adult return rate (%) of natural summer steelhead in the Umatilla River, 1995-2006.

		Nι	ımber of	returns t	o TMFD ^a	
Outmigration year	Smolt yield	S1	S2	S3	Total	Smolt-to-adult return rate (%)
1995	54,361	614	223	0	837	1.54
1996	73,361	614	426	0	1,040	1.42
1997	22,221	709	282	34	1,025	4.61
1998	59,182	1,858	1,293	0	3,151	5.32
1999	46,530	1,244	979	72	2,295	4.93
2000	81,759	2,642	1,373	0	4,015	4.91
2001 ^b	33,844	672	460	0	1,131	3.34
2002 ^b	77,016	1,641	0	0	1,641	2.13
Mean ^c	50,217	1,249	630	13	1,892	3.53

^a Partitioned by age structure (S1, S2 and & S3) using adult scale analysis.

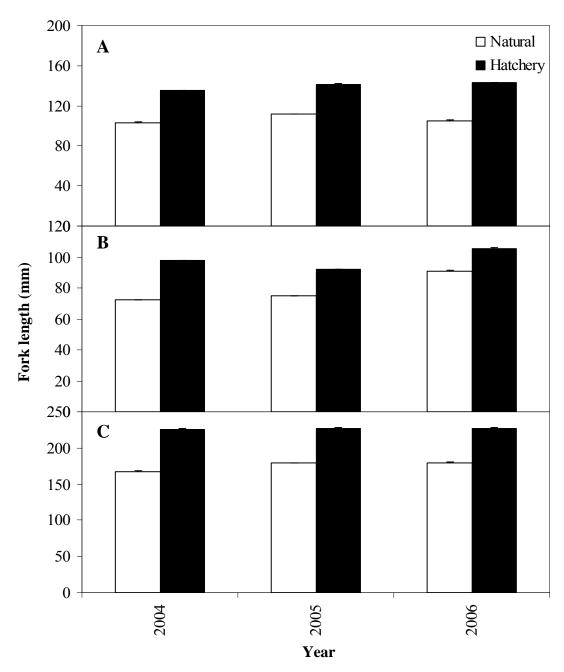
b Mean fecundity of natural females estimated as 3,979 for S1 and 6,965 for S2 fish. c Smolt yield was adjusted for age structure.

^d Dataset incomplete. Estimates do no include age 3 and 4 emigrants.

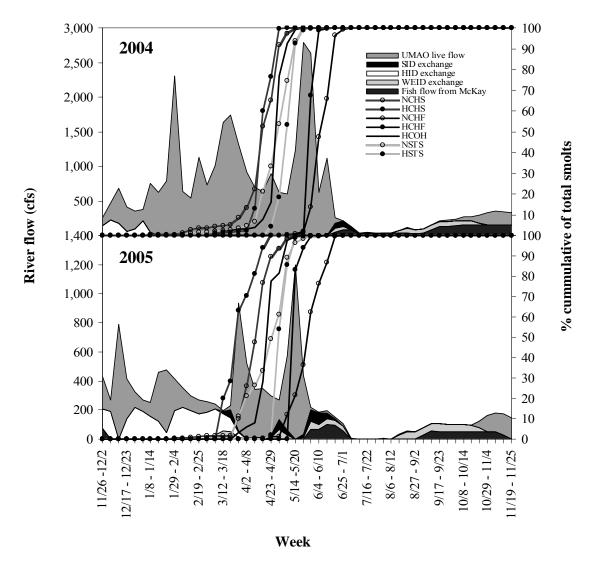
^eIncludes BY 1993-2004

^b Dataset incomplete. Estimates do not include adult return data for 2004 and 2005.

^c Includes smolt outmigration years 1995-2002.



Appendix Figure A-1. Comparison of mean annual fork length (+1 SE) for natural and hatchery spring Chinook salmon (A), fall Chinook salmon (B), and summer steelhead (C), outmigration years 2004-2006. All years were significantly different (P < 0.05).



Appendix Figure A-2. Weekly river flow by proportion of source (water exchange, stored water, or live flow) and cumulative emigration timing of natural (NCHS) and hatchery (HCHS) spring Chinook salmon, natural (NCHF) and hatchery (HCHF0) subyearling fall Chinook salmon, hatchery (HCOH) coho salmon, and natural (NSTS) and hatchery (HSTS) summer steelhead in the Umatilla River, 2004-2005.

APPENDIX B

Three Mile Falls Dam Adult Ladder Detector Design Plans and Specifications

Appendix B-1 Three Mile Falls Dam Design Description Narrative

I Demolition Plan

The demolition work for this project consists of removing two existing concrete weirs – specifically the second and fourth weirs downstream from the fish viewing window. The weirs will be removed using a concrete saw capable of cutting flush with the walls and the floor of the ladder. The concrete weir pieces will then be lifted from the ladder after they have been sawed free.

When the weirs are removed, the remaining fish ladder structure will have reduced strength to resist the implied load related to the soil loading on the outside of the ladder wall to the east. The soil loading effectively imposes a bending moment on the floor slab of the fish ladder. A note is included on the demolition plan requiring the contractor to either remove the soil load by excavation or to shore the east ladder wall temporarily. Once the floor of the ladder is repaired and the pre-cast concrete weirs and cast-in-place concrete cap beam are in place, the new weirs including the cap beam will make the composite fish ladder structure considerably stronger (with regard to resisting the soil load) than the original structure

During operation, the PIT tag antennas sit inside 10 inch deep cavities in the floor of the fish ladder. The electro-magnetic field (EMF) emitted by the antenna will intersect with the floor slab. To prevent the existing steel reinforcing in the floor slab from interfering with the antenna's EMF, the floor of the ladder will be saw cut to remove existing steel reinforcing within the floor within a three-foot radius of the antennas. This distance was stipulated at the initial site visit and technically extends to and beyond the nearest existing ladder wall adjacent to the weir. In this area the concrete removal will end where the floor meets the existing ladder wall. The three foot radius from the PIT tag antenna has been laid out and used to define a rectangular area for concrete removal. The perimeter of the concrete removal area will be saw cut without overcutting at the corners. This will result in more structurally integrity at the reentrant corner of the existing concrete compared to removal where over-cutting at the corners is allowed.

Previous PIT tag projects have revealed that nearby metal grading can be a source of noise for the PIT tag antenna tag detection. Apparently, vibrations, wind, and miscellaneous transient loads on the grating create conditions where electrical contacts are made and then broken repeatedly. If these contacts change the resistance of or result in a new circuit to ground, the EMF field may be significantly affected by the phenomena. To prevent this from occurring at the Three Mile Falls project, we will require the contractor to remove some of the existing steel grating near the antennas and replace it with non-conductive fiberglass grating.

Field photos taken at the project show an existing structural support running longitudinal with the ladder and supporting the walkway grating at mid-span. Unfortunately, there are no as-built drawings showing this member. This member presently passes over the top of the existing weirs.

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This member will need to be modified during construction to accommodate the two new cast-inplace cap beams which will have a sloped finish surface matching the sloped surface of the existing grating. Presumably a section of the beam will be cut out in the field with a cutting torch and a connection to the new CIP cap beam made at the proper location. A detail of this connection is included in the contract drawings with warnig to the contractor that the size of the existing member is not known and the detail should be checked closely to assure that it is viable.

II Floor of Fish Ladder

To simplify the construction of the PIT tag antennas, the sloped floor of the fish ladder will be modified slightly in the immediate vicinity of the antenna to allow for a more uniform and orthogonal antenna geometry. The differential between the original ladder floor and the modified floor is on the order of 1.75 inches.

1/4" x 3" x 2'-8" Fibergrate pultruded fiberglass plate will be used to control the size of the gap between the in-place antenna and the opening in the floor of the fish ladder. The plate will be field trimmed/installed to allow the screws to be located as required to secure the plate such that a 1/8" gap is provided.

The antenna will bear on two blocks of HDPE which will secured to the bottom of the cavity in the floor of the fish ladder. HDPE can be purchased as 1" x 2" bar from McMaster-Carr.

HDPE blocks upstream and downstream of the antenna will also lock the lower end of the antenna in place so that it does not vibrate as the water flows through the opening in the antenna.

III Recess in Floor

Recesses in the floor of the fish ladder were not utilized at the Ice Harbor Dam and Lower Granite Dam PIT Tag projects. The recesses at Three Mile Falls Dam are necessary so that the invert of the antenna opening (the top of the bottom horizontal leg of the antenna) matches the existing slot geometry.

Since the antenna physically intersects the floor of the ladder, a zone of intense EMF (emitted by the antenna) is projected into the concrete floor close to the antenna. For this reason, the steel-reinforced floor must be removed and replaced with glass fiber reinforced polymer (GFRP) reinforced concrete. GFRP reinforcing bars are essentially fiberglass members with strengths and applications similar to steel reinforcing. To accommodate the specified depth at the recess in the floor, the floor of ladder is thickened from 12 inches to 1'-4".

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To prevent the antenna from effectively grinding against the concrete floor or being high-centered on a piece loose grit in the bottom of the recess, HDPE bearings are used on the bottom of the recess to protect the antenna by supporting it slightly above the floor. Similarly, HDPE bearings are used on the vertical surfaces upstream and downstream to support the portion of the antenna cantilevered below the fiberglass guides above the recess. The HDPE bearings can be machined as necessary to make the antenna plumb and true.

To minimize grit and debris that could accumulate in the recess, 1/4" thick fiberglass plate is used upstream and downstream from the antenna to minimize the gap between the concrete and the antenna.

IV Precast Concrete Weir Segments

The precast weir segments will be cast lying flat, presumably at the contractor's fabrication shop. The concrete specified in the drawings is low-slump 4,000 psi 28-day compressive strength concrete. The concrete properties are speced as follows and are based on typical quality concrete readily available locally: W/C ratio 0.45 max, ASTM C-33 aggregate grading number 57 or 67, slump 3-5 inch, and entrained air at 5-6%.

Using precast concrete rather than cast-in-place concrete was utilized very effectively for similar PIT tag work at Ice Harbor and Lower Granite Dams recently. The construction contractor, Knight Construction, reports that this was a very efficient and practical way to construct the weirs and was far superior to casting the walls in-place upright within the confines of the fish ladder.

The vertical reinforcing (vertical when the weirs are in the installed orientation) will be grade 60 steel conforming to ASTM 615. The horizontal reinforcing will be GFRP fiberglass. By having steel reinforcing in one direction and non-conductive GFRP reinforcing in the other, the reinforcing will not create conductive loops and will not have the ability to cause arcing between the horizontal and vertical reinforcing bars when they are exposed to the EMF of the antenna.

The precast weirs – specifically the vertical steel reinforcing - will be anchored to the floor of the ladder using #6 hooks in the new concrete floor and #6 dowels in the existing concrete floor. Lenton interlock couplers cast in the precast weir segments will tie the floor reinforcing to the new weir. (http://www.erico.com/products/InterLock.asp) This was the system used to secure the precast concrete segments at Lower Granite and Ice Harbor dams. Special requirements including testing for continuity to ground are stipulated on the drawings to prevent the steel reinforcing and #6 dowels from forming a circuit to ground with the existing fish ladder floor reinforcing.

The horizontal reinforcing in the precast weir segments will be #4 GFRP reinforcing and is included in the design primarily for shrinkage and crack control. This GFRP reinforcing is designed in accordance with the GFRP manufacturer's guidance including lap lengths and bend

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radii. The reinforcing will be made into the shapes (bends) required at the factory by shaping the fiberglass fibers and then injecting and curing the vinyl ester resin. Hughes Brothers – makers of Aslan FRP fiberglass rebar – has been consulted for the design of the reinforcing for this project.

The forms for new concrete will use 3/4" chamfers at exposed corners. This will make a neat corner and will prevent rough edges in the ladder that could de-scale fish.

The outboard edges of the weirs segments will include a ½" nominal gap to accommodate any irregularities or one-of-plumbness in the existing fish ladder wall and also allow the contractor to precisely place the weir segments while meeting the prescribed clearance & tolerance at the opening between the pre-cast segments at the slot opening. The gap at the existing wall will ultimately be sealed using ½" closed-cell polyethylene backer rod and E6100 sealant. This sealant was specified by a Fisheries person that Tara White contacted. It is a fish-friendly nontoxic, non-sag sealant available from Eclectic Products that adheres to dry concrete and cures in 48-72 hours.

The precast weir segments will require a number of threaded inserts that will be utilized to secure the fiberglass antenna guides to the weir segments at the slot opening. The contractor will precisely and carefully drill holes in the forms and secure the threaded inserts (F-42 Ferrule Inserts) in the proper locations prior to casting the segments. The size and spacing of the inserts is based on accommodating the load case of a complete blockage of the fish ladder at the weir and the blocking material being completely supported by the antenna and or the guides. This is an extremely unlikely situation, so a low factor of safety was used to determine the size and spacing of the threaded inserts.

