



LAWRENCE
LIVERMORE
NATIONAL
LABORATORY

Regulation of Electrochemically Generated Ferrous Ions from an Iron Cathode for Pd-Catalytic Transformation of MTBE in Groundwater

P. Liao, S. Yuan, M. Chen, M. Tong, W. Xie, P.
Zhang

April 5, 2013

Environmental Science & Technology

Disclaimer

This document was prepared as an account of work sponsored by an agency of the United States government. Neither the United States government nor Lawrence Livermore National Security, LLC, nor any of their employees makes any warranty, expressed or implied, or assumes any legal liability or responsibility for the accuracy, completeness, or usefulness of any information, apparatus, product, or process disclosed, or represents that its use would not infringe privately owned rights. Reference herein to any specific commercial product, process, or service by trade name, trademark, manufacturer, or otherwise does not necessarily constitute or imply its endorsement, recommendation, or favoring by the United States government or Lawrence Livermore National Security, LLC. The views and opinions of authors expressed herein do not necessarily state or reflect those of the United States government or Lawrence Livermore National Security, LLC, and shall not be used for advertising or product endorsement purposes.

1 **Regulation of Electrochemically Generated Ferrous Ions from an**
2 **Iron Cathode for Pd-Catalytic Transformation of MTBE in**
3 **Groundwater**

4 *Peng Liao*[†], *Songhu Yuan*^{*†}, *Mingjie Chen*[‡], *Man Tong*[†], *Wenjing Xie*[†], *Peng Zhang*[†]

5 [†]State Key Lab of Biogeology and Environmental Geology, China University of
6 Geosciences, 388 Lumo Road, Wuhan 430074, P. R. China

7 [‡]Atmospheric, Earth and Energy Division, Lawrence Livermore National Laboratory,
8 P.O. Box 808, L-184, Livermore, CA 94550, United States

9 _____

10 *E-mail: yuansonghu622@hotmail.com; Phone: +86-27-67848629;

11 Fax: +86-27-67848629.

12

13 **RECEIVED DATE (to be automatically inserted after your manuscript is**
14 **accepted if required according to the journal that you are submitting your paper**
15 **to)**

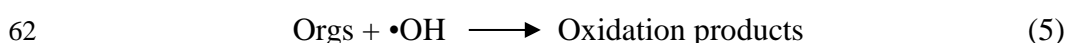
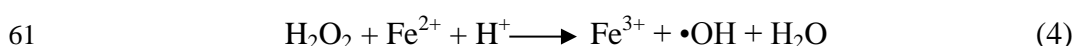
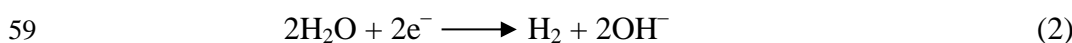
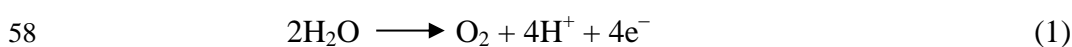
21 **ABSTRACT** An effective approach is developed to regulate the generation of Fe^{2+}
22 from an iron cathode in a three-electrode system. The Fe^{2+} is then used for the
23 Pd-catalytic transformation of methyl *tert*-butyl ether (MTBE) in simulated
24 groundwater. Under conditions of pH 3, a total current of 50 mA, and 1 g/L Pd/ Al_2O_3 ,
25 20 mg/L MTBE was completely transformed within 60 min in an undivided
26 electrolytic cell using the iron cathode, with 14 mg/L of accumulated Fe^{2+} . Fe^{2+}
27 accumulation follows pseudo-first-order kinetics and the rate is regulated by electric
28 current and groundwater pH, giving the relation of $k = 5.3 \times 10^4 \cdot 10^{-2\text{pH}} - 7.25 \cdot I^2 -$
29 $8.8 \times 10^{11} \cdot 10^{2\text{pH}-28}$, where k is the rate constant of Fe^{2+} accumulation (min^{-1}) and I is
30 the current (mA). In a modified three-electrode column using iron as the first cathode,
31 the localized acidic and oxidizing conditions developed automatically in the iron
32 cathode zone by partitioning the current between the two cathodes, leading to
33 controllable generation of Fe^{2+} for MTBE transformation. The stable transformation
34 of MTBE in a long-term study suggests that this method is a sustainable approach to
35 supplying Fe^{2+} for Pd-catalytic transformation of organic contaminants in
36 groundwater.

37

38 **INTRODUCTION**

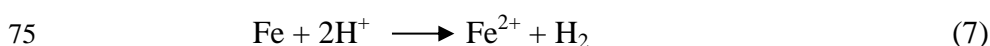
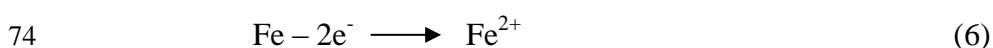
39 Methyl *tert*-butyl ether (MTBE) has been manufactured almost exclusively for use
40 as a fuel additive since the 1970s.¹⁻³ MTBE has been found in various environments,
41 particularly in groundwater, because of its high solubility under ambient conditions
42 (48 g/L).¹ MTBE was reclassified by the USEPA as “carcinogenic in humans by all

43 routes of exposure” based on findings that indicated side effects of headaches,
 44 vomiting, diarrhea, fever, cough, muscle aches, and skin and eye irritation.^{1,4} Current
 45 processes that have been proposed for remediation of MTBE in groundwater include
 46 phase transfer,⁵ in situ chemical oxidation (ISCO),⁶ advanced oxidation technologies
 47 (AOTs),⁷⁻⁹ and biodegradation.^{2,10,11} Electrochemical processes have attracted
 48 increasing interest in groundwater remediation as it is simple and versatile.¹²⁻¹⁵
 49 Recently, a hybrid electrolysis and Pd-catalytic oxidation process, in which H₂O₂ was
 50 produced from the combination of electro-generated H₂ and O₂ on a Pd catalyst under
 51 acidic conditions (eq. 1–3), has been shown to be highly effective in treating a wide
 52 range of recalcitrant contaminants in wastewaters and groundwater (eqs. 4 and
 53 5).^{12,13,16} The concentration of Fe(II) significantly affects the efficiency of the
 54 contaminant transformation,^{12,13,16} and a certain amount of Fe(II) (13.7 mg/L) shifts
 55 trichloroethylene (TCE) transformation from hydrodechlorination to hydroxyl
 56 oxidation.¹² However, in situ application of this process is limited because of a lack of
 57 Fe(II) (ca. 10 mg/L), especially in iron-defective aquifers.



63 Fe(II) is normally supplied by the addition of ferrous salts,^{17,18} sacrificed iron
 64 anode,^{19,20} and iron-containing minerals in Fenton-based processes.^{21,22} The addition

65 of ferrous salts complicates the implementation process. Sacrificial iron anodes can be
66 effective (eq. 6) but produce a lot of iron sludge.²³ When iron is used as the anode, it
67 corrodes and no O₂ is produced. Therefore, there will be a lack of O₂ for H₂O₂
68 production as O₂ is a necessary reactant. Although iron-containing materials, such as
69 Fe₃O₄ and zero-valency iron (ZVI), can be mixed with Pd to supply Fe(II), the
70 reaction will deplete the materials,^{24,25} and precipitates on the Pd surface will decrease
71 the catalytic activity and clog the aquifer.²⁶⁻²⁸ Therefore, it is crucial to seek a new
72 cost-effective strategy to supply Fe(II) for the hybrid electrolysis and Pd-catalytic
73 oxidation process.



76 During the process, localized acidic conditions in the vicinity of Pd automatically
77 develops in a modified three-electrode system as there are more H⁺ ions produced at
78 the anode than OH⁻ ions produced at the adjacent cathode.^{13,16} One anode and two
79 cathodes aim to sustain water electrolysis, thus supplying H₂, O₂, H⁺ and OH⁻ for
80 H₂O₂ production on the Pd surface (eqs. 1–3).^{13,16} Therefore, it is not feasible to
81 change the electrochemical reactions at the three electrodes. However, the chemical
82 reactions at the electrodes will still occur regardless of whether or not electricity is
83 applied. For example, chemical corrosion of iron occurs naturally, though it is
84 particularly intensive under acidic conditions (eq. 7).²⁹ When the iron is used as the
85 first cathode in the three-electrode system,¹³ the chemical corrosion is enhanced under
86 the automatically developed acidic conditions. However, the corrosion is inhibited by

87 a negative applied current that protects the cathode.^{30–32} Consequently, using iron as
88 the first cathode does not change the electrochemical reactions (eqs. 1–3), but supplies
89 Fe(II) due to chemical corrosion under acidic conditions (eq. 7). The production of
90 Fe(II) from the iron cathode can be theoretically regulated by conditioning the
91 cathodic current.

92 In this study, an undivided electrolytic cell with an iron cathode is used to regulate
93 the generation of Fe²⁺ at different values of pH and current for Pd-catalytic
94 transformation of MTBE in simulated groundwater. The kinetics and mechanism of
95 Fe²⁺ accumulation are also explored. The relation of Fe²⁺ accumulation with pH and
96 cathodic current is modeled. A modified three-electrode column using an iron cathode
97 is then configured to regulate Fe²⁺ accumulation by varying the cathodic current
98 applied on the iron cathode for transformation of MTBE. The long-term performance
99 of this column is ultimately evaluated.

100

101 **EXPERIMENTAL**

102 **Chemicals.** MTBE (99.8%) and *tert*-butyl alcohol (TBA, 99.5%) were purchased
103 from Sigma-Aldrich. *Tert*-butyl formate (TBF, 99.5%), acetone (AC, 99.5%),
104 formaldehyde (FA, 37%), and formic acid (99.5%) were supplied by Aladdin
105 Chemistry Co. Ltd. Palladium on alumina powder (5% wt. Pd, Shanxi Kaida
106 Chemical Ltd, China), with an average particle size of 6 μm, was used as the catalyst
107 in the batch experiments. Palladium on alumina pellets (0.5% wt. Pd, Shanxi Kaida
108 Chemical Ltd) 4 mm in size was used in the column experiments. Iron plate (S45C

109 type, Wuhan Steel Processing Co., Ltd, China) and mixed metal oxides (MMO, IrO₂
110 and Ta₂O₅ coating on titanium diamond mesh, Shanxi Kaida Chemical Ltd) with
111 dimensions of 5.0 cm diameter and 1.7 mm thickness were used as the electrodes.
112 Seventeen holes (4.1 mm in diameter) were evenly distributed through the iron plate
113 electrodes. Prior to the experiments, the iron electrode was polished with coarse
114 emery cloth, etched by diluted HCl solution (5 wt %), and washed with deionized
115 water. Deionized water (18.2 mΩ·cm) obtained from a Millipore Milli-Q system was
116 used in all the experiments. All the chemicals used in this study were above analytical
117 grade.

