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Y/ER-105/V1 & D2

**Remedial Investigation Work Plan for Bear Creek Valley
Operable Unit 1 (S-3 Ponds, Boneyard/Burnyard, Oil
Landfarm, Sanitary Landfill I, and the Burial Grounds,
Including Oil Retention Ponds 1 and 2)
at the Oak Ridge Y-12 Plant, Oak Ridge, Tennessee**

Volume 1. Main Text

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DOE/OR/01-1161/V1&D2
Y/ER-105/V1&D2

Energy Systems Environmental Restoration Program
Y-12 Environmental Restoration Program

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Oak Ridge, Tennessee 37831-8169
managed by
MARTIN MARIETTA ENERGY SYSTEMS, INC.
for the
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ACRONYMS

AEA	Atomic Energy Act
ALARA	as low as reasonably achievable
ANSI/ASME	American National Standards Institute/American Society of Mechanical Engineers
AOC	area of contamination
ARAR	applicable and relevant or appropriate requirement
BCHR	Bear Creek Hydrologic Regime
BCK	Bear Creek Kilometer
BCV	Bear Creek Valley
BG	Burial Ground
BMAP	Biological Monitoring and Abatement Program
BNI	Bechtel National, Inc.
BOD	Biochemical Oxygen Demand
BRA	baseline risk assessment
BYBY	Boneyard/Burnyard
CAA	Clean Air Act
CAMU	Corrective Action Management Unit
CEC	cation exchange capacity
CERCLA	Comprehensive Environmental Response, Compensation, and Liability Act
<i>CFR</i>	<i>Code of Federal Regulations</i>
CLP	Contract Laboratory Program
CMS	Corrective Measures Study
COCs	contaminants of concern
COPC	chemicals of potential concern
CPR	cardiopulmonary resuscitation
CSFs	carcinogenic slope factors
CWA	Clean Water Act
DCGs	Derived Concentration Guides
DNAPL	dense nonaqueous phase liquid
DOE	U.S. Department of Energy
DOE-ORO	U.S. Department of Energy Oak Ridge Operations Office
DOT	U.S. Department of Transportation
DQO	Data Quality Objective
EDTA	ethylenediaminetetracetic acid
EM	electromagnetic induction
Energy Systems	Martin Marietta Energy Systems, Inc.
EPA	U.S. Environmental Protection Agency
EQR	rinsates
ER	Environmental Restoration
ERA	ecological risk assessment
ET	East Tributary
FBL	field blanks
FDA	U.S. Food and Drug Administration
FFA	Federal Facilities Agreement

FFCA	mixed-waste Federal Facilities Compliance Agreement
FID	flame ionization detector
FR	<i>Federal Register</i>
FS	Feasibility Study
FSA	field staging area
GC/MS	gas chromatograph/mass spectrophotometer
GM	Geiger-Mueller
GWQAR	groundwater quality assessment report
HAZMAT	hazardous waste and hazardous materials
HCDA	Hazardous Chemical Disposal Area
HSWA	Hazardous and Solid Waste Amendments
HWDU	Hazardous Waste Disposal Unit
ICP	inductively coupled plasma
IDLH	Immediately Dangerous to Life or Health
IDW	investigation-derived waste
IP	integration point
IRIS	Integrated Risk Information System
K	erodibility factor
LCS	laboratory control sample
LDRs	land disposal restrictions
LLW	low-level waste
LLWDDD	Low-Level Waste Disposal, Development, and Demonstration
LNAPL	light nonaqueous phase liquid
M&TE	measuring and test equipment
MCL	maximum contaminant level
MCLG	maximum contaminant level goal
MLE	most likely exposure
MOU	Memorandum of Understanding
MS	matrix spike
MSD	matrix spike duplicate
MSDS	Material Safety Data Sheet
MTRs	minimum technology requirements
NAAQSs	National Ambient Air Quality Standards
NCP	National Contingency Plan
NEPA	National Environmental Policy Act
NERP	National Environmental Research Park
NESHAPs	National Emission Standards for Hazardous Air Pollutants
NFI	no further investigation
NPDES	National Pollutant Discharge Elimination System
NPL	National Priorities List
NPRM	Notice of Proposed Rulemaking
NRC	U.S. Nuclear Regulatory Commission
NRDA	Natural Resource Damage Assessment
NT	North Tributary
ODW	Office of Drinking Water
OREIS	Oak Ridge Environmental Information System
ORGDP	Oak Ridge Gaseous Diffusion Plant
ORNL	Oak Ridge National Laboratory

ORP	Oil Retention Pond
ORR	Oak Ridge Reservation
OSHA	Occupational Safety and Health Administration
OU	operable unit
OVA	organic vapor analyzer
PA	Preliminary Assessment
PAH	polycyclic aromatic hydrocarbon
PARCC	precision, accuracy, representativeness, completeness, and comparability
PCB	polychlorinated biphenyl
PCE	perchloroethene (or tetrachlorethene)
PGDP	Paducah Gaseous Diffusion Plant
PHSO	Project Health and Safety Officer
PID	photoionization detector
PORTS	Portsmouth Gaseous Diffusion Plant
PPE	personal protective equipment
PRG	preliminary remediation goal
QA	quality assurance
QAPjP	Quality Assurance Project Plan
QC	quality control
RA	Remedial Action
RAOs	remedial action objectives
RCRA	Resource Conservation and Recovery Act
RD	Remedial Design
RFA	RCRA Facility Assessment
RfCs	reference concentrations
RfDs	reference doses
RFI	RCRA Facility Investigation
RI	Remedial Investigation
RME	reasonable maximum exposure
ROD	Record of Decision
RPD	relative percent difference
RT	retention time
SAIC	Science Applications International Corporation
SAP	sampling and analysis plan
SARA	Superfund Amendments and Reauthorization Act
SDWA	Safe Drinking Water Act
SEG	Scientific Ecology Group
SEN	Secretary of Energy Notice
SHSO	Site Health and Safety Officer
SI	Site Investigation
SL1	Sanitary Landfill I
SMCLs	secondary maximum contaminant levels
SOW	statement of work
SRM	standard reference material
ST	South Tributary
SWMU	solid waste management unit
SWSA	Solid Waste Storage Area

TAL	Target Analyte List
TBC	to be considered
TBL	trip blank
TCA	<i>Tennessee Code Annotated</i>
TCE	trichloroethene
TCL	Target Compound List
TCLP	Toxicity Characteristic Leaching Procedure
TDEC	Tennessee Department of Environment and Conservation
TDS	total dissolved solids
TICs	tentatively identified compounds
TLD	thermoluminescent dosimeter
TLV	Threshold Limit Value
TOC	total organic carbon
TRU	transuranic
TSCA	Toxic Substances Control Act
TSD	treatment, storage, and disposal
TUs	Temporary Units
TVA	Tennessee Valley Authority
TWRA	Tennessee Wildlife Resources Agency
UEFPC	Upper East Fork Poplar Creek
USC	United States Code
USGS	U.S. Geologic Survey
USLE	Universal Soil Loss Equation
VOA	volatile organic analysis
VOC	volatile organic compound
WAG	Waste Area Grouping
WDC	Waste Disposal Coordinator
WOM	White Oak Mountain
WT	West Tributary
WTSD	Waste Treatment, Storage, and Disposal

EXECUTIVE SUMMARY

The enactment of the Resource Conservation and Recovery Act (RCRA) in 1976 and the Hazardous and Solid Waste Amendments to RCRA in 1984 created management requirements for hazardous waste facilities. The facilities within the Oak Ridge Reservation (ORR) were in the process of meeting the RCRA requirements when ORR was placed on the Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA) National Priorities List (NPL) on November 21, 1989. To ensure that all necessary steps were undertaken to protect the public health and the environment during the investigation, evaluation, and remediation of all Solid Waste Management Units (SWMUs) and Operable Units (OUs) on the ORR, the Department Of Energy Oak Ridge Operations (DOE-ORO) Office in Oak Ridge, Tennessee, the Tennessee Department of Environment and Conservation (TDEC), and the Environmental Protection Agency have formulated a Federal Facility Agreement (FFA) for ORR. The FFA outlines all deliverables, review times, and schedules to ensure that activities are undertaken in a timely manner.

Under RCRA, actions typically follow the RCRA Facility Assessment/RCRA Facility Investigation/Corrective Measures Study implementation process. Under CERCLA, actions typically follow the Preliminary Assessment/Site Investigation/Remedial Investigation (RI)/Feasibility Study (FS)/remedial design/remedial action process. A goal of the FFA is to coordinate response actions under CERCLA with closure, postclosure care, and corrective measures under way or planned under RCRA and applicable state laws, in such a manner as to maximize flexibility and minimize duplication of investigative actions. The development of this document incorporates the requirements under both CERCLA and RCRA into an RI work plan. Further studies of the sites in Bear Creek Valley (BCV) OU 1 will follow the CERCLA process.

BCV OU 1 is one of three source OUs within BCV. It is comprised of five sites that are located on the northern side of BCV to the west of the Y-12 Plant. All sites in this OU are outside the Y-12 Plant boundary. The five sites are the following: the S-3 Site; the Oil Landfarm; the Boneyard/Burnyard, which includes the Hazardous Chemical Disposal Area (HCDA); Sanitary Landfill I (SL 1); and the Bear Creek Burial Grounds (BG-A North, -A South, -B, -C East, -C West, -D, -E, and -J), which includes the Oil Retention Ponds (ORPs) 1 and 2, the Walk-in Pits, and the Uranium Vaults.

The S-3 Site, Oil Landfarm, HCDA, and the Burial Grounds have had some previous investigation under the Y-12 Plant RCRA Program. No investigation has been carried out for the Boneyard/Burnyard or SL 1.

S-3 Site. The S-3 Ponds were constructed in 1951 and consisted of four unlined surface impoundments covering a total area of roughly 1.5 ha (3.7 acres) with a storage capacity of 38 million L (10 million gal). During operation, up to 21,000 L/d (5500 gal/d) of effluent was pumped into the ponds through a pipeline from the Y-12 Plant. The S-3 Site underwent RCRA closure in 1988 and was paved to create a parking lot. Groundwater below the S-3 Site is known to be contaminated with nitrate, metals, organics, and radioactive chemicals. Waste remaining below the RCRA cap is known to contain metals and radioactive chemicals.

Oil Landfarm. The Oil Landfarm was used for the land application of waste oils and coolants that contained beryllium compounds, depleted uranium, polychlorinated biphenyls (PCBs), and chlorinated hydrocarbons. Disposal operations were discontinued in 1982. Final RCRA closure of the site was certified by TDEC in December 1990. Soils at this site are contaminated with PCBs and uranium; groundwater is contaminated with organics.

Boneyard/Burnyard. The Boneyard/Burnyard consists of ~3.2 ha (8 acres) that were used from 1943 to 1970 as a disposal site for waste from the Y-12 Plant. Burning and disposal of debris and sanitary, metallic, chemical, and radioactive wastes are known to have occurred. The site has been abandoned and is predominantly covered with grassy vegetation. The southeastern portion of this site is overlain by the HCDA. The HCDA covers ~0.8 ha (2 acres) and was used to burn or neutralize liquid and gaseous wastes from 1975 until 1981. The HCDA is covered with a RCRA cap. Groundwater at the Boneyard/Burnyard is contaminated with radioactive chemicals. The nature of soil contamination is not known.

SL 1. SL 1 received various types of nonhazardous wastes from the Y-12 Plant. Waste disposal at SL 1 was carried out from 1962 to 1982, and the site was graded, capped with clay, and closed in 1983 under a TDEC-approved closure plan according to TDEC rules and regulations for sanitary landfills. No hazardous materials have been recorded at this site.

Burial Grounds. The Burial Grounds are located on the southern slope of Pine Ridge ~3 km (2 miles) west of the Y-12 Plant. They comprise a series of waste disposal units covering an area of ~140 ha (350 acres). Each waste disposal unit consists of a series of trenches excavated to between 4.3 and 7.6 m (14 and 25 ft) below grade. The trenches received a variety of hazardous and nonhazardous solid and liquid wastes from the Y-12 Plant. All hazardous waste disposal operations were discontinued in 1981. All trenches known to have received RCRA hazardous material have been capped, or are in the process of being capped, as part of RCRA closure. The Burial Grounds include ORPs 1 and 2, the Walk-in Pits, and the Uranium Vaults. The ORPs were constructed to intercept seepage from the Burial Grounds trenches. Both ponds were capped and RCRA-closed in 1990. The Walk-in Pits were used to dispose of solid or drummed liquid wastes including pyrophoric saw fines of uranium and thorium and shock-sensitive chemicals. Details of wastes delivered to the Burial Grounds are contained in an inventory of Y-12 Plant records. It is proposed that this comprehensive list be used to characterize wastes currently in place. The Uranium Vaults were constructed to store enriched uranium waste but were never used for that purpose. A wide range of contaminants may have been disposed of in the Burial Grounds, and groundwater shows the highest levels of organic contamination in BCV. The nature and extent of soil contamination outside capped areas resulting from groundwater transport is not known.

The development of this RI work plan is based on the foundation set by the Data Quality Objective (DQO) process. The purpose of the DQO process is to ensure that data of appropriate and known quality are obtained and are adequate for the intended uses. During the DQO process, data users were identified and were involved at the beginning of the work plan development to identify data types and uses needed for the RI/FS. Historical data for each of the sites were reviewed to develop a list of potential contaminants of concern and to develop conceptual models detailing potential exposure pathways. From this

preliminary work, the scope of the investigation was narrowed, specific RI goals were defined, and the project objectives were developed.

The intent and scope of the work plan are to assemble all data necessary to facilitate selection of remediation alternatives for the sites in BCV OU 1 such that the risk to human health and the environment is reduced to acceptable levels based on agreements with regulators. The goal is to develop a final Record Of Decision (ROD). For source OUs such as BCV OU 1, data acquisition will not be carried out in a single event, but will be carried out in three stages that accommodate the schedule for developing an ROD and provide data to the final site-wide ROD. The three stages for a source unit are as follows:

Stage 1 Assemble sufficient data to support decisions such as the need for removal actions, whether to continue with the RI process, or whether no further action is required.

If the decision is made to continue the RI/FS process, then:

Stage 2 Assemble sufficient data to allow for a ROD that reduces risks to the human health and the environment.

Stage 3 Provide input from the source OU that allows a valley-wide ROD to be issued for the integrator OU.

It is anticipated that the third stage of the process will rely on postremediation performance monitoring data collected at the source OU.

One goal of the RI work plan will be to ensure that sampling operations required for the initial stage are not repeated at later stages. The overall goals of this RI are to define the nature and extent of contamination so that the impact of leachate, surface water runoff, and sediment from the OU 1 sites on the integrator OU can be evaluated, the risk to human health and the environment can be defined, and the general physical characteristics of the subsurface can be determined such that remedial alternatives can be screened. Specific objectives to meet these goals follow:

- To define the contaminants of concern at each site.
- To quantify the contaminant concentrations in waste and soils at each site and the spatial distributions of contaminants of concern.
- To assess the contribution made by contamination in soils and wastes to the concentrations of dissolved and particulate phases in groundwater and surface water.
- To support the FS in the screening of appropriate alternatives during the remedy selection.
- To provide technically and legally defensible data on which to base credible decisions regarding risk levels and cleanup alternatives to allow a ROD to be issued for the source OU.
- To provide data input to the site-wide baseline risk assessment that will be carried out to issue a ROD for the integrator OU.

The BCV OU 1 RI work plan was prepared following the Remedial Investigation Work Plan Annotated Outline DOE/OR/01-1077, Rev 0. This work plan contains summaries of geographical, historical, operational, geological, hydrological, and ecological information specific to sites in BCV OU 1. The potential for release of contaminants to receptors through various media is addressed, and a sampling and analysis plan is presented to obtain the objectives of the RI. Proposed sampling activities include geophysical and radiological surveys, shallow soil borings, soil borings, piezometer and monitoring well installation, groundwater sampling, surface water sampling, and stream sediment sampling. A field sampling investigation plan, a health and safety plan, and a waste management plan are also included in the work plan.

1. INTRODUCTION

1.1 REGULATORY INITIATIVE

The Oak Ridge Reservation (ORR) is comprised of three major installations: the Oak Ridge National Laboratory (ORNL), the Oak Ridge Y-12 Plant, and the Oak Ridge K-25 Site [formerly referred to as the Oak Ridge Gaseous Diffusion Plant (ORGDP)]. These installations were constructed in the early to mid-1940s by the Atomic Energy Commission [predecessor of the U.S. Department of Energy (DOE)] as research, development, and process facilities in support of the Manhattan Project. These installations, along with the Paducah Gaseous Diffusion Plant (PGDP) in Paducah, Kentucky, and the Portsmouth Gaseous Diffusion Plant (PORTS) in Piketon, Ohio, are currently administered by the DOE Oak Ridge Operations (DOE-ORO) office in Oak Ridge, Tennessee.