A number of #6 dowels will be cast into the top edge of the precast weir segments to carry shear from the weir segments into the cast-in-place cap beam. The location of the dowels is detailed to prevent potential interference with the longitudinal reinforcing as well as with the end dowels in the cap beam.

V Weir Cap Beam

The cast-in-place concrete cap beam on top of the new weirs will be very similar to the cap beams on top of the weirs at the recently constructed Ice Harbor Dam and Lower Granite Dam PIT Tag projects. These projects are mentioned in section 6.0 of the SOW. (Incidentally, these facilities are reported to be the most successful existing PIT tag detection systems administered by PSMFC.)

The cap beam above the modified weirs effectively protects the top leg of the antenna from impacts from floating debris during high water levels. It is conceivable that the replacement weirs could be constructed without the concrete cap beam on top, saving some costs, however the cap beam strengthens the new weirs considerably which will dampen potentially deleterious vibrations and, as mentioned above, the cap beam does provide protection to the antenna. If the

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antenna is constructed so that the top leg of the antenna is beneath the cast-in-place cap beam, a removable section of grating will be fabricated with fiberglass angles and will be lowered into the antenna guide above the antenna. This assembly will provide protection from floating debris in the event that high water over-tops the antenna. The gap between the top of the antenna and the bottom of the cap beam is only a couple of inches. It is unlikely that this gap will pass debris of any size and it is presumed that water will rarely if ever reach this height in the ladder.

The top surface of the cap beam will match the existing steel grating elevation and will slope at 10H:1V. As a minimum, the thickness of the concrete cap beam is as required to provide sufficient beam strength to resist the design head on the weir as well as concentrated dead and live load from the adjacent grating supports.

The top surface of the cap beam could be located immediately beneath the grating with the grating running over the cap beam. This would require more grating as well as openings in the grating to allow access to the antenna. With this in mind, the top of the proposed concrete cap beams will instead match the top of the existing grating.

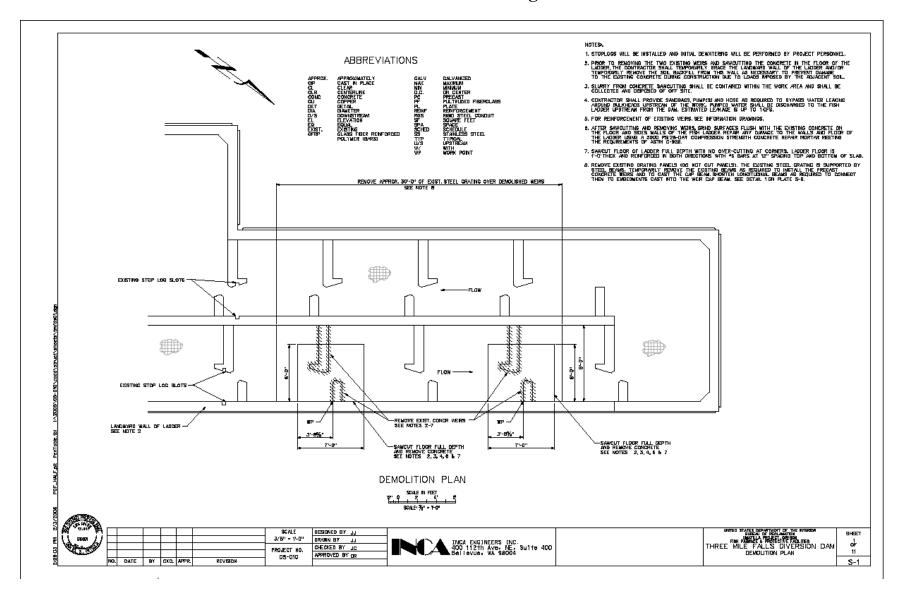
The reinforcement of the cap beam is similar to the Ice Harbor and Lower Granite Dam projects with regular steel bars used for the longitudinal reinforcing and non-conducting GFRP reinforcing used for the transverse reinforcement (hoops). This prevents loops of conductive reinforcing from being created within the EMF field generated by the antenna. The longitudinal reinforcing is doweled into the existing fish ladder walls – which are 12" thick per the as-built drawings. The dowelled bars will be set in epoxy per the Hilti North American Product Technical Guide. Precautions will be taken to prevent the dowelled-in reinforcing from making contact with existing steel reinforcement. This will prevent possible arcing (and resulting EMF noise) at the contact between the new reinforcing and the existing reinforcing.

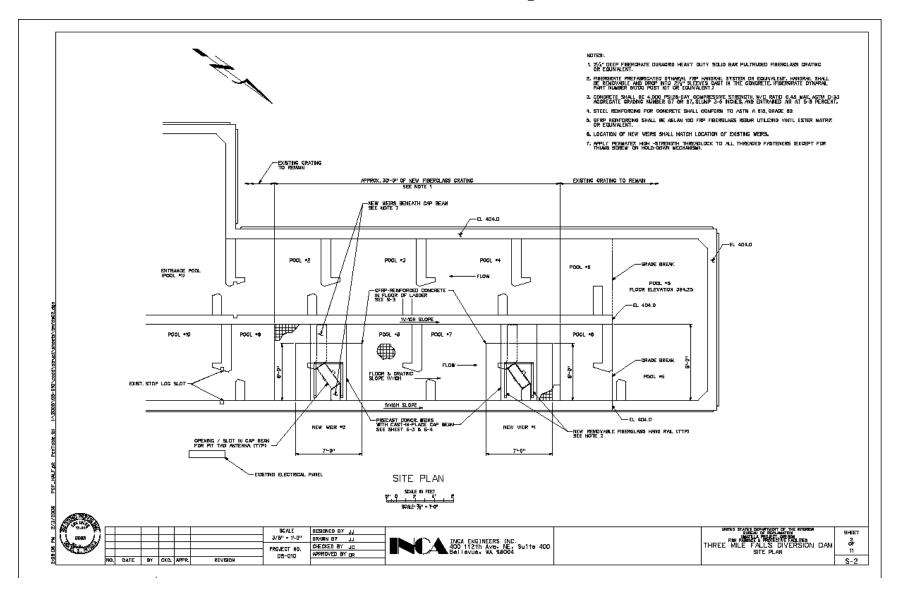
The cap beam will have an opening for the antenna similar to the Ice Harbor and clamps the antenna into the operating position since the antennas are slightly positively buoyant. This clamping effect will also prevent the antenna from moving or vibrating within the guides.

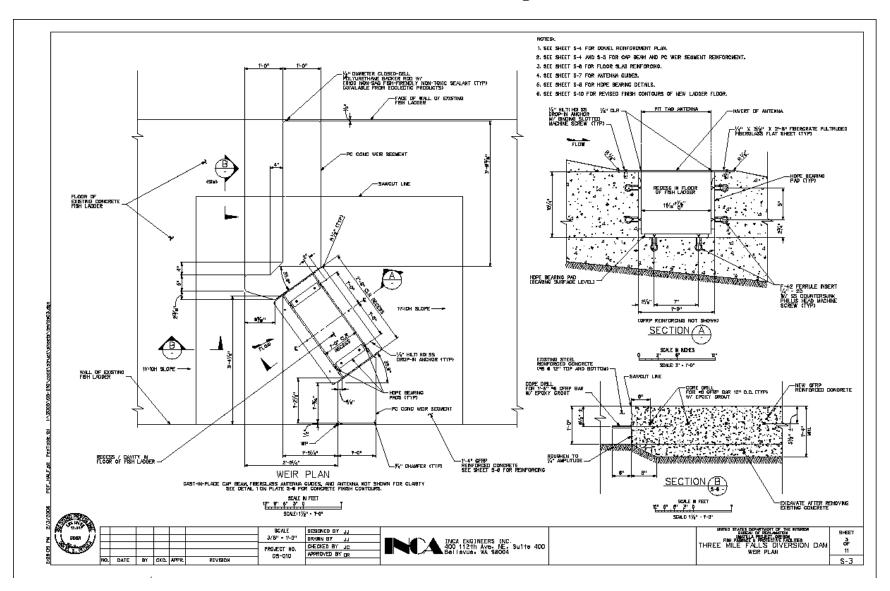
The 4-inch wide slots adjoining the opening for the antenna in the cap beam are intended to allow PSMFC to easily extract the antenna if the need arises without dewatering the ladder. PSMFC (or others) would drop a lifting strap down the 1" x 4" upstream slot, the flow will push the strap beneath the top leg of the antenna, and the strap would be snagged using a wire hook (ie. coat-hanger) and pulled up through the 3" x 4" downstream slot.

The openings in the cap beams for the antennas do not require covers like what is in place at the Ice Harbor and Lower Granite Dam projects. Personnel do not need to be able to walk directly above the antenna since at Three Miles Falls Dam the entire ladder area is covered with grating. The opening as well as the cables from the antenna (a tripping hazard) is protected through the use of removable fiberglass handrail surrounding the opening.

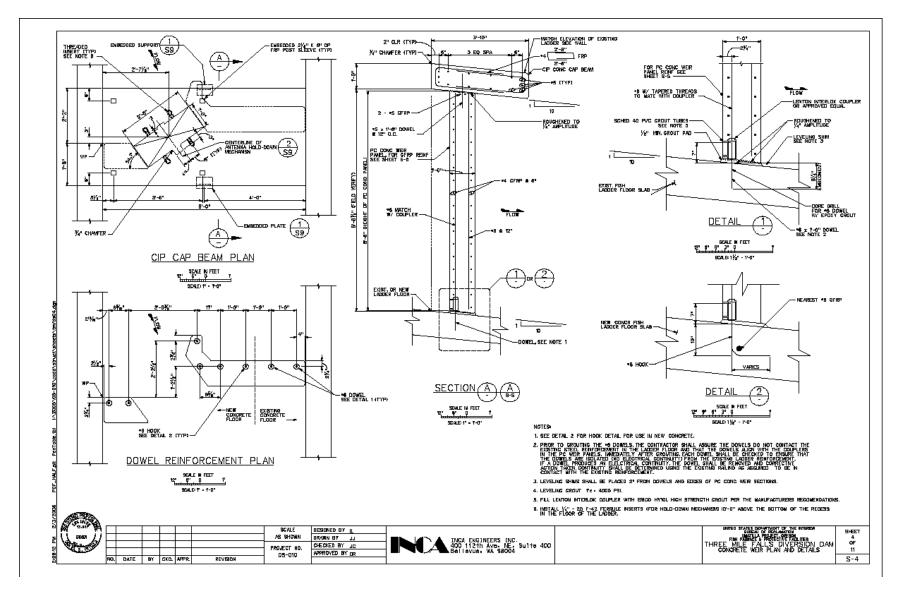
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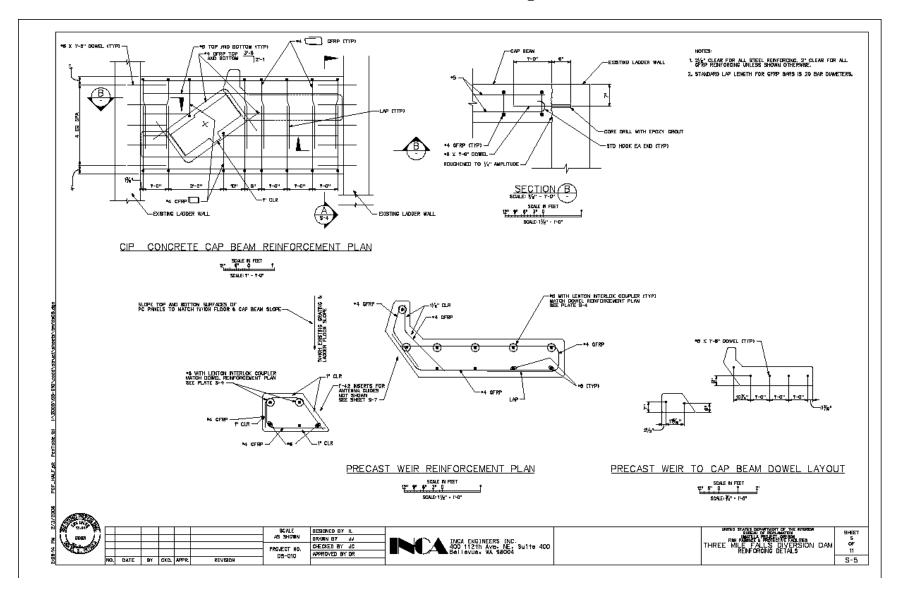