118 **Batch Experiments.** An undivided acrylic electrolytic cell was used to evaluate
119 Fe²⁺ generation for MTBE transformation at ambient temperatures (25 ± 1 °C) (Figure
120 S1 in the Supporting Information). An MMO mesh and an iron plate were
121 respectively used as the anode and cathode, with 40 mm spacing in parallel positions.
122 For each test, 400 mL of 20 mg/L MTBE solution was transferred into the cell, and 10
123 mM Na₂SO₄ and 1 g/L Pd/Al₂O₃ (5% wt. Pd) were attained by the addition of specific
124 masses of Na₂SO₄ and Pd/Al₂O₃ powder. Aqueous pH was adjusted to 3.0 by 1 M
125 H₂SO₄. For comparison, an MMO mesh was used as the cathode in the presence of 10
126 mg/L Fe²⁺. The reactor was immediately sealed and stirred at 600 rpm using a
127 Teflon-coated magnetic stirring bar. A constant electric current of 50 mA (2 mA/cm²)
128 was applied with a cell potential of about 4 V. The aqueous solution was sampled at
129 pre-determined time intervals for analysis of MTBE and its transformation products,
130 iron species, pH, H₂O₂, and •OH concentration. All experiments were carried out in

131 duplicate.

132 **Column Experiments.** A vertical acrylic tube column (5.0 cm inner diameter × 30
133 cm length) was used to simulate one-dimensional flow conditions (Figure 1). Three
134 electrodes were mounted in the upward sequence of anode (MMO), cathode 1 (iron
135 plate) and cathode 2 (MMO). 4.5 g of Pd/Al₂O₃ pellets were supported by cathode 1
136 to produce a pellet bed monolayer. The remaining space in the column was packed
137 with 4 mm glass beads with a porosity of 0.46. The total and pore volume (PV) were
138 588 and 268 mL, respectively. Solution was pumped through the bottom of the
139 columns at a rate of 2.5 mL/min using a peristaltic pump (Luxi, model HL-2, China).
140 The flow rate was maintained at 2.5 mL/min, resulting in Darcy's velocity of 1.83
141 m/day. The total current was kept at 80 mA by a DC power supply (GPC-3060D,
142 Taiwan Goodwill Instrument). By adjusting the rheostat, different currents were
143 partitioned between cathodes 1 and 2 so that the localized pH in the Pd vicinity could
144 be regulated. Simulated MTBE-contaminated groundwater (10 mg/L) with 5 mM
145 Na₂SO₄ and 0.5 mM CaSO₄ deoxygenated solution (~ 840 μS/cm) was stored in a 5 L
146 gas-tight collapsible bag. The column was flushed by 2 PVs of contaminated
147 groundwater before switching on the power supply. Control experiments show that
148 the adsorption of MTBE on Pd/Al₂O₃ pellets and glass beads were negligible during
149 operation. At regular time intervals, 1 mL of groundwater was sampled from six ports
150 (see Figure 1 for locations) and measured for MTBE, pH, and iron species
151 concentration.

152 **Chemical Analysis.** MTBE, TBF, TBA, and AC concentrations were analyzed

153 using a gas chromatograph (GC, Shimadzu 2014C, Japan) equipped with a flame
154 ionization detector (FID), a capillary column (30 m × 0.32 mm × 0.5 μm) and a
155 headspace concentrator (DK-3001A, Beijing Zhongxing Huili Technology
156 Development Co., Ltd.). The column temperature was held at 35 °C for 2 min, heated
157 from 35 to 60 °C at a rate of 5 °C/min, and then cooled to 35 °C at a rate of 10 °C/min.
158 The injector and detector temperatures were set at 220 and 250 °C, respectively. 2 mL
159 of the aqueous sample was collected and injected into a 20-mL headspace vial sealed
160 with septa. The equilibrium temperature in the vial was set to 105 °C. The detection
161 limits for all compounds were 0.1 mg/L. Gas concentrations were calculated by
162 Henry's law. FA was analyzed by an LC-15C HPLC (Shimadzu) equipped with a UV
163 detector and an XDB-C18 column (4.6 × 50 mm) after derivatization with
164 2,4-dinitrophenylhydrazine.³³ The mobile phase used a mixture of acetonitrile and
165 water (60:40, v/v) at 1 mL/min, with the detection wavelength at 355 nm. Carboxylic
166 acids were also measured by HPLC. The mobile phase used a mixture of phosphoric
167 buffer (pH 2.7) and methanol at 99:1 (v:v) at 0.8 mL/min, and a detection wavelength
168 of 210 nm. The detection limits for FA and carboxylic acids were 1 μM.

169 H₂O₂ concentration was determined at 405 nm by a spectrometer (UV-1800 PC,
170 Shanghai Mapada Spectrum Instrument Co., Ltd) after coloration with TiSO₄.³⁴ The
171 •OH levels were determined using dimethyl sulfoxide (DMSO) trapping and HPLC
172 methods described in previous literature.³⁵ Iron species were measured at 510 nm
173 using the 1,10-o-phenanthroline analytical method.³⁶ The corrosion potentials of the
174 iron electrode at various pH and at 50 mA were measured on a CS150 electrochemical

175 workstation (Wuhan CorrTest Instrument, China) using a Pt electrode as the counter
176 electrode and a saturated calomel electrode (SCE) as the reference electrode. The
177 electrolyte used for measurement was 10 mM NaSO₄.

178

179 **RESULTS AND DISCUSSION**

180 **Generation of Fe²⁺ from Iron Cathode for MTBE Transformation.** [Figure 2a](#)
181 shows the variation in the MTBE concentration in an undivided electrolytic cell using
182 an iron cathode under conditions of pH 3, 50 mA and 1 g/L Pd/Al₂O₃. The control
183 experiment using the iron cathode without an applied current or Pd catalyst does not
184 present any significant transformation of MTBE. In the presence of Pd/Al₂O₃, MTBE
185 (20 mg/L) was completely transformed in 60 min following pseudo-first-order
186 kinetics with a rate constant of 0.040 min⁻¹. For comparison, the experiment also used
187 an MMO cathode in the presence of 10 mg/L Fe²⁺, which is the average concentration
188 accumulated in the cell using the iron cathode. This showed a slower MTBE
189 transformation with a pseudo-first-order rate constant of 0.036 min⁻¹. These results
190 prove that it is feasible to generate Fe²⁺ ions from an iron cathode for Pd-catalytic
191 transformation of organic contaminants in groundwater.

192 The concentration of Fe²⁺ increases to a plateau at about 14 mg/L within 60 min of
193 using the iron cathode, and decreases from the initial 10 mg/L to 8 mg/L using the
194 MMO cathode ([Figure 2b](#)). The solution pH increased from 3.0 to 3.3 using the iron
195 cathode. A parallel test using the MMO cathode had minimal variation of the pH. The
196 pH increase is attributed to the consumption of H⁺ ions by chemical corrosion of the

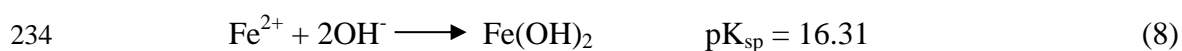
197 iron cathode (eq. 7). The concentration of Fe^{2+} released from chemical corrosion due
198 to a rise in the pH was calculated to be 14 mg/L within 60 min, which is in good
199 agreement with the measured cumulative concentration. In addition, the corrosion
200 potential (Eh) of the iron electrode under conditions of pH 3 and 50 mA was measured
201 to be -0.440 V (Figure S2a), falling in the corrosion region (Figure S2b). As a result,
202 chemical corrosion of the iron cathode is responsible for Fe^{2+} generation under the
203 protection of cathodic current.

204 The cumulative H_2O_2 concentration reaches approximately 14 mg/L within 60 min
205 in the presence of Pd/ Al_2O_3 and absence of Fe^{2+} (Figure S3). The cumulative
206 concentration of $\bullet\text{OH}$ radicals (oxidation potential: 2.8V/SHE) increases for both
207 systems using the iron and MMO cathodes (Figure S4). Good correlation between
208 MTBE transformation rate constants and $\bullet\text{OH}$ generation rate constants (Figure S5)
209 indicates that $\bullet\text{OH}$ was responsible for MTBE transformation ($k_{\text{MTBE}\cdot\text{OH}} = 1.6 \times 10^9$
210 $\text{M}^{-1} \text{s}^{-1}$).¹² The combination of H_2O_2 and Fe^{2+} produces $\bullet\text{OH}$ radicals for MTBE
211 oxidation (eqs. 4 and 5).^{1,17,18} The ratio of H_2O_2 to Fe^{2+} is critical for the
212 transformation efficiency.^{17,37} H_2O_2 and Fe^{2+} were generated gradually using the iron
213 cathode. In the cell using the MMO cathode, Fe^{2+} concentration was in excess relative
214 to the H_2O_2 concentration during the first stage, but was defective in the later stage. In
215 this regard, using an iron cathode is advantageous over the external addition of Fe(II)
216 in terms of Fe^{2+} supply.