During the construction and operation of these research, development, and process facilities, the associated decontamination, maintenance, and fabrication processes resulted in the generation of various hazardous and radioactive waste by-products. Hazardous waste treatment, storage, and disposal (TSD) facilities were created at each of the DOE-ORO installations to handle such by-products. Some of these facilities continue to receive hazardous wastes, while others have been decommissioned. All DOE-ORO TSD facilities are currently subject to the requirements of several laws:

- **Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA, also referred to as Superfund)**—created in 1980 to establish a program to identify sites [operable units (OUs)] from which environmental releases of hazardous substances might occur or have occurred. At such sites, Superfund promotes the evaluation of damage to natural resources, ensures cleanup by the responsible party or the government, and creates a claims procedure for parties involved in site cleanup and natural resource reclamation. Sites identified by CERCLA are evaluated and then placed on the National Priorities List (NPL), if appropriate. The ORR was listed on the NPL in the December 1989 *Federal Register* (54 FR 48184).
- **Resource Conservation and Recovery Act (RCRA)**—created in 1976 as a hazardous waste management system that mandates permitting currently operating TSD facilities. Under RCRA, these TSD facilities are referred to as solid waste management units (SWMUs). RCRA defines an SWMU as any discernible waste management unit at a RCRA facility from which hazardous waste or hazardous constituents might migrate, irrespective of whether the unit was intended for the management of solid or hazardous waste. Such units include any area at a facility at which hazardous waste or hazardous constituents have been routinely and systematically released.
- **Hazardous and Solid Waste Amendments (HSWA)**—created as amendments to RCRA (1984) that provide the U.S. Environmental Protection Agency (EPA) with the authority to enforce corrective actions by broadening the scope of the RCRA Corrective Action Program. In addition to evaluating and correcting releases to the uppermost aquifer from regulated RCRA units, HSWA promotes the cleanup of continuing releases to any media resulting from waste management units and practices at RCRA facilities. Among the most significant provisions of the HSWA are the following:

- **Section 3004(u): Corrective Action for Continuing Releases.** Section 3004(u) states that for permits issued after November 8, 1984, corrective action is required for releases of hazardous waste or constituents from any SWMU at any TSD facility seeking a permit for permanent operation, regardless of when waste was placed in the unit. Thus, corrective actions apply to releases presently occurring as well as past releases.
- **Section 3004(v): Corrective Action Beyond the Facility Boundary.** Section 3004(v) authorized EPA to require that corrective action be taken by the facility owner or operator for releases that have migrated beyond the facility boundary (off site). Such action should be taken where necessary to protect human health and the environment unless the owner/operator demonstrates to the satisfaction of the administrator that permission to undertake such action was denied.
- **Superfund Amendments and Reauthorization Act (SARA)**—created in 1986 as a 5-year extension of the Superfund/CERCLA program to clean up hazardous releases at uncontrolled or abandoned hazardous waste sites.
- **National Environmental Policy Act (NEPA)**—created in 1968 to direct federal agencies to consider the impacts of their actions (e.g., construction, remediation) on the environment as a part of all decision-making processes.

In anticipation of the ORR being listed on the NPL, DOE-ORO, the Tennessee Department of Environment and Conservation (TDEC), and EPA formulated a Federal Facility Agreement (FFA) for the ORR. The FFA was designed to ensure that all necessary steps were undertaken to protect the public health and the environment during the investigation, evaluation, and remediation of all SWMUs/OU's. A goal of the FFA is to coordinate response actions under CERCLA and applicable state laws, in such a manner as to maximize flexibility and minimize duplication of investigative actions (Energy Systems 1992a). The FFA also outlines all deliverables, review times, and schedules to ensure that activities are undertaken in a timely manner.

DOE and TDEC have formulated an oversight agreement/agreement-in-principle. As stated in this agreement, TDEC is the lead agency for the state of Tennessee. The purposes of this agreement are to ensure compliance with applicable federal, state, and local environmental laws and to ensure the citizens of Tennessee that their health, safety, and environment are being protected through a program of independent monitoring and oversight by the state.

Each of these agreements provides a measure of oversight for all environmental restoration at ORR installations. PGDP and PORTS are not listed on the NPL and therefore have no FFA. (They do, however, have agreements-in-principle with their respective states and regional EPA offices.)

1.2 DOE-ORO ENVIRONMENTAL RESTORATION PROGRAM

The mission of the DOE-ORO Environmental Restoration (ER) Program is "to eliminate or reduce to prescribed safe levels the risks to the environment or to human health and safety posed by inactive and surplus ORO-managed sites and facilities that have been contaminated by radioactive, hazardous, or mixed wastes" (Energy Systems 1991). Because

ORR was placed on the NPL in December 1989, all remedial activities including characterization, alternative selection, and implementation must meet the requirements of RCRA, CERCLA, and NEPA. Therefore, the DOE ER Program decided to address the requirements of all applicable laws in a series of "integrated" documentation for each site.

To facilitate consistency in the preparation of these integrated documents at all DOE-ORO facilities, the managing and operating contractor for DOE-ORO facilities established an ER Division. In addition to central staff who oversee the main program areas, ER site program managers were appointed at each of the DOE-ORO facilities. These site program managers oversee the day-to-day operations of the facility-specific programs and interface with central ER Division staff.

1.3 Y-12 PLANT ER PROGRAM

A major objective of CERCLA investigations is to characterize the nature and extent of contamination at uncontrolled hazardous waste sites. HSWA requires assessments to be conducted at all former and current SWMUs at RCRA-permitted facilities. For SWMUs that are suspected to be the source of contaminant releases to the environment, a RCRA Facility Investigation (RFI) is required to define the nature and extent of the release. Descriptions of most of the individual SWMUs at the Y-12 Plant are provided in a report entitled *Solid Waste Management Unit Information for the Y-12 Plant RCRA 3004(u) Facility Assessment* and in its supplements (Welch and Poore 1987; Welch 1987; Murphy 1989).

Under RCRA, actions typically follow the RCRA Facility Assessment (RFA)/RFI/Corrective Measures Study (CMS) implementation process. Under CERCLA, actions typically follow the Preliminary Assessment (PA)/Site Investigation (SI)/Remedial Investigation (RI)/Feasibility Study (FS)/remedial design/remedial action process. The development of this document incorporates the requirements under both CERCLA and RCRA in an RI work plan. Further studies of the sites in Bear Creek Valley (BCV) Operable Unit 1 (OU 1) will follow the CERCLA process.

1.4 FACILITY OU STRATEGY

The ORR consists of ~14,300 ha (35,300 acres) of federally owned land in Oak Ridge. Under its RCRA HSWA permit, DOE has begun the RFI process at more than 500 SWMUs at the ORR. Under the present strategy, OUs are defined as sites for which an RI/FS will be performed. The RI/FS will include any necessary removal actions or interim actions. Many of the remediation units on the ORR were grouped not only according to their proximity but also by common physical and hydrological parameters.

Because the hydrogeologic regime of the ORR is complex, and because numerous sources contribute to groundwater contamination within a geographical area, more timely investigations can be achieved by separating the sources of contamination from the groundwater and surface water for investigation and remediation. To permit effective evaluation of the cumulative impact of releases from multiple sources of contamination, an approach has been adopted for the ORR that separates studies of the groundwater and surface water from studies of the sources. This was done by assigning waste area groupings (WAGs) as source OUs and the hydrogeologic regimes that are common to more than one

source OU as integrator OUs. This approach will result in more immediate attention [removal action/interim action and/or Records of Decision (RODs)] for the source OUs while investigation continues into remediation alternatives for the integrator OUs (Energy Systems 1992a). Both source and integrator OUs will achieve a ROD (Fig. 1.1).

Data acquisition for each OU in an integrator unit will not be carried out in a single event but may be carried out in the following three stages (Fig. 1.2).

- Stage 1—Assemble sufficient data to support making intermediate decisions. Some questions that would need to be answered are the following: Can the site be designated as “Site Evaluation Accomplished”? Should a focused effort be implemented to support a finding of no-further-investigation-required? Is a removal action required or appropriate? Is a focused FS relevant to the site? Should the waste unit proceed to Stage 2 of the CERCLA process?
- Stage 2—Assemble sufficient data to select an acceptable source remediation and obtain a source unit ROD.
- Stage 3—Compile valley-wide information including postremediation source unit monitoring to obtain a ROD for the entire hydrologic regime.

The premise of the staged approach is that the remediation of a CERCLA OU is a complex process, starting with the identification of those sources that warrant investigation on the ORR and continuing through the proposed plan for implementing a remedial action. To implement the entire process on all potential sources at the Y-12 Plant would be impossible. *Therefore, the staged approach is designed to provide mechanisms for withdrawing a site from the process as soon as information is available that suggests it is not necessary for the site to proceed through the entire RI/FS process.*

Each stage supports a different level of decision making and, therefore, would require different types and amounts of data. However, a goal of the RI work plan will be to ensure that sampling operations required for the initial stage are not repeated at later stages. Some of the individual sites within BCV OU 1 have been studied extensively, and Stage 1 may require review, evaluation, and validation of existing data. For other BCV OU 1 sources, additional Stage 1 data may be required. Stage 2 is the complete RI and the data requirements are those sufficient to develop an ROD. It is anticipated that the third stage of the process will rely on postremediation performance monitoring data collected at the source.

One of the primary goals of the Y-12 Plant strategy is to integrate source OUs and integrator OUs. The primary responsibilities of the source OU are as follows:

- to identify the nature and extent of the source;
- to identify and quantify (with models provided by the integrator unit) potential releases and transport mechanisms from the source to the integrator OU, including leaching, transport through the unsaturated zone to the groundwater, movement through the stormflow system to the groundwater, surface water/sediment, and runoff to surface water/sediment;
- to identify and quantify (if possible and necessary) potential releases to the air;

Integrator OU

Source OU 1

Source OU 2

* No Further Investigation

Fig. 1.1. The relationships between the integrator unit and source units in the same hydrological system.

Stages of the Source OU RI/FS

* Baseline Risk assessment

Fig. 1.2. The staged RI process for a source OU.

- to collect and provide to the integrator OU information (e.g., runoff and stormflow data) showing the connection and possible integration points between a specific source and the integrator OU;
- to identify and quantify current and future potential on-site risks to humans associated with the source;
- to identify terrestrial ecological effects associated with the source. To gain an understanding of potential risks to species with a small habitat range that are generally low in the food chain and can be connected easily to a specific source (e.g., earthworms);
- to identify potential wetlands within the source OU boundaries and effects on wetlands species;
- to use risk/ARAR-based groundwater PRGs, developed by the integrator OU, to identify whether current or potential future releases could pose a risk to future on-site receptors and current and future off-site receptors. These PRGs will be compared with detected or modeled concentrations to help determine if releases from the source need to be remediated;
- to use human and ecological-based surface water PRGs, developed by the integrator OU, to discern if current or potential future releases from the source need to be addressed by remediation.

The primary responsibilities of the integrator unit are as follows:

- to identify contamination currently in the environmental media that should be addressed immediately;
- to gain an understanding of the primary flow mechanisms within and between the Nolichucky and Maynardville groundwater formations;
- to gain an understanding of the interaction between the Nolichucky and Bear Creek and Upper East Fork Poplar Creek (UEFPC);
- to gain an understanding of the interaction between groundwater/surface water and the Clinch River;
- to identify points in the environmental media (surface water and groundwater) where high levels of contamination are entering the media (integrator points) to help prioritize the cleanup of the source units;
- to gain an understanding of and provide the tools for quantifying (if possible) contaminant transport through the groundwater system (e.g., identify sorption properties and site-specific water-to-soil distribution coefficients for primary chemicals of concern);
- to develop and provide to the source OUs conceptual and (if possible) quantitative models that should be used to describe unsaturated zone flow, stormflow zone flow, and surface water runoff to ensure that source OUs are using consistent techniques for estimating chemical migration to the integrator;
- to provide source OUs groundwater preliminary remediation goals (PRGs) that can be used to make remediation decisions about the source. These PRGs should be based on risk and applicable or relevant and appropriate requirements (ARARs) and may account for potential risks to both a future on-site receptor and a downgradient off-site receptor. PRGs for the off-site receptor should eventually account for dilution/attenuation that

occurs as hazardous substances move through the groundwater. Estimated and measured concentrations of hazardous substances being released from the source unit to the groundwater will be compared to these PRGs to determine if remediation of the source is necessary;

- to provide source OUs surface water PRGs that can be used to make remediation decisions about the source. These PRGs should be risk-based and should account for potential risks to ecological and human receptors. Estimated or measured concentrations of hazardous substances being released from the source to the integrator will be compared to these PRGs to help determine if remediation of the source is necessary;
- to gain an understanding of the off-site risks associated with hazardous substances currently detected in groundwater, surface water, and sediments;
- to gain an understanding of potential future risk associated with continued off-site migration of hazardous substances currently detected in groundwater and sediments (e.g., to learn how this contamination may act as a source for future risk even if source control measures are implemented on source OUs);
- to gain an understanding of potential risks to aquatic species and food chain impacts associated with contamination in surface water and sediments;
- to gain an understanding of the potential risks to terrestrial species that have a large habitat range and that are higher in the food chain (e.g., deer and predator species).

Five integrator OUs have been identified for the ORR: Bethel Valley, Melton Valley, the K-25 Site area, BCV OU 4, and UEFPC OU 1. ORNL WAGs have not yet been assigned to OUs. There have been 14 source OUs identified at the K-25 Site and 8 source OUs identified at the Y-12 Plant. In addition, three combined source and integrator OUs have been identified: Chestnut Ridge at the Y-12 Plant, Lower East Fork Poplar Creek (LEFPC), and ORNL WAGs 11 and 13 with the Oak Ridge Associated Universities (ORAU) site at Freels Bend.

The Y-12 Plant source OUs have been grouped not only by their proximity but also by common physical and hydrogeologic parameters (Table 1.1). Extensive study of the relationship between groundwater movement and contaminant transport at the Y-12 Plant indicates that groundwater can be subdivided into two distinct hydrogeologic regimes: UEFPC and BCV. This subdivision is based on topography, surface water drainage, and groundwater flow patterns. It provides a basis for unifying monitoring efforts at the Y-12 Plant and for tailoring monitoring efforts to the hydrogeologic characterization of each regime. The strategy used to delineate the boundaries of each regime is included in *Comprehensive Groundwater Monitoring Plan for the DOE Y-12 Plant, Oak Ridge, Tennessee* (Geraghty and Miller 1990).

BCV OU 1 is one of two source OUs within BCV (Fig. 1.3; refer also to Fig. 2.3 in Chap. 2). The integrator OU for BCV is BCV OU 4, which consists of surface water and groundwater from Bear Creek and its tributaries, as well as floodplain soils and sediments. The boundaries of BCV OU 4 are from the crest of Chestnut Ridge to the crest of Pine Ridge and from the surface water divide near the S-3 Site to the groundwater divide west of State Highway 95 between Bear Creek and Grassy Creek. Source OUs in BCV are as follows:

Table 1.1. Y-12 Plant OUs

OU	Type	Sites of concern
UEFPC OU 1	Integrator	Groundwater, surface water, and sediment within the Y-12 Plant
UEFPC OU 2	Source	Abandoned Nitric Acid Pipeline
UEFPC OU 3	Source	S-2 Site, Building 81-10, Coal Pile Trench, Machine Coolant Storage Tanks, Waste Coolant Processing Facility, Tanksite 2063, Salvage Yard Drum Deheader, Salvage Yard Waste Oil Storage Tanks, Salvage Yard Oil/Solvent Drum Storage Area, and Salvage Yard Scrap Metal Storage Area
BCV OU 1	Source	S-3 Ponds; SL 1; HCDA/Boneyard/Burnyard; Oil Landfarm; Burial Grounds [includes BG-A (North and South), -B, -C, -D, -E, and -J; ORPs 1 and 2; the Walk-in Pits; and the Uranium Vaults]
BCV OU 2	Source	Rust Spoil Area, Spoil Area 1, SY-200 Yard
BCV OU 4	Integrator	Bear Creek Groundwater and Surface Flow Regime and Floodplain Sediments
Chestnut Ridge OU 1	Source	Chestnut Ridge Security Pits Hazardous Waste Disposal Units
Chestnut Ridge OU 2	Source	Filled Coal Ash Pond, Upper McCoy Branch
Chestnut Ridge OU 3	Source	United Nuclear Corporation Disposal Site
Chestnut Ridge OU 4	Source	Rogers Quarry and Lower McCoy Branch
LEFPC OU	Integrator/Source	LEFPC Floodplain and Creek

- BCV OU 1, which is comprised of the S-3 Site; the Oil Landfarm; the Boneyard/Burnyard, including the Hazardous Chemicals Disposal Area (HCDA); Sanitary Landfill I (SL 1); and the Bear Creek Burial Grounds [(BG)-A North, BG-A South, BG-B, BG-C East, BG-C West, BG-D, BG-E, and BG-J)], including Oil Retention Ponds (ORPs) 1 and 2, the Walk-in Pits, and the Uranium Vaults.
- BCV OU 2, which includes the Rust Spoil Area, Spoil Area 1, and the SY-200 Yard.