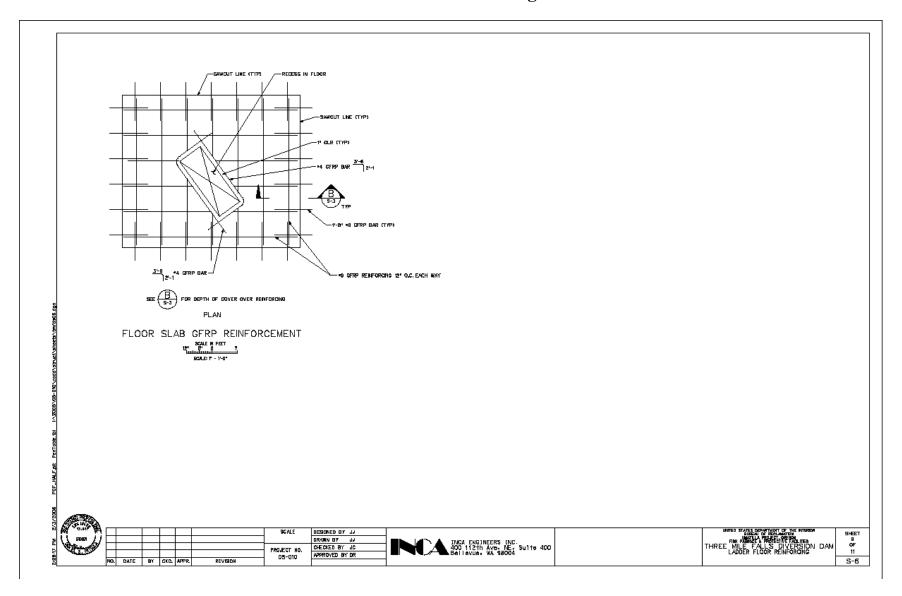




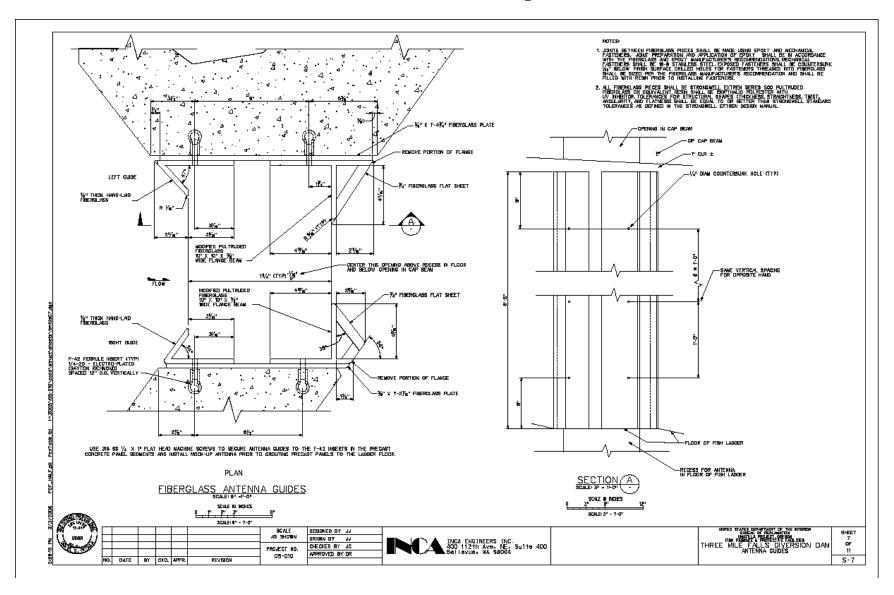
Appendix B-2 Construction Drawings

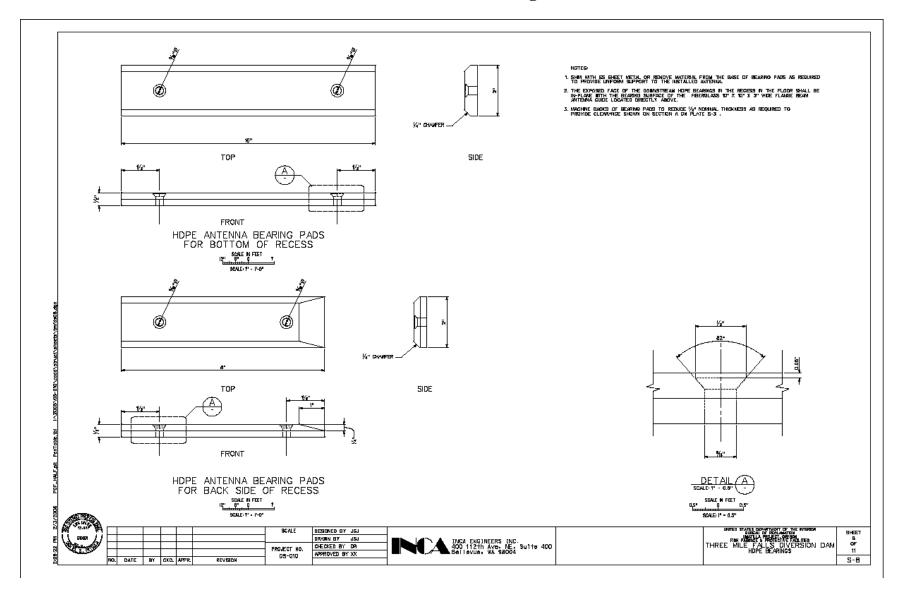


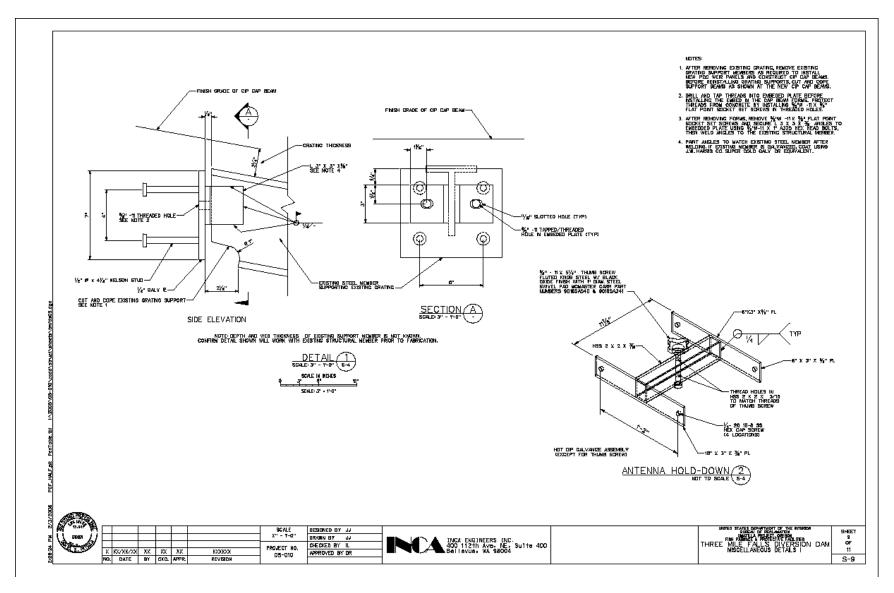


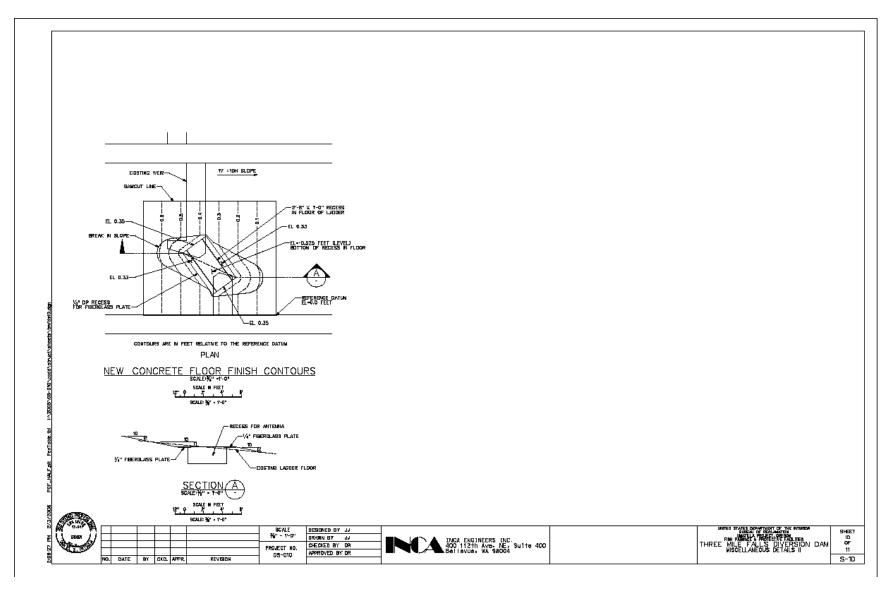


Appendix B-2 Construction Drawings

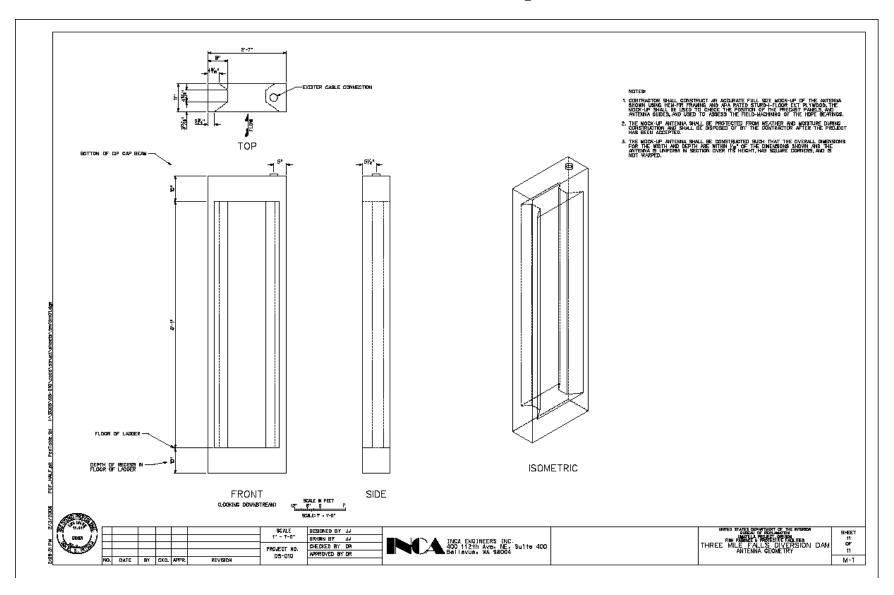








Appendix B-2 Construction Drawings



Appendix B-3

Three Mile Falls Diversion Dam Ladder Modifications for PIT Tag Detection

Calculations

prepared for:

Oregon Department of Fish & Wildlife

Fisheries Research 80866 Hwy 395N Hermiston, OR 96838

prepared by:

INCA Engineers, Inc. 400 - 112th Ave. SE, Suite 400 Bellevue, WA 98004 (Inca Job No.: 05-010)

February 206



CERTIFICATE OF ENGINEER

The work contained herein was prepared under the supervision and direction of the undersigned.



Daniel S. Patsula, P.E. Vice President

ENGINEERS INC. Contract/Client				Job No05 - 0/0 Vol./Sheet No of
Phase/Subject				
Design Topic			 	
		ecked By	 _ Date _	Page No

THREE MILE FALLS DAM. PIT TAG

1. DESIGN DATA: DI-D2

2, CIP CAP BEAM 1-8

3. WEIR 9-17

4. REFERENCES RI-RIO

DI **Brian Twitchell**

⊏rom:

ent:

To: Cc: Subject:

Jeff Juel [j.juel@incainc.com] Wednesday, September 21, 2005 8:42 AM Brian Twitchell Dan Russell (E-mail) 05-010 - Three mile falls dam fish ladder grating loading criteria

Brian,

The cap beam on the new weirs may/will need to support the end of a steel beam at mid-span. The existing steel beam supports the adjacent grating. I found a reference to the walkway grating in a document this morning. The design load for the grating is 150 psf uniform live load or 1000 lb point load.

We need to consider this loading when designing the cap beam. See $\mbox{\it me}$ if you have questions.

Jeff

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			Sheet No. 1 Job Number 05-010	of1
INCA ENGIN TITLE		Falls Dam PIT Tag		D2
SUBJECT	Water Level	in Fish Ladder		
MADE BY	JSJ	DATE 9/21/2005	CHECKED BY	DATE

From Designer's Operating Criteria dated 06-17-94

	Headwater	Tailwater
Flow Condition	Elevation	Elevation
High Flow	405.47	395.23
Winter High Flow	405.12	395.06
Low Flow	404.78	394.79
No Flow	404.30	394.67

Assume that the water level in the fish ladder is defined by an equal head drop at each of the project fish ladder's ten pools and that the water levels at the entrance pool and exit pool match the tailwater and headwater elevations respectively.