217 During the transformation of MTBE using the iron cathode, the primary
218 transformation intermediates identified are TBF, followed by AC, TBA, FA, and

219 formic acid (Figure S6). The relatively low mass balance of carbon (65%) suggests
 220 the production of CO₂.¹² Accumulation of intermediates is due to their low reaction
 221 rates with •OH, which are one or two orders of magnitude lower than that of MTBE.³⁸
 222 It should be noted that the toxicities of the intermediates are lower than that of
 223 MTBE.³⁹ The intermediates identified in this study are similar to those reported in
 224 MTBE oxidation by anodic Fenton⁴⁰ and UV/H₂O₂.⁸

225 **Kinetic Model for Fe²⁺ Accumulation.** When electricity is applied to the iron
 226 cathode, accumulation of Fe²⁺ depends on the current and solution pH. The decreasing
 227 solution pH increases the rate of chemical corrosion of the iron cathode (eq. 7).²⁹ The
 228 increasing current increases the electron density on the iron cathode, which suppresses
 229 the release of electrons from the iron corrosion (reverse reaction of eq. 6).^{30,42} The
 230 accumulation of Fe²⁺ also decreases at high pH due to precipitation (eq. 8). If the
 231 reductive regeneration of Fe²⁺ on the iron cathode is neglected, the rate of Fe²⁺
 232 accumulation using an iron cathode in the absence of Pd/Al₂O₃ can be expressed as eq.
 233 9.



$$235 \quad \frac{d[\text{Fe}(\text{II})]}{dt} = k_{\text{H}^+} \cdot [\text{H}^+]^2 \cdot [\text{Fe}] - k_{\text{e}^-} \cdot [\text{n}_{\text{e}^-}]^2 \cdot [\text{Fe}^{2+}] - k_{\text{OH}^-} \cdot [\text{OH}^-]^2 \cdot [\text{Fe}^{2+}] \quad (9)$$

236 Where k_{H^+} , k_{e^-} and k_{OH^-} are the rate constants for Fe²⁺ generation by H⁺ and [n_{e-}] is the
 237 electron density concentration.

$$238 \quad [\text{H}^+] = 10^{-\text{pH}} \quad (10)$$

$$239 \quad [\text{OH}^-] = 10^{\text{pH}-14} \quad (11)$$

$$240 \quad [\text{n}_{\text{e}^-}] = \text{dn}/\text{dt} = (\text{L} \cdot \text{I}/\text{N}_{\text{A}})/\text{V} \quad (12)$$

241 Where L is the quantity of electricity per unit time (6.24×10^{18}), N_A is the Avogadro
 242 constant (6.02×10^{23}), and V is the electrolyte volume (0.4 L in this study). Since the
 243 concentration of Fe (expressed by surface area) is linearly related with Fe^{2+} (Figure
 244 S7), eq. 9 can be re-written as

$$\begin{aligned}
 245 \quad \frac{d[Fe(II)]}{dt} &= (k_H^{+'} \cdot 10^{-2pH} - k_e^- \cdot k' \cdot I^2 - k_{OH}^- \cdot 10^{2pH-28}) \cdot [Fe^{2+}] \\
 246 \quad &= k \cdot [Fe^{2+}] \qquad \qquad \qquad (13)
 \end{aligned}$$

247 where k ($= k_H^{+'} \cdot 10^{-2pH} - k_e^- \cdot k' \cdot I^2 - k_{OH}^- \cdot 10^{2pH-28}$) is the final accumulation rate of
 248 Fe^{2+} , and k' ($= C^2/N_A^2 \cdot V^2$) is a constant ($= 6.7 \times 10^{-10}$). The accumulation of Fe^{2+}
 249 conforms to pseudo-first-order kinetics, and the rate constant can be quantitatively
 250 regulated by electric current and solution pH.

251 **Effects of Current and pH on Fe^{2+} Accumulation.** Figure 3 displays the variation
 252 of Fe^{2+} concentration at different currents and pHs using the iron cathode. The values
 253 of $k_H^{+'}$, k_e^- , and k_{OH}^- were calculated to be 5.3×10^4 , 1.08×10^{10} , and $8.8 \times 10^{11} \text{ M}^{-1}$
 254 s^{-1} at pH 2 without electricity application, at pH 3 and 50 mA current, and at pH 5 and
 255 50 mA current, respectively.⁴³ At pH 3, the accumulation of Fe^{2+} at different currents
 256 approximately follows pseudo-first-order kinetics. A quantitative relationship between
 257 k and current was obtained as $k = 0.053 - 7.25 \cdot I^2$ when pH = 3 is set in eq. 13. It can
 258 be seen that the k value decreases quadratically with increasing the current. The rate
 259 constant of Fe^{2+} accumulation decreases from 0.052 min^{-1} without an applied current
 260 to 0.005 min^{-1} at 80 mA (inset in Figure 3a). When the current is higher than 120 mA,
 261 Fe^{2+} accumulation is minimal because of the low k values (close to zero). It can be
 262 calculated that the accumulation of Fe^{2+} is insignificant when the current is higher

263 than 85 mA.

264 **Figure 3b** shows that the accumulation of Fe^{2+} at 50 mA with different pHs also
265 follows pseudo-first-order kinetics, giving the quantitative relationship of $k = 5.3 \times$
266 $10^4 \cdot 10^{-2\text{pH}} - 8.8 \times 10^{11} \cdot 10^{2\text{pH}-28} - 3.6 \times 10^3$ when $I = 50$ mA in eq. 13. The rate
267 constant of Fe^{2+} accumulation decreases from 0.022 to 0.00001 min^{-1} as the solution
268 pH increases from 2 to 7 (inset in **Figure 3b**). Similar results were found for currents
269 of 0, 20, and 80 mA (**Figure S8**). This suggests that high pH inhibits the corrosion of
270 the iron cathode at a specific current (**Figure S2**).²⁹ When the solution pH was
271 elevated to 5.7, Fe^{2+} did not accumulate significantly, which is consistent with the fact
272 that minimal Fe^{2+} accumulates at pH 7 at any applied current (**Figure S8**). The
273 relatively high correlation coefficients ($R^2 > 0.934$) suggest that the kinetic model (eq.
274 13) is reasonable for Fe^{2+} accumulation.

275 The accumulation of Fe^{2+} at different currents and pHs is calculated and depicted in
276 **Figure 3c**. A quantitative relationship between current and pH is found to be $\text{pH} =$
277 $4.12 - \log(I)/2$ ($I \neq 0$) when k is set at zero in eq. 13. When the pH value is above the
278 curve, no Fe^{2+} accumulates (region 1 in **Figure 3c**). In contrast, when the pH value is
279 below the curve, the value of k is negative, implying that the rate of Fe^{2+} consumption
280 is faster than that of generation (region 2 in **Figure 3c**).

281 **Electrochemically Regulated Fe^{2+} Generation in a Three-electrode Column.**

282 For this study, a modified three-electrode column¹³ using an iron cathode (**Figure 1**)
283 was configured to regulate the localized pH around the iron cathode for generation of
284 Fe^{2+} by adjusting the current partition between cathodes 1 (iron cathode) and 2. When

285 the current applied on the iron cathode increased from 30 to 50 mA, the pH in the
286 vicinity of iron cathode increased slightly from 3.0 to 3.5 (Figure 4a). With the current
287 further raised to 80 mA, the localized pH remarkably increases to 8. An exponential
288 relationship (eq. 14) between the pH and current is observed for the iron cathode zone
289 (inset in Figure 4a). This monotone function indicates that the localized pH is
290 determined by the current applied on the iron cathode, which differs from the batch
291 experiments where the pH is adjusted manually.

$$292 \quad \text{pH} = 1.3 \cdot \exp(I_{\text{iron}}) \quad (30 \leq I \leq 80 \text{ mA}) \quad R^2 = 0.973 \quad (14)$$

293 As the localized pH is determined by the current on the iron cathode, accumulation
294 of Fe^{2+} can be exclusively regulated by the current by substituting eq. 14 into eq. 13.
295 Figure 4b shows that the steady-state Fe^{2+} concentration along the column decreases
296 from maximum values in the iron cathode zone to the low values in both sides of the
297 iron cathode. It should be noted that minimal Fe^{2+} were observed around cathode 2
298 because of the neutral or alkaline conditions.²³ Cumulative Fe^{2+} concentrations in the
299 vicinity of the iron cathode at different current partitions follow in the order of 30:50 >
300 40:40 > 50:30 > 60:20 > 80:0. This was confirmed by a slight decrement as the
301 current on the iron cathode increased from 30 to 50 mA (inset in Figure 4b), followed
302 by a dramatic decrease to a low value below the detection limit as the current
303 increased to 80 mA. Decreasing the current on the iron cathode speeds up Fe^{2+}
304 generation. However, excess Fe^{2+} may consume H_2O_2 and quench $\bullet\text{OH}$,^{12,13,16-18,23} and
305 a low current on the iron cathode can lead to an alkaline effluent. It is important to
306 keep the current on the iron cathode at appropriate levels for generating the required

307 dosage of Fe^{2+} .

308 **MTBE Transformation in the Column.** In the three-electrode column using an
309 iron cathode, MTBE (10 mg/L) transformation attains steady state after 150 min of
310 operation (1.4 PVs) in a typical column (Figure S9). As shown in Figure 4c, MTBE
311 can be efficiently transformed through the Pd zone. With the five ratios of current
312 partition, the removal rates of MTBE are in the sequence 50:30 > 30:50 > 40:40 >
313 60:20 > 80:0, implying the critical role of current partition on MTBE transformation.
314 The anodic current generates O_2 and H^+ (eq. 1), and the cathodic current partitioned
315 on the iron (first cathode) generates H_2 and OH^- (eq. 2). As the quantity of H^+
316 produced at the anode is more than that of OH^- produced at iron (first cathode), and
317 the reflux of OH^- from cathode 2 is complex, a low pH in local Pd zone was
318 automatically developed, which contributes to the production of Fe^{2+} (eq. 7), H_2O_2 (eq.
319 3) and $\bullet\text{OH}$ (eq. 4). The lowest transformation of MTBE at the current ratio of 80:0 is
320 likely result of the minimal generation of $\bullet\text{OH}$ from Fe^{2+} and H_2O_2 under alkaline
321 conditions (pH 8).²³ The low efficiency of transformation at the current ratio of 60:20
322 is a result of the low concentrations of Fe^{2+} and H_2O_2 at pH 5.¹⁶ At the current ratio of
323 30:50, the higher Fe^{2+} concentration generated competes with MTBE for $\bullet\text{OH}$.^{17,18,23}
324 Although similar efficiencies were obtained for current ratios of 50:30 and 40:40, the
325 effluent is alkaline (Figure 4a) and the energy consumption is higher for the latter
326 ratio (Table S1).

327 **Long-term Performance for MTBE Transformation.** The long-term MTBE
328 transformation performance of the three-electrode column using an iron cathode is

329 also evaluated. [Figure 5](#) shows that MTBE transformation can be sustained at a stable
330 level of around 70% for 20 days. About 30–40% MTBE is transformed in the anodic
331 zone (Ports 1 and 2), which may be caused by the fact that certain Fe^{2+} and H_2O_2
332 produced in the iron cathode zone diffuses into the anode zone under the localized
333 acidic condition. This finding suggests that the reaction zone was extended by
334 convection-dispersion effects. The localized pH in the iron cathode zone decreases
335 from 6.5 to 3.5 in the first 8 h, remaining almost constant until 10 days, after which it
336 decreased gradually to 3.3 ([Figure S10](#)). The decline of pH in the iron cathode zone is
337 beneficial to the accumulation of Fe^{2+} and H_2O_2 ,¹² which enhances MTBE
338 transformation. It is noted that the effluent pH dramatically decreases at 8 h from 8.5
339 to 6.5 ([Figure S10](#)). Since the transformation of MTBE does not show any significant
340 decrease after 20 days of operation, a relatively long longevity can be proposed for the
341 transformation of MTBE in the three-electrode column using an iron cathode.