1.5 INTENT AND SCOPE OF THE RI WORK PLAN

BCV OU 1 is the focus of this work plan. The intent of the work plan is to assemble all data necessary to select remediation alternatives for the sites in BCV OU 1 in order to reduce the risk to human health and the environment to acceptable levels based on agreements with the regulators. Because BCV OU 1 is a source OU that is associated with an integrator OU, there are certain constraints on the media that are to be addressed for remediation by the RI/FS and on the scheduling of the RI/FS process. Groundwater and surface water associated with BCV OU 1 will not be addressed as media to be remediated in the BCV OU 1 RI/FS process; risks to off-source and off-site receptors via groundwater and surface water pathways will be addressed in the BCV OU 4 (integrator OU) Baseline Risk Assessment (BRA).

Fig. 1.3. Relationships between sites, OUs, and source and integrator OUs in BCV.

Floodplain soils will be addressed by the RI for BCV OU 4 and will not be included in the work plan for BCV OU 1 unless they are within the BCV OU 1 site boundaries.

The development of data needs in the RI work plan entailed development of the site conceptual models and lists of likely removal actions/interim actions and likely remedial alternatives that will be evaluated by the RI. Likely removal actions/interim actions are as follows:

- Removal or over packing of drums or other containers.
- Excavation of highly contaminated soils, especially those that are radioactive.
- Drainage controls (i.e., capture and treatment) of contaminated shallow groundwater).
- Drainage controls (i.e., capture and treatment) of contaminated seeps.
- Fences, warning signs, or other security measures.

Likely remedial alternatives are as follows:

- No action.
- Source containment.
- Hydraulic isolation.
- Removal/treatment.

Based on the site conceptual model, the activities described in this work plan are designed to support these remedial alternatives and likely removal actions.

The schedule of RI/FS activities in BCV OU 1 will follow the staged approach outlined in the facility OU strategy for the Y-12 Plant outlined in Sect. 1.4. The goal is to develop a ROD for all of the OUs in BCV, including the integrator OU.

1.6 SPECIAL PROBLEMS

Special problems for BCV OU 1 are as follows:

- Wastes in BG-B and the Walk-in Pits include pyrophoric saw fines. Wastes in the Walk-in Pits include shock-sensitive chemicals, in addition to pyrophoric saw fines. Thus, intrusive investigative methods at these sites are specifically excluded for health and safety reasons. A project is currently under way to cap these areas with a RCRA-style cap.
- Parts of BG-A North, BG-A South, BG-C West, the Oil Landfarm, the HCDA, and the S-3 Ponds were closed in place as landfills under RCRA between 1988 and 1990. Closure was conducted pursuant to Tennessee Rule 1200-1-11-.05(14)(e). With the exception of the HCDA, each of these sites is covered with an engineered cap, and the types of wastes deposited at these sites are documented; therefore, no samples will be taken that call for penetrating a cap. Waste in trenches will be characterized using a statistical sample of Y-12 Plant records that describes masses of wastes delivered to the Burial Grounds.

1.7 BCV OU 1 PROJECT OBJECTIVES

The overall RI/FS goal for BCV OU 1 is to select cost-effective remediation alternatives that are protective of human health and the environment, that maintain protection over time, and that minimize the volume of untreated waste generated. The objective of the RI is to provide sufficient information to assess the risks to human health and the environment and to support the development, evaluation, and selection of appropriate response alternatives. To resolve these problems and to meet the project objectives, the following questions need to be posed concerning the environment and contamination at the OU:

- What are the volume and nature of waste materials at the various sites?
- What are the current nature and extent of soil contamination at each site?
- What are the expected nature and extent of soil contamination at each site in the future?
- What are the chemicals of potential concern (COPC) within the site boundaries?
- Which sites contain contaminants in forms and concentrations that adversely affect human receptors and exceed health-based criteria or ARARs [contaminants of concern (COCs)]?
- Which sites act as contaminant sources that contribute unacceptable levels of contaminants to the integrator OU (via groundwater or surface water)?
- What are the present and potential future pathways of contaminant migration away from each waste site?

A staged approach to the RI work plan was outlined in Sect. 1.4. Each stage of the RI has different objectives; however, an objective of the work plan is to ensure that data acquired during each stage of the RI is comprehensive and that no repetition of sampling of any media at any location is required as a result of changing objectives. The objective of each sample detailed in the sampling plan is to provide sufficient information to support the final remediation of the site.

The following objectives will be met in order to assess the risks contributed by this OU to the overall risk to human health and the environment in BCV and to allow a ROD to be issued:

- to determine if the wastes pose an on-site risk to human health and the environment at the present time.
- to determine if the wastes will pose an on-site risk to human health and the environment in the future.
- to determine if the wastes pose an off-site risk to human health and the environment at the present time.
- to determine if the wastes will pose an off-site risk to human health and the environment in the future.
- to determine if the wastes and/or waste derivatives are in compliance with ARARs.

- to support the screening of appropriate alternatives during the remedy selection and provide technically and legally defensible data on which to base credible decisions regarding risk levels and cleanup alternatives.
- to carry out a BRA in order to obtain an ROD for the source unit.

1.8 DATA QUALITY OBJECTIVES

Data quality objectives (DQOs) have been developed to meet the project objectives outlined in Sect. 1.7. The DQO development process for BCV OU 1 is outlined in Fig. 1.4; the relationship between the DQOs and the overall work plan is shown in Fig. 1.5. Table 1.2 details the regulatory drivers that are applicable to BCV OU 1 under CERCLA and the subsequent project objectives, DQOs, data uses, and data needs that are derived from interpretation of the regulatory drivers.

The DQO process for BCV OU 1 were worked through with TDEC at a workshop held on February 16 through 23, 1994. The following discussion of DQOs is derived from the discussions and decisions made at this workshop.

1.8.1 Statement of the Problem

1.8.1.1 Site background

Six sites at three discrete areas in BCV have been included in BCV OU 1 to be evaluated as contaminant source OUs. The site name, unit number, and unit status for each site in BCV OU 1 are presented in Table 1.3. Chapters 2 and 3 present detailed descriptions of each site. A brief description of each site follows:

S-3 Ponds. The S-3 Ponds were constructed in 1951 and consisted of four unlined surface impoundments covering a total area of ~1.5 ha (3.7 acres) with a storage capacity of 38 million L (10 million gal). During operation, up to 21,000 L/d (5500 gal/d) of effluent was pumped to the pond via a pipeline from the Y-12 Plant. The S-3 Site underwent RCRA closure between 1988 and 1990, with TDEC accepting final closure on November 15, 1990. The site was paved to create a parking lot. Primary contaminants are nitrate, uranium, metals, and organic solvents.

Oil Landfarm. The Oil Landfarm was used from 1973 to 1982 for the land application of waste oils and coolants that contained beryllium compounds, depleted uranium, polychlorinated biphenyls (PCBs), and chlorinated hydrocarbons. Final RCRA closure of the site was accepted by TDEC on November 15, 1990.

Boneyard/Burnyard. The Boneyard/Burnyard consists of ~3.2 ha (8 acres) used from 1943 to 1970 as a disposal site for waste from the Y-12 Plant. Burning and disposal of debris and sanitary, metallic, chemical, and radioactive wastes are known to have occurred. The site has been abandoned and is predominantly a gravel area with grassy vegetation. The southeastern portion of this site is overlain by the HCDA. The HCDA covers ~0.8 ha (2 acres) and was used to burn or neutralize liquid and gaseous wastes from 1975 until 1981. The HCDA is covered with a RCRA-type cap, and the final closure was accepted by TDEC in November 1990.

Fig. 1.4. The DQO development process.

Fig. 1.5. The relationship between DQOs and the overall work plan for BCV OU 1.

Table 1.2. Regulatory drivers, project objectives, DQOs, data uses, and data needs for BCV OU 1

Regulatory driver (NCP—40 CFR Part 300 Subpart E)	Project objectives	Questions posed by the DQO process	Determine data needs to satisfy DQO requirements	
			Data uses	Data needs
<p>§ Sect. 300.430 Remedial investigation/feasibility study and selection of remedy</p> <p>(a)(i) Program goal. The national goal of the remedy selection process is to select remedies that are protective of human health and the environment, that maintain protection over time, and that minimize untreated waste</p> <p>(d) (2) ... field investigations to assess the following factors:</p> <p>(i) Physical characteristics of the site, including important surface features, soils, geology, hydrogeology, meteorology, and ecology</p> <p>(ii) Characteristics or classifications of air, surface water, and groundwater</p> <p>(iii) The general characteristics of the waste, including quantities, state, concentration, toxicity, propensity to bioaccumulate, persistence, and mobility</p> <p>(iv) The extent to which the source can be adequately identified and characterized</p> <p>(v) Actual and potential exposure pathways through environmental media</p> <p>(vi) Actual and potential exposure routes (for example, inhalation and ingestion)</p>	Determine if the waste unit poses an unacceptable risk to human health and the environment	Do wastes contribute to current on-site worker risk?	Evaluate current worker exposure pathways/risk: inhalation, radiation, dermal contact, incidental ingestion	Current air concentrations, contaminant concentrations in soil and waste, exposure rates
		Does groundwater or surface water at the source exceed ARARs?	Evaluate source contribution to current off-site migration and off-site exposure pathways: exposure to groundwater or surface water	Groundwater and surface water concentrations in current leachate from waste unit; water ARARs at point that source feeds integrator OU
		Do wastes contribute to on-site or off-site ecological risks?	Evaluate ecological effects both on site and off site; evaluate propensity to bioaccumulate and persistence in the environment	Concentrations in soil; concentrations in vegetation; terrestrial species—population survey/body burden; water ARARs at point where source feeds integrator OU
		Will wastes contribute to future on-site or off-site risk?	Evaluate future on-site risk assuming multiple potential land uses—commercial and residential—and future ecological risk	Contaminant waste concentrations; future mass flux (contaminant volumes, contaminant leachate concentrations)
		Is it likely that groundwater or surface water at the source will exceed ARARs in the future?	Evaluate potential future groundwater/surface water flux and risk off site	Contaminant waste concentrations, contaminant concentrations in leachate, ARARs at point that source feeds integrator OU; nature of containment
(e)(9)(iii)(B) <i>Compliance with ARARs</i> . The alternatives shall be assessed to determine whether they attain applicable or relevant and appropriate requirements under federal environmental laws and state environmental or facility siting laws	Determine if the remediation alternatives are in compliance with ARARs	Are contaminant concentrations greater than regulatory levels?	Compare contaminant concentrations to ARARs	Contaminant concentrations—in waste unit, surrounding environment, and leachate

SL 1. SL 1 received various types of nonhazardous waste from the Y-12 Plant from 1968 to 1982. The site was graded, capped with clay, and closed in 1983 under a TDEC-approved closure plan according to TDEC rules and regulations for sanitary landfills.

Burial Grounds. The Burial Grounds are located on the southern slope of Pine Ridge ~3 km (2 miles) west of the Y-12 Plant with various portions operational from 1955 to 1993. They consist of a series of waste disposal units covering an area of ~140 ha (350 acres). Each waste disposal unit consists of a series of trenches excavated to between 4.3 and 7.6 m (14 and 25 ft) below grade. The trenches received a variety of hazardous and nonhazardous solid and liquid wastes from the Y-12 Plant. All hazardous waste disposal operations were discontinued in 1981, and all radiological waste disposal operations ceased by mid-1991. The last disposal to the Burial Grounds was sewage sludge from Oak Ridge disposed in Pit A-18 in 1993. All trenches known to have received RCRA hazardous material have been capped, or are in the process of being capped, as part of RCRA closure. The Burial Grounds include ORPs 1 and 2, the Walk-in Pits, and the Uranium Vaults, as well as BG-A (North and South), -B, -C, -D, -E, and -J. The ORPs were constructed to intercept seepage from the Burial Grounds trenches (-A North). Both ponds were capped and closed under RCRA in 1990. The Walk-in Pits were used to dispose of solid or drummed liquid wastes, including pyrophoric uranium and thorium saw fines and shock-sensitive chemicals and are currently being capped. The Uranium Vaults have not been mentioned in previous environmental

Table 1.3. Remedial units of BCV OU 1

Remedial unit	Unit no.	Status
S-3 Ponds	T-004	Closed under RCRA in 1988; TDEC accepted 11/90
ORPs 1 and 2	T-008, T-009	Closed under RCRA in 1990; TDEC accepted 12/90
Oil Landfarm	T-014	Closed under RCRA in 1990; TDEC accepted 11/90
Burial Grounds	D-024	BG-A (North and South) closed under RCRA; TDEC accepted 12/90. WIPs currently undergoing closure. Other sites not RCRA
SL 1	D-101	Closed in 1983
HCDA	D-024-HC	Closed under RCRA; TDEC accepted 11/90
Boneyard/Burnyard		Not a RCRA site

reports of the Burial Grounds, as they were only recently identified, and there are no historical records concerning use of the vaults after they were constructed. A wide range of contaminants may have been disposed of in the Burial Grounds, and a sample of Y-12 Plant

records has been used to characterize the volumes and masses of the various waste components.

1.8.1.2 Problem statement

From a scrutiny of earlier investigations, the BCV OU 1 sites have been identified as having contamination within their boundaries that might adversely affect human health, the environment or contribute unacceptable contaminant levels to media in the integrator OU. The problem statement for BCV OU 1 is as follows:

During the operation of the Y-12 Plant since the 1940's, hazardous and radioactive materials have been disposed of at the various sites in Bear Creek Valley that make up Operable Unit 1 (The S-3 Site, Boneyard/Burnyard, Oil Landfarm, Sanitary Landfill 1, and Burial Grounds). Considerable volumes of solid hazardous and radioactive materials are known to be present in trenches at the Burial Grounds and, probably, at the Boneyard/Burnyard. Hazardous liquids are known to have been disposed of at each site except SL 1, and have migrated away from the sites into surrounding groundwater and surface water. Soils within, and adjacent to, each of these sites have probably been contaminated during disposal operations.

Some sites have undergone closure under RCRA and have multilayer caps over the disposal areas. Exposure to the disposed materials and/or contaminated soils may be a threat to human health and the environment. Leaching of contaminants from disposed materials and contaminated soils may contribute to levels of contamination above decision criteria from the integrator OU (OU 4) in groundwater and surface water below, and adjacent to, the sites. Erosion of contaminated soils may transport contaminants to surface water bodies in the form of sediments. The primary problems associated with sites in OU 1 are the potential for unacceptable risk due to the following:

- *exposure to disposed hazardous and/or radioactive materials for human health and the environment,*
- *exposure to contaminated soils for human health and the environment,*
- *leaching of disposed hazardous and/or radioactive materials to impact groundwater and surface water,*
- *leaching of contaminated soils to impact groundwater and surface water, and*
- *erosion of contaminated soils resulting in contaminated stream sediments.*

1.8.2 Decisions To Be Made

The ultimate decisions for each site, then, are whether or not the site must be remediated and what might be the most effective remedial action. The decision to remediate will be based on the results of the risk assessment, and the amount and type of remediation will be based on the nine criteria of CERCLA and compliance with NEPA. The CERCLA criteria, which are grouped into three categories are Threshold Criteria, Primary Balancing Criteria, and Modifying Criteria. The nine criteria within these categories are as follows:

- ***Threshold Criteria***
 - Overall protection of human health and the environment
 - Compliance with ARARs
- ***Primary Balancing Criteria***
 - Long-term effectiveness and performance
 - Reduction of toxicity, mobility, or volume through treatment
 - Short-term effectiveness
 - Implementability
 - Cost
- ***Modifying Criteria***
 - State acceptance
 - Community acceptance

1.8.2.1 Preliminary and interim decisions

For the sites in Bear Creek Valley, translating the overall RI/FS process into stages allows decisions to be made in a timely order, dependant upon the amount of available information. Decisions that are made for each site as early a possible during data collection in the RI are:

1. If evidence for an immediate hazard/risk at a site exists, then investigate whether to instigate rapid removal actions/interim actions and proceed with the RI/FS concurrently to this action taking place. If immediate actions are required then determine the appropriate actions to take. If no evidence for an immediate hazard exists, then proceed with the RI/FS with no removal action until data is sufficient to provide the acceptable decision error.
2. If the BRA indicates unacceptable risk and characterization data is sufficient to provide the desired decision error then continue with the FS and determine what remedial actions needs to be taken at each site.
3. If the BRA indicates acceptable risk and data sufficient to provide the desired decision error then no further investigation will be needed and develop the proposed plan.

