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## **ASSESSMENT OF GENOTOXIC ACTIVITY OF PETROLEUM HYDROCARBON-BIOREMEDIATED SOIL**

Grażyna Płaza<sup>1</sup>, Grzegorz Nałęcz-Jawecki<sup>2</sup>, Krzysztof Ulfig<sup>1</sup>, Robin L. Brigmon<sup>3</sup>

<sup>1</sup> - Institute for Ecology of Industrial Areas, 40-844 Katowice, Kossutha 6 Street, Poland

<sup>2</sup> - Department of Environmental Health Sciences, Medical University, Banacha 1 Street,  
02-097 Warszawa, Poland

<sup>3</sup> - Savannah River National Laboratory, Aiken, South Carolina, 29808, USA

## **Abstract**

The relationship between toxicity and soil contamination must be understood to develop reliable indicators of environmental restoration for bioremediation. Two bacterial rapid bioassays: SOS chromotest and umu-test with and without metabolic activation (S-9 mixture) were used to evaluate genotoxicity of petroleum hydrocarbon-contaminated soil following bioremediation treatment. The soil was taken from an engineered biopile at the Czor Polish oil refinery. The bioremediation process in the biopile lasted 4 years, and the toxicity measurements were done after this treatment. Carcinogens detected in the soil, polyaromatic hydrocarbons (PAHs), were reduced to low concentrations (2 mg/kg dry wt) by the bioremediation process. Genotoxicity was not observed for soils tested with and without metabolic activation by a liver homogenate (S-9 mixture). However, umu-test was more sensitive than SOS-chromotest in the analysis of petroleum hydrocarbon-bioremediated soil. Analytical results of soil used in the bioassays confirmed that the bioremediation process reduced 81% of the petroleum hydrocarbons including PAHs. We conclude that the combined test systems employed in this study are useful tools for the genotoxic examination of remediated petroleum hydrocarbon-contaminated soil.

**Keywords:** genotoxicity, petroleum hydrocarbons, bioremediation, biopile, PAHs

## 1. Introduction

The success of bioremediation is determined by the metabolic potential of microorganisms to detoxify/utilize organic contaminants or mineralize them to CO<sub>2</sub>, H<sub>2</sub>O and biomass. The process is highly dependent on biodegradability and bioavailability of contaminants as well as environmental parameters including pH, temperature, and nutrient availability (ref). Bioremediation is an attractive, environmentally friendly, and relatively cost-effective alternative to conventional physicochemical soil and water treatment techniques (Bouwer *et al.*, 1994; Barthe, 1986). There are a variety of *in situ* and *ex situ* bioremediation approaches, including biopiling, bioreactors, bioventing, biosparging and landfarming that can be applied to optimize biodegradation (Brigmon *et al.* 2002, Hazen, 1997; Bouwer *et al.*, 1994).

Considering different existing bioremediation strategies and characteristics of the lagoon materials at the Czechowice-Dziedzice refinery, the *ex situ/on site* bioremediation in biopile was selected to clean-up soil heavily contaminated with petroleum waste (Plaza et al, 2003). As a result of bioremediation no secondary waste stream is created, contaminants are safely treated *in situ* therefore minimizing handling and associated health hazards, and the reuse of the previously contaminated site without restriction is possible.

A number of rapid test systems to detect environmental carcinogens have been devised for soil and water analyses. The umu-test and SOS chromotest have been widely employed for soil testing with successful results (Reifferscheid et al. 1991). The SOS chromotest described by Quilardet *et al.* (1982) uses one of the SOS (define) genes, *sfi A* focused to *lac Z* on the chromosome of *E. coli*, and overcomes the drawback in the Ames test. The umu-test is based on the use of genetically engineered bacteria, *Salmonella typhimurium* TA 1535 pSK 1002. The umu-test has been standardized and validated by German DIN (DIN 38 415 T3) and on the international level by ISO (ISO/DIS 13829, 2000). In this project genotoxicity assessment of remediated hydrocarbon-contaminated soil was performed with the SOS-chromotest and umu-tests.