342 **Implications for Groundwater Remediation.** In this study, an iron cathode is
343 employed in an in situ supply of Fe^{2+} for Pd-catalytic transformation of MTBE in
344 simulated groundwater. Kinetic analysis and experimental results show that the
345 accumulation of Fe^{2+} can be quantitatively related to the cathodic current applied on
346 the iron and the solution pH. In a modified three-electrode column using an iron
347 cathode, both the pH and Fe^{2+} accumulation in the Pd vicinity are determined by the
348 current applied on the iron cathode, and MTBE transformation can be sustained at
349 high levels for relatively long periods of time. Fe^{2+} concentration can be exclusively
350 regulated by the current being applied on the iron cathode, providing a flexible

351 approach to supplying Fe^{2+} for Pd-catalytic transformation of organic contaminants in
352 an iron-defective aquifer. The findings of this study may also indicate a quantitative
353 supply of Fe^{2+} for the conventional electro-Fenton process, i.e., using one iron
354 cathode and one O_2 reduction cathode. The effective transformation of organic
355 contaminants including dyes,¹⁶ TCE,¹² toluene,¹³ phenol¹³ and MTBE was
356 successfully achieved by the hybrid electrolysis and Pd-catalytic oxidation process.
357 However, a quantitative description of the transformation kinetics is still lacking.
358 Future work will look to develop a one-dimensional reactive transport model that will
359 describe the transformation process and guide laboratory and field remediation design.

360

361 **Supporting Information Available**

362 Additional information: Figures S1–S10, and Tables S1. This material is available
363 free of charge via the Internet at <http://pubs.acs.org>.

364

365 **ACKNOWLEDGEMENTS**

366 This work was supported by the Natural Science Foundation of China (NSFC, No.
367 41172220), and the State Key Laboratory of Biogeology and Environmental Geology,
368 China University of Geosciences (No. GBL11204).

369 Prepared by LLNL under Contract DE-AC52-07NA27344.

370 **REFERENCES**

371 (1) Cooper, W. J.; Cramer, C. J.; Martin, N. H.; Mezyk, S. P.; Shea, K. E. O.,
372 Somntag, C. V. Free radical mechanisms for the treatment of methyl tert-butyl ether

373 (MTBE) via advanced oxidation/reductive processes in aqueous solutions. *Chem. Rev.*
374 **2009**, *109*, 1302–1345.

375 (2) Hunkeler, D.; Butler, B. J.; Aravena, R.; Barker, J. F. Monitoring
376 biodegradation of methyl tert-butyl ether (MTBE) using compound-specific carbon
377 isotope analysis. *Environ. Sci. Technol.* **2001**, *35*, 676–681.

378 (3) Acero, J. L.; Handerlein, S. B.; Schmidt, T. C.; Suter, M. J. F.; Gunten, U. V.
379 MTBE oxidation by conventional ozonation and the combination ozone/hydrogen
380 peroxide: Efficiency of the processes and bromate formation. *Environ. Sci. Technol.*
381 **2001**, *35*, 4252–4259.

382 (4) Cooper, W. J.; Tobien, T.; Nickelsen, M. G.; Adams, J. W.; Shea, K. E. O.;
383 Bartels, D. M.; Wishart, J. F.; Tornatore, P. M.; Newman, K. S.; Gregoire, K.
384 Radiation chemistry of methyl tert-butyl ether in aqueous solution. *Environ. Sci.*
385 *Technol.* **2004**, *38*, 3994–4001.

386 (5) Yazaydin, A. O.; Thompson, R. W. Molecular simulation of the adsorption of
387 MTBE in silicalite, mordenite, and zeolite beta. *J. Phys. Chem. B* **2006**, *110*, 14458–
388 14462.

389 (6) Tsitonaki, A.; Petri, B.; Crimi, M.; Mosbaek, H.; Siegrist, R. L.; Bjerg, P. L. In
390 situ chemical oxidation of contaminated soil and groundwater using persulfate: A
391 review. *Crit. Rev. Environ. Sci. Technol.* **2010**, *40*, 55–91.

392 (7) Johnson, D. C.; Shamamian, V. A.; Callahan, J. H.; Denes, F. S.; Manolache, S.
393 O.; Dandy, D. S. Treatment of methyl tert-butyl ether contaminated water using a
394 dense medium plasma reactor: A mechanistic and kinetic investigation. *Environ. Sci.*

395 *Technol.* **2003**, *37*, 4804–4810.

396 (8) Stefan, M. I.; Mack, J.; Bolton, J. R.; Corporation, C. C. Degradation pathways
397 during the treatment of methyl tert-butyl ether by the UV/H₂O₂ process. *Environ. Sci.*
398 *Technol.* **2000**, *34*, 650–658.

399 (9) Wu, T. N. Electrocatalytic oxidation of methyl tert-butyl ether (MTBE) in
400 aqueous solution at a nickel electrode. *Chemosphere* **2007**, *69*, 271–278.

401 (10) Waul, C.; Arvin, E.; Schmidt, J. E. Modeling the competitive effect of
402 ammonium oxidizers and heterotrophs on the degradation of MTBE in a packed bed
403 reactor. *Water Res.* **2008**, *42*, 3098–3108.

404 (11) Waul, C.; Arvin, E.; Schmidt, J. E. Model description and kinetic parameter
405 analysis of MTBE biodegradation in a packed bed reactor. *Water Res.* **2008**, *42*,
406 3122–3134.

407 (12) Yuan, S. H.; Mao, X. H.; Alshawabkeh, A. N. Efficient degradation of TCE in
408 groundwater using Pd and electro-generated H₂ and O₂: A shift in pathway from
409 hydrodechlorination to oxidation in the presence of ferrous ions. *Environ. Sci.*
410 *Technol.* **2012**, *46*, 3398–3405.

411 (13) Yuan, S. H.; Chen, M. J.; Mao, X. H.; Alshawabkeh, A. N. A three-electrode
412 column for Pd-catalytic oxidation of TCE in groundwater with automatic
413 pH-regulation and resistance to reduced sulfur compound foiling. *Water Res.* **2013**, *47*,
414 269–278.

415 (14) Mao, X. H.; Ciblak, A.; Amiri, M.; Alshawabkeh, A. N. Redox control for
416 electrochemical dechlorination of trichloroethylene in bicarbonate aqueous media.

- 417 *Environ. Sci. Technol.* **2011**, *45*, 6517–6523.
- 418 (15) Mao, X. H.; Yuan, S. H.; Fallahpour, N.; Ciblak, A.; Howard, J.; Padilla, I.;
419 Loch-caruso, R.; Alshawabkeh, A. N. Electrochemically induced dual reactive barriers
420 for transformation of TCE and mixture of contaminants in groundwater. *Environ. Sci.*
421 *Technol.* **2012**, *46*, 12003–12011.
- 422 (16) Yuan, S. H.; Fan, Y.; Zhang, Y. C.; Tong, M.; Liao, P. Pd-catalytic in situ
423 generation of H₂O₂ from H₂ and O₂ produced by water electrolysis for the efficient
424 electro-Fenton degradation of rhodamine B. *Environ. Sci. Technol.* **2011**, *45*,
425 8514–8520.
- 426 (17) Burbano, A. A.; Dionysiou, D. D.; Suidan, M. T.; Richardson, T. L. Oxidation
427 kinetics and effect of pH on the degradation of MTBE with Fenton reagent. *Water Res.*
428 **2005**, *39*, 107–118.
- 429 (18) Burbano, A. A.; Dionysiou, D. D.; Suidan, M. T. Effect of
430 oxidant-to-substrate ratios on the degradation of MTBE with Fenton reagent. *Water*
431 *Res.* **2008**, *42*, 3225–3239.
- 432 (19) Wang, Q.; Lemley, A. T. Kinetic model and optimization of 2,4-D
433 degradation by anodic fenton treatment. *Environ. Sci. Technol.* **2001**, *35*, 4509–4514.
- 434 (20) Kong, L. J.; Lemley, A. T. Kinetic modeling of 2,4-dichlorophenoxyacetic
435 acid (2,4-D) degradation in soil slurry by anodic fenton treatment. *J. Agric. Food*
436 *Chem.* **2006**, *54*, 3941–3950.
- 437 (21) Joo, S. H.; Feitz, A. J.; Sedlak, D. L.; Waite, T. D. Quantification of the
438 oxidizing capacity of nanoparticulate zero-valent iron. *Environ. Sci. Technol.* **2005**, *39*,

439 1263–1268.

440 (22) Navalon, S.; Alvaro, M.; Garcia, H. Heterogeneous fentoncatalysts based on
441 clays, silicas and zeolites. *Appl. Catal. B Environ.* **2010**, *99*, 1–26.

442 (23) Brillas, E.; Sires, I.; Oturan, A. Electro-Fenton process and related
443 electrochemical technologies based on Fenton's reaction chemistry. *Chem. Rev.* **2009**,
444 *109*, 6570–6631.

445 (24) Chaplin, B. P.; Reinhard, M.; Schneider, W. F.; Schuth, C.; Shapley, J. R.;
446 Strathmann, T. J.; Werth, C. J. Critical review of Pd-based catalytic treatment of
447 priority contaminants in water. *Environ. Sci. Technol.* **2012**, *46*, 3655–3670.

448 (25) He, F.; Zhao, D. Y. Hydrodechlorination of trichloroethene using stabilized
449 Fe-Pd nanoparticles: Reaction mechanism and effects of stabilizers, catalysts and
450 reaction conditions. *Appl. Catal. B Environ.* **2008**, *84*, 533–540.

451 (26) Luo, H. P.; Jin, S.; Fallgren, P. H.; Colberg, P. J. S.; Johnson, P. A.
452 Prevention of iron passivation and enhancement of nitrate reduction by electron
453 supplementation. *Chem. Eng. J.* **2010**, *160*, 185–189.

454 (27) Henderson, A. D.; Demond, A. H. Long-term performance of zero-valent iron
455 permeable reactive barriers: A critical review. *Environ. Eng. Sci.* **2007**, *24*, 401–423.

456 (28) Alonso, F.; Beletskaya, I. P.; Yus, M. Metal-mediated reductive
457 hydrodehalogenation of organic halides. *Chem. Rev.* **2002**, *102*, 4009–4091.

458 (29) Kim, S.; Kim, J.; and Moon, I. Parameter-based model for the forecasting of
459 pipe corrosion in refinery plants. *Ind. Eng. Chem. Res.* **2011**, *50*, 12626–12629.

460 (30) Sudrabin, L. P.; Marks, H. C. Cathodic protection of steel in contact with

461 water. *Ind. Eng. Chem.*, **1952**, *44*, 1786–1791.