Flow Condition	Headwater Elevation	Tailwater Elevation	Total Head Drop	Drop per Pool
High Flow	405.47	395,23	10.24	1.024
Winter High Flow	405.12	395.06	10.06	1.006
Low Flow	404.78	394.79	9.99	0.999
No Flow	404.30	394.67	9.63	0.963

Pools #8 and #10 are immediately upstream from the two proposed new weirs.

	Water	r Level
Flow Condition	Pool #8	Pool #10
High Flow	402.398	405.47
Winter High Flow	402.102	405.12
Low Flow	401.783	404.78
No Flow	401.411	404.30

Also of use is the floor elevation in the fish ladder at the proposed new wiers. The floor slopes at 1V:10H, and the floor elevation at the 180 degree turn in the ladder is 394.25.

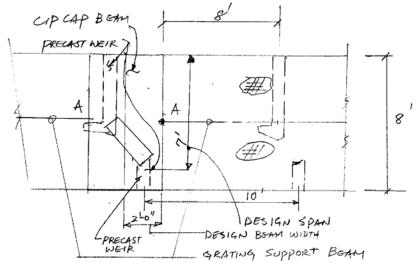
The ladder pools are 10 feet long and the weirs to be replaced are 10 feet and 30 feet from the toe of the floor slope.

Floor elevations at new weirs ---> 395.25 and 397.25

	Water Depth		
Flow Condition	Pool #8	Pool #10	
High Flow	7.15	8.22	
Winter High Flow	6.85	7.87	
Low Flow	6.53	7.53	
No Flow	6.16	7.05	

(File, New, Calc Sheet.xit)

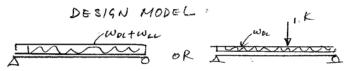
ENGINEERS INC. Contract/Client THREE MILE: FALLS DAM PIT TA	Job No. 05-010 Vol./Sheet No of
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Design Topic CIP CAP BEAM Made By IL Date 9-29-05 Checked By JC D	Date 10/6/05 Page No. /



GRATING DESIGN LOAD, 150 PSE OR 1000 K ASSUME CAP BEAM THICKNESS = 1.5

BASICALLY, CAP BEAM IS SUPPOPED BY PRECAST WEIRS VERTICALLY. IT IS DOWELLED INTO WALLS TO SUPPORT UNBALANCED LOADS.

ASSUME DESIGN SPAN LENGTH = 7'
DESIGN WIDTH = (3,833)/2 = 1,92' x2'



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DEAD LOAD

GRATING DEAD LOAD

ASSUME TRIBUTARY WIDTH = 9/24,5 (CONSERV.)

WEG = 0.005 x 4,5 = 0.023 K/,

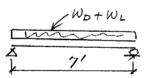
CAP BEAM DEAD LOAD :

LIVE LOAD

- · UNIFORM ON GRATING & CAP BEAM (150 PSF)

 WL = . 15 (2'+8') = 0,9 */,
- · IF CONCENTRATED LOAD, IK.
 LOCATES AT MIDDLE OF SPAN.

CASE 1.



$$W = W_0 + W_L$$

$$W_u = 1.4(0.023 + 0.30) + 1.7(0.9)$$

$$= 0.45 + 1.53$$

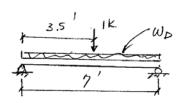
$$= 1.98 \frac{k}{1}$$

$$M_u = \frac{w_u \ell^2}{8} = \frac{1.98 \times 7}{8} = 12.13$$

$$V_u = \frac{1.98 \times 7}{2} = 6.93 \frac{k}{1}$$

$$G_0 VERNS$$

CASE 2



$$W_{u} = 1.4 (0.023 + 0.30) = 0.45 \%$$

$$P_{u} = 1.0 \times 1.7 = 1.7 \times$$

$$M_{u} = \frac{0.45 \times 7^{2}}{8} + \frac{1.7 \times 7}{4} = 2.76 + 2.98 = 5.74 \%$$

$$V_{u} = \frac{0.45 \times 7}{2} + 1.7 = 3.28 \times$$

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ENGINEERS INC.	Vol./Sheet No of
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Made By /L Date 9-29-05 Checked By Date Date	10/6/02 PAGE NO 4
REINFORCEMENT: (FOR RIGHT HALF ONLY 9" IS EFFECTIVE	, 2'WDE)
DESIGN REINFORCEMENT WITH HYDRA	tulic FACTOR 1,3
Mu = 1.3×12,13 = 15.77 & 16 1K	
Tu= 1,3× 6,93 = 9	
H= 12"	
d = 7.75''	
fe'= 4 KSI fy = 60 KSI	//
SPACE TO PUT REINFORCE MENT ABOUT	T 9" (FLEX.)
By CHART AS = 0.5 M 1.	# / 6/2 2-#3
ACI Asmin = = = = bind = 3 4000 ×9 ×7.75 =	0,22 IN /FT O.K
TGMP. STEEL. As = 0.0018 Ac= 0.0018x 12 x 12 =	
SHEAR = THY #(E) @ 12" DOWEL S	HEAR FRICTION
Vn = Av fy M USE #6 DOWEL	T TWITE COM
11- AVT9 ME TO DOWEL	. 10 Existing Conc.
= 0,31×60×(0,6×) NOT INTER	HONAL KOUGHENGD
= 0,31×60×0,6 = 11,16 E	
\$ Vn = 0.85 × 11.16 = 9.5 × > Vu	0, 5

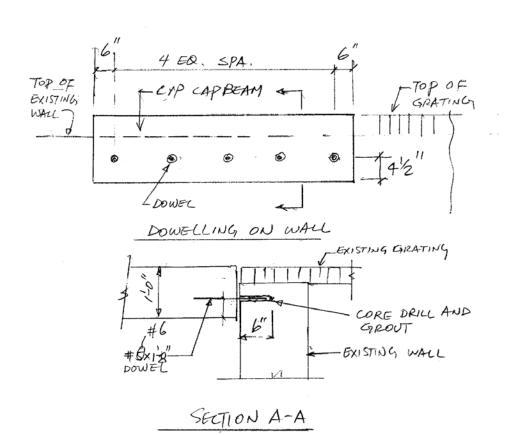
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EMBEDDED LONGTH FOR HORIZONTAL DOWEL.

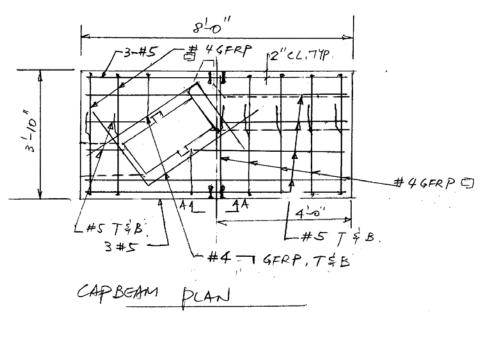
6" EMBEDMENT IS ADEQUATE

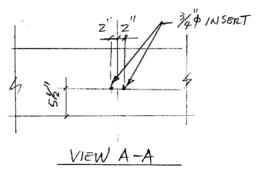
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Contract/Client THREE MILE FALLS DAM PT TAG
Phase/Subject
Design Topic OP CAP BEAM Made By 1 Date 10-3-05 Checked By 0 Date 19/6/05 Page No. 6
Page No
CHECK VERTICAL SHEAR: Vuinax = 6,93 K
Vuinax = 0,13
PVc = 2 \$ feb = 2x0,85 \ 4000 = 108 PS1
D=MIN WIDTH AT OPENING FOR INSTRUMENT = 9"
Ac = b · d = 9 × 9,5 = 85,5 1N2
\$ \(\bar{V}_{n} = \frac{108 \times 85.5}{1000} = 9.23 \times > \bar{V}_{u max} 0. K
(: Provide Min Shear Steel) < 2 Vin n.G
ADD ONE #5 EACH SIDE AS SHEAR FRICTION
OPENING
#5 SHEAR FRICTION REPAR

ENGINEERS INC. Contract/Client 7HRBE MILE FALLS DAM PIT	Job No. <u>05 - 010</u> Vol./Sheet No of TAG
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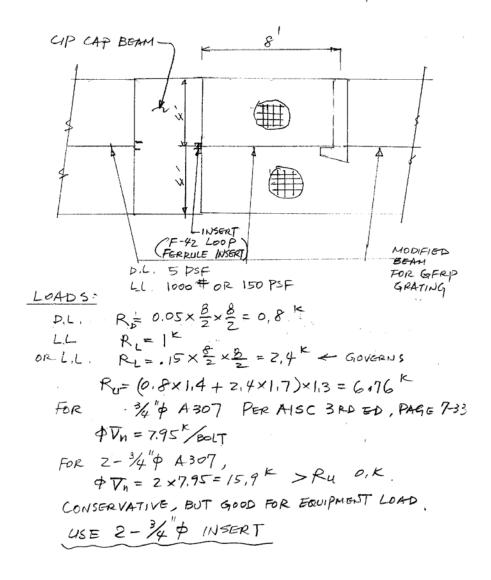
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SIZE OF INSERT FOR BEAM ATTACHING



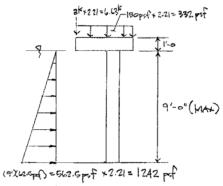
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Contract/Client THREE MILE FALLS DAM PIT TAG		_
hase/Subject DESIGN LOADS / ASSUMPTIONS		_
Design Topic		_
Aade By BRUS Date 9/05 Checked By 14 Date	9-30-05 Page No. 10	

DESIGN ASSUMPTIONS

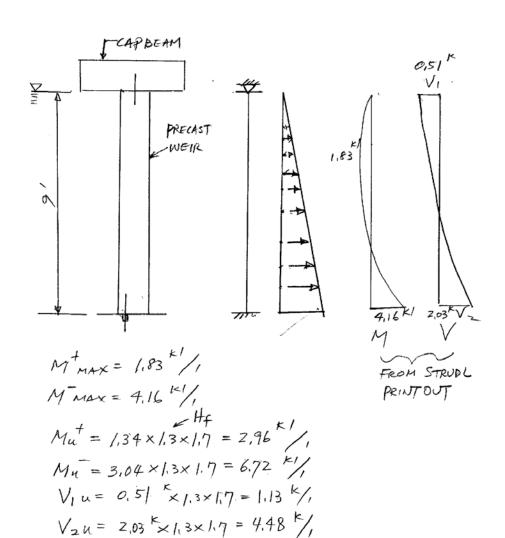
- · MAHMUM WEIL HEIGHT = 9'-0" (TO BOTTOM OF NEW CIP CONC. WILLOWAY)
- · DESIGN FOR FULL HYDROCTATIC PRESCUE AN ONE SIDE (WEIL IS BLOCKED)
- · WEIR THICKNESS = 12"
- · NEW CIP CONC. WALKWAY THICKNESS = 12"
- · FISH LADDER COVERED WITH GRATING (150 pef LL or 1,000 LL)
- · BEAM AT CENTER OF WEIR TO SUPPORT GRATING
- ATTACH BELLY TO NEW CIP CONE. WHICKING!