## 2. Materials and methods

### 2.1. Field description

More than a century of sulphuric acid-based oil refining process by the CZOR refinery and associated waste disposal has produced an estimated 120,000 tons of acidic, highly weathered, petroleum sludge and wastewater. This waste has been deposited into three open waste lagoons, 3 meters deep and covering 3.8 hectares. One of the waste lagoons (0.3 hectare) was chosen for aerobic biopile demonstration (Plaza et al. 2003). The waste from this lagoon was removed, and heavily petroleum contaminated soil was engineered into a biopile for bioremediation. The biopile was constructed with actively and passively aerated sections in 1997 (Plaza et al. 2004). The project focus was on evaluation of novel environmental restoration technologies and research for effective bioremediation of heavily contaminated petroleum waste soils. This was accomplished by comparing bioremediation processes and the contaminant removal rates of both easily biodegradable light petroleum hydrocarbons and more recalcitrant polyaromatic hydrocarbons (PAHs) under active vs. passive aeration. The application efficiently employed cost-effective additives, including mineral NPK fertilizers, the surfactant Rokafenol N8, and employing indigenous microbial consortium for enhanced biodegradation of soil hydrocarbons in the biopile.

The sampling strategy was similar to that applied in a previous project (Altman *et al.*, 1997; IETU, 1999). The total number of soil sampling locations at the biopile was twenty-three. Soil was sampled (ca. 1 kg) from shallow (ca. 30-40 cm) and deep (ca. 80-100 cm) layers. The soil sampling locations in the biopile are presented in Figure 1.

## 2.2. Characterization of soil samples

Each soil sample was examined for the following physicochemical parameters: TPH – (total petroleum hydrocarbons), non-polar aliphatic hydrocarbons; TPOC – (total petroleum organic carbon); POLAR - (polar aliphatic petroleum compounds); total PAHs – (polycyclic aromatic hydrocarbons), 16 PAHs according to EPA (EPA Method 3620, 1992; EPA Method 8440, 1995); pH in H<sub>2</sub>O; pH in 1 M KCl; conductivity; N-NH<sub>4</sub>, N-NO<sub>2</sub>, N-NO<sub>3</sub>, PO<sub>4</sub>, Ca, Mg, C<sub>TOT</sub>, N<sub>TOT</sub>, P<sub>TOT</sub>, P<sub>2</sub>O<sub>5</sub>, K<sub>2</sub>O, and metals were measured according to ISO standards. Soil microbial activity was measured by the dehydrogenase activity according to Alef and

Nannipieri (1996). This method is based on the use of TTC (triphenyltetrazolium chloride)) as an artificial electron acceptor.

### 2.3. *Extraction of soil samples*

All soil samples from each location were mixed and split for the two bioassays. The remediated soil extraction procedure was performed as described by McDaniel *et al.* (1993). Five grams of soil was extracted for 16 h in a Soxhlet extractor with methylene chloride. A rotary evaporator was then used to reduce the extract volume to 5 ml. One ml of DMSO (dimethyl sulfoxide)) was then added and the resultant mixture evaporated under reduced pressure to 1 ml. The exchange solvent – DMSO was selected as it is not toxic and genotoxic to the bacteria tested.

### 2.4. *Genotoxicity bioassays*

The SOS-chromotest with *Escherichia coli* PQ37 was performed on the soils in 20 ml sterile glass test tubes with and without S9 metabolic activation, according to a modified method described by Quillardet and Hofnung (1985). Umu test with *Salmonella typhimurium* TA1535/pSK1002 was carried out on the soils according to the method of Oda *et al.* (1985). The incubation mixture consisted of 5, 10 µl (SOS) or 20, 30 µl (umu) extracts, log-phase bacterial culture and S9 mix containing 4% of S9. These bioassay results can be influenced by sample color so two different procedures were applied to eliminate this bias. In case of the umu test the bacterial suspension was diluted 10-fold with the fresh medium after 2 h of exposure, followed by an additional incubation period of 2 h. In the SOS-chromotest the bacterial suspension was centrifuged and rinsed twice with fresh medium. All soil samples were tested in triplicate.

## 3. **Results**

Results of soil physical-chemical analyses of the biopile soil used in these bioassays are presented in Table 1. The soil characterization results demonstrate contaminant concentrations necessary to determine test endpoint differences. The soil pH was 6.13-7.18

with a mean of 6.6. The metal concentrations varied within natural concentration ranges determined for Polish soils. The mean TPH, TPOC and PAH concentrations were 6.05 g/kg d.w., 7.38 g/kg d.w., and 2.21 mg/kg d.w., respectively. The hydrocarbon concentrations decreased 81% during bioremediation in this biopile. By the end of the 20<sup>th</sup> month of bioremediation, the TPH and PAH concentrations were reduced below the Polish risk guidelines (PIOS, 1994 not in ref list). A restored green area having lush vegetation including grass and trees now covers the biopile surface. After 48 months the active bioremediation process was terminated and monitored natural attenuation (MNA) was selected as the follow-up bioremediation strategy for this site. Decreases of petroleum hydrocarbon concentrations during the bioremediation treatment are presented in Table 2. It is calculated that 120 tons of petroleum hydrocarbons were removed during this period (IETU, 1999). The significant decrease of petroleum hydrocarbons and corresponding high dehydrogenase activity were noted during the first year of the bioremediation process (Plaza et al 2003a).