462 (31) Pedefferri, P. Cathodic protection and cathodic prevention. *Constr. Build.*
463 *Mater.* **1996**, *10*, 391–402.

464 (32) Metwally, I. A.; Al-Mandhari, H. M.; Nadir, A. G. Factors affecting
465 cathodic-protection interference. *Eng. Anal. Bound. Elem.* **2007**, *31*, 485–493.

466 (33) Guillard, C.; Charton, N.; Pichat, P. Degradation mechanism of *t*-butyl methyl
467 ether (MTBE) in atmospheric droplets. *Chemosphere* **2003**, *53*, 469–477.

468 (34) Eisenberg, G. Colorimetric determination of hydrogen peroxide. *Ind. Eng.*
469 *Chem. Anal. Ed.* **1943**, *15*, 327–328.

470 (35) Tai, C.; Peng, J. F.; Liu, J. F.; Jiang, G. B.; Zou, H. Determination of
471 hydroxyl radicals in advanced oxidation processes with dimethyl sulfoxide trapping
472 and liquid chromatography. *Anal. Chim. Acta* **2004**, *527*, 73–80.

473 (36) Komadel, P.; Stucki, J. W. Quantitative assay of minerals for iron(II) and
474 iron(III) using 1,10-phenanthroline. III. A rapid photochemical method. *Clays Clay*
475 *Miner.* **1988**, *36*, 379–381.

476 (37) Arnold, S. M.; Hickey, W. J.; Harris, R. F. Degradation of atrazine by
477 Fenton's reagent: Condition optimization and product quantification. *Environ. Sci.*
478 *Technol.* **1995**, *29*, 2083–2089.

479 (38) Buxton, G. V.; Greenstock, C. L.; Helman, W. P.; Ross, A. B. Critical review
480 of rate constants for reactions of hydrated electrons, hydrogen atoms and hydroxyl
481 radicals ($\bullet\text{OH}/\bullet\text{O}^{\bullet}$) in aqueous solution. *J. Phys. Chem. Ref. Data* **1988**, *17*, 513–886.

482 (39) US EPA. *Technologies for treating MTBE and other fuel oxygenates*, 2005,

483 <http://www.epa.gov/tio/download/remed/542r04009/542r04009.pdf>.

484 (40) Hong, S.; Zhang, H.; Duttweiler, C. M.; Lemley, A. T. Degradation of methyl
485 tertiary-butyl ether (MTBE) by anodic Fenton treatment. *J. Hazard. Mater.* **2007**, *144*,
486 29–40.

487 (41) Liang, C.; Guo, Y. Y.; Chien, Y. C.; Wu, Y. J. Oxidative degradation of MTBE
488 by pyrite-activated persulfate: Proposed reaction pathways. *Ind. Eng. Chem. Res.*
489 **2010**, *49*, 8858–8864.

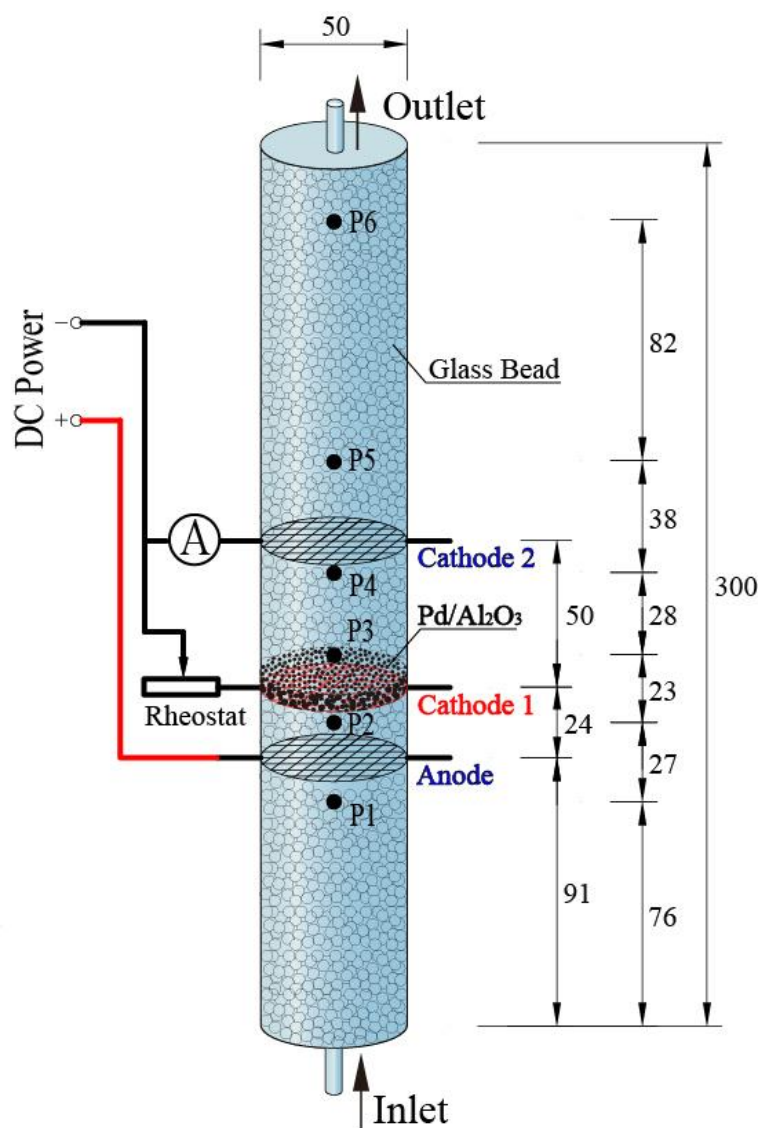
490 (42) Uhlig, H. H.; Revie, R. W. *Corrosion and Corrosion Control*, John Wiley:
491 New York, 1985.

492 (43) Bergendahl, J. A.; Thies, T. P. Fenton's oxidation of MTBE with zero-valent
493 iron. *Water Res.* **2004**, *38*, 327–334.

494 (44) Fontes, D. E.; Mills, A. L.; Hornberger, G. M.; Herman, J. S. Physical and
495 chemical factors influencing transport of microorganisms through porous-media. *Appl.*
496 *Environ. Microbiol.* **1991**, *57*, 2473–2481.

497

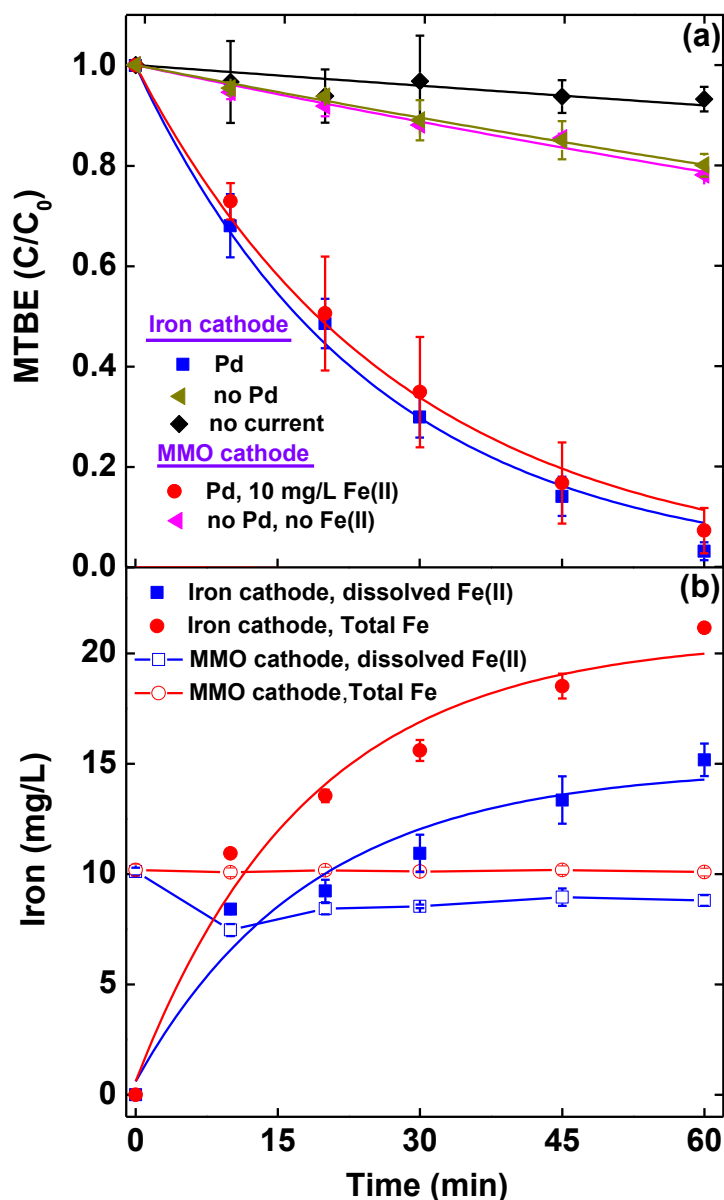
498



500

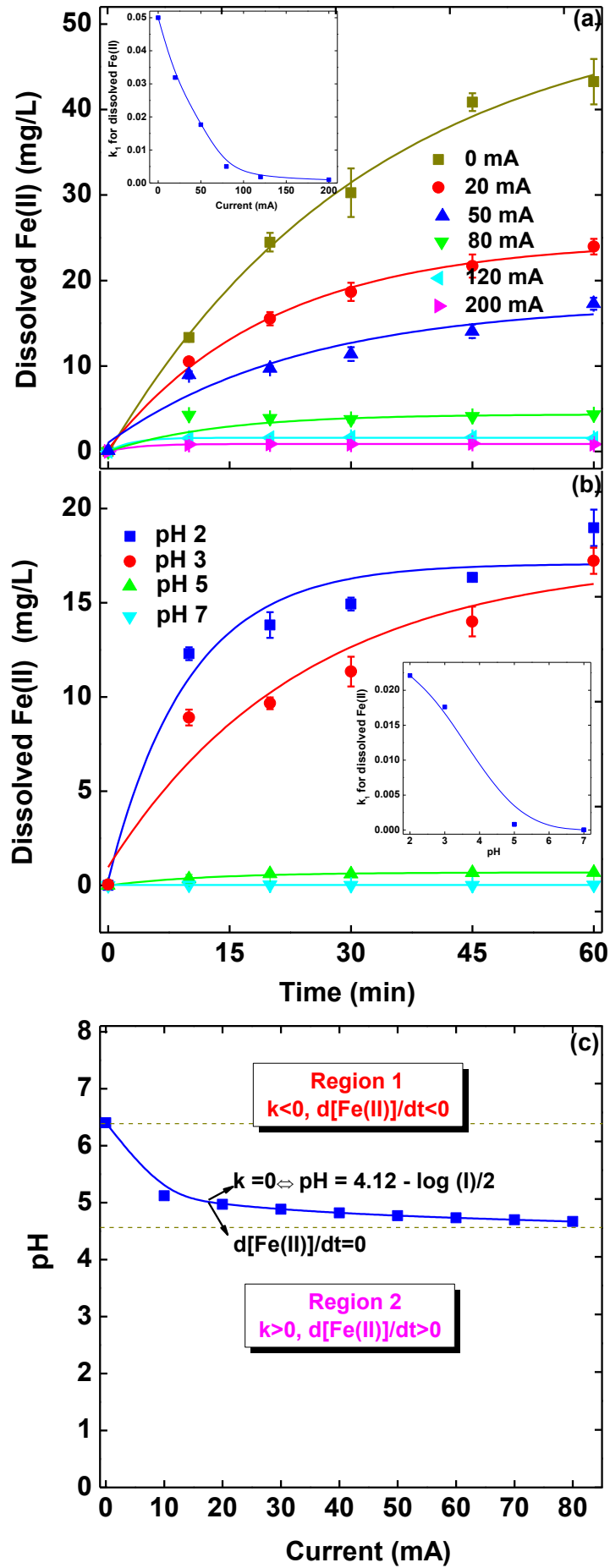
501 **Figure 1** A schematic of the modified three-electrode column using an iron cathode.