1.8.2.2 Questions addressed to make the remedial action decision

To complete the RI/FS process, and ensure that sufficient data is assembled to be able to make the decisions, the following questions need to be posed individually for each site:

- Question 1. Does exposure to the disposed hazardous materials and/or contaminated soils pose unacceptable risks to human health and the environment at any site under current or future land-use scenarios?
- Question 2. Does leachate from contaminated soils and/or disposed materials below the caps contaminate groundwater at any site such that contaminant concentrations in groundwater exceed decision criteria from the integrator OU (OU 4)?

- Question 3. Does leachate from contaminated soils and/or disposed materials outside capped areas contaminate groundwater at any site such that contaminant concentrations in groundwater exceed decision criteria from the integrator OU (OU 4)?
- Question 4. Does surface runoff or groundwater discharge from any site cause surface water to exceed decision criteria from the integrator OU (OU 4) in surface water bodies at, or adjacent to, that site?
- Question 5. Does soil erosion from any site pose unacceptable risks to human health and the environment at any site under current or future land-use scenarios?
- Question 6. What parameters are required to assess the likely remedial technologies?

1.8.3 Inputs to the Decision

Each question posed above needs to be addressed in order to make the remedial action decision has a series of data inputs. For BCV OU 1 the general data inputs to these questions are listed in Table 1.4, and Table 1.5 lists data in Table 1.4 that are input each question.

Table 1.4 General data inputs for BCV OU 1 remedial action decision

Input #	Input	Work Plan Location	Summary
1	Composition of the hazardous materials remaining in the waste site	Sect. 2.3 and Appendix A	RCRA closure documentation provides inventory of waste materials in the Burial Grounds and analyses of sludges at the S-3 site.
2	Nature and extent of soil contamination adjacent to the waste trenches and former ponds	Sect. 3.4.2 and Appendix C	Missing data to be obtained from RI Activities. Including statistical parameter: variance, maximum, minimum, mean, and upper 95% confidence concentrations for COPCs, frequency of detects, uncertainty, sample distribution, detection limits, and QC samples
3	Background Soil Constituents Data	Table 3.4	ORR Background Soil Characterization Project (DOE/OR/01-1175)
4	COPCs	Table 5.1	Based on the risk assessment data evaluation process
5	List of analytes sampled for and detection limits	Table 8.2	
6	Human Health Conceptual Site Model (and risk/HI numbers)	Sect. 3.9.4 and Figs 3.34 & 3.35.	Exposure Assessment (media and pathways); toxicity Assessment (slope factors and reference doses); present and future land-use

7	Soil PRGs and risk based action levels	Section 5.2 and Table 5.7	For residential and industrial land-use
8	Sediment PRGs and risk based action levels	For Soil PRGs see Section 5.2 and Table 5.7	For residential and industrial land-use. Sediment PRGs from the ecological risk assessment also required
9	Ecological site conceptual model	Sect. 3.9.4.3 and Figs 3.36 and 3.37	Includes all chemical specific exposure parameters plant-uptake parameters, phytotoxicity factors, and bioaccumulation models
10	Assessment endpoints for the site ecological model	Table 6.4	
11	Ecological population surveys	Sect. 3.7	Missing data to be obtained from RI activities
12	Characterization of the soil lithology	Sect. 3.4.1.2	Missing data to be obtained from RI Activities
13	A chemical conceptual model for the hazardous materials and/or soil	TBD	For example PHREEQE, WATEQ, EQ3/EQ6
14	Contaminant Fate and Transport Conceptual Site Model	Preliminary Model in Sect. 3.9	Model will be improved with data from the RI activities
15	Characterization of the groundwater and surface water flow systems	Current models in Sects. 3.5 (surface water) and 3.6 (groundwater)	Missing data include: the depth to groundwater in relation to the depth of the waste trenches and the effect of RCRA caps on groundwater recharge
16	Identify the fate and transport/leaching model	TBD	Such as SESOIL, HELP, RISKPRO, MULTIMED. Also refer to a sludge leaching model for the S-3 Site by Lee and Hyder, 1987 (Y/TS-263)
17	Identify the chemical model	TBD	Such as PHREEQE, EQ3/EQ6. Also refer to the sludge leaching model by Lee and Hyder, 1987 (Y/TS-263)
18	Physical, chemical, geochemical, and hydrological soil parameters to required to complete the fate and transport model	Mostly missing	Missing data to be obtained from RE Activities. Parameters such as chemical specific Kds, soil CEC, permeability, soil moisture curves, evapotranspiration, and recharge
19	Chemical parameters required for the hazardous materials chemical model	Mostly missing	Missing data to be obtained from RI Activities. Parameters such as mineral stability constants and oxidation potentials

20	Comprehensive site description	Sects. 2 and 3. Missing data to be obtained from RI Activities	Include habitat surveys, land-use surveys, slopes and elevations
21	Physical composition of media	Sect. 3. Some data is missing	Missing data to be obtained from RI Activities
22	Extent and reliability of the RCRA caps	Figs 2.4, 2.5, 2.7, and 2.8. Reliability data missing	Missing data to be obtained from RI Activities
23	Groundwater ARARs	Section 5, Table 5.2	ARARs will be updated according to regulatory changes
24	Groundwater decision criteria from OU 4	Missing	Available after BCV OU 4 RI BRA is completed
25	Surface Water ARARs	Section 5, Table 5.2	ARARs will be updated according to regulatory changes
26	Surface water decision criteria from OU 4	Missing	Available after BCV OU 4 RI BRA is completed
27	Soil erodibility factors	Tables 3.6 and 3.9	These constants are K factors for use in the Universal Soil Loss Equation (USDA, 1978). Refer to Section 3.4.3
28	Identify the soil erosion and sediment transport models	TBD	For example the universal soil loss equation

Table 1.5 Index of inputs to questions addressed to make the remedial action decision

Question to be addressed	Input numbers from Table 2.4
Question 1. Does exposure to the disposed hazardous materials and/or contaminated soils pose unacceptable risks to human health and the environment at any site under current or future land-use scenarios?	1, 2, 3, 4, 5, 6, 7, 9, 10, 11, 12, 13, 20, 22
Question 2. Does leachate from contaminated soils and/or disposed materials below the caps contaminate groundwater at any site such that contaminant concentrations in groundwater exceed decision criteria from the integrator OU (OU 4)?	1, 3, 4, 7, 12, 13, 14, 15, 16, 17, 18, 19, 20, 22, 23, 24
Question 3. Does leachate from contaminated soils and/or disposed materials outside capped areas contaminate groundwater at any site such that contaminant concentrations in groundwater exceed decision criteria from the integrator OU (OU 4)?	2, 3, 4, 7, 12, 13, 14, 15, 16, 17, 18, 19, 20, 22, 23, 24
Question 4. Does surface runoff or groundwater discharge from any site cause surface water to exceed decision criteria from the integrator OU (OU 4) in surface water bodies at, or adjacent to, that site?	2, 3, 4, 7, 12, 14, 15, 16, 18, 20, 22, 25, 26

Question 5. Does soil erosion from any site pose unacceptable risks to human health and the environment at any site under current or future land-use scenarios?	2, 3, 4, 7, 8, 12, 15, 20, 22, 27, 28
Question 6. What parameters are required to assess the likely remedial technologies?	1, 2, 3, 4, 6, 7, 9, 12, 20, 21, 23, 24, 25, 26

1.8.3.1 Data users

An RI conducted under CERCLA is designed to characterize the site environment and the nature and extent of contamination by using historical data to the fullest extent and by acquiring new data through field sampling operations. The data obtained by the RI should allow the subsequent participants of the process to accomplish their tasks. For instance, data on the contaminant concentrations permit the risk assessors to calculate the chronic toxicity and the risk of excess lifetime cancers under certain exposure assumptions. The same data allow the planners and engineers in the FS to screen remedial alternatives and make recommendations regarding cleanup options.

The following is a description of the data users for BCV OU 1 and their responsibilities:

- **RI Contractor**—solicits input and concurrence from the other participants for data needs and designs an investigation to obtain those data; evaluates the data and determines the nature and extent of contamination at each site as well as the potential risk to human and ecological receptors.
- **FS Contractor**—uses the results of the RI to assess the magnitude of the cleanup task and develops and evaluates the remedial alternatives in a systematic manner to make recommendations for the RODs.
- **Integrating Contractor (Energy Systems ER Division)**—organizes the overall approach to the RI/FS and evaluates the results and conclusions to ensure uniformity with the FFA and regulators; coordinates the contractors for the RI, FS, Engineering Design, and Construction Management tasks.
- **Lead Agency (DOE-ORO ER Division)**—has primary responsibility for the success of the Environmental Restoration Program; initiates and oversees the entire process while closely coordinating input and approval from the other FFA participants.
- **FFA Participants (DOE, EPA, and TDEC)**—have concurrent approval authority on any recommendations or initiatives and have regulatory responsibilities to ensure public health and safety and to respond to public needs. DOE has lead agency oversight on the approach, schedule, and budget for the environmental programs.
- **General Public and Public Interest Groups**—participate in making policy decisions by providing input on basic considerations such as the definition of acceptable risk and what constitutes a worthwhile benefit.

1.8.4 Boundaries of the Study

The boundaries for the sites in BCV OU 1 are defined by the geographic limits for each site and by the site conceptual mode. Geographic boundaries are not seen as fixed, but will be updated during the RI as more data becomes available to evaluate a site. The conceptual boundaries are those between the source OU and the integrator OU.

For the S-3 site, the geographic boundaries of the disposed materials are the dimensions of the former S-3 Ponds. However, the conceptual site model for the S-3 site includes possible down-gradient migration of contaminants in groundwater and precipitation of contaminants causing secondary contamination in soils adjacent to the former pond boundaries. The extent of secondary soil contamination is not known at this time; therefore, reasonable boundaries on the study have been set to allow for sampling of soils adjacent to the ponds. The current boundaries are shown on Fig. 1.6. These boundaries are: to the south, the main stem of Bear Creek; the boundary of the West End Treatment Plant to the west; and roads at the west end of the Y-12 Plant as the east and north boundaries.

For the BYBY, OLF, and SL 1, the geographic boundaries of the disposed materials and soil contamination are not known at this time and contamination that may be attributable to the BYBY has been reported outside previously determined limits. It is probable that secondary contamination from more than one of these potential sources may be commingled; therefore, reasonable boundaries on the study have been set to allow for sampling of soils adjacent to the sites and to include all three sites within one boundary. There are no boundaries identified between the OLF, BYBY, and SL 1. The current boundaries are shown on Fig. 1.7. These boundaries are: to the south, the main stem of Bear Creek and Bear Creek Valley Road; the north-south OLF access road to the west; the RCRA chemical; storage area and salvage yard to the east, and the BCBG access road to the north.

For the BCBG, the geographic boundaries of the disposed materials are the dimensions of the trenches in the disposal areas. However, the conceptual site model for the BCBG includes possible down-gradient migration of contaminants in groundwater and precipitation of contaminants causing secondary contamination in soils adjacent to the burial grounds and the present caps. The extent of secondary soil contamination is not known at this time; therefore, reasonable boundaries on the study have been set to allow for sampling of soils adjacent to the ponds. The current boundaries are shown on Fig. 1.8. These boundaries are: to the south, the main stem of Bear Creek; the BCBG access road and the BCBG fence to the west; the BCBG access road and NT-6 to the east, and the BCBG fence and the upper limits of BG-A north on Pine Ridge to the north.

The conceptual boundaries are those between the source site and the integrator OU. Remediation of contaminated groundwater and surface water is not within the scope of this RI/FS; however, sampling of these media will be carried out because the study of a source unit with the aim of evaluating the extent of off site migration of contaminants requires an integrated investigation of all media involved. The boundary of the study lies at the groundwater table or at adjacent surface water bodies. These boundaries are, however, limits on remediation and not on sampling.

Fig. 1.6. S-3 Site proposed study boundary.

Fig. 1.7. Boneyard/Burnyard proposed study boundary.

Fig. 1.8. Burial Grounds proposed study boundary.

1.8.5 Decision Rule

The decision rule to be followed for cleanup of the Y-12 Plant and its related source and integrator OUs is described in the *Oak Ridge Reservation Site Management Plan for the Environmental Restoration Program* (Energy Systems 1992a). In this decision-rule approach, the purpose is to provide for early remedial actions as information becomes available to support remedy selection while still facilitating a final remedy. The intent is to achieve control of contaminant sources that may be affecting the integrator OU and that constitute unacceptable risks via direct contact and exposure pathways. The ultimate intent is to determine if each source OU contributes to unacceptable risks at the site and in the integrator OU. The decision rule is incorporated in the facility OU strategy outlined in Fig. 1.1.

The decision-rule approach for use within the BCV OU 1 source OU is described in Sects. 1.4 and 1.5 and is shown in Figs. 1.9 and 1.10. The initial approach will lead to a recommendation of no further investigation, a removal action/interim action, or a continuation of the CERCLA process leading to a ROD. Once the investigation is complete, an RI report will be issued that describes the results of the site characterization activities as well as the baseline risk assessment.

Following the RI, a limited number of alternatives will be developed in the FS for remediation of the site to ensure long-term protection of human health and the environment. An analysis of alternatives will then be performed using the nine criteria established (Sect. 1.8.2). Based on the analysis of all alternatives studied in the FS and the evaluation of the alternatives against these criteria, DOE will recommend preferred alternatives for cleanup. After review and comment on the plan by the regulators and after those comments have been addressed and resolved, the plan will be made available for public comment. Initially, comments will be addressed and incorporated in a ROD followed by remedial design/remedial action for that action. Ultimately, the source OUs in Bear Creek Valley, together with the integrator OU, will develop a final valley-wide BRA to be able to issue a ROD for the whole valley.

1.8.6 Uncertainty Constraints

At this stage in the RI/FS process, decision errors are used to determine the sample size that is needed to provide input to and to resolve the decisions. This is done by using standard statistical formula, for example those published in *Methods for Evaluating the Attainment of Cleanup Standards. Volume 1. Soils and Solid Media*. (EPA 230/02-89-042). The formula used to calculate the number of samples required is selected dependent upon the sampling strategy. Formula differ for random sampling, stratified random sampling or sequential sampling. Limits on decision errors are entered into these formula along with the estimated variance in contaminant distribution, the required resolution, and the action levels for COPCs. This exercise needs to be carried out for each COPC at each site.

At this time, the list of COPCs is extensive and variance in contaminant concentration distributions is not known. Conservative values for variance could be assumed at this time; however, using estimates based on conservative values of variance may result in large numbers of samples being taken and analyzed. This would not be a cost-effective approach to sampling sites in this OU. The strategy of this Work Plan is to carry out sampling in two phases. The

Fig. 1.9. Decision rule for BC

Fig. 1.10. Decision rule for

Does erosion of surface soils result in contamination to sediments in adjacent water bodies?

Phase 1 sampling

Does existing data quality and model input requirements?

NO

Additional sampling for soil parameters in phase 2

Form soil loss sediment transport model

NO
Surface runoff not considered in the FS/ROD

QUESTION 6

What parameters are required to assess likely remedial alternatives?

Phase 1 sampling

Does existing data meet quality requirements?

NO

Additional sampling for soil parameters in phase 2

YES

Input to the ROD

first phase will involve sufficient sampling to obtain better estimates of contaminant variance and to reduce the COPC list to those contaminants that can be demonstrated to be present as secondary contamination in soils adjacent to the waste sites. Phase 2 will be designed to determine the extent of the more focused list of contaminants and will employ rapid field screening method where possible.

On completion of Phase 1 sampling, the limits on decision errors can be specified. At this time the format of Phase 2 sampling will be determined, and the required number of samples will be calculated to satisfy the limits on decision errors for the focused COPC list. It is proposed that a meeting will be held with all parties of the FFA after completion of Phase 1 at each site and the DQO process for Phase 2 sampling will be completed at this meeting.