The initial air injection into the biopile after construction stimulated significant reductions of contaminants (> 50%). This reduction was likely due to increased bioavailability of the contaminants after soil mixing and addition of wood chips to allow aeration. While some of this initial reduction may have been due to volatilization, the increase in biopile biomass was correlated with soil organic contaminant reduction demonstrating biodegradation (Plaza et al. 2003). The addition of fertilizers (NPK) and surfactant – Rokafenol N8 together with air and leachate recirculation accelerated the bioremediation process. A biopile is a relatively simple and low maintenance operation. The technology provides operating conditions that minimize fugitive air emissions and maximize biodegradation rates.

The use of bioassays combined with chemical analyses at this site has yielded reliable results for risk analyses (Plaza et al, 2004a). These results demonstrate that active bioremediation can reduce the toxicity of contaminated soils using different genotoxicity tests. Bioremediation improves the soil quality by safely reducing petroleum hydrocarbons as measured by site physico-chemical parameters and specific microbiological activity (Plaza et al. 2004b).

The results of genotoxicity tests with and without metabolic activation are presented in Table 3. The sample is considered genotoxic if  $\beta$ -galactosidase induction coefficient is higher than 1.5. Induction factor (IF) is defined as the ratio of specific activity of  $\beta$ -galactosidase at a given sample concentration to specific  $\beta$ -galactosidase activity. The SOS-chromotest IF

coefficient for the tested soils was 1.09 with values ranging between 0.86-1.63. Only one IF coefficient was  $> 1.5$  out of the 23 samples, probably due to soil toxicity. Decreased phosphatase activity was also observed in this particular soil sample. Similar results were obtained in the bioassay with metabolic activation (S-9 mixture). Mean value of IF was 1.18 with ranged values between 1.00 – 1.20. Umu-test responses to the soil samples were generally similar to the SOS-chromotest. The values of IR (Define) coefficients for all test samples were low, but in tests with metabolic activation the values of IR were slightly higher. Mean values of IR coefficient were 1.02 and 0.86 with and without metabolic activation, respectively. The biopile soil was originally contaminated by potentially genotoxic compounds such as benzo(a)pyrene, dibenzo(a,h)anthracene, benzo(b)fluoranthene and benzo(k)fluoranthene). Bioassay results utilizing these PAHs, although known to be genotoxic, can vary with test organisms depending on the bioavailability (Alexander and Alexander, 2000). However PAH concentration was reduced by the time of this testing and did not cause apparent genotoxic effects here. The umu-test is considered more sensitive than SOS-chromotest. For the positive controls (2-aminoanthracene with S9 activation) the umu-test was 5-fold more sensitive than the SOS-chromotest.

#### **4. Discussion**

A biopile is a relatively simple and low maintenance operation. The technology provides operating conditions that minimize fugitive air emissions and maximize biodegradation rates. Bioremediation improves the soil quality by safely reducing petroleum hydrocarbons as measured by site physico-chemical parameters and specific microbiological activity (Płaza et al. 2003). The use of bioassays at this site combined with chemical analyses has proven to yield reliable results for genotoxic risk analyses. These results confirm the effect of active bioremediation upon reducing the genotoxicity of contaminated soils in SOS chromotest and umu-test.

This study demonstrated that genotoxic risk from treated biopile soil was not observed. The results indicated some differences between these two bioassays. Both tests were more sensitive after enzymatic activation of the soil samples. This demonstrates that the soil contaminants causing genotoxic effects, petroleum hydrocarbons and PAHs, were



transformed by enzymatic activity. Soil sample processing, for example extraction technique, is critical to contaminant bioavailability and measurements of genotoxicity (Alexander and Alexander, 2000). The results obtained from the assays are in accordance with chemical analyses. Based on these results, this bioremediation demonstration has successfully reduced the genotoxicity from the soil contaminants in the biopile.