502 P1 to P6 refers to six sample ports. The unit of dimensions is in millimeters.

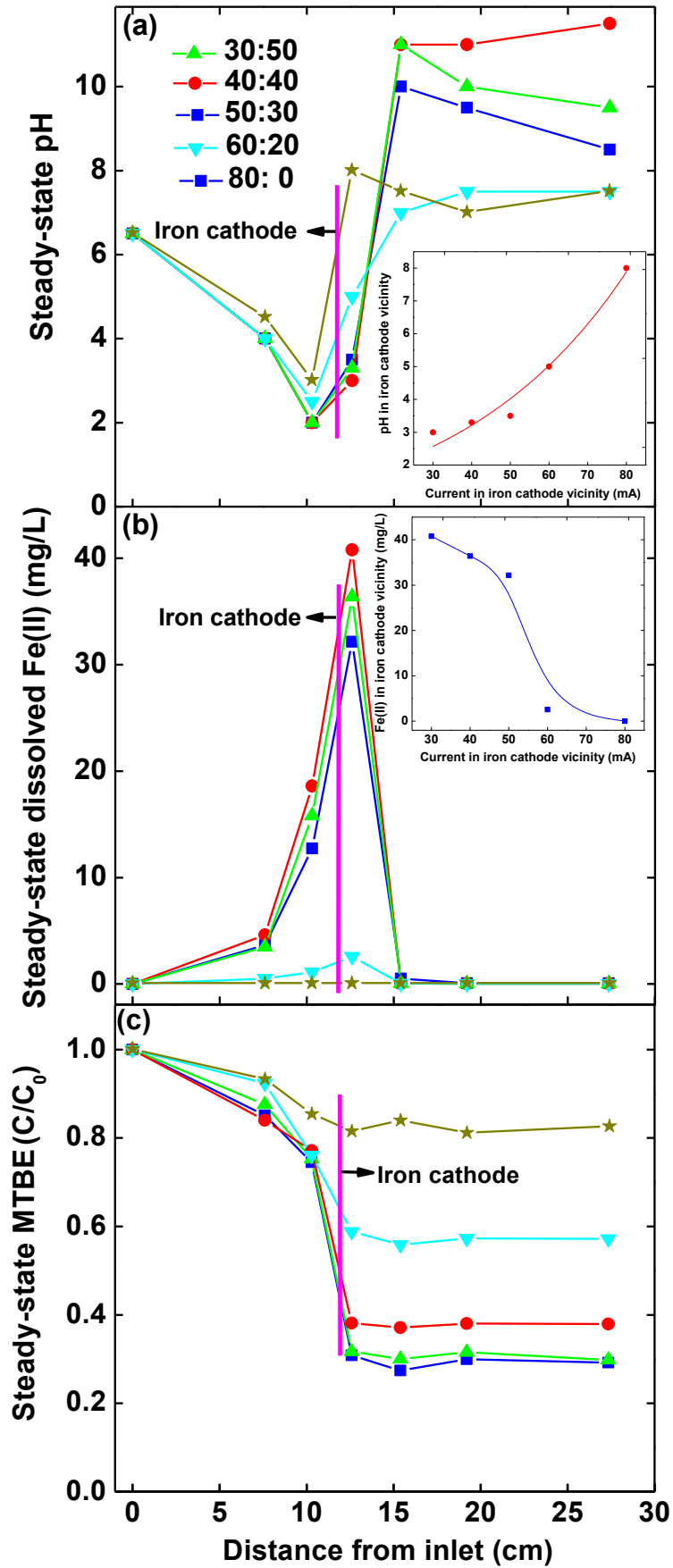


503

504 **Figure 2** (a) Transformation of MTBE and (b) variations of iron species concentration
 505 in the electrolytic cell. The reaction conditions using the iron cathode are based on 20
 506 mg/L initial MTBE concentration, pH 3.0, 1 g/L Pd/Al₂O₃ and 10 mM Na₂SO₄
 507 background electrolyte. The reaction conditions using the MMO cathode are the same
 508 as using the iron cathode except that 10 mg/L Fe²⁺ was added. Curves refer to
 509 pseudo-first-order kinetic fittings. Error bars indicate 95% confidence intervals.



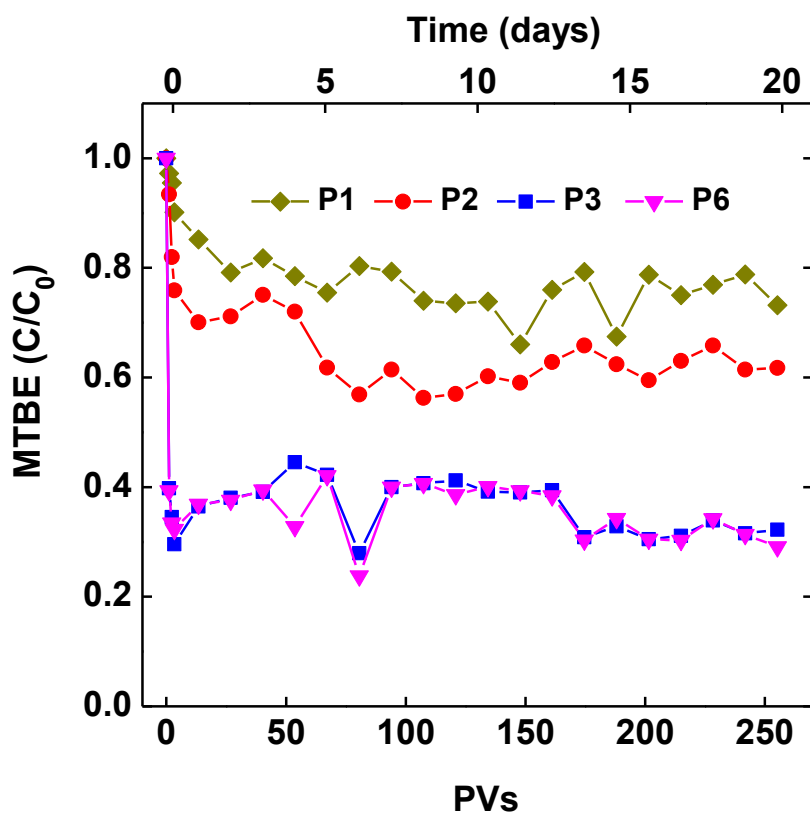
511 **Figure 3** Effects of (a) current and (b) pH on accumulation of Fe^{2+} using the iron
512 cathode, and (c) relationship of current and pH for accumulation of Fe^{2+} . Unless
513 otherwise specified, the reaction conditions are based on pH 3, 50 mA current and 10
514 mM Na_2SO_4 background electrolyte. Curves refer to pseudo-first-order kinetic fittings.
515 The insets in (a) and (b) illustrate the pseudo-first-order rate constants of Fe^{2+}
516 accumulation versus current and pH, respectively. Error bars indicate 95% confidence
517 intervals.



518

519 **Figure 4** Effects of current partition on the steady-state variations of (a) MTBE, (b)

520 pH and (c) dissolved Fe^{2+} concentration along the column. The inset in (b) illustrates
 521 the linear regression between pH and current in the vicinity of iron cathode. The inset
 522 in (c) denotes the linear regression between Fe^{2+} concentration and current in the
 523 vicinity of iron cathode. Operation conditions are based on 10 mg/L initial MTBE
 524 concentration, 2.5 mL/min and a total current of 80 mA with different partitions in
 525 cathodes 1 and 2.
 526



527
 528 **Figure 5** Long-term performance of the three-electrode column using an iron cathode
 529 for MTBE transformation. P1, P2, P3 and P6 denote different sampling ports (see
 530 [Figure 1](#) for details). Operation conditions are based on 10 mg/L initial MTBE
 531 concentration, 2.5 mL/min and a total current of 80 mA with 50 and 30 mA
 532 partitioned in cathodes 1 and 2, respectively.