Two sets of data will be evaluated to make the final decisions for the various sites within BCV OU 1. The historical data from multiple past studies will be weighed for their reliability and usefulness and incorporated in a site-specific screening-level risk assessment that will be used to make decisions in Stage 1 of the RI. Considerations that give weight to the usefulness of historical data are the degree of documentation available both for sample collection and for laboratory analysis, the physical coverage of sampling over the site, and the spectrum of analytes determined in the laboratory analysis. The criteria for determining the usefulness and reliability of historical data are very site specific. The usefulness of data refers to whether the data can be used to meet project DQOs. In Chap. 6 of the work plan, the data requirements are outlined, and where historic data exists that can meet these requirements, the data are identified in the site action plan (Chap. 7). More recent studies (circa 1990), in general, support a broader range for uses of the data than did older sampling events. Uncertainty is usually inherent when using any historical data because of incomplete documentation of quality assurance (QA) procedures. Exact sampling locations and methods may not be well known, and documentation in the form of field logbooks may not be available for answering questions or supporting field decisions affecting sample integrity. Laboratory analysis may have been conducted under an unknown level of quality control (QC) and statement of work, thereby undermining confidence in precision, accuracy, and sensitivity levels.

The approach taken for BCV OU 1 is that if the preliminary hazard evaluation indicates that a problem exists or data gaps prevent a complete hazard evaluation, then an investigation will be conducted to fill data gaps and carry out validation of historical data for which QA is lacking through confirmatory sampling. If the preliminary hazard evaluation does not indicate that a problem exists, then the quality of the historical data will be evaluated in order to support an NFI, which may involve sampling to verify these data. Where historical data are deemed to have insufficient QA controls, the sampling plan will propose additional samples to confirm values in this data set, allowing the data to be used in the RI.

Any data derived from the sampling proposed in this work plan will be collected and documented in such a manner as to be technically and legally defensible. Specific methods to ensure defensibility are discussed in Chap. 8, the Quality Assurance Project Plan (QAPjP). Sampling designs will be set forth to determine the nature and extent of contamination and to characterize the distribution of the contaminants. Sampling will be performed according to Energy Systems Environmental Surveillance Procedures (ESPs) (Kimbrough et al. 1990), and, in general, laboratory methods will be performed according to the EPA-Contract Laboratory Program (CLP) statement of work (SOW) protocols. Sampling procedures will be documented with field logbooks and chain-of-custody records. Each laboratory analysis will have a data package documenting the procedures followed for the analysis. Documentation

at all stages of sampling and analysis will include details of procedures used to minimize cross-contamination and to confirm accuracy and precision. In cases where decisions do not require that specific contaminants be analyzed to the level of accuracy defined by the CLP protocols, other more cost-effective analytical methods may be used. For example, NAA may be used for analysis of metals in soils where the quantification required by the DQOs is not below 100 ppm.

Uncertainties exist for both the historical and proposed new data sets. The primary uncertainties are related to the physicochemical nature of the contamination and the transport medium. The heterogeneity of the waste and soil mediums, the leachability of the waste materials, the mobility of specific contaminants, interaction between contaminants, and the physical media (affecting fate and transport) will impart a degree of uncertainty to the results. Two objectives will be most affected. First, because of the heterogeneity of the waste and soils, precise characterization of the wastes and of soil contaminant concentrations is difficult. Field duplicates can estimate the variance of the soil contaminants within a specified level of confidence. Second, the contribution of the contaminants from the wastes and the soil to the groundwater cannot be accurately estimated. Bench-scale leaching studies are inaccurate in duplicating the environmental conditions; geochemical models cannot account for unknown/unquantified variables; and partitioning estimates based on concentrations observed in the soil and those observed in the groundwater are highly inaccurate. The results of this investigation must be considered in light of these uncertainties, and decisions must be made accordingly. Statistical methods will be used to calculate uncertainties resulting from media heterogeneity, and uncertainties will be minimized, where possible, by employing statistically based sampling plans. A special subset of uncertainty exists for the analysis of Y-12 Plant records that has been carried out to characterize the waste materials in the Bear Creek Burial Grounds (Energy Systems 1993a). Approximately 50,000 individual disposals to the Burial Grounds have been tabulated from the Y-12 Plant records. The disposal records cover three main types of disposal practices: (1) blanket disposals that account for about two-thirds of the total waste in the Burial Grounds, (2) nonroutine hazardous disposals that account for a small portion of the total wastes, and (3) other disposals that account for about one-third of the total waste. Volumes or masses of wastes have been determined using, in order of completeness and usability of the data, the following sources:

- **Y-12 Maintenance Division Monthly Activity Reports.** These reports are available from January 1957 to the present (with roughly 1 month missing from the files per year).
- **Requests for Hazardous/Classified Disposals (UCN-2109 and UCN-6538).** Approximately 11,300 UCN-2109s and 1,600 UCN-6538s have been entered into a computer data base.
- **Kardex Blanket Files.** Approximately 600 cards with 27,000 entries from 1964–1980 covering waste disposal under 170 blankets have been analyzed.
- **Tanker Logs.** These records cover ~1,800 liquid disposals on 25 different blankets between 1971 and 1980. These logs have been transcribed into a computer data base.
- **Vehicle Trip Logs.** These logs cover blanket, hazardous/classified, and other deliveries from 1960 to 1980. Most of the information in vehicle trip logs is already recorded in the previous files and records, and the logs themselves are not standardized so they are difficult to evaluate. For these reasons, roughly 8% of vehicle trip logs (covering 290 days) were sampled and examined to statistically estimate the total deliveries for this time period. This analysis has been verified by comparison to the ongoing SWIMS program.

- **Waste Management Inventory—Manifest System.** Since 1980, all disposal information has been placed on a computerized data base. Over 13,000 disposals to the Bear Creek Burial Grounds have been recorded by this system. Since 1980, disposal records are complete, and the quantities and compositions of waste disposed of in the Burial Grounds since that time can be determined precisely.

The quantity and composition of ~80% of the blanket disposals mentioned in the Monthly Activity Reports have been tabulated. The blanket disposals account for ~two-thirds of the total waste in the Burial Grounds; thus, 53% of the waste materials in the Burial Grounds has been tabulated using these records. It is assumed that the waste documented by the 20% of blanket records that have not been tabulated are similar to those recorded in the 80% of records that have been tabulated. Therefore, through the use of these blanket disposal records, the wastes in the Burial Grounds can be characterized.

The total number of UCN-2109s for hazardous/classified disposals mentioned in the Monthly Activity Reports corresponds closely to the number of records tabulated, so the records of this type are essentially complete. The hazardous/classified deliveries account for a small fraction of total deliveries to the Burial Grounds; however, the composition of many of these deliveries is of greater potential interest.

For the estimated one-third of all Bear Creek deliveries that were neither hazardous nor blanket, the Vehicle Trip Logs have been used to determine the quantity and makeup of the wastes. For these materials, a random sample of work days and a statistical analysis of the sample have been carried out and are detailed in the waste inventory report (Energy Systems 1983). Using this random sample, the volume and composition of this remaining one-third of the wastes have been determined with ranges at specified levels of uncertainty.

In summary the existing analysis of the waste delivery records and document activities at the Burial Grounds provides an inventory of the waste materials with a measurable degree of uncertainty. This provides a better characterization of waste materials than can be obtained by sampling all trenches in the Burial Grounds and with less risk posed to RI personnel (Energy Systems, 1983, Parts 1-10).

Uncertainty that exists concerning the extent to which contaminants will leach from soils will be evaluated using soil leaching models such as HELP, SESOIL, and MULTIMED. The contaminant source information for these models will be derived from sampling or from waste inventory data. The soil and aquifer parameters will be derived from sampling data. The natural variability in these data will be accommodated by running these models using a Monte-Carlo simulation. This will allow evaluation of the mean leachate concentrations.

1.8.7 Design Optimization

The objective of the BCV OU 1 sampling design is to characterize the nature, extent, and distribution of contaminants within the site in a statistically valid and legally defensible manner. A statistically based sampling design is used during the RI because it allows the results to be extrapolated from the set of samples collected in BCV OU 1 to the entire site under investigation in a scientifically and technically valid fashion. This sampling design will allow the RI investigators to estimate contaminant concentrations means and associated confidence intervals to determine whether or not a contaminant exceeds background levels or poses a risk to human health and the environment. In addition, it will provide the FS

investigators with an approximation of media effect above the levels of preliminary remediation goals and delineate areas to be screened during the actual cleanup operations. While providing these data, the sampling design also minimizes worker health risk, cost, and time by focusing on the most effective methods for satisfying the DQOs.

As is discussed in detail in Chap. 3, a number of previous studies at the sites in OU 1 have shown that potential COCs currently exist in the soils, sediments, surface water, and groundwater in the vicinity of BCV OU 1. Specifically, PCB, uranium, and cadmium contamination has been found in surface soils and sediments; nitrate, uranium, aluminum, and cadmium contamination has been found in surface waters; and contaminant plumes of VOCs and nitrate have been found in groundwater samples.

In order to determine the magnitude of the contamination and the relative location of the contaminant sources (trenches, pits, ponds) for these various contaminants, a two-phased sampling design is proposed based on the DQOs previously discussed. During the first phase of the RI investigation, the nature and magnitude of contaminant concentrations will be determined in OU 1. During the second phase, the distribution and extent of the major COCs (as identified during Phase 1) will be determined. Although described here as a two-phase approach, the intent of this work plan is to carry out all field work in one field activity. Decision rules laid out in this work plan will allow rapid evaluation of the sampling required for Phase 2. The size of the OU will allow concurrent sampling and data evaluation.

1.8.7.1 Phase 1 sampling design

A stratified and systematic random sampling design has been used during Phase 1 of the RI. As will be discussed in Chap. 7, BCV OU 1 is divided into a number of different strata (waste areas): the S-3 Site, the Oil Landfarm, the Boneyard/Burnyard, and the various waste areas (trench groupings) within the Bear Creek Burial Grounds. Within each of the BCV OU 1 strata, a number of potential sampling points have been selected that are located immediately downgradient from each of the (capped areas) or located directly on uncapped areas. A systematic random sample will then be selected for each area with the number of samples selected being proportional to the relative size of each area.

After the samples for the Phase 1 RI have been collected and analyzed, a five-part decision rule will be employed, using the EPA's *Risk Assessment Guidance for Superfund, Vol. I: Human Health Evaluation Manual, Part A* (EPA 1989d) to determine whether the concentration in the samples collected in BCV OU 1 exceeds established detection limits, background concentrations (if known), established action levels, or preliminary remediation goals. Specifically, potential COCs will be considered for further evaluation if any one of the following applies:

- The analyte was positively detected in at least one sample.
- The analyte concentration was measured at levels significantly higher than naturally occurring levels (background) of the same analyte.
- The analyte concentration exceeded established action levels.

- The analyte was identified but the exact concentration was unknown (i.e., J-qualified data) or was detected at levels significantly higher than the same chemicals detected in associated blank samples.
- The analyte concentration was higher than established preliminary remediation goals.

A systematic process, following this decision rule, will be used to narrow the list of COPCs. COPCs will be associated where possible with specific media (soil, groundwater, surface water, or air) and specific sites within OU 1. This will allow considerable streamlining of sampling and analytical techniques, as described below in Phase 2 sample design.

1.8.7.2 Phase 2 sampling design

The specific sampling design to be used in Phase 2 of the RI will largely depend on the results obtained in Phase 1. As with Phase 1, a statistically based sampling design will be used to characterize the distribution and extent of the COPCs determined in Phase 1 using guidance from the soil sampling Quality Assurance User's Guide (Barth et al. 1989).

Since the list of contaminants to be investigated in Phase 2 should be smaller than the list of contaminants analyzed in Phase 1, many more samples can be collected to characterize the extent of contamination. A series of rapid field screening techniques that involve rapid sample collection and analysis techniques (including hydropunch sampling, field GCs, NAA, and field XRF) will be used to quickly identify locations of potential contamination. The results from these sample collections and surveys will be verified by subsequent analytical laboratory analysis for statistically selected subsets of samples.

1.9 SCHEDULE

The schedule for the BCV OU 1 RI is provided in Fig. 1.11. It is proposed that the Phase 1 sampling as is outlined in this work plan be completed within 9 months of Notice to Proceed (NTP). During this time soil data should be made available from Phase 1 sampling to design Phase 2 sampling. It is envisioned that before Phase 1 is completed, Phase 2 will have begun, thus ensuring continuous field work in one operation.

1.10 QUALITY ASSURANCE

1.10.1 Program Planning and Implementation

The overall planning and implementation of QA and QC activities for the BCV OU 1 RI are described in Chap. 8 and are governed by the *Environmental Restoration Division Quality Assurance Program Plan*, ES/ER/TM-4, Rev. 2 (Energy Systems 1992b), which incorporates the guidance in *Interim Guidelines and Specifications for Preparing Quality Assurance Project Plans*, QAMS 005/80 (EPA 1980b) and Energy Systems procedures; the American Society of Mechanical Engineers document *Quality Assurance Program Requirements for Nuclear Facilities*, NQA-1 (ANSI/ASME 1986); and applicable DOE orders.

Fig. 1.11. Schedule for the BCV OU 1 RI/FS through ROD.

1.10.2 Project-Specific Planning

Project-specific QA planning is described in detail in Chaps. 7 and 8. All QAMS 005/80 QA/QC and ER Quality Program Plan requirements are addressed in this work plan. A project-specific locator page is presented as Table 1.6.

Table 1.6. Cross reference of EPA QAMS 005/80 and ES/ER/TM-4/R2 elements with BCV OU 1 RI work plan sections

EPA QAMS 005/80	ES/ER/TM-4/R2 element	Location in RI work plan
1. Title Page	4.1.2 QA Project Plan Document Control	Title Page; Chaps. 8, 9, and 10
2. Table of Contents	4.1.2 QA Project Plan Document Control	Contents
3. Project Description	1.6.2.2.1 Project Description	Executive Summary
4. Project Organization and Responsibility	1.5 Organization and Program Responsibilities	Sect. 7.2
5. QA Objectives	6.1.1. QA Objectives for Measurement Data	Sect. 1.8, 8.1
6. Sampling Procedures	5.3.1 Sampling Procedures 6.1.2 Sampling Procedures	Sect. 7.4.1, 8.2
7. Sample Custody	5.3.2 Sample Custody	Sect. 8.4
8. Calibration Procedures and Frequency	5.4 Control of Measuring and Testing Equipment	Sect. 8.5
9. Analytical Procedures	5.3.3 Analytical Procedures 6.1.6 Specific Routine Procedures	Sect. 8.3
10. Data Reduction, Validation, and Reporting	6.1.3 Analytical Procedures	Sect. 8.7
11. Internal QC Checks	8.2 Inspection, Testing, and Operating Status	Sect. 8.8
12. Performance and System Audits	9.3 Field Audits and Surveillances 9.4 Analytical Laboratory Audits and Surveillances 10.1 Audits and Surveillances 10.1.3 Field and Laboratory Audits and Surveillances	Sect. 8.10
13. Preventive Maintenance	5.4 Control of Measuring and Testing Equipment	Sect. 8.6

Table 1.6 (continued)

EPA QAMS 005/80	ES/ER/TM-4/R2 element	Location in RI work plan
14. Specific Routine Procedures Used to Assess Data Precision, Accuracy, Representativeness, Completeness, and Comparability	6.1.5 Internal Quality Control Checks and Frequency	Sect. 8.9
15. Corrective Actions	3.2 Control of Nonconforming Items	Sect. 8.11
16. QA Reports to Management	1.6.2.2.2 QA Reports to Management 3.2 Control of Nonconforming Items	Sect. 8.12

2. HISTORY AND CURRENT CONDITIONS

BCV OU 1 consists of the S-3 Site, the Boneyard/Burnyard, the Oil Landfarm, SL 1, and the Burial Grounds source areas in BCV. This chapter summarizes the available information on the nature, quantity, and concentration of contaminants in each of the source areas (formerly identified as SWMUs) located in BCV OU 1. In the absence of direct evidence (measurements) of the contents of each source area, the nature and concentration of contaminants in receiving media (surface water, soil/sediment, groundwater) immediately adjacent to each source area will be summarized. The purpose is to identify those contaminants that may have been or may be released from each source area.

2.1 GEOGRAPHIC INFORMATION

The DOE ORR is located within the corporate limits of the city of Oak Ridge in eastern Tennessee, ~16 km (10 miles) southeast of the Cumberland Mountains and 113 km (70 miles) northwest of the Great Smoky Mountains. Figure 2.1 shows the city's location approximately 20 miles northwest of Knoxville, Tennessee, which is east of Nashville, Tennessee, and north of Atlanta, Georgia, at the intersection of I-40 and I-75. The area is serviced by Interstates 40 and 75, which intersect in nearby Knoxville, Tennessee, east of Oak Ridge. The reservation, 14,300 ha (35,300 acres) of federally owned land, houses three facilities (the K-25 Site, ORNL, and the Y-12 Plant), which are managed for DOE by Martin Marietta Energy Systems, Inc. (Energy Systems).