The combination of two bioassays used here for examination of petroleum hydrocarbon-contaminated soil genotoxic activity after bioremediation was comparable to data from chemical testing. The findings that contaminants were significantly reduced matched the genotoxicity findings. The bioassays covered a wide range of mutagenic reference substances and complement each other (Quillardet and Hofnung, 1993). The results demonstrated that the umu and SOS test systems with and without metabolic activation were suitable for evaluation of soil genotoxicity. All the assays used are potentially useful in the detection of genotoxicity. However, each test showed different sensitivities to remediated soil. Thus, we suggest, as other authors have done, a battery approach for biological evaluation of genotoxicity (DECHEMA, 1995; Weisburger, 1999; Ehrlichmann et al., 2000; Bundy, 2000) or toxicity (Baun et al., 2000) of environmental samples. There is a need to develop environmentally acceptable endpoints for soil quality and an integrated approach to estimate ecological risk. The use of biological endpoints to evaluate contaminated sites remediated to regulatory cleanup levels should be incorporated into ecological risk assessment. These and other soil genotoxicity bioassays could be useful tools for development of predictive models for risk-based corrective actions (Citterio et al., 2002). These endpoints could be included in regulatory guidelines for assessing remedial effectiveness for both engineered and MNA environmental restoration programs.

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**Table 1.** Physico-chemical characteristics of the biopile soil.

<b>1. BIOPILE 1</b>							
	Unit	Mean	S.D.	V.C.	Median	MIN	MAX
pHH <sub>2</sub> O	-	6,62	0,42	0,06	6,63	6,13	7,18
pHKCL	-	7,18	0,44	0,06	7,15	6,50	7,18
COND	MS/cm at 25°C	0,75	0,24	0,32	0,71	0,44	1,16
HU	%	4,35	0,76	0,18	4,24	3,37	5,43
NNO <sub>3</sub>	mg/kg d.w.	1,10	1,27	1,15	0,54	0,02	2,77
NNO <sub>2</sub>	m/kg d.w.	0,01	0,00	0,00	0,01	0,01	0,01
NNH <sub>4</sub>	m/kg d.w.	59,94	11,98	0,20	55,80	50,70	87,17
C <sub>TOT</sub>	% d.w.	1,90	1,20	0,63	1,95	0,29	3,52
N <sub>TOT</sub>	% d.w.	0,90	0,65	0,72	1,11	0,11	1,64
C:N	-	8,41	11,55	1,37	1,85	0,24	32,00
P <sub>TOT</sub>	% d.w.	0,07	0,01	0,14	0,06	0,06	0,08
PO <sub>4</sub>	mg/kg d.w.	18,00	8,38	0,47	17,75	9,00	29,00
P <sub>2</sub> O <sub>5</sub>	mg/100g d.w.	16,45	2,85	0,17	16,60	12,16	19,81
K <sub>2</sub> O	mg/100g d.w.	14,23	3,28	0,23	13,75	11,20	21,90
S <sub>TOT</sub>	% d.w.	0,08	0,05	0,67	0,07	0,04	0,20
SSO <sub>4</sub>	mg/kg d.w.	0,04	0,02	0,57	0,03	0,02	0,08
Ca	mg/kg d.w.	3991,88	2186,64	0,55	3597,50	1875,00	8980,00
Mg	mg/kg d.w.	611,13	259,36	0,42	537,00	262,00	1018,00
TPH	g/kg d.w.	6,05	4,54	0,75	6,01	0,74	13,77
TPOC	g/kg d.w.	7,38	5,45	0,74	7,34	1,06	16,96
PAHS	mg/kg d.w.	2,22	1,28	0,58	1,87	1,00	5,24
Cd	mg/kg d.w.	2,09	0,00	0,00	2,09	2,09	2,09
Pb	mg/kg d.w.	62,06	25,92	0,42	53,95	39,80	121,60
Zn	mg/kg d.w.	112,86	18,04	0,16	111,45	86,10	137,30
Cu	m/kg d.w.	27,45	7,07	0,26	25,25	21,80	44,10
Cr	mg/kg d.w.	34,40	10,83	0,31	30,45	26,00	60,00
Ni	mg/kg d.w.	45,68	10,83	0,24	27,15	22,40	169,60
Co	mg/kg d.w.	8,24	1,48	0,18	8,10	6,60	11,50
Mn	mg/kg d.w.	274,78	45,06	0,16	282,95	222,50	339,90
Fe	mg/kg d.w.	18599,38	3882,72	0,21	17456,50	14878,00	27492,00
Hg	mg/kg d.w.	1,34	0,34	0,25	1,40	0,90	1,80
As	mg/kg d.w.	7,64	1,63	0,21	7,05	6,00	10,60