 BY STEAR LOTO

 (4'X5'X150psf)= 3k STEAR LOTO
- ·WEIR FIXED TO FISH LADDER FLOOR; AN CONNECTION AT WHICHWAY
- · DESIGN WITH SINGLE LOAD FACTOR METHOD = My = (1.3) 1.7) M = 2.21M



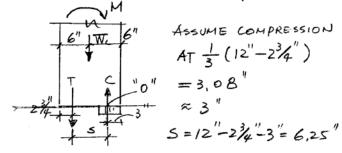
ENGINEERS INC. Contract/Client THREE MILE FALLS DAM PIT TA	Job No of of
Phase/Subject Design Topic WETR	
Made By 14 Date 9-29-05 Checked By VV	Date 10/6/95 Page No 1/



	Job No. 05 - 0/0 Vol./Sheet No of
ENGINEERS INC. Contract/Client THREE MILE FALLS DAM PIT TAG	VOLUMENTO.
Phase/Subject	
Design Topic WETR	2/1-/
Made By/L Date _9-29-05 Checked By Date	10/6/05 Page No. /2
DESIGN FOR ML = 2,96 /FT	
h=12"	
d = 9.5'' $f_c' = 4 \times 51$ $f_n = 60 \times 51$	/s ² /
te = 4 " ty = 60 " s" As = Asmin = 200 bwd/fy = 200x #6	12×9.5/60,000 = 0,38'
#6 @	e 12"
DESIGN FOR Ma = 6,72 K/,	
h = 12''	
$d = 9.5''$ $fc' = 4^{ks1} fy = 60^{ks1} As yeq$	1 = 0,25 /N2 (CHART)
As = Asmin = 200 bud/fy = 0,38 1/1	
#6@12"	
DESIGN FOR VIU,	A
FORCE SMALL, USE #5 @ 12"	Lower
DESIGN FOR VZU = 4.48 ASSUME	
\$ \forall \forall = 0,85 \text{\$1.44-0.25} \text{\$1.44-0.25} \text{\$2.5}	>> Vzu 0, K
As REQUIRED — FOR FLEXURAL	

ENGINEERS INC. Contract/Client THREE MILE FALLS DAM PLT TAG	Job No of of
Phase/Subject	
Design Topic WETRS Made By Date 9-30-05 Checked By Date Date	10/6/05 Page No. /3
DOWELING TO EXISTING CONCRE	TE

1. TENSION FORCE DUE TO BENDING:



WEIGHT OF WEIR = WE = .15 × 1 × 1 × 9 = 1.35 / FT (IGNOR CAP)

$$M_{u} = 6.72$$

$$d = 12'' - 2.75'' = 9.25''$$

$$Q = \frac{.44 \times 60}{0.85 \times 40 \times 12} = 0.65''$$

TENSION REINFORCING REQUIRED,
$$As_{1eg'd} = \frac{6.72x|2 - |.35(6"-3")}{0.9x 60 \times (9.25 - \frac{0.65}{2})} = 0.1521N^{2}$$

#6 IS PROVIDED, As = 0.44 IN

EXCESSIVE REINFORING WILL RESIST SHEAR FORCE

2. SHEAR FORCE

3. TOTAL ASTER'S AVINGE 0,24 IN CAN = 144

EMBOD, LONGTH, $\ell = (\frac{\rho_1 24}{0.44}) \times 15'' = 8.18'' \times 8'4''$ BRIDGE MANUAL

ENGINEERS INC. Contract/Client THREE MILE: FALLS	DAM	Job No of of OF TAG
Phase/Subject		,
Design Topic WEIR		
Made By 14 Date 9-30-05 Checked By	de	Date 10/6/ page No. 14

FROM CATALOG TRW NELSON STUD

EMBEDMENT PROPERTIES OF HEADED STUDS

TABLE 15 FULL EMBEDMENT SHEAR CAPACITY

OF HEADED ANCHORS

34" \$ × 636 STUD FEQUIVLENT TO #6

Suc = 19.99 K FOR fe = 3 KS) >> Vu = 23,3 K FOR fe = 4 KS) = 4.48 K

(For comparisons

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Fri Sep 30 09:30:02 2005

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Reading password file C:\Program Files\GTStrudl\28\password28.pwd CI-i-audfile, Command AUDIT file FILE0930.aud has been activated

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*** G T S T R U D L ***
RELEASE DATE
January 2005
                    VERSION
28.0
                                        COMPLETION NO. 4643
**** ACTIVE UNITS - LENGTH WEIGHT ANGLE TEMPERATURE TIME
**** ASSUMED TO BE INCH POUND RADIAN PAHRENHEIT SECOND
```

av and proprietary to the GTSTRUDL OWNED BY AND PROPRIETARY TO THE GEORGIA TECH RESEARCH CORPORATION RELEASE DATE VERSION COMPLETION NO. January 2005 28.0 4643 ·

**** ACTIVE UNITS - LENGTH WEIGHT ANGLE TEMPERATURE TIME
**** ASSUMED TO BE INCH POUND RADIAN FAHRENHEIT SECOND { 10} > UNIT KIP PT SEC RADIAN (11) > JOINT COORDINATES (12) > 1 0 0 0;2 0 -9 0 (13) > STATUS SUPPORT JOINS 1 2 (14) > 5 (15) > JOINT RELEASE (16) > 1 MOMINT Z (17) > 5 (18) > TUPE SPACE PRANE (19) > MERGER INCIDENCE (20) > 1 1 2

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21] > S
22] > UNIT INCH
23] > MEMBER PROPERTIES PRISMATIC
24) > 1
25] > $
25] > S
24) > 1
25] > E 3605 ALL MEMBERS
27] > E 3605 ALL MEMBERS
28] > FOR 15 0.2 ALL MEMBERS
29] > UNIT FT
30] > DENDITY 0.15
31] > SINACTIVE MEMBERS 51 to 53
32] > $
33] > DEAD LOADING 'DL' 'STRUCTURE WEIGHT' DIRECT -Y ALL MEMBERS
34) > $
35] > UNIT KIPS FEET
36] > LOADING 'MYDEO' '10 PT WATER PRESSURE'
37] > MEMBER LOAD
38] > 1 FORCE X GLOBAL LINEAR WA 0.0 WB 0.563 LA 0.0 LB 9.0
39] > $
39] > $
31] > $
31] > STOMM LOADING 'TOTAL-1' 'UNFACTORED (DEADL + UNIF)' FROM -
42] > $
42] > $
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43] > STUPPMESS ANALYSIS
44] > $
45] > STUPPMESS ANALYSIS
45 POR MEMBERS 0.00 SECONDS
46 FOR COMSISTMENTY CHECKS FOR 1 MEMBERS 0.00 SECONDS
( 42) > 5 TIPPNESS ANALYSIS
THE FOR CONSISTENCY CHOCKS FOR 1 MEMBERS
THE FOR CONSISTENCY CHOCKS FOR 1 MEMBERS
THE TOR BEARWHICHT REDUCTION
THE TO GENERATE 1 ELEMENT STIF, MATRICES
THE TO ASSESSEL THE STIPPNESS MATRIX
THE TO PROCESS 2 JOINTS
THE TO PROCESS 2 JOINTS
THE TO PROCESS 2 JOINT DISPLACEMENTS
THE TO PROCESS 2 JOINT DISPLACEMENTS
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441} >
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           E FOR FRANCIS CHECK

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46| > 5

46| > SUNITINCH

47| > SOUTHIT ORDERED

48| > SOUTHIT ORDERED

48| > SOUTHIT DECIMAL 4

59| > S. IST DISPLACEMENT JOINTS ALL

50| > $

51| > SLOAD LIST 'BARGE' 'TOTAL'

52| > UNIT FT KIPS

53| > OUTHUT DECIMAL 2

54| > OUTHUT BEY NEMBER

55| > LADA LIST 'HYDRO'

56| > SECT FR DS 0.0 0.1 MEMB 1

57| > LIST SECTION FORCES MEMBER 1
             .........
            *RESULTS OF LATEST ANALYSES*
              PROBLEM - NONE
                                                                            TITLE - NONE GIVEN
              ACTIVE UNITS PEET KIP RAD DEGF SEC
               INTERNAL MEMBER RESULTS
                 MEMBER SECTION FORCES
                                                                    ---
                            LOADING HYDRO
                                                                                                      10 PT WATER PRESSURE
                                                                 Z BENDING
                                                      AXIAL Y SHEAR

0.0000000E+00 0.9067001
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0.00000008+00 -0.5852383

0.00000008+00 0.2432168

0.00000008+00 1.436496

0.00000008+00 3.040203
                0.900
 ( 58) > LIST REACTIONS
```

RESULTS OF LATEST ANALYSES

PROBLEM - NONE TITLE - NONE GIVEN

ACTIVE UNITS FEET KIP RAD DEGP SEC

RESULTANT JOINT LOADS SUPPORTS

JOINT	LOAI	ING /X	Y PORCE	z force	X MOMENT	TMEMON Y	Z MOMENT
1	GLOBAL HYI	ORO -0.51	0.00	0.00	0.00	0.00	0.00
2	GLOBAL HYI	DRO -2.03	0.00	0.00	0.00	0.00	3.04
{ 6	99) > SLOAD LIST 50) > 51) > 51) >	ALL					



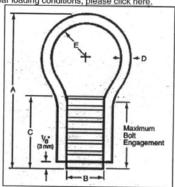


F-42 Loop Ferrule Insert

Several features make the F-42 Insert particularly useful to the precaster who must make provision for the anchorage of wall panels, the suspension of ceilings, sprinkler systems, plumbing and heating pipes and ductwork, etc.

The end of the ferrule opposite the tapped end is closed and will not permit entry of fresh concrete. The wire loops provide anchorage suitable for the bolt size required.

For insert capacities under shear loading conditions, please click here.



Nominal Bott Diameter and	Safe Working Load	Minimum Edge	Maximum Bolt	Va '9'	1.	Dimensions		
Threads per Inch	Tension		Engagement	. A	В	C	D	E
W* -20	650 lbs. 2.9 kN	5" 125 mm	% 19 mm	2¾* 70 mm	3/16" 14 mm	11/4" 32 mm	243° 6.2 mm	1/2" 13 mm
36" - 16	2,000 lbs. 8.9 kN	5" 125 mm					.243° 6.2 mm	
1/2" - 13	2,000 lbs. 8.9 kN	5° 125 mm					243' 6.2 mm	
56"-11	2,400 lbs. 10.7 kN	5" 125 mm	11/6" 29 mm	3½* 89 mm	7/4" 22 mm	15%* 41 mm	262* 6.7 mm	34° 19 mm
94" - 10	2,400 lbs. 10.7 kN	5" 125 mm	11/6" 29 mm	3½° 89 mm	Date Street Control of the Street Control of	Total Control of the	.262° 6.7 mm	-
₩-0	5,300 lbs. 23.6 kN	8" 200 mm	1%° 29 mm	6° 152 mm				
1*-8	5,300 lbs. 23.6 kN	8" 200 mm	11/4° 29 mm	6" 152 mm	136" 35 mm	15%* 41 mm	3751 0.5 mm	11/4 OO ma

SWL's are based on 1/2* (13 mm) setback from face of concrete.

SWL's provide a safety factor of approximately 3 to 1 in 3,000 psi (21 MPa) normal weight concrete.

Specify: (1) quantity, (2) type, (3) bolt dia.,(4) finish.

Example: 1,000 pcs., F-42 Loop Ferrule Insert, 1/2" (13 mm) dia., electro-plated.

DB Grout Sieeve | Formliners | General/Technical Info | Lifting Systems | Lifting/Handling Inserts | Misc Inserts | Misc Products | NC Threaded Inserts | Product Application Info

http://www.daytonrichmond.com/products/precast/f42.html

9/23/2002

Tension Development Length R 2 of Straight Uncoated Deformed Bars



Table 5.1-5 - Tension Development Lengths, Is for Grade 60 Uncoated Bars for = 3.000 psi; Normal Weight Com

f'c = 3,000 psi; Normal Weight Concrete					
	1	ď	Bar Spa Article	ified for ucing Per 8.25.3.1 ng ≥ 6")	
Bar Size	Top Bars Ft In.	Others Ft In.	Top Bars Ft In.	Others Ft In.	
3	1′-5″	1′-0″	1'-2"	1'-0"	
4	1′-5″	1′-0″	1'-2"	1'-0"	
5	1′-9″	1′-3″	_1'-5"	1'-0"	
6	2′-3″	1′-8″	1'-10"	1'4"	
7	3′-1″	2′-3″	2′-6″	1'-9"	
8	4'-1"	2'-11"	3'-3" 7	2'-4"	
9	5'-2"	3′-8″	4'-1"	2'-11"	
10	6'-6"	4′-8″	5'-3"	3'-9"	
_11	8′-0″	5′-9″	6'-5" (4'-7"	
₹14	10'-11"	7′-10″	8′-9″	6'-3"	
18	14'-1"	10'-1"	11'-3"	8'-1"	

Table 5.1-6 — Tension Development Lengths, & for Grade 60 Uncoated Bars

f's #4,000 psi; Normal Weight Concrete

I = 4,000 psi; Normal Weight Concrete					
	la		Bar Spa Article	fied for cing Per 8.25.3.1 ng ≥ 67	
Bar Size	Top Bars Ft In.	Others Ft In.	Top Bars Ft In.	Others Ft In.	
3	1′-5″	1'-0"	1'-2"	1′-0″	
- 4	1'-5"	1'-0"	1'-2"	1'-0"	
, 5	1′-9″	1′-3″	1′-5″	30505	
6	2'-2"	1'-6"	1'-9"	1'-3"	
7	2′-8″	1'-11"	2'-2"	(1'-7")	
8	3'-6"	2'-6"	2'-10"	2'-0"	
9	4'-6"	3′-2″	3′-7″_	2'-7"	
10	5′-8″	4'-1"	4′-6″	3′-3″	
11	6'-11"	5'-0"	5'-7"	4'-0"	
14	9′-5″	6′-9″	7'-7"	5′-5″	
18	12′-3″	8′-9″	9'-9"	7'-0"	