The Y-12 Plant contains ~324 ha (800 acres) and is located adjacent to the corporate center of the city of Oak Ridge. The plant occupies a valley between Chestnut Ridge, to the south, and Pine Ridge, to the north of the plant. The plant complex, built predominantly in the mid-1940s, is roughly divided into two portions; the western portion was previously devoted to DOE weapons-manufacturing activities, and the eastern portion is devoted to ORNL research programs, general plant operations, and maintenance activities.

Extensive geographical information concerning the Y-12 Plant, its facilities, and the surrounding area is given in Welch (1989). The Y-12 Plant is located in the Valley and Ridge Physiographic Province, which lies between the Cumberland Mountains to the northwest and the Great Smoky Mountains to the southeast. This province is characterized by roughly parallel, alternating ridges and valleys. The main facilities of the Y-12 Plant lie on the BCV floor, with auxiliary facilities located along Pine Ridge and Chestnut Ridge.

Bear Creek originates east of the former S-3 Ponds near the western margin of the Y-12 Plant. BCV OU 1 is located immediately west of the main Y-12 Plant facilities and extends ~3.2 km (2 miles) west down BCV along the north side of Bear Creek and Bear Creek Valley Road (Fig. 2.2). From east to west, the BCV OU 1 consists of the S-3 Site, the Oil Landfarm, the Boneyard/Burnyard, SL 1, and the Burial Grounds (Fig. 2.2).

The BCV OU 1 source areas are located within the upper portion of the Bear Creek watershed and several of the creek's tributaries flow through BCV OU 1. In this document, as in previous studies (Turner et al. 1991), monitoring and sampling sites are referred to by "creek kilometers." Sites are designated as "BCK," for "Bear Creek Kilometer," plus the kilometer value (e.g., BCK-12.36). As indicated in the footnote to Table 2.1, tributaries are

Fig. 2.1. Location of Oak Ridge.

Fig. 2.2. Location of BCV OU 1 in re

LEGEND:

SCALE: 1" = 2500'

BEAR CREEK VALLEY OPERABLE UNIT 1 RI WORK PLAN

Science Applications
International Corporation

REVISION	DRAWN BY:	CHKD. BY:	DATE:
0	D. EVANS		5-20-93
1	D. EVANS		8-27-93

XREFERENCES	PLOT FILES
93044/XREF/BASEMAP 93044/XREF/SITEDASH	93044/PLOT/325BC1-1

FIG 2.2

FIGURE #

93044\DWGS\325BC1-1

CAD FILE #

0

ROT ANGLE

Table 2.1. Upper Bear Creek kilometer (mile) locations

Kilometer	Mile	Landmark*
12.87	8.00	Headwaters near S-3 Ponds
12.46	7.74	Road (gravel) crossing between "lagoons"
12.39	7.70	Confluence with tributary NT-1
12.38	7.69	Confluence with spring SS-1
11.97	7.44	Bear Creek Road crossing
11.96	7.43	Confluence with tributary NT-2
11.84	7.36	Road (gravel) crossing at east end of Oil Landfarm
11.68	7.26	Confluence with spring SS-2
11.67	7.25	Confluence with spring SS-3
11.59	7.20	Confluence with tributary NT-3
11.28	7.01	Road (gravel) crossing at west end of Oil Landfarm
11.15	6.93	Confluence with tributary NT-4
10.94	6.80	Confluence with tributary ST-1
10.78	6.70	Confluence with tributary NT-5
10.40	6.46	Confluence with tributary NT-6
10.27	6.38	Road (gravel) crossing at entrance to Burial Grounds
10.14	6.30	Confluence with spring SS-4
9.99	6.21	Confluence with tributary NT-7
9.98	6.20	Road (gravel) crossing at west end of Burial Grounds
9.45	5.87	Confluence with tributary NT-8
9.42	5.86	Road (asphalt) crossing (Bear Creek Rd)
9.41	5.85	Confluence with spring SS-5
9.06	5.63	Confluence with tributary NT-9
8.67	5.39	Confluence with tributary NT-10

*Naming convention: NT = northern tributary, ST = southern tributary, SS = south side spring.

designated as NT, ST, WT, or ET plus a value representing the tributary number counting from the headwaters. BCK OU 1 is located between BCK 12.87 and BCK 9.41 and is transected by Tributaries NT-1 through NT-7 (Fig. 2.3). Table 2.1 lists creek landmarks referenced to kilometers or miles within the upper portion of Bear Creek. Thus, NT-4 is the fourth tributary originating on the north side of Bear Creek. Figure 2.3 shows some important locations on the upper section of Bear Creek using this site naming system.

2.2 BACKGROUND INFORMATION

The Oak Ridge Y-12 Plant was built by the U.S. Army Corps of Engineers in 1943 as part of the Manhattan Project. The original mission of the Plant was to separate the

Fig. 2.3. Upper Bear Creek Hydrologic Regime.

fissionable isotope of uranium (^{235}U) using an electromagnetic separation process. After World War II, this process was discontinued in favor of a more economical gaseous diffusion process conducted at the nearby Oak Ridge K-25 Site (Welch 1989).

Since its inception, the Y-12 Plant has developed into a sophisticated manufacturing and developmental engineering facility. Current manufacturing activities at the Y-12 Plant include chemical processing of lithium and uranium compounds and precision fabrication of components from lithium, uranium, and many other materials. Metallurgical and machine shop operations are conducted at the plant to support these activities.

For the purpose of this RI work plan, the S-3 Ponds, Boneyard/Burnyard, Oil Landfarm, SL 1, and Burial Grounds are considered to be the primary contaminant sources in BCV OU 1 (Fig. 2.2). The source areas and the possible contaminant types associated with each are summarized in Table 2.2. The general contaminant types are based on a combination of operational information detailed in Sect. 2.3 and the results of site-specific sampling that are detailed in Sects. 3.4, 3.5, and 3.6.

Much of the background information pertaining to the Y-12 Plant (and ORR in general) has remained classified since the beginning of the Manhattan Project. The current agreement with DOE is that classified material will be declassified to the extent possible. In some cases declassified summaries of classified material prepared in an appropriately secure way will be issued. (Such summaries have already been carried out for the Burial Ground inventories.) Material that remains classified will be available for review by appropriately cleared personnel and will be entered into the on-site record. Classified material will not be entered into the off-site record.

2.3 OPERATIONAL INFORMATION

The following subsections contain descriptions of the wastes managed in each source area as well as information about their generation, disposal methods, and any subsequent treatment or cleanup activities at the source area. Where monitoring data have been obtained that are specific to a source area, those results are also included.

2.3.1 S-3 Ponds

The S-3 Ponds were located adjacent to the west end of the Y-12 Plant and were a part of the former S-3 Waste Management Area (Fig. 2.4). A chronological list of events at the S-3 Ponds is provided in Table 2.3. Constructed in 1951, these were four unlined impoundments covering an area of roughly 120×120 m (400×400 ft). The original pond excavations penetrated residual soil and fill materials but, based on the depth of bedrock in wells surrounding the site, did not extend down to the bedrock. Before closure in 1988, the ponds were ~ 5 m (17 ft) deep and contained variable amounts of sludge ranging from 0.6 to 1.5 m (2 to 5 ft) thick. The sludges were produced by the neutralization and in situ denitrification of wastewater in the ponds. While in operation, each pond had a storage capacity of 9.5 million L (2.5 million gal). The site of the S-3 Ponds was closed with a multilayered RCRA cap and covered with asphalt to create a parking lot.

Accurate records of the types and amounts of waste disposed of in the S-3 Ponds are not available. However, three different sources of waste (generally classified as toxic, corrosive,

Fig. 2.4. Site map for the

Table 2.2. Summary of possible contaminant types from BCV OU 1 source areas

Source areas	Possible contaminants
S-3 Ponds	Metals—Ba, Pb, Cr
	Radionuclides—U, Pu, Tc
	Nitrate
	VOCs
Boneyard/Burnyard	Radionuclides—U
	VOCs
	Pesticides/PCBs
	Other unknowns
Oil Landfarm	Metals—Be, Cr, As, Pb, Hg
	Radionuclides—U
	VOCs
	PCBs
SL 1	Unknown
Burial Grounds	Metals—Be, As, Cr, Pb
	Radionuclides—U, Th
	VOCs
	PCBs

and radioactive) are known to have been disposed of in the S-3 Ponds. The three sources are (1) liquid wastes generated by Y-12 Plant operations and other facilities, (2) sludges generated from the in situ neutralization and biodegradation of the pond waters and sludges generated from the treatment of acid raffinate, and (3) contaminated sediments from upper Bear Creek. Table 2.4 is a generic summary of the waste materials known to have been disposed of in the S-3 Ponds.

Waste products were discharged into the ponds from tank trucks, small containers, and an underground nitric acid pipeline, which was abandoned and plugged in 1983. The wastes consisted primarily of nitric acid, nitrate waste, pickling and plating wastes, and miscellaneous liquid wastes associated with routine cleanup operations (e.g., mop water). Other wastes included dilute acid, machine coolants, caustic solutions, and acids with a pH of less than 2.0. During the mid-1950s, uranyl nitrate solutions containing small amounts of plutonium were discharged into the ponds. Depleted uranium in nitric acid solution and technetium contained in raffinate and condensate also may have been placed in the ponds while they remained in operation. All waste discharges into the ponds were terminated in March 1983.

Table 2.3. Chronological list of events at the S-3 Ponds

Year	Event	Reference
1951	The S-3 Ponds were constructed. They consisted of four unlined surface impoundments with a total storage capacity of 38 million L (10 million gal)	Energy Systems 1992c
Mid-1950s	Uranyl nitrate solutions containing small amounts of plutonium were discharged into ponds	
1951-1983	Primary waste disposed of in S-3 Ponds consisted of nitric acid, nitrate waste, pickling and plating wastes, and miscellaneous liquid wastes associated with routine cleanup operations. Other wastes included dilute acid, machine coolants, depleted uranium, technetium, caustic solutions, biodegradation sludges, and acids with pH of <2.0	
1976	The nitric acid recovery system became operational	
1983	Annual liquid waste entering the ponds was reported to be ~10.3 million L (2.7 million gal)	
1983	The underground pipeline used to transport waste to the S-3 Ponds was plugged and abandoned	Energy Systems 1987a
March 1983	All waste discharges into the ponds were terminated	
1983	Characterization monitoring at the S-3 Ponds began in response to a Memorandum of Understanding (MOU) signed by DOE, EPA, and TDEC	Jeter 1983 and Turner et al. 1991
1983	In situ treatment of wastewater in the ponds began	
May-November 1983	Liquid in all four ponds was neutralized	
May 1983-September 1984	Biodegradation of liquids occurred. Nitrate levels in the pond water decreased to ~100 ppm	
September 1984	In situ treatment of wastewater was completed. Nitrate levels were maintained below 50 ppm	
December 1984	S-3 Ponds HWDU ^a was granted interim status under TN Rule 1200-1-11-.07(3)	

Table. 2.3 (continued)

Year	Event	Reference
1985	Supernatant treatment to remove trace metals and organics began. The supernatant was pumped to the S-3 Ponds Treatment Facility where processes that included pH adjustment, precipitation, coagulation, flocculation, carbon absorption, filtration and chlorination were conducted. The treated effluent was discharged to EFPC in accordance with an NPDES ^b permit	
January 1986	Assessment monitoring was initiated at the S-3 Ponds	
Spring 1986	Supernatant treatment processes were completed	
March 1987	The formalized Groundwater Quality Assessment Plan for the site was submitted to TDEC	Geraghty and Miller 1987
1988	Sediment from the two "lagoons" in upper Bear Creek was placed in the southwest cell of the S-3 Ponds. The sediment was characterized in place and after being disposed (Turner and Kamp 1984). Data are reported in McMahon 1988	EMD HSEAD 1988
February 1988	RCRA Closure Plan was revised	EMD HSEAD 1988
1988	The S-3 Ponds were closed with a multi-layered RCRA cap. Currently, the site is covered with asphalt and used as a parking lot	Stone 1990
November 15, 1990	Final closure of the S-3 Ponds was accepted by TDEC	
September 30, 1991	TDEC issued a RCRA postclosure permit for the S-3 Ponds	

^aHWDU = Hazardous Waste Disposal Unit.

^bNPDES = National Pollutant Discharge Elimination System.

Influent discharge rates varied throughout the period of operation, but the amount of waste was significantly reduced when a nitric acid recovery system became operational in 1976. In 1983, the annual quantity of liquid waste entering the ponds was reported to be about 10.3 million L (2.7 million gal). The annual loss from the ponds caused by evaporation was estimated to be one-half to two-thirds the annual input from direct rainfall, or about 11.4 million L (3 million gal) (Energy Systems 1987a); the remainder, plus waste inputs into the ponds, is assumed to have infiltrated to the groundwater.

Table 2.4. Waste disposed of in the S-3 Ponds^{ab}

Substance	Classification
Liquid waste	
Nitrate wastes	Corrosive
Machine coolants	Toxic
Caustic solutions	Corrosive
Pickling wastes	Toxic
Plating wastes	Toxic
Mop water	Toxic
Nitric acid	Corrosive
Miscellaneous dilute acids	Corrosive
Miscellaneous acids (pH<2)	Corrosive
Depleted uranium (in nitrate solutions)	Radioactive
Plutonium (in uranyl nitrate solutions)	Radioactive
Technetium (in raffinate and condensate)	Radioactive
Miscellaneous radiological waste ^c	Radioactive
Biodenitrification sludges	Toxic
Contaminated Bear Creek sediments	Toxic

^aThe annual volume of waste was 2.7 million gal in 1983. This volume is reduced from pre-1976 quantities.

^bThe ponds were operated from 1951 to 1984.

^cSee complete list in Appendix A.

2.3.1.1 S-3 Ponds treatment and closure

In situ treatment of wastewater in the S-3 Ponds began in 1983 and continued until September 1984. Between May and November 1983, liquids in all four of the ponds were neutralized. Biodenitrification of the liquids occurred from May 1983 to September 1984, causing nitrate levels in the pond water to decrease to ~100 parts per million (ppm). Nutrients continued to be added to the ponds to maintain the bioreaction, and aerators agitated the sludge to denitrify the residuals (Table 2.5). Subsequently, nitrate levels were maintained below 50 ppm in the pond water.

After denitrification, the contents of the ponds were allowed to settle, and the supernatant was pumped to the S-3 Ponds Treatment Facility for removal of trace metals and organics. A flexible hose attached to the suction side of the pump allowed the decanting of remaining small pockets of liquid on the surface of the sludge. The treatment steps for the supernatant included pH adjustment, precipitation, coagulation, flocculation, carbon adsorption, filtration, and chlorination. Treated effluent was discharged to EFPC in accordance with a National Pollutant Discharge Elimination System (NPDES) permit. The supernatant treatment began in 1985 and was completed by the spring of 1986.

Table 2.5. Chemical additions to the northeast quadrant (S-3 Ponds)

Date	Chemical addition
4/26/88	38,000 L (10,000 gal) of acetic acid and lime slurry
4/28/88	19,000 L (5000 gal) of acetic acid and lime slurry
4/29/88	38,000 L (10,000 gal) of acetic acid and lime slurry
5/04/88	408 kg (900 lb) of lime slurry
5/09/88	318 kg (700 lb) of lime slurry
5/11/88	1270 kg (2800 lb) of lime slurry
5/12/88	1451 kg (3200 lb) of lime slurry
5/18/88	2280 kg (600 gal) of acetic acid and 1361 kg (3000 lb) of lime slurry
5/19/88	3.8 L (1 gal) of triethyl phosphate and 907 kg (2000 lb) of lime slurry
5/23/88	1361 kg (3000 lb) of lime slurry
5/24/88	2280 (600 gal) of acetic acid and 2359 kg (5200 lb) of lime slurry
5/25/88	380 L (100 gal) of acetic acid and 3.8 L (1 gal) of triethyl phosphate

Source: Adapted from Stone 1990.

Immediately before structural stabilization and filling for closure of the S-3 Ponds in 1988 (Energy Systems 1988), sediment that had accumulated in two "lagoons" in upper Bear Creek adjacent to the S-3 Ponds was placed in the southwest cell of the S-3 Ponds. Turner and Kamp (1984) characterized this material in place and each truckload of sediment placed in the ponds was also characterized by sampling and laboratory analyses. The data from the 1988 sampling and characterization have been tabulated in Appendix A.