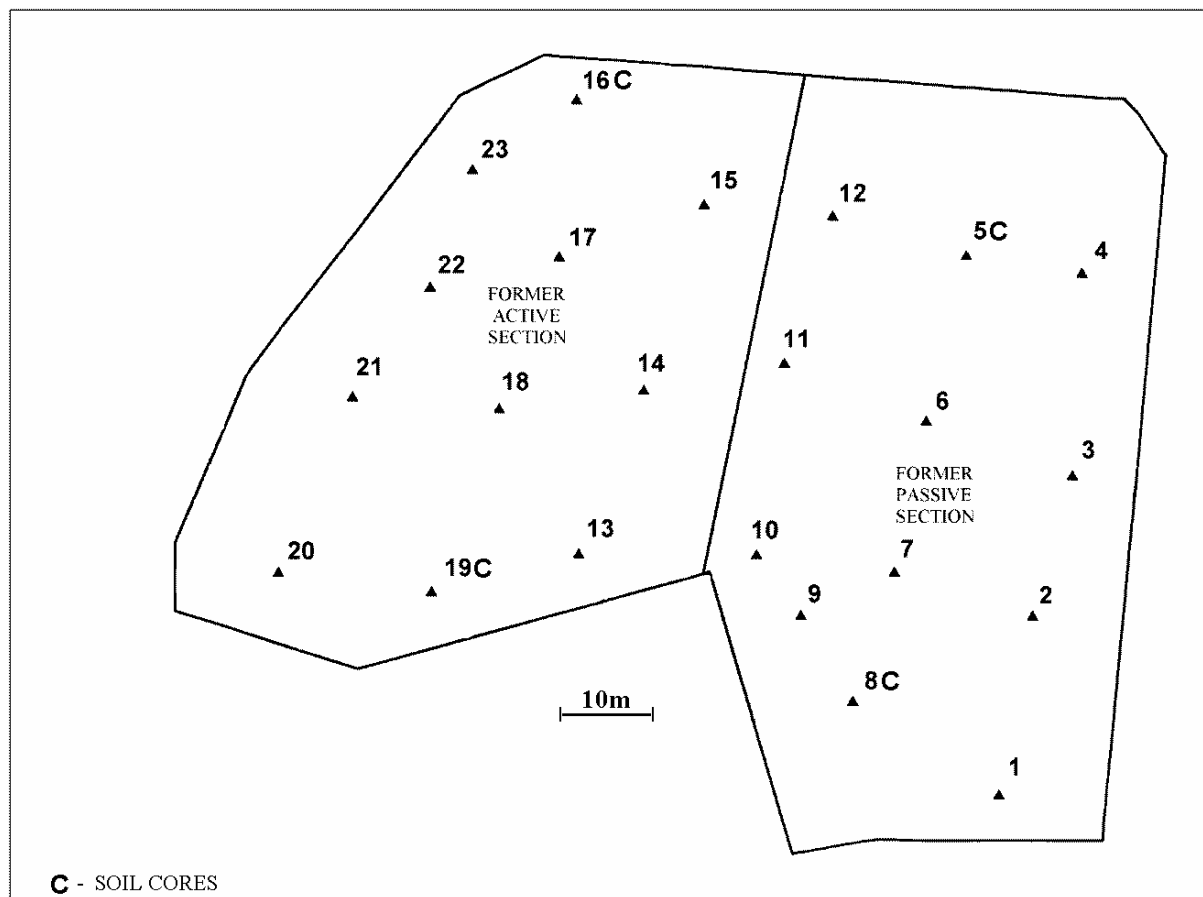
S.D. – standard deviation  
V.C. – variation coefficient

**Table 2.** Changes in biopile petroleum hydrocarbons concentrations and microbial activity (means for years) during bioremediation.

Years	<b>TTC</b>	<b>TPH</b>	<b>TPOC</b>	<b>POLAR</b>	<b>PAHs</b>
	mg TPF/g d.w.	g/kg d.w.	g/kg d.w.	g/kg d.w.	mg/kg d.w.
1	6.05	35.25	50.06	14.81	6.97
2	22.46	17.04	28.03	10.99	5.21
3	20.07	14.17	17.03	2.15	3.44
4	41.57	7.93	9.65	1.93	2.44

**Table 3.** Genotoxicity results of the biopile soil

	<b>unit</b>	<b>Mean</b>	<b>S.D.</b>	<b>V.C.</b>	<b>Median</b>	<b>MIN</b>	<b>MAX</b>
<b>SOS-chromotest</b>	IF	1,09	0,25	0,21	1,12	0,86	1,63
<b>SOS-chromotest with S-9 activation</b>	IF	1,18	0,07	0,07	1,06	1,00	1,20
<b>umu-test</b>	IR	0,86	0,07	0,09	0,89	0,74	0,93
<b>umu-test with S-9 activation</b>	IR	1,02	0,07	0,07	1,00	0,93	1,16



**Figure. 1.** Biopile sections and soil sampling locations