Table 5.1-7 - Tension Development Lengths, & for Grade 60 Uncoated Bars f'c = 5,000 psi; Normal Weight Concrete

ld Modified for Bar Spacing Per Article 8.25.3.1 Ы (Spacing ≥ 6") Top Bars Top Bars Ft. - In. Bar Others Others Size Ft. - In. Ft. - In. Ft. - In. 3 1'-5" 1'-0" 1'-2" 1'-0" 4 1'-5" 1'-0" 1'-2" 1'-0" 5 1'-9" 1′-3″ 1'-5" 1'-0" 2'-2" 1'-6" 6 1'-9" 1′-3″ 7 2'-6" 1′-9″ 2'-0" 1'-5" 8 3′-2″ 2'-3" 2'-6" 1'-10" 9 4'-0" 2'-10" 3'-2" 2'-4" 10 5'-1" 3′-8″ 4'-1" 2'-11" 6'-3" 4′-5″ 5'-0" 11 3'-7" 14 8′-5″ 6'-1" 6'-9" 4'-10" 18 10'-11" 7'-10" 6'-3"

Table 5.1-8 — Tension Development Lengths. 4 for Grade 60 Uncoated Bars

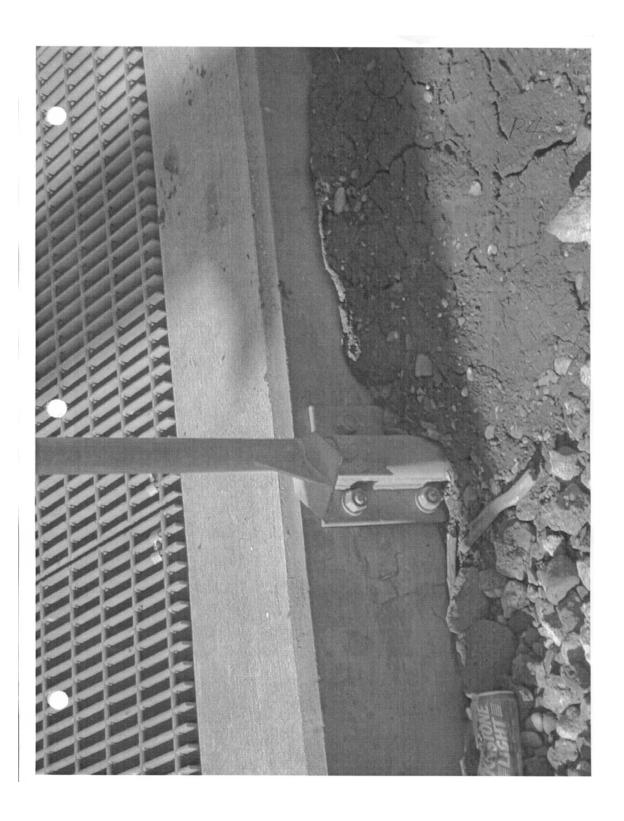
	$f_c = 6,000 \text{m}$	si; Normal V	Veight Concre	ete 🧍
	la .		Bar Spa Article	fied for cing Per 8.25.3.1 g ≥ 6")
Bar Size	Top Bars Ft In.	Others Ft In.	Top Bars Ft In.	Others Ft In.
3	1′-5″	1'-0"	1'-2"	1′-0″
4	1′-5″	1′-0″	1'-2"	1′-0″
5	1′-9″	1'-3"	1′-5″	1'-0"
6	2'-2"	1′-6″	1′-9″	1'-3"//
7	2'-6"	1'.9"	2′-0″	12-5"
. 8	2.11	2'-1" 2-	6 25-40	1-8"
. 9	3′-8″	2-7-3	2-11	2'-1"
10	4′-8″	3'-4"	3′-8″	2'-8"
11	5′-8″	4'-1"	4′-6″	3'-3"
14	7′-9″	5'-6"	6'-2"	4'-5"
18	10′-0″	7'-2"	8′-0″	5'-9"
	经产品的	Sign strategy	1000	THE RESERVE THE PERSON NAMED IN

Reference AASHTO Standard Specifications for Highway Bridges, Interims 1991, Article 8.25.

5.1-A4









R5

Nelson® embedment properties of headed studs

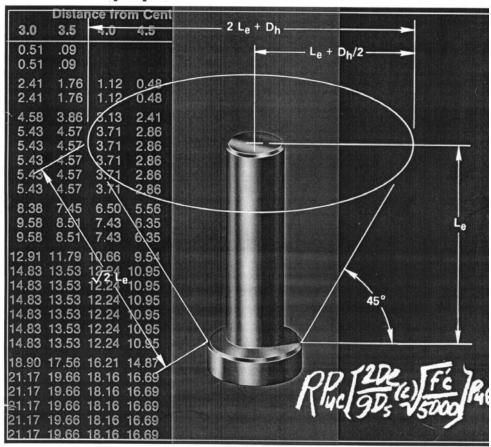


Table 15. Full Embedment Shear Capacities Of Headed Anchors

RL

	A.W. (2.)				C	oncrete Shear C	apacity (Suc) I	Ops	
Anchor (1.)	Length	H/Ds	Sue (3.)		al Concrete (14			ghi Concrete (
Size	(in.)	(No. of Dia.)	(Kips)	1'c = 3000 psi	1'c = 4000 psi	fic = 5000 psi	1'c = 3000 psi	f'c = 4000 psi	f'c = 5000 psi
% × 21%s % × 4%	2% 4	10.25 16.00	2.65 2.65	2.22	2.59 2.59	2.65 2.65	1.85 1.85	2.16 2.16	2.41 2.41
%×4% %×6%	4 6	10.67 16.00	5.96 5.96	4.98 4.98	5.79 5.79	5.96 5.96	4.14 4.14	4.83 4.83	5.41 5.41
% × 2% % × 3%	3 /	4.00 6.00	10.60 10.60	8.87 8.87	10.33	10.60 10.60	7.39 7.39	8.61 8.61	9.64 9.64
%×4% %×5%	5716	8.00	10.60	8.87 8.87	10.33	10.60 10.60	7.39 7.39	8.61 8.61	9.64 9.64
1/2 × 6 1/4 1/4 × 8 1/4	8	12.00 16.00	10.60 10.60	8.87 8.87	10.33	10.60 10.60	7.39 7.39	8.61 8.61	9.64 9.64
% x 21%6 % x 6%s % x 8%s	2½ 6¾ 8	4.00 10.20 12.80	16.56 16.56 16.56	13.89 13.89 13.89	16.19 16.19 16.19	16.56 16.56	11.57 11.57 11.57	13.49 13.49 13.49	15.10 15.10 15.10
7, x 37,6 7, x 317,6 7, x 47,6	3 3½ 4	4.00 4.67 5.33	23.86 23.86 23.86	19.99 19.99 19.99	23.3 23.3 23.3	23.86 23.86 23.86	16.67 16.67	19.42 19.42 19.42	21.73 21.73 21.73
1/4 × 5/16	6	8,00 9,33	23.86 23.8 1	19.99	23.3	23.86 23.86	16.67 16.67	19.42 19.42	21.73 21.73
% × 7 % 8 % × 8 % 8	8	10.67	23.86 23.86	19.99 19.99	23.3 23.3	23.86 23.86	16.67 16.67	19.42 19.42	21.73 21.73
% × 31%,6 % × 4%,6 % × 5%,6 % × 6%,8	3½ 4 5 6	4.00 4.57 5.71 6.86	32.47 32.47 32.47 32.47	27.19 27.19 27.19 27.19	31.69 31.69 31.69 31.69	32.47 32.47 32.47 32.47	22.66 22.66 22.66 22.66	26.41 26.41 26.41 26.41	29.55 29.55 29.55 29.55
% × 7% 6 % × 8% e	7 8	8.00 9.14	32.47 32.47	27.19 27.19	31.69 31.69	32.47	22.66 22.66	26.41 26.41	29.55 . 29.55

NOTES (1.) Stock Anchor Sizes

(3.) Sue = Ultimate Stud Embedded Shear Strength Sue = .9 A₅f₅, where Suc > Sue, Sue Controls

2 Spacing For Development Of Full Shear Capacity.

There are two basic failure modes for studs bject to pure shear forces. In the first, the ncrete capacity exceeds the anchor capacity and failure occurs in the anchor. The second failure mode occurs when the anchor capacity exceeds the concrete capacity. From the Ollgaard, Slutter and Fisher investigation of the concrete and is preceded by localized crushing ahead of the stud, bending in the stud and cracking extending at an angle from under the stud head behind the stud to the concrete — steel interface.

This failure is somewhat different from the large conical type failures that occur in tension loading, and is relatively unaffected by stud length or stud spacing as compared with tension loading.

Spacing to develop full shear capacity is influenced by the following factors with Case B. assuming higher relative importance.

- A. Spacing between anchors in a group or with regard to boundary conditions on anchors without a free edge in the direction of the shear force.
- B. Spacing between anchors and distance from a free edge of anchors at an edge subject to shear force.

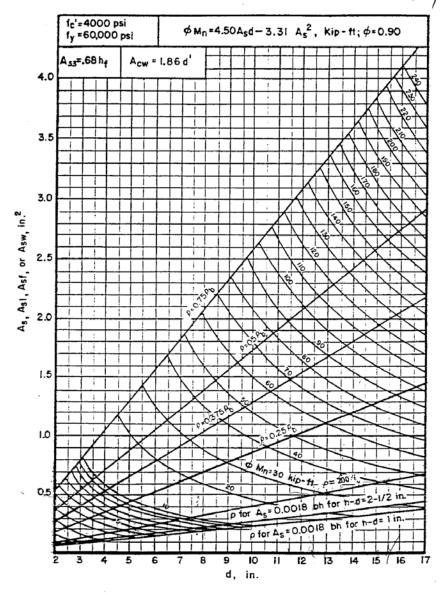
As long as the anchor has no free edge in the direction of the shear force, Case A. applies, and spacings are governed by the H/Ds ratio. A spacing equal to the ratio of 4.0 is satisfactory to develop the full potential shear capacity of a headed anchor. Table 16., which follows, shows the full spacing requirements.

Free edge conditions in the direction of the shear force are covered in Section 5.3.2.

F L E > J R E 8.6.1

FLEXURE 8.6.1-Design moment strength ϕM_n for slab sections 12 in. wide

Reference: ACI 318-83, Sections 7.12, 8.4.1, 8.4.3, 9.3.2, 10.2, 10.3.1–10.3.3, 10.5.1, and 10.5.3, and Commentary Sections 10.3.1(A)(1) and 10.3.2



 Whenever reinforcement continuity is required, lapped splices should be used. The length of lap splices varies with concrete strength, type of concrete, bar grades, size, surface geometry, spacing, and concrete cover. Details of lapped splices should be in accordance with the project specifications. Mechanical connections are not yet available.

6.3—Quality control and inspection

Quality control should be carried out by lot testing of FRP bars. The manufacturer should supply adequate lot or production run traceability. Tests conducted by the manufacturer or a third-party independent testing agency can be used.

All tests should be performed using the recommended test methods cited in the literature. Material characterization tests that include the following properties should be performed at least once before and after any change in manufacturing process, procedure, or materials:

- Tensile strength, tensile modulus of elasticity, and ultimate strain;
- Fatigue strength;
- Bond strength;
- · Coefficient of thermal expansion; and
- · Durability in alkaline environment.

To assess quality control of an individual lot of FRP bars, it is recommended to determine tensile strength, tensile modulus of elasticity, and ultimate strain. The manufacturer should furnish upon request a certificate of conformance for any given lot of FRP bars with a description of the test protocol.

PART 4—DESIGN RECOMMENDATIONS

CHAPTER 7—GENERAL DESIGN CONSIDERATIONS

The general design recommendations for flexural concrete elements reinforced with FRP bars are presented in this chapter. The recommendations presented are based on principles of equilibrium and compatibility and the constitutive laws of the materials. Furthermore, the brittle behavior of both FRP reinforcement and concrete allows consideration to be given to either FRP rupture or concrete crushing as the mechanisms that control failure.

7.1—Design philosophy

Both strength and working stress design approaches were considered by this committee. The committee opted for the strength design approach of reinforced concrete members reinforced with FRP bars to ensure consistency with other ACI documents. In particular, this guide makes reference to provisions as per ACI 318-95, "Building Code Requirements for Structural Concrete and Commentary." These design recommendations are based on limit states design principles in that an FRP reinforced concrete member is designed based on its required strength and then checked for fatigue endurance, creep rupture endurance, and serviceability criteria. In many instances, serviceability criteria or fatigue and creep rupture endurance limits may control the design of concrete members reinforced for flexure with FRP bars (especially aramid and glass FRP that exhibit low stiffness).

The load factors given in ACI 318 are used to determine the required strength of a reinforced concrete member.

7.2—Design material properties

The material properties provided by the manufacturer, such as the guaranteed tensile strength, should be considered as initial properties that do not include the effects of long-term exposure to the environment. Because long-term exposure to various types of environments can reduce the tensile strength and creep rupture and fatigue endurance of FRP bars, the material properties used in design equations should be reduced based on the type and level of environmental exposure.

Equations (7-1) through (7-3) give the tensile properties that should be used in all design equations. The design tensile strength should be determined by

$$f_{fu} = C_E f^*_{fu} \tag{7-1}$$

where

f_{fit} = design tensile strength of FRP, considering reductions for service environment, psi;

C_E = environmental reduction factor, given in Table 7.1 for various fiber type and exposure conditions; and

 f_{fu} = guaranteed tensile strength of an FRP bar defined as the mean tensile strength of a sample of test specimens minus three times the standard deviation (f_{fu}^*) = $f_{u,\sigma v} - 3\sigma$), psi.

The design rupture strain should be determined as

$$\varepsilon_{fu} = C_E \varepsilon^*_{fu} \tag{7-2}$$

where

 ε_{fjt} = design rupture strain of FRP reinforcement; and

guaranteed rupture strain of FRP reinforcement defined as the mean tensile strain at failure of a sample of test specimens minus three times the standard deviation $(\varepsilon^*_{fi} = \varepsilon_{u,ave} - 3\sigma)$.

The design modulus of elasticity will be the same as the value reported by the manufacturer.

The environmental reduction factors given in Table 7.1 are conservative estimates depending on the durability of each fiber type and are based on the consensus of Committee 440. Temperature effects are included in the C_E values. FRP bars, however, should not be used in environments with a service temperature higher than the T_g of the resin used for their manufacturing. It is expected that with continued research, these values will become more reflective of actual effects of environment. The methodology regarding the use of these factors, however, is not expected to change.

7.2.1 Tensile strength of FRP bars at bends—The design tensile strength of FRP bars at a bend portion can be determined as

$$f_{fb} = \left(0.05 \cdot \frac{r_b}{d_b} + 0.3\right) f_{fu} \le f_{fu}$$
 (7-3)



Table 7.1—Environmental reduction factor for various fibers and exposure conditions

Exposure condition	Fiber type	Environmental reduction factor C _E
_	Carbon	1.0
Concrete not exposed to earth	Glass	0.8
	Aramid	0.9
-	Carbon	0.9
Concrete exposed to earth and weather	Glass	0.7
, souther	Aramid	0.8

where

 f_{fb} = design tensile strength of the bend of FRP bar, psi;

 $r_b = \text{radius of the bend, in.;}$

db = diameter of reinforcing bar, in.; and

 f_{fit} = design tensile strength of FRP, considering reductions for service environment, psi.

Equation (7-3) is adapted from design recommendations by the Japan Society of Civil Engineers (1997b). Limited research on FRP hooks (Ehsani, Saadatmanesh, and Tao 1995) indicates that the tensile force developed by the bent portion of a GFRP bar is mainly influenced by the ratio of the bend radius to the bar diameter, r_b/d_b , the tail length, and to a lesser extent, the concrete strength.

For an alternative determination of the reduction in tensile strength due to bending, manufacturers of bent bars may provide test results based on test methodologies cited in the literature.

CHAPTER 8—FLEXURE

The design of FRP reinforced concrete members for flexure is analogous to the design of steel-reinforced concrete members. Experimental data on concrete members reinforced with FRP bars show that flexural capacity can be calculated based on assumptions similar to those made for members reinforced with steel bars (Faza and GangaRao 1993; Nanni 1993b; GangaRao and Vijay 1997a). The design of members reinforced with FRP bars should take into account the mechanical behavior of FRP materials.

8.1—General considerations

The recommendations given in this chapter are only for rectangular sections, as the experimental work has almost exclusively considered members with this shape. In addition, this chapter refers only to cases of rectangular sections with a single layer of one type of FRP reinforcement. The concepts described here, however, can also be applied to the analysis and design of members with different geometry and multiple types, multiple layers, or both, of FRP reinforcement. Although there is no evidence that the flexural theory, as developed here, does not apply equally well to nonrectangular sections, the behavior of nonrectangular sections has yet to be confirmed by experimental results.

8.1.1 Flexural design philosophy—Steel-reinforced concrete sections are commonly under-reinforced to ensure yielding of steel before the crushing of concrete. The yielding of the steel provides ductility and a warning of failure of

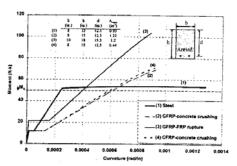


Fig. 8.1—Theoretical moment-curvature relationships for reinforced concrete sections using steel and FRP bars.

the member. The nonductile behavior of FRP reinforcement necessitates a reconsideration of this approach.

If FRP reinforcement ruptures, failure of the member is sudden and catastrophic. There would be limited warning of impending failure in the form of extensive cracking and large deflection due to the significant elongation that FRP reinforcement experiences before rupture. In any case, the member would not exhibit ductility as is commonly observed for under-reinforced concrete beams reinforced with steel rebars.

The concrete crushing failure mode is marginally more desirable for flexural members reinforced with FRP bars (Nanni 1993b). By experiencing concrete crushing, a flexural member does exhibit some plastic behavior before failure.

In conclusion, both failure modes (FRP rupture and concrete crushing) are acceptable in governing the design of flexural members reinforced with FRP bars provided that strength and serviceability criteria are satisfied. To compensate for the lack of ductility, the member should possess a higher reserve of strength. The suggested margin of safety against failure is therefore higher than that used in traditional steel-reinforced concrete design.

Experimental results (Nanni 1993b; Jaeger, Mufti, and Tadros 1997; GangaRao and Vijay 1997a; Theriault and Benmokrane 1998) indicated that when FRP reinforcing bars ruptured in tension, the failure was sudden and led to the collapse of the member. A more progressive, less catastrophic failure with a higher deformability factor was observed when the member failed due to the crushing of concrete. The use of high-strength concrete allows for better use of the high-strength properties of FRP bars and can increase the stiffness of the cracked section, but the brittleness of high-strength concrete, as compared to normal-strength concrete, can reduce the overall deformability of the flexural member.

Figure 8.1 shows a comparison of the theoretical momentcurvature behavior of beam cross sections designed for the same strength ϕM_n following the design approach of ACI 318 and that described in this chapter (including the recommended strength reduction factors). Three cases are presented in addition to the steel reinforced cross section: two sections reinforced with GFRP bars and one reinforced with

RIO

Physical Properties

I. Tensile Stress, Nominal Diameter & Cross Sectional Area, Modulus of Elasticity:

	7	1	7		0100	000000	mi riica,	Modulus	of Elasticit
Bar	Size	Cross Sectional Area		Nominal	Dia.	Tensile	Strength	Modulus of	Elasticity
mm	inches		in2	mm2	in2	MPa	ksi	GPa	psi 106
6	#2	33.23	0.0515	6.35	0.25	825		40.8	5.92
9	#3		0.131	9.53	0.375	760	110	40.8	5.92
	#4		0.224	12.70	0.5	690	100	40.8	5.92
	#5		0.337	15.88	0.625	655	95	40.8	5.92
			0.458	19.05	0.75	620	90	40.8	5.92
			0.593	22.23	0.875	586	85	40.8	5.92
			0.834	25.4	1	550	80	40.8	5.92
32	#10	807.34	1.251	31.75	1.25	517	75		5.92

29,000,000

Hughes Brothers reserves the right to make improvements in the product and/or proces may result in benefits or changes to some physical-mechanical characteristics. The dherin is considered representative of current production and is believed to be reliated best available characterization of the product as of January, 2000.

Cross Sectional Area

The cross sectional area of the rebar may be determined by immersing a sample in water and measuring the volume displacement of the piece. Cross sectional area may also be calculated using the nominal diameter. When calculating the cross sectional area, the cross section is assumed to be a circle.

Nominal Diameter

The nominal diameter of the rebar is the average diameter and assumes the shape of the rebar is a circle.

Tensile Strength

Tensile strength values shown are determined as the average failure load divided by the cross sectional area based on nominal bar diameter. Tensile stress varies as diameter increases due to shear lag which develops between the fibers in the larger sizes.



Results of destructive tensile testing performed at Penn State University

Typical Stress/Strain Curve for GFRP Rebar

http://www.hughesbros.com/RebarDes.html

7/31/02

M	\mathbf{A}			Sheet No. Job Number	1 05-010	of _	1
INCA ENGINE TITLE	FAS AC Three Mile Falls D	iversion D	am.				
SUBJECT	Grating Support						
MADE BY	JSJ	DATE	1/30/2006	CHECKED BY	ш	DATE	1/30/2006

The existing steel grating over the ladder is supported by steel beams. The construction drawings provided by the client do not explicitly show the orientation, spacing or size of these support members.

Per the field visit photos, the bearing bars of the grating run cross-wise with the fish ladder.

Per a design document related to this project, the design load for the existing steel grating is 150 psf uniform live load or 1,000 lb point load.

As a worst case scenario, assume that the main support for the grating runs longitudinal with the grating with a single beam located half way between the ladder side walls. An angle bracket most-likely supports the grating where it meets the side walls of the ladder. To be conservative, this angle will be ignored.

This beam will be interupted by the demolition work as well as by the installation of the new weir cap beams.

Where the beam is interupted, the structural element will be cleanly cut, an end plate welded to the beam, and holes drilled for securing the end plate to the cap beam via F-42 threaded inserts cast into the cap beam.

Assume the beam span between supports is presently 10 feet. (The spacing between the weirs.) (The beam span will be reduced by approximately 1.5 feet due to the width of the new cap beam.)

The grating area between supports is 68 SF Resulting Live Load = 10,200 lb

Assume half of the resulting live load must be carried by the new connection, 4 - 1/2" diam x 4 1/8" Nelson studs, and (2) 5/8" diam. bolt(s). (The other half of the load goes to the other end of the support beam and whatever presently supports it.)

Number of bolts per connection 2
Shear load per threaded cap bolt 2.6 kips

A307 galvanized cap bolts will be used

From AISC Table 1-D, the allowable shear (in KIPS) for threaded parts in single shear is as follows:

Size	Allowable Shear	
5/8	3.1	
3/4	44	

A pair of 5/8" bolts will suffice.

Four Nelson studs will be welded to a rectangular base plate cast onto the faces of the cap beam at mid-span. The Nelson studs will be located/separated vertically to avoid conflicts with the reinforcing.

The vertical separation between Nelson Studs can be 3.5 inches without interfering with the horizontal #5 reinforcing. (See New Weir.dgn model - weir section)

(File, New, Calc Sheet.xk) Page 1



Sheet No. 1 of 1

Job Number 05-010

CHECKED BY

From Designer's Operating Criteria dated 06-17-94

	Headwater	Tailwater
Flow Condition	Elevation	Elevation
High Flow	405.47	395.23
Winter High Flow	405.12	395.06
Low Flow	404.78	394.79
No Flow	404.30	394.67

These calculations account for the variation in the water depth at each weir slot and qunatify the flow and water depth at each weir.

Headwater Elevation 405.47

Weir #10/#9

MADE BY

JSJ

Distance from grade break 40.15 ft (grade break at pool 5) Floor elevation at slot 398.265 ft (1V:10H floor slope)

Water depth (just upstream) 7.205 ft U/S Pool elevation 405.47 ft

Flow through slot cfs (from discharge function)

DATE 10/24/2005

Head loss at slot 0.5 ft

Top of weir elevation 408.5 ft (from as-builts 30-100-5603)

Weir #9/#8

Distance from grade break 30.25 ft (grade break at pool 5) Floor elevation at slot 397.275 ft (1V:10H floor slope)

Water depth (just upstream) 6.705 ft U/S Pool elevation 404.97 ft

Flow through slot cfs (from discharge function)

Head loss at slot 0.5 ft

Top of weir elevation 408.5 ft (from as-builts 30-100-5603)

Weir #8/#7

Distance from grade break 20.1 ft (grade break at pool 5) Floor elevation at slot 396.2575 ft (1V:10H floor slope)

Water depth (just upstream) 397.275 ft U/S Pool elevation 404.47 ft

Flow through slot cfs (from discharge function)

Head loss at slot 0.5 ft

Top of weir elevation 405.5 ft (from as-builts 30-100-5603)

Weir #7/#6

Distance from grade break 9.9 ft (grade break at pool 5) Floor elevation at slot 395.24 ft (1V:10H floor slope) Water depth (just upstream) 10.23 ft U/S Pool elevation 403.97 ft

Flow through slot cfs (from discharge function) Head loss at slot 0.5 ft

Top of weir elevation 404.5 ft (from as-builts 30-100-5603)

Weir #6/#5

Distance from grade break 0.0 ft (grade break at pool 5) 394.25 ft (1V:10H floor slope) Floor elevation at slot Water depth (just upstream) 11.22 ft U/S Pool elevation 403.47 ft Flow through slot cfs (from discharge function)

Head loss at slot 0.5 ft Top of weir elevation 403.5 ft (from as-builts 30-100-5603)

Weir #5/#4

Distance from grade break 10.45 ft (grade break at pool 5) Floor elevation at slot 393.205 ft (1V:10H floor slope) Water depth (just upstream) 12.265 ft U/S Pool elevation 402.97 ft Flow through slot cfs (from discharge function) 0.5 ft Head loss at slot Top of weir elevation 403.5 ft (from as-builts 30-100-5603)

Weir #4/#3

20.9 ft (grade break at pool 5) Distance from grade break Floor elevation at slot 392.16 ft (1V:10H floor slope) 13.31 ft

Water depth (just upstream) U/S Pool elevation 402.47 ft

Flow through slot cfs (from discharge function) 0.5 ft Head loss at slot 403.5 ft (from as-builts 30-100-5603) Top of weir elevation



Sheet No.	1	of	1
Job Number	05-010		

Three Mile Falls Diversion Dam - Existing Weir Elevations

SUBJECT MADE BY

(Based on Ready for Construction Drawings)

JSJ DATE 1/26/2006

CHECKED BY BT

DATE 1/26/2006

These calculations define height of the clear opening of the antennas.

New Wier #1

Distance from grade break to downstream face of weir

Small (left) weir panel 8.56 feet Large (right) weir panel 10 feet

New Wier #2

Distance from grade break to downstream face of weir

Small (left) weir panel 28.56 feet Large (right) weir panel 30 feet

Grating elevation at grade break Ladder floor elevation at grade break 404.00 feet (Plan 5603) 394.25 feet (Sect L 5608)

9.75 feet

(9'-9")

difference The slope of the existing fish lader floor and grating is 1V:10H. References to Ready for Construction Drawing (typical)

New Weir #1

Elevation of existing grating finish grade directly above downstream face of weir panels

Small (left) weir panel 404.86 feet Large (right) weir panel 405.00 feet

Elevation at floor of fish ladder where face of weir panel meets the floor of the fish ladder

Small (left) weir panel 395.11 feet 395.25 feet Large (right) weir panel

Top elevation of existing weir panels 404.5 feet (Plan 5603)

New Weir #2

Elevation of existing grating finish grade directly above downstream face of weir panels

Small (left) weir panel 406.86 feet 407.00 feet Large (right) weir panel

Elevation at floor of fish ladder where face of weir panel meets the floor of the fish ladder

Small (left) weir panel 397.11 feet 397.25 feet Large (right) weir panel

Top elevation of existing weir panels 406.5 feet (Plan 5603)

2/12/2007 Antenna Height Calca



Sheet No.	1	of	1
Job Number	05-010		

Three Mile Falls Diversion Dam - Existing Weir Elevations

SUBJECT MADE BY

(Based on Ready for Construction Drawings) JSJ

DATE 1/26/2006

CHECKED BY BT

DATE 1/26/2006

Antennas:

The antenna shall be fabricated so that the top of the antenna shall be below the top of the existing slot / weir panels

inches

The antennas shall be identical at each of the two locations

Antenna #1

Elevation of top surface of antenna 404.17 feet height of horizontal legs of antenna 10 inches Elevation of floor of ladder 395.25 feet Height of clear opening of antenna 8.08 feet (8'-1")

Antenna #2

Elevation of top surface of antenna 406.17 feet height of horizontal legs of antenna 10 inches 397.25 feet Elevation of floor of ladder Height of clear opening of antenna 8.08 feet

Antenna hold-down drop-in anchor elevation

The antenna hold-down strut/mechanism will be installed level in the opening in the cap beam.

The vertical difference between the screw pad surface and the

center of the holes in the end plates is 3 inches

Vertical distance above floor of recess to

center of the holes in the end plates is 10.0 feet

Height of Antenna Guide Fiberglass Pieces

(For New Weir #2)

Horizontal distance from grade break 26.8 feet (measured from demo plan cadd model) Elevation of bottom of guides 397.28 feet (elevation of top edge of recess in floor)

Elevation of line where right panel meets underside of CIP cap beam 405.86 feet

Difference 8.58 feet

(8'-6")

Fastener Spacing

cumulative 8 12 92 101 9

2/12/2007 Antenna Height Calca

= (8'-5")

	V			Sheet No. Job Number	 of _	2
INCA ENGINE		lile Falls Dam				
SUBJECT	Fibergla	ss Fastener Spacing				
MADE BY	JSJ	DATE	9/23/2005	CHECKED BY	DATE	

These calculations apply to the allowable shear to secure the fasteners used with the fiberglass antenna guide pieces.

On 9/23/05 I spoke with Ethan Love of Fibergrate at (254) 977-1310. Ethan has assisted with the fiberglass design and fabrication for previous INCA PIT tag projects. Ethan provided the following guidance for designing fasteners for securing the downstream Fiberglass guide to the precast concrete:

For fiberglass material:

Ultimate shear 4,500 psi in-plane

Ult. Bearing stress 30,000 psi

Edge spacing 4d Spacing 5d

We will require that some fasteners be countersunk into the face of the fiberglass guides. Countersinking is necessary to prevent the antenna from hanging up on the fasteners when lowered into position.

Depth of recess to top of countersink 0.05 inch

Per Ethan, I can use 1/2 of the cross section area of the countersink plus the diameter x depth for the remainder of the hole to give the effective shear bearing area.

Thickness of fiberglass piece 3/8 Inch

Safety Factor 1 (This designed accommodates an extremely unlikely

non-catastrophic load case therefore a low

safety factor is used.)

McMaster - Carr makes a 18-8 stainless steel phillips undercut flat head machine screw in 1/4" - 20 size. This fastener requires a minimal depth of countersink.

The bearing area to be used for calculating the shear for this fastener in 0.375 inch thick material is:

 Countersunk area
 0.0359 in²

 Hole / shaft area
 0.0545 in²

 Ultimate shear force
 325.9 lbs

 Allowable Shear Force
 325.9 lbs

McMaster - Carr also makes a 18-8 stainless steel phillips undercut flat head machine screw in 3/8" - 16 size.

The bearing area to be used for calculating the shear for this fastener in 3/8" thick material is:

 Countersunk area
 0.0729 in²

 Hole / shaft area
 0.0615 in²

 Ultimate shear force
 440.8 lbs

 Allowable Shear Force
 440.8 lbs

Check the shear of the SS fastener to see if that governs.

Ultimate shear stress on 18-8 stainless fastener (anealed)

Allowable shear force on 3/8" diam 18-8 stainless fastener

2,982 lbs

Allowable shear force on 1/4" diam 18-8 stainless fastener

5,301 lbs

Clearly the strength of the fiberglass controls.

(File, New, Calc Sheet.xk) Page 1

	$\mathbf{\Lambda}$			Sheet No. Job Number	 of _	2
INCA ENGINE		ile Falls Dam				
SUBJECT	Fiberglas	s Fastener Spacing				
MADE BY	JSJ	DATE	9/23/2005	CHECKED BY	DATE	

For the design loading on these fasteners, assume that the water level is to the top of the cap beam on one side, and that the pool is dry on the other.

Also assume that the blockage at the weir slot only spans the antenna opening and none of the load is carried by the adjoining concrete precast weir segments (this is very unlikely).

The shearing force on the fasteners is greatest near the bottom of the weir slot due to increasing head pressure with depth.

The lateral force resulting from a head differential acting on the weir is as follows:

Depth of water upstream 9 ft
Depth of water downstream 0 ft
Width of slot opening (concr. to concr.) 2.867 ft

The portion of the antenna that extends into the recess in the floor of the ladder will be supported against lateral forces by the HDPE bearings. This will absorb the lateral force on the antenna. Assume that the fiberglass guides are required to resist the lateral force beginning 1.5 feet above the floor of the ladder.

Assume 1/4" -20 fasteners used.

Number of feet above the floor of the ladder: Vertical spacing between fasteners Number of fasteners at each vertical interval	1.5 ft 12 in 4	
Loading per fastener	312.5 lb	OK - allowable shear = 325.9 lbs
Number of feet above the floor of the ladder:	4 ft	
Vertical spacing between fasteners	18 in	
Number of fasteners at each vertical interval	4	
Loading per fastener	312.5 lb	OK - allowable shear = 325.9 lbs
Number of feet above the floor of the ladder:	5.5 ft	
Vertical spacing between fasteners	24 in	
Number of fasteners at each vertical interval	4	
Loading per fastener	291.7 lb	OK - allowable shear = 325.9 lbs

(File, New, Calc Sheet.xt) Page 2

Material Name	Young's	Shear	Poisson	Density, ρ	Tensile	Tensile	Shear	Shear	CTE	Source
	Modulus,	Modulus,	Ratio, v		Strength	Strength	Strength	Strength		
	E (Msi)	G (Msi)		(lb/in ³)	Ultimate	Yield	Ultimate	Yield	(10 ⁻⁸ in/in °F)	
					(Ksi)	(Ksi)	(Ksi)	(Ksi)	F)	
Steel, Stainless, 18-8 (AISI 301), Annealed	29	11.2	0.27	0.286	75	30	27	15.6		Military Handbook 5E (1987), S-Basis (minimum aerospace property levels)
Steel, Stainless, 18-8 (AISI 301), 1/2 Hard	26	10.5	0.27	0.286	151	105	72	41		(1987), B-Basis (90% of properties exceed value, 95% confidence)
Steel, Stainless, 18-8 (AISI 301), Full Hard	26	10.5	0.27	0.286	185	142	98	56.6		(1987), B-Basis (90% of properties exceed value, 95% confidence)

Appendix B-4 Three Mile Falls Dam Adult Fish Ladder Site Visit Photos 04 March 2005

The following photos were taken during an on-site visit to Three Mile Falls Dam in March of 2005



Photo 1. (DSCN1339) LP Panelboard at east bank of Three Mile Falls Dam (TMFD).



Photo 2. (DSCN1340) Close-up view of LP Panelboard at east bank of TMFD.



Photo 3. (DSCN1341) Panelboard label at east bank adult fish ladder at TMFD.



Photo 4. (DSCN 1342) Panelboard key at east bank adult fish ladder at TMFD.



Photo 5. (DSCN 1343) Northfacing view of walkway along the east bank adult fish ladder at TMFD. Taken directly above the adult counting window.



Photo 6. (DSCN 1344) View of vertical slots within the adult fish ladder at TMFD.

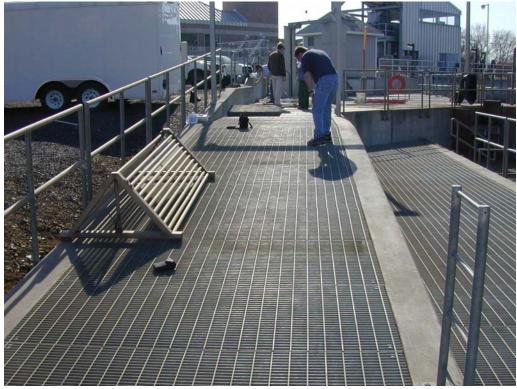


Photo 7. (DSCN 1345) Southfacing view of walkway on east bank adult fish ladder of TMFD.



Photo 8. (DSCN 1346) Proposed antenna location #1 at second weir downstream of the counting window, TMFD.



Photo 9. (DSCN 1347) Proposed antenna location #2 at fourth weir downstream of the counting window, TMFD.

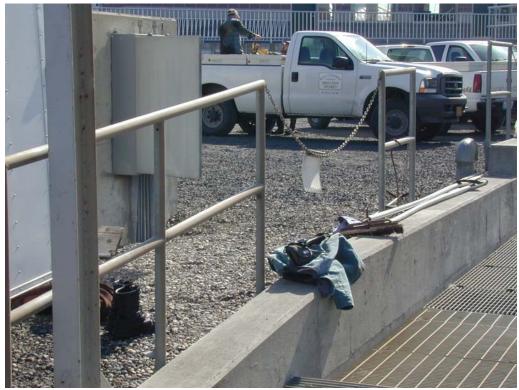


Photo 10. (DSCN 1348) Southfacing view of the electrical panel box and walkway along east bank adult fish ladder, TMFD.



Photo 11. (DSCN 1349) View of handrail attachments, TMFD.



Photo 12. (DSCN 1350) View of handrail attachment along ladder wall.



Photo 13. (DSCN 1351) South facing view of walkway and corner of electronics room trailer at east bank adult fish ladder of TMFD.