Water and sludges in the S-3 Ponds have been extensively characterized on several occasions (Jeter 1983; Kimbrough and Turner 1987). Although the pond water chemistry varied from pond to pond and from year to year, in general, pond water chemistry remained highly acidic, dominated by the nitrate anion, with extremely high concentrations of the major cationic constituents (calcium, magnesium, potassium, aluminum) and moderately high concentrations of trace metals. The few organics analyses available for the pond waters indicate that concentrations of organics could have been highly variable over the period of operation. Sludges in the ponds were characterized both before and following neutralization/biodenitrification (Jeter 1983; Kimbrough and Turner 1987). Appendix A gives selected results of investigations of the composition of water and sludges in the S-3 Ponds that remained after neutralization.

The S-3 Ponds Hazardous Waste Disposal Unit (HWDU) was granted interim status under TN Rule 1200-1-11-.07(3) in December 1984 and is therefore subject to the groundwater monitoring requirements described in TN Rule 1200-1-11-.05. Before this date, groundwater monitoring at the S-3 Ponds was conducted in accordance with two DOE monitoring programs: historical and characterization monitoring. Characterization monitoring

at the S-3 Ponds began in 1983 in response to the Memorandum of Understanding (MOU) signed by DOE, EPA, and TDEC and continued until 1986. Assessment monitoring was initiated at the S-3 Ponds in January 1986. However, a formalized Groundwater Quality Assessment Plan for the site was not submitted to TDEC until March 1987 (Geraghty and Miller 1987).

TDEC granted final closure of the S-3 Ponds in November 1990. The site was closed as an HWDU and is subject to postclosure care under TN Rule 1200-1-11-.05(14)(e). TDEC issued a RCRA postclosure permit for the S-3 Ponds HWDU on September 30, 1991. The permit contains a technical definition of the boundaries or vertical line of compliance of the regulated unit as well as maximum contaminant levels for regulated contaminants at specified monitoring points. The permit was appealed by DOE and Energy Systems on October 30, 1991.

2.3.2 Oil Landfarm

The Oil Landfarm (Fig. 2.5) is a unit that was included in the former Oil Landfarm Waste Management Area. A chronological list of the activities at the site is included in Table 2.6. The Oil Landfarm is located ~2.4 km (1.5 miles) west of the Y-12 Plant, north of SL 1 and Bear Creek Road (Fig. 2.2). Landfarming activities ended in 1982, after 9 years of operation during which ~3.8 million L (1 million gal) of liquid wastes were disposed (Turner et al. 1991). The oils and coolants deposited in the Oil Landfarm are known to have been contaminated with beryllium compounds, depleted uranium, PCBs, tetrachloroethene, and 1,1,1-trichloroethane. In 1990, the Oil Landfarm was closed as a landfill with a multilayered engineered cap pursuant to TN Rule 1200-1-11-.05(14)(e).

The facility was used for the biological degradation of waste oil and machine coolants through landfarming, a process involving the application of waste oils and coolants to nutrient-adjusted soil during the dry months of the year (April to October). After application, the plots were cultivated frequently to maintain aerobic conditions to enhance biodegradation of the wastes. When possible, waste was not applied immediately before or after periods of precipitation.

Wastes placed in the Oil Landfarm included the following:

- Waste oils.
- Beryllium-contaminated oils.
- Coolants.
- Mop waters.
- Tanker oils from the K-25 Site.
- Wastes from cooling towers and the Burial Grounds.
- Unidentified miscellaneous liquid wastes.

Fig. 25. Site map for the Oil Landfarm Area, including th

LEGEND:

	CLAY CAPPED DISPOSAL UNIT
	RCRA CAPPED DISPOSAL UNIT
	UNCAPPED DISPOSAL UNIT

BEAR CREEK VALLEY OPERABLE UNIT 1 RI WORK PLAN

SAIC
Science Applications
International Corporation

REVISION	DRAWN BY:	CHKD. BY:	DATE
0	D. EVANS		5-20-93
1	D. EVANS		8-27-93

XREFERENCES	PLOT FILES
93044/XREF/BASEMAP 93044/XREF/MANGAREA	93044/PLOT/325BC1-4

FIG 25	93044\DWG\325BC1-4	0
FIGURE #	CAD FILE #	ROT. ANGLE

Table 2.6. Chronological list of events at the Oil Landfarm

Year	Event	Reference
1973	Landfarming activities began at the site. Waste oils and coolants that contained beryllium compounds, depleted uranium, PCBs, and chlorinated organic compounds, mop waters, tanker oils from the K-25 Site, wastes from cooling towers and the Burial Grounds, and unidentified miscellaneous liquid wastes were disposed of at the site	Energy Systems 1991
1974-1975	57,000 L (15,000 gal) of oil skimmed from the surface of the two Oil Retention Ponds in BG-A were disposed of in the Landfarm	
1975	Historical monitoring at the site began as part of DOE monitoring programs	
1979	Analyses of waste were implemented before application to the Landfarm. The analytical results for oil samples collected from the surface of one of the ponds in the Burial Grounds indicated the presence of PCBs	
1979	Oils were analyzed for uranium, beryllium, thorium, and PCBs. Permissible concentration levels of 5 mg/L were established for PCBs. Waste oils containing greater than 5 mg/L with uranium concentration levels greater than permissible limits were placed in storage for incineration at the K-25 Site	
1981	Analyses of waste were expanded to include chlorinated hydrocarbons with a maximum permissible concentration of 3%	
1982	Landfarm activities ceased. Approximately 3.8 million L (1 million gal) of liquid waste were disposed of during the operational history. A total of 2.2 million L (583,500 gal) of waste coolant and 1.6 million L (430,800 gal) of waste oil were reported to be disposed of at the site (Union Carbide 1984)	Turner et al. 1991
1983	Groundwater characterization monitoring began in response to the MOU signed by DOE, EPA, and TDEC	
1983	Core samples from perimeter monitoring wells were analyzed for organic contaminants	
September 1983	Composite soil samples were collected from plots 1-12, 13-18, and 19-24 and analyzed for VOCs, acid and base-neutral extractables, inductively coupled plasma metals, and pesticide/PCBs	

Table 2.6 (continued)

Year	Event	Reference
1983-1984	Water and sediment samples were collected at tributaries NT-3 and NT-4. Levels of VOCs, phenols, and beryllium were all below detection limits	Turner et al. 1991
December 1984	The Oil Landfarm was granted interim status under TN Rule 1200-1-11.07(3)	
1985	A total of 200 soil samples were collected in a randomized block design to delineate areas containing PCB levels $>25 \mu\text{g/g}$. Samples were also analyzed for uranium	Herbes 1988
January 1986	DOE characterization monitoring ended	
January 1986	Interim status assessment monitoring was initiated at the site in lieu of detection monitoring as required under TN Rule 1200-1-11-.05(6)(a)4 due to the detection of contaminants in groundwater	
1986-1987	Groundwater samples from three monitoring wells bordering the Oil Landfarm were analyzed for total and dissolved toxic metals and gross radioactivity	
1987	DOE historical monitoring at the site ended	
February 1988	RCRA Closure Plan was revised	EMD HSEAD 1988
November 1990	The Oil Landfarm was closed as a landfill with a multilayered engineered cap pursuant to TN Rule 1200-1-11.05(14)(e)	Bohrman 1989
December 1990	Final RCRA closure for the site was certified by TDEC	Energy Systems 1992c

Because operating instructions called for different types of oils to be applied to different plots, both the composition and volume of liquid waste applied varied from plot to plot (BNI 1983). Thus, over 50% of the total amount of beryllium-contaminated oil disposed of over the lifetime of the Oil Landfarm was administered to Plot 1/2, with most of the remainder being divided nearly equally between Plots 7/8 and 9/10. Tanker oils were applied most heavily to Plots 3/4 and 5/6; coolants were applied most heavily to Plots 3/4, 5/6, 11/12, 13/14, and 15/16. (See Appendix A for a map of plot locations in the Oil Landfarm.) In general, the east plots received more total liquid waste application than the west plots, which received more than the north plots.

Summaries have been compiled for the types and amounts of waste oils (Table 2.7) and waste coolants (Table 2.8) disposed of in the plots during the operating period. While waste oils typically consisted of crankcase, hydraulic, gear, vacuum pump, mineral, and machine coolant oils, the totals for 1974 and 1975 included $\sim 57,000 \text{ L}$ (15,000 gal) of oil skimmed from the surface of the two ORPs located in Y-12 Plant BG-A. The machine coolants included nonsoluble oils and water-soluble mixtures consisting of $\sim 5\%$ oil and 95% water.

Table 2.7. Summary of waste oil disposals at the Y-12 Plant Oil Landfarm

Calendar year	Origin	No. of disposals	Volume (gal)	Active plots (acres)	Description
1973	Y-12	11	42,800	0.4	Waste oil
1974	Y-12	14	53,300	1.0	Waste oil
1975	Y-12	4	14,600	1.0	Waste oil
1976	Y-12	9	25,600	1.0	Waste oil
	NR ^a	5	17,700	1.0	No record
1977	NR	5	14,000	1.4	No record
1978	Y-12	12	36,000	1.7	Waste oil
	ORGDP	1	1,000	1.7	Mineral oil
1979	Y-12	4	16,500	2.6	Waste oil
	ORGDP	16	63,900	2.6	Waste oil
1980	Y-12	22	72,400	2.6	Waste oil
	Y-12	5	17,800	2.6	Mineral oil
	ORGDP	7	22,300	2.6	Waste oil
1981	NA ^b	NA	NA	2.6	No disposals
1982	Y-12	9	32,900	2.6	Waste oil
TOTAL		124	430,800		

^aNR = no record.

^bNA = not applicable.

Source: Union Carbide 1984.

Waste disposed of in 1980 included 1000 gal of dye penetrant, which could act as potential artifacts in future dye-tracer studies.

The typical composition of waste oils and coolants generated at the Y-12 Plant is summarized in Appendix A. The values are considered to be representative of the oily wastes disposed of in the Oil Landfarm plots. As expected, the waste oils exhibit lower densities and higher carbon contents than the oil/water coolant mixtures. Uranium contents of the oily wastes ranged from ~20 to 80 ppm. The data further indicate that waste oils contained high levels of calcium, iron, sodium, silicon, phosphorous, chlorine, zinc, and lead. Analyses of machine coolants show a similar composition, with the exception of barium, which is present in the coolants at levels 10 to 50 times higher than in the oils.

Before 1979, waste oils and coolants were not specifically analyzed for contaminants before application to the Oil Landfarm. In 1979, analytical results for samples of oil that were collected from the surface of one ORP in the Burial Grounds and applied to the Oil Landfarm indicated the presence of PCBs. Thereafter, oils were analyzed for uranium, beryllium, thorium, and PCBs. A maximum permissible concentration level of 5 mg/L was established by the Y-12 Plant for PCBs in waste oils to be landfarmed. Oils containing >5 mg/L of PCBs were shipped for commercial disposal if uranium concentrations were less

Table 2.8. Summary of waste coolant disposals at the Y-12 Plant Oil Landfarm

Calendar year	No. of disposals	Volume (gal)	Active plots (acres)	Description
1973	12	34,800	0.9	Coolants
	8	22,500	0.9	Coolants and mop waters
	2	9,700	0.9	Water contaminated with nitrate
1974	22	104,000	1.6	Coolants
	6	21,000	1.6	Coolants and mop waters
1975	21	97,500	2.0	Coolants
1976	25	107,000	2.0	Coolants
	3	9,400	2.0	Water
1977	23	114,000	1.6	Coolants
1978	15	62,600	1.4	Coolants
1979	NA ^a	NA	1.4	No disposals
1980	1	1,000	1.4	Dye penetrant
1981	NA	NA	1.4	No disposals
1982	NA	NA	1.4	No disposals
TOTAL	138	583,500		

^aUnion Carbide 1984.

Source: Union Carbide, 1984.

than permissible limits established for this removal. If uranium concentrations were greater than the established permissible limits, waste oils containing >5 mg/L of PCBs were placed in storage for future incineration at the K-25 Site Toxic Substances Control Act (TSCA) Incinerator.

In December 1981, the analyses of wastes were expanded to include chlorinated hydrocarbons. A maximum permissible concentration of 3% was established for chlorinated solvents in the waste oils. Since 1981, sampling and analysis of waste oils for chlorinated solvents have shown that landfarmed waste oils have contained concentrations of <3% chlorinated hydrocarbons.

The Oil Landfarm HWDU was granted interim status under TN Rule 1200-1-11-.07(3) in December 1984 and is therefore subject to the groundwater monitoring requirements described in TN Rule 1200-1-11-.05. Before this date, groundwater monitoring at the Oil Landfarm was conducted in accordance with two DOE monitoring programs. Historical monitoring at the site was conducted from 1975 to 1987 as part of a program to identify groundwater contamination sources at the Y-12 Plant. Characterization monitoring began in

1983 in response to the MOU signed by DOE, EPA, and TDEC and continued until 1986. The MOU required DOE to initiate an investigation of groundwater contamination in the BCV Waste Disposal Area. Characterization monitoring was performed to evaluate the extent of contamination in soils, stream sediments, groundwater, and surface water.

Groundwater quality data collected during historical and characterization monitoring verified the presence of contaminants in groundwater at the Oil Landfarm. Therefore, interim status assessment monitoring was implemented at the site in lieu of detection monitoring, as required under TN Rule 1200-1-11-.05(6)(a)4. Assessment monitoring was initiated at the Oil Landfarm in January 1986.

TDEC acceptance of certification of final closure of the Oil Landfarm was granted in November 1990. The site was closed as an HWDU and is subject to postclosure care under TN Rule 1200-1-11-.06(7)(g) and (h). Pursuant to TN Rule 1200-1-11-.06(1)(c), if final rule administrative disposition of the Part B Permit application is made, groundwater monitoring requirements under TN Rule 1200-1-11-.06 will apply to this site.

2.3.3 Boneyard/Burnyard

The Boneyard/Burnyard consists of three sites: the Boneyard, the HCDA, and the Burnyard. The Boneyard is located east of the Oil Landfarm (Fig. 2.5) and consists of a series of unlined earthen trenches. This area was an active waste-disposal site from 1943 to 1970. The wastes have been characterized as ranging from ignitable and possibly radioactive to inert. They are known to have included organics, metals, debris, acids, and beryllium; the total quantity of material is unknown. Table 2.9 provides a chronological list of the activities at this site and the Burnyard.

Magnesium chips were disposed of in the southwest corner of the Boneyard, south of the gravel road. The metal chips were placed in burn pans in the unlined earthen trenches. Ignitable solvents were used to initiate the burn. The residue remaining in the trenches was covered with soil and compacted until the trenches were filled. The trenches then were covered with topsoil and seeded with grass. The remaining land in the Boneyard was used to dispose of construction spoil material such as concrete and rebar.

The Burnyard consisted of two trenches ~90 m (300 ft) long and 12 m (40 ft) wide. The site is no longer visible because the HCDA was built on top of it and this in turn has been capped (Fig. 2.5). The Burnyard functioned as an active waste site from 1943 to 1968. It is estimated that the site received ~90 m³ (350 yd³) per month [1000 m³ (4000 yd³) per year] of sanitary refuse from plant operations, including solids, liquids, and sludges. The waste materials may have contained empty pesticide containers, metal shavings, solvents, oils, and lab chemicals. The wastes were placed in unlined earthen trenches and burned. Oils and other flammable liquids, possibly transformer oils containing PCBs, were used to start and sustain combustion. Therefore, dioxin/furan contamination is a possibility because they are produced when PCBs are burned at relatively low temperatures. When filled, the trenches were covered with soil. No collection or treatment systems, other than burning, have been used on this site (Energy Systems 1987b).