533

Supporting Information

534

535

536 **Regulation of Electrochemically Generated Ferrous Ions from an** 537 **Iron Cathode for Pd-Catalytic Transformation of MTBE in** 538 **Groundwater**

539 *Peng Liao[†], Songhu Yuan^{*†}, Mingjie Chen[‡], Man Tong[†], Wenjing Xie[†], Peng Zhang[†]*

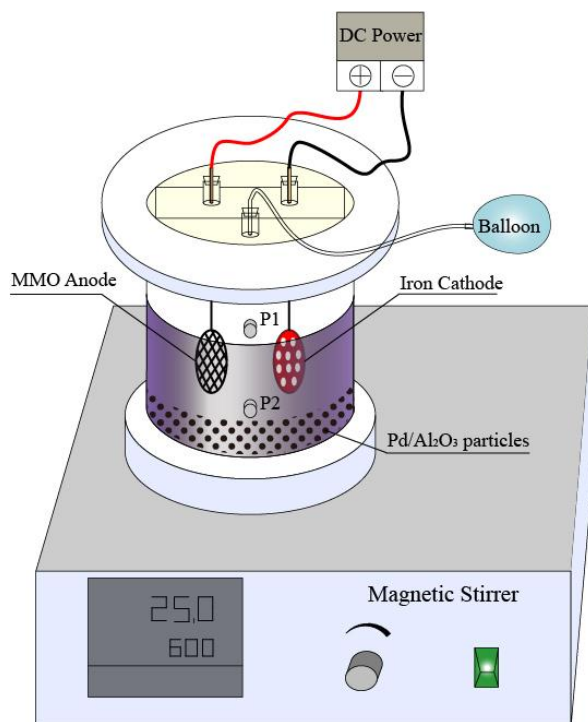
540 [†]State Key Lab of Biogeology and Environmental Geology, China University of
541 Geosciences, 388 Lumo Road, Wuhan 430074, P. R. China

542 [‡]Atmospheric, Earth and Energy Division, Lawrence Livermore National Laboratory,
543 P.O. Box 808, L-184, Livermore, CA 94550, United States

544

545 Supporting information includes 10 figures and 1 table.

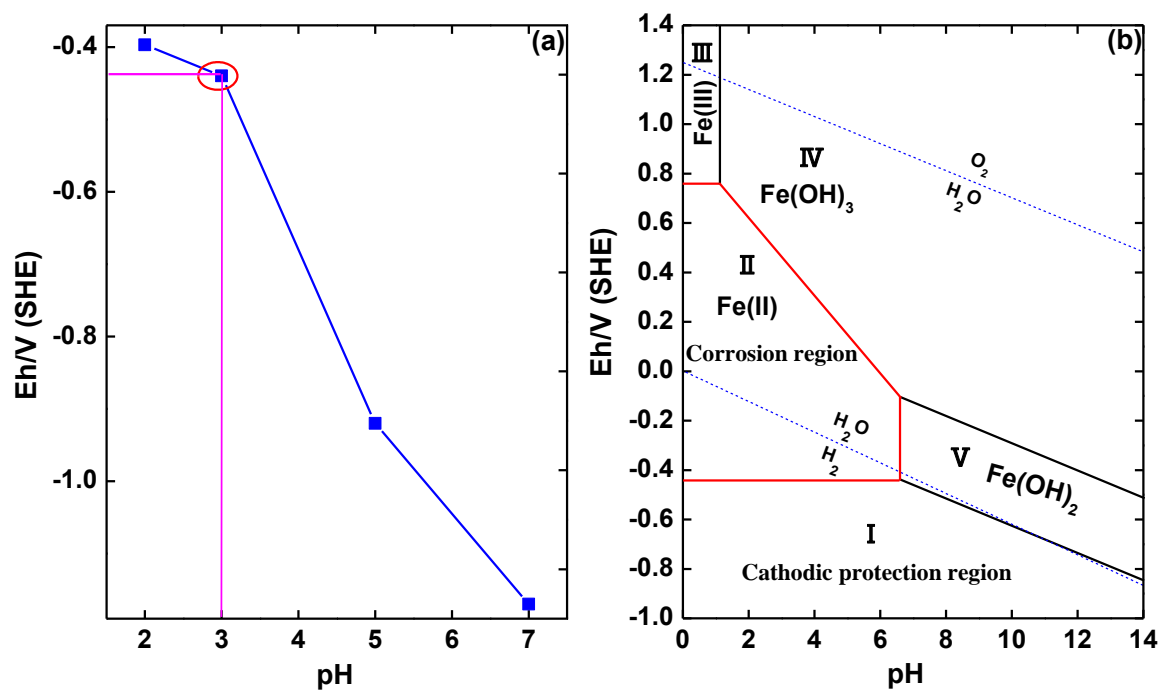
546



547

548 **Figure S1** Experimental setup used for MTBE degradation.

549

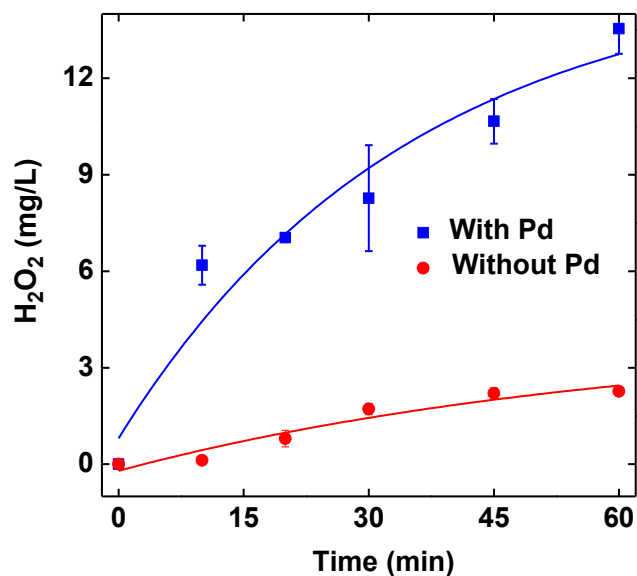


550

551 **Figure S2** (a) The corrosion potentials of the iron electrode at different pHs, and (b)

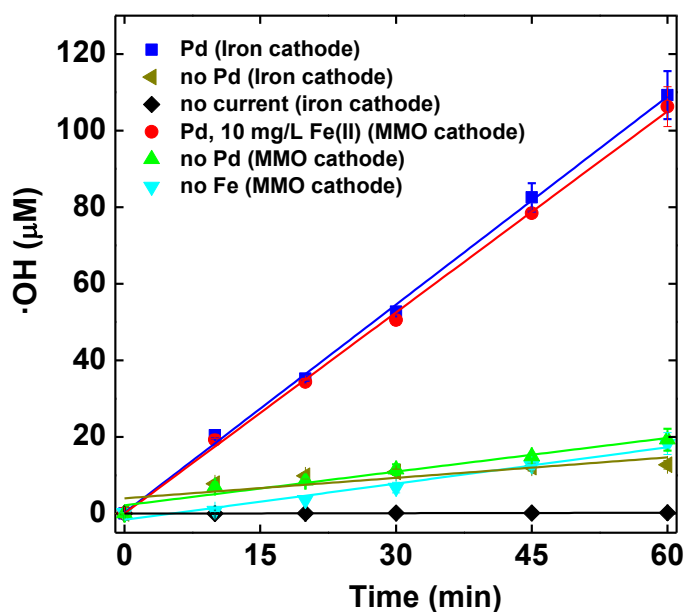
552 Eh-pH diagram for the Fe-H₂O system at 298 K.¹

553



554

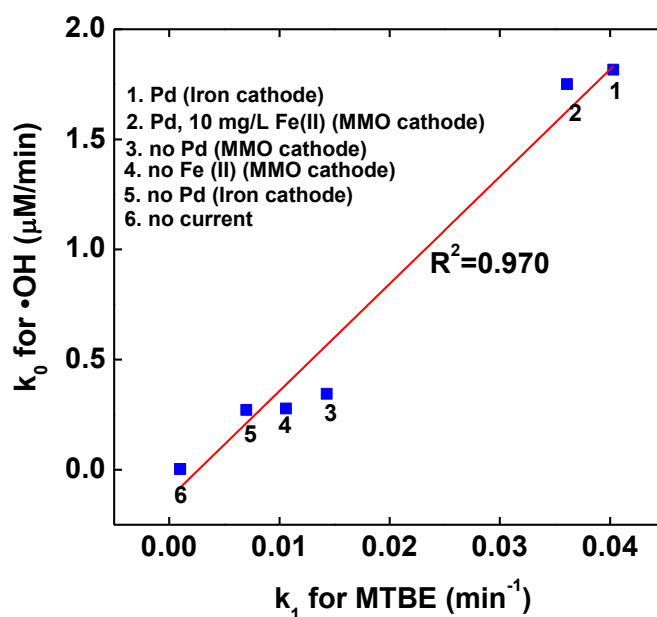
555 **Figure S3** H₂O₂ production using the MMO cathode in the batch experiment. The
 556 reaction conditions are based on 20 mg/L initial MTBE concentration, pH 3.0, 1 g/L
 557 Pd/Al₂O₃ and 10 mM Na₂SO₄ background electrolyte. Curves refer to
 558 pseudo-first-order kinetic fittings. Error bars indicate 95% confidence intervals.



559

560 **Figure S4** Cumulative •OH concentrations under different conditions. The reaction

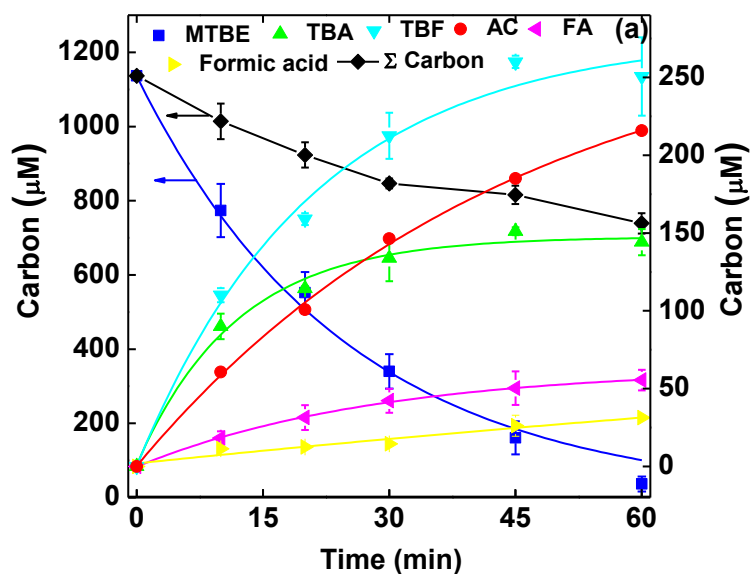
561 conditions are the same as in Figure 2 in the manuscript. Lines refer to
562 pseudo-zero-order kinetic fittings, which is given by $C_t = k_0t$, where t is the reaction
563 time (min), k_0 is the zero-order rate constant ($\mu\text{M}/\text{min}$), and C_0 and C_t are the
564 concentrations (μM) at times of $t = 0$ and $t = t$, respectively. Error bars indicate 95%
565 confidence intervals.



566

567 **Figure S5** Correlation of $\bullet\text{OH}$ production rate constants with MTBE transformation
568 rate constants.