Table 2.9. Chronological list of events at the Boneyard/Burnyard

Year	Event	Reference
1943	The Boneyard was used as a disposal site for waste from the Y-12 Plant. Burning and disposal of debris and sanitary, metallic, chemical, and radioactive wastes are known to have occurred. In the southwest corner of the Boneyard, metal chips were placed in pans in unlined earthen trenches and burned. The residue remaining in the trench was covered with topsoil and compacted until the trench was filled. The trench was then covered with topsoil and seeded with grass	Energy Systems 1991
1943	The Burnyard received $\sim 1000 \text{ m}^3$ (4000 yd^3 /year) of sanitary refuse from plant operations including solids, liquids, and sludges. The wastes were placed in two unlined earthen trenches and burned. When filled, the trenches were covered with soil	
1968	Disposal of waste at the Burnyard was discontinued	
1970	Disposal of waste at the Boneyard site was discontinued	
1975–1981	A Hazardous Chemical Disposal Area was established on ~ 2 acres in the southeastern portion of Boneyard/Burnyard on top of the Burnyard. The area was used to burn or neutralize liquid and gaseous wastes. The material was broadly characterized as ignitable, reactive, corrosive, toxic, highly flammable, or in some instances inert	Energy Systems 1991 Energy Systems 1987b
1985	An investigation conducted by Bechtel National, Inc., on sediment samples collected down stream of the Boneyard indicated levels of mercury, uranium, several polynuclear aromatic hydrocarbons, and phthalates	Energy Systems 1987b
1989	The Hazardous Chemical Disposal Area was covered with a RCRA-type cap	Energy Systems 1991

The HCDA is part of the Boneyard/Burnyard area; it is located directly southwest of the Boneyard, on top of the former Burnyard site (Fig 2.5). The site received solid, liquid, and gaseous waste materials from 1975 to 1981. It is estimated that during this period $<3.8 \text{ m}^3$ (5 yd^3) of waste per year were deposited (Turner et al. 1991). The material was broadly characterized as ignitable, reactive, corrosive, toxic, highly flammable, or in some instances inert (Energy Systems 1987b).

Generally, the HCDA received wastes that posed safety hazards within the Y-12 Plant. The material came from two sources: gas cylinders with leaking or damaged valves and lab chemicals considered to be reactive or explosive. Gas cylinders containing noncorrosive gases were allowed to leak into the atmosphere or were bled off to expedite the process. Those containing corrosive gases were bled through neutralizing slurries. Empty gas cylinders either were destroyed or transported to another location for repair. The lab chemicals included acids, bases, organics, water-reactive compounds, and explosive compounds such as picric acid,

benzoyl peroxide, and ether. Bottles of chemicals were broken under water spray in a concrete vessel that was open to the atmosphere. After the explosion or chemical reaction had taken place, the effluent was discharged into a small unlined surface impoundment and allowed to percolate through the soil. The chemical residue remaining in the concrete vessel was removed periodically and transported to the Burial Grounds (Energy Systems 1987b).

In 1989, the HCDA was covered with a RCRA-type cap (Energy Systems 1991).

2.3.4 SL 1

SL 1 is located ~1.3 km (0.8 mile) west of the Y-12 Plant just north of Bear Creek and immediately south of the Oil Landfarm as shown on Fig. 2.5. A chronological list of events at this site is included in Table 2.10. The site was used between 1968 and 1980 for the disposal of combustible and decomposable solid wastes. The landfill was permitted for materials such as paper and cardboard, plastics, rubber, wood, brush, animal bedding, organic garbage, textile products, and asphalt roofing materials. Although administrative controls were used to exclude the disposal of toxic chemicals and other contaminated materials, it is possible that some of these materials were disposed of in the landfill (Energy Systems 1987c).

Trenches were excavated to depths of ~6 m (20 ft) and backfilled to ~4.5 m (15 ft) above grade. Approximately 27,000 m³ (105,000 yd³) of refuse were disposed of at the landfill. The landfill was closed in 1985 by grading to promote drainage, capping with 0.6 m (2 ft) of clay and topsoil, and establishing a vegetation cover (Energy Systems 1987b).

2.3.5 Burial Grounds

The Burial Grounds are located ~3.2 km (2 miles) west of the Y-12 Plant (Fig. 2.6). The Burial Grounds consist of several principal waste-disposal units designated as BG-A, -B, -C, -D, -E, and -J; the Walk-in Pits; and the Uranium Vaults. Collectively, these units are referred to in this document as the Burial Grounds. Table 2.11 provides a chronological list of events that are associated with this source area.

Each waste-disposal unit consists of a series of trenches used for disposal of liquid and solid wastes. The trenches are between 4.2 and 7.5 m (14 and 25 ft) deep (Figs. 2.7 and 2.8). Appendix B gives the dimensions of the Burial Grounds trenches. In BG-A North and South, perforated standpipes were installed vertically into some trenches for liquid-waste disposal; rock and gravel were backfilled around the standpipes for support and maximum rate of drainage (Fig. 2.7).

The first trench in BG-A was excavated in August 1955 for the disposal of solid wastes. In July 1959, the Y-12 Plant began using this facility for the disposal of certain types of liquid waste. Before 1959, liquids generated at the Y-12 Plant from floor-cleaning operations (mop water) had been placed in the burial facilities at ORNL. Since that time, several types of wastes have been disposed of in the Burial Grounds; Table 2.12 lists the major types of wastes. The actual quantity and identity of materials are uncertain, and other materials may have been disposed of that are not listed in any inventory (Turner et al. 1991).

Table 2.10. Chronological list of events at SL 1

Year	Event	Reference
1968-1980	The site was used for the disposal of combustible and decomposable solid wastes. A total of ~27,000 m ³ (35,315 yd ³) of refuse was disposed of at the landfill.	Energy Systems 1987b
1982	Waste disposal at the site was terminated.	
1985	The site was closed by grading, capping with 0.6 m (2 ft) of clay, and establishing a vegetative cover.	Energy Systems 1987b

BG-A operated from 1959 to 1979, and throughout that period several million gallons of oil and mop water were disposed of by discharge either directly into oil disposal pits in BG-A North and South, discharged into standpipes in the burial trenches in BG-A South, or burned in an open tank in BG-A.

Before 1961, liquid and solid wastes were placed in unlined trenches within the Burial Grounds. In 1961, a surface tank was installed in BG-A to hold waste oils and coolants. The accumulations were burned, and liquids that did not burn were drained into adjacent trenches. An estimated 684,000 L (180,000 gal) of oils and coolants were disposed of by this method between 1961 and 1968, when the practice was terminated. Oils burned at this site probably contained PCBs; therefore, dioxin/furan contamination is a possibility.

In January 1969, a new procedure was adopted; the oils and coolants were poured into standpipes installed in the trenches. There are little data concerning specifications of these standpipes. From historic photographs, they appear to be 18-24 in. in diameter and probably extend to the base of the trenches (maximum 20 ft). Between 1970 and 1981, volatile solvents were also disposed of by pouring into rock-covered areas at the southeastern end of the BG-A South trenches. This area was later called the isolation area (Fig. 2.7). During this period, ~380,000 L (100,000 gal) of solvents were disposed of in this manner. Table 2.13 summarizes the solvents disposed of during the 12-year period. By 1976, most of the waste oils and coolants were being disposed of at the Oil Landfarm, and in 1979, oil disposal ceased in the Burial Grounds (Turner et al. 1991).

An estimated 2.3 million L (600,000 gal) of mop water were disposed of annually in BG-A between 1971 and 1978; the quantities disposed of before 1971 were not recorded. Disposal of mop water in BG-A was terminated in 1979; after that date, mop water was placed in the S-3 Ponds (Turner et al. 1991).

BG-B was opened in 1962 for the disposal of depleted uranium metal and oxides. BG-C was opened in 1962 for the disposal of beryllium, beryllium oxide, thorium, and solid waste contaminated with these materials; other materials contaminated with enriched uranium were also disposed of in BG-C. BG-D was used after 1968 for the disposal of depleted uranium metals and oxides after BG-B had reached its capacity. An area of the Burial Grounds referred to as the Walk-in Pits was used from 1966 to 1981 for the disposal of chemicals and uranium metal saw fines (Fig. 2.8). Since 1981, the Walk-in Pits have been used solely for the disposal of uranium metal saw fines (Turner et al. 1991).

Fig. 2.6. Site map for the I

LEGEND:

- CLAY CAPPED DISPOSAL UNIT
- RCRA CAPPED DISPOSAL UNIT
- UNCAPPED DISPOSAL UNIT

BEAR CREEK VALLEY OPERABLE UNIT 1 RI WORK PLAN

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Science Applications
International Corporation

REVISION	DRAWN BY:	CHKD BY	DATE:
0	D. EVANS		5-20-93
1	D. EVANS		8-27-93

XREFERENCES	PLOT FILES
93044/XREF/BASEMAP 93044/XREF/MANGAREA	93044/PLOT/325BC1-5

FIG. 2.6 93044\DWGS\325BC1-5 0
FIGURE # CAD FILE # ROT ANGLE

Table 2.11. Chronological list of events at the Burial Grounds

Year	Event	Reference
August 1955	Disposal at the Burial Grounds began with the disposal of solid waste in Unit A. The Burial Grounds consisted of waste disposal units A, B, C, D, E, and J. Each unit consisted of a series of trenches, 4.3 to 7.6 m (14 to 25 ft) deep, that were used for the disposal of liquid and solid wastes	Turner et al. 1991
July 1959	Large volumes of solvent-contaminated water and oils began being disposed of in BG-A	Turner et al. 1991
1961	A surface tank was installed in BG-A to hold waste oils and coolants; accumulations were burned. Liquids that did not burn were drained into trenches. An estimated 684,000 L (180,000 gal) of oil and coolants were disposed of by this method	
1962	BG-B was opened and used for disposal of depleted uranium metal and oxides	Turner et al. 1991
1962	BG-C was opened for the disposal of beryllium, beryllium oxide, thorium, and solid waste contaminated with these materials. Other materials contaminated with enriched uranium were also disposed of in BG-C	Turner et al. 1991
1966	The Walk-In Pits were established in the Burial Grounds. The area was used to dispose of chemicals and uranium metal saw fines	Turner et al. 1991
1968	Burning of waste oils and coolants was terminated in BG-A	
1968	BG-B reached waste disposal capacity	
1968	BG-D was opened. The area was used to replace BG-B and received depleted uranium metals and oxides waste	
1969-1971	Oil was first observed seeping from the ends of trenches in BG-A into Tributary NT-7 and from the soil surface above the oil disposal pits	
January 1969	Oil and coolants were poured into standpipes that were installed in the trenches	
1970-1981	Approximately 380,000 L (100,000 gal) of waste solvents were poured into rock piles and into waste-filled trenches in the southern part of BG-A	Turner et al. 1991
1971-1978	An estimated 2.3 million L (600,000 gal) of mop water were disposed of annually in BG-A	Turner et al. 1991
May 1971	ORP 1 was constructed with dimensions of 55 m (180 ft) long, 34 m (110 ft) wide, and 2 m (6 ft) deep	Bailey 1979
May 1972	ORP 2 was constructed on Tributary NT-6 at the northeast corner of BG-A	Bailey 1979
1974-1975	Approximately 57,000 L (15,000 gal) of oil were skimmed and removed from the surfaces of both oil retention ponds for disposal on the Oil Landfarm	

Table 2.11 (continued)

Year	Event	Reference
1975	Since 1975, no significant accumulation of oil has occurred on surface of ORP 2	
1975-1987	Historical monitoring at the site began as part of a program to identify groundwater contamination sources at Y-12 Plant	
1979	There were 68,400 L (18,000 gal) of oil removed from ORP 1	
1979	The disposal of mop water in BG-A was terminated	Turner et al. 1991
1979	Oil disposal ceased in the Burial Grounds	
1980	Trenches and disposal areas in the Burial Grounds that received hazardous wastes were designated and given interim status	
1980, 1982	A series of soil cores was collected along transects from the perimeter of ORP 1 for PCB analysis	
1981	After 1981, the Walk-in Pits were used for disposal of uranium metal saw-fines, organics, and laboratory wastes	
1983	Water and sediment samples collected from sites in NT-6 and NT-7 were analyzed for VOCs, metals, cyanide, phenols, and a number of conventional water quality parameters. Several VOCs principally including 1,1-dichloroethane, tetrachloroethylene (PCE), trans-1,2-dichloroethylene, and trichloroethylene (TCE) were found in the sediments of ORP 1 and in both water and sediments of NT-7 downstream of the pond. VOCs were not detected in a water sample collected north of ORP 1, which indicated that the VOC source differed from that of PCBs	Turner et al. 1991
September 1983	Sediment samples collected from ORP 1 and NT-7 downstream of the pond did not show elevated levels of toxic metals, cyanide, or phenols in comparison to the control site on NT-7 upstream of the Burial Grounds	
September 1983	Samples from ORP 1 were found to contain PCBs ranging from 3 to 31 mg/L	
1983, 1984	Concentrations of PCBs in NT-7 channel sediments at the confluence with Bear Creek were reported to range from 5.2 to 28.2 mg/g	McCauley 1984b
December 1984	The Burial Grounds were granted interim status and therefore became subject to the groundwater monitoring requirements described in TN Rule 1200-1-11-.05	

Table 2.11 (continued)

Year	Event	Reference
1985-1988	Extensive sediment sampling throughout the tributary system showed that PCBs were present in concentrations of several hundred milligrams per gram in surface sediments of NT-7 between ORP 1 and the visible area of oil seepage at the west end of Burial Trench A-14 but declined below 25 mg/g by a depth of 60 cm (24 in.)	
October 1985	Five cores were collected from the bottom of ORP 1 and subdivided into 15-cm (6-in.) intervals for analysis. Concentrations of PCBs in surface sediments ranged as high as 710 mg/g but declined to 3 mg/g or less at a depth of 67.5 cm (27 in.)	Herbes 1988
February 1988	RCRA Closure Plan was revised	EMD HSEAD 1988
1989	980 m ³ (3,842 yd ³) of sludge and soils was excavated from ORP 1, ORP 2, and NT-7. An engineered cap was constructed over each pond and NT-7 under an approved RCRA closure plan. A new channel was constructed for NT-7, and this tributary was rerouted ~15 m (50 ft) west of its original course	Collins 1990a
1989	BG-A and BG-C west were closed in place as a landfill and covered with an engineered cap under RCRA	Collins 1990b
November 1990	TDEC granted approval for the final closure of BG-A	
1993-1994	BG-A and WIPs covered with an engineered cap under RCRA	

BG-E and BG-J were used primarily for the disposal of uranium. Through 1983, BG-E had received a total of ~145 m³ (190 yd³) of uranium. In addition to the uranium, disposals made to BG-J through 1983 were estimated to include ~22 m³ (29 yd³) of debris and ~10 m³ (14 yd³) of inorganic salts (Turner et al. 1991).

Another type of disposal area worth mentioning is the Walk-in Pits (BG-WI1, BG-WI2, and BG-WI3). Unlike other trenches that have vertical walls and are filled by dump-truck operations, the Walk-in Pits have a sloped entrance ramp leading to the bottom of the pit. This allows disposals to be made more carefully, and, in fact, the Walk-in Pits have been used for many disposals of hazardous materials.

2.3.5.1 Burial Grounds waste inventory

An inventory of waste materials delivered to the Burial Grounds between the 1950s and 1980 was created in 1983 (Energy Systems 1983). Since 1980, the Maintenance Division has placed all disposal information in a computer data base known as the Waste Management Inventory—Manifest System.

Fig. 2.7. Locations of trenches, oil disposal pits, sta in BG-A North and BG-A

LEGEND:

- SEEP
- STANDPIPE
- CLAY CAPPED DISPOSAL UNIT
- RCRA CAPPED DISPOSAL UNIT
- UNCAPPED DISPOSAL UNIT

BEAR CREEK VALLEY OPERABLE UNIT 1 RI WORK PLAN

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XREFERENCES	PLOT FILES
93044/XREF/BASEMAP 93044/XREF/MANGAREA 93044/XREF/TRENCH	93044/PLOT/325BC1-6

FIG. 2.7 93044\DWGS\325BC1-6 0
FIGURE # CAD FILE # ROT. ANGLE

pipes, and the isolation area
outh.

Fig. 2.8. Locations of trenches in BC

LEGEND:

- SEEP
 STANDPIPE
 RCRA CAPPED DISPOSAL UNIT
 UNCAPPED DISPOSAL UNIT

SCALE: 1" = 150'

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1	D. EVANS		8-27-93

XREFERENCES	PLOT FILES
93044/XREF/BASEMAP 93044/XREF/MANGAREA 93044/XREF/TRENCH	93044/PLOT/325BC1-7

FIG. 2.8

93044\DWGS\325BC1-7

0

FIGURE #

CAD FILE #

ROT. ANGLE

Table 2.12. Partial summary of waste materials deposited in the Burial Grounds

Constituent	Burial Grounds				
	A	B	C	D	Walk-in Pits
Heavy metals ^a	P ^b	P	P	P	P
Oils and coolants	P	P ^c	P	P ^c	P
Salts	P	NI ^d	P	NI	P
Debris	P	NI	P	NI	P
Solvents	P	NI	P	NI	P
Ethylenediaminetetracetic acid (EDTA)	P ^e	NI	NI	NI	NI
Asbestos	P ^f	NI	P	NI	P
Material contaminated with radioisotopes	P	P	P	NI	P ^g
Mop water	P	NI	NI	NI	NI

^aIncludes beryllium and uranium.

^bP = listed in the records as having been deposited.