569

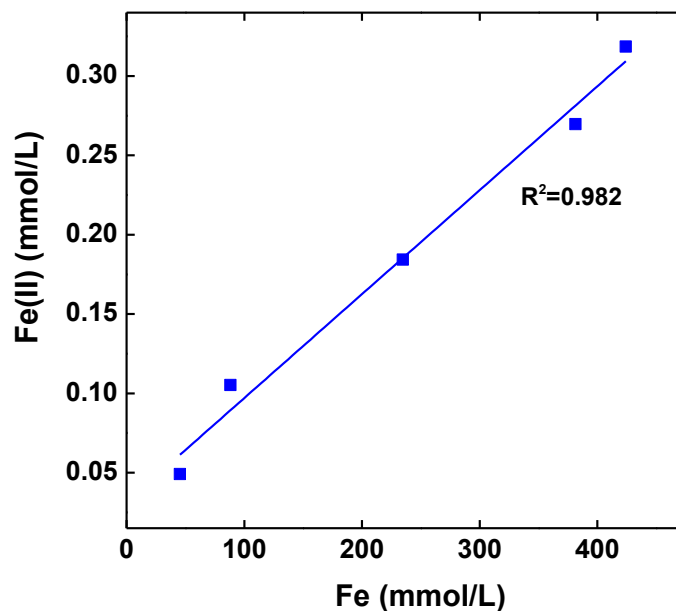


570

571 **Figure S6** Transformation profiles of MTBE using the iron cathode. The reaction
 572 conditions are the same as in Figure 2 in the manuscript.

573

574



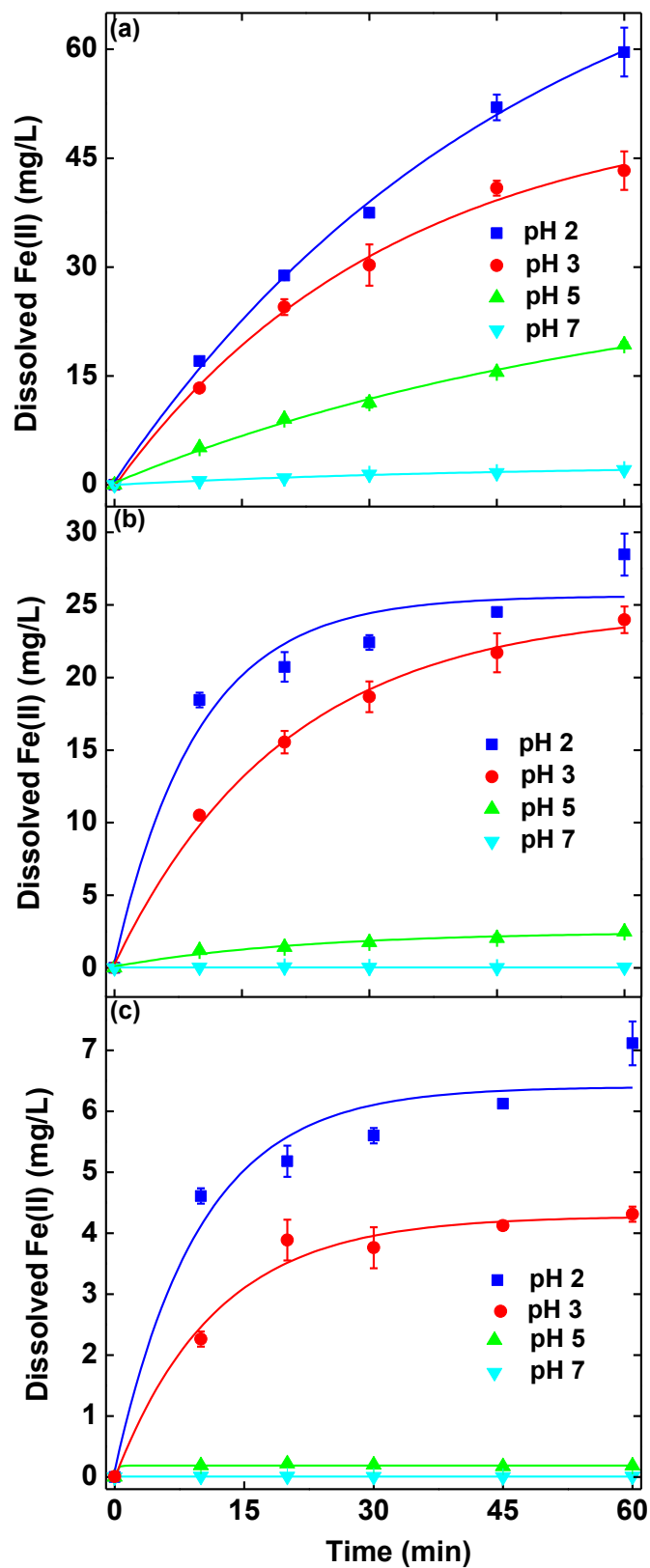
575

576 **Figure S7** Correlation of Fe(II) concentration with Fe concentration. The reaction
 577 conditions are based on pH 3, 50 mA current, 10 mM Na₂SO₄ background electrolyte,

578 and 60 min reaction. The different Fe concentrations are expressed by the different

579 surface areas of the iron electrode immersed in the solution.

580



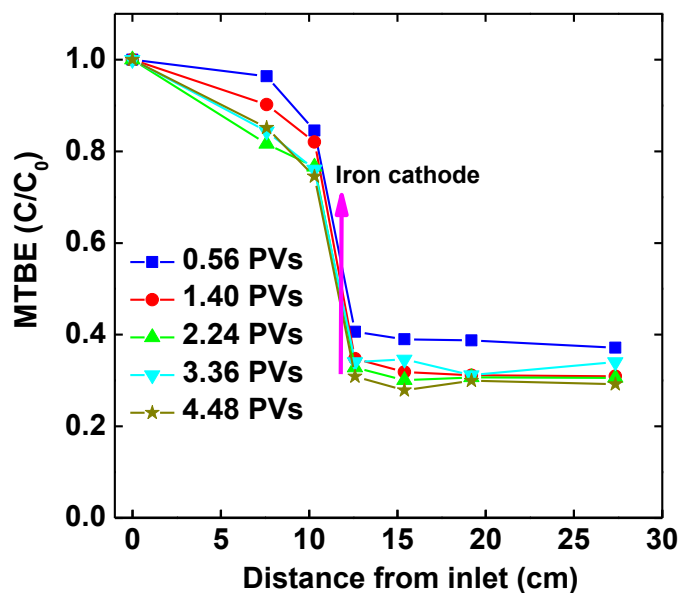
581

582 **Figure S8** Effect of pH on the accumulation of Fe(II) using the iron cathode at (a) 0,

583 (b) 20 and (c) 80 mA. The reaction conditions are based on 10 mM Na₂SO₄

584 background electrolyte. Curves refer to pseudo-first-order kintic fittings. Error bars
585 indicate 95% confidence intervals.

586



587

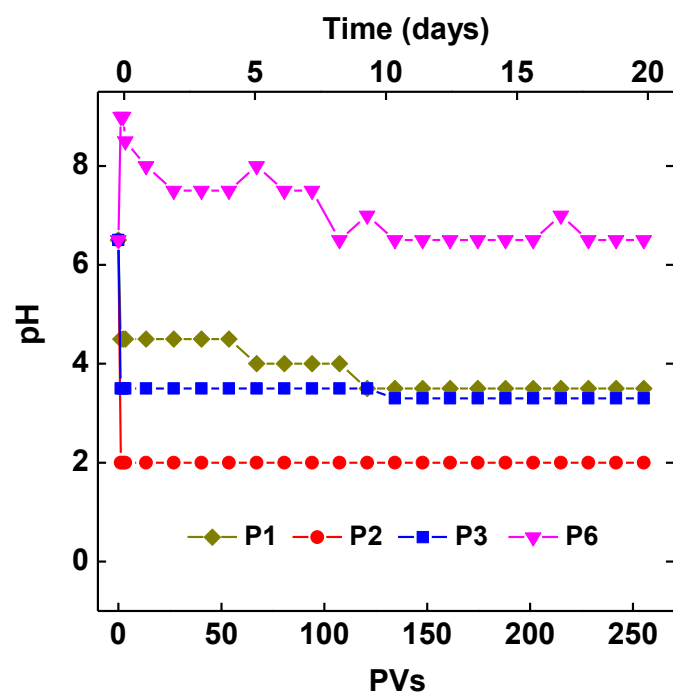
588 **Figure S9** Typical variations of MTBE concentrations versus time in the column.

589 Operation conditions are based on 10 mg/L initial MTBE concentration, 2.5 mL/min

590 and a total current of 80 mA with 50 and 30 mA partitioned in cathodes 1 and 2,

591 respectively.

592



593

594 **Figure S10** Variations of pH in the modified three-electrode column using an iron

595 cathode. P1, P2, P3 and P6 denote different sample ports (See [Figure 1](#) for details).

596 Operation conditions are the same as in [Figure 5](#) in the manuscript.

Table S1 Parameters and Results Associated with the Column Experiments.

No. ^a	Column description	Steady-state pH		Transformation efficiency ^b	Cell voltage	Energy consumption
		Iron cathode vicinity	Effluent	(%)	(V)	(kW•h/g _(MTBE))
C1	Pd/Al ₂ O ₃ , no current	6.5–7	6.5–7	4	—	—
C2	no Pd/Al ₂ O ₃	3.5–4.5	8.5–9	35	23.8–29.3	1.8–2.2
C3	Pd/Al ₂ O ₃	3.5–4	8–8.5	70.9	22.7–28.8	0.9–1.1
C4 ^c	Pd/Al ₂ O ₃	3–3.5	11–11.5	62.1	31.7–33.4	1.2–1.3
C5 ^c	Pd/Al ₂ O ₃	3.5–4	9.5–10	70.2	30.5–31.6	1.3–1.4
C6 ^c	Pd/Al ₂ O ₃	5–5.5	7–7.5	42.8	29.4–31.5	1.8–2.0
C7 ^c	Pd/Al ₂ O ₃	7–8	7.5–8	19	17.2–18.7	1.7–1.8
C8 ^d	Pd/Al ₂ O ₃	3.5–3.2	6.5–7	70	25.6–30.2	1.0–1.2

^aFor C1–C8, 5 mM Na₂SO₄ and 0.5 mM CaSO₄ were dissolved in groundwater as background electrolytes. ^bTransformation efficiency refers to total removal percentages of MTBE in the columns under steady state. ^cThe cathode current partitions between cathodes 1 and 2 from C4 to C7 are

30:50, 40:40, 60:20, and 80:0, respectively. ^dA much longer period of 20 d is lasted. Unless otherwise specified, the operation conditions are based on 10 mg/L initial MTBE concentration and 2.5 mL/min, and the total current of 80 mA is applied in cathodes 1 (50 mA) and 2 (30 mA).

597 **References**

598 (1) Uhlig, H. H.; Revie, R. W. *Corrosion and Corrosion Control*, John Wiley:

599 New York, 1985.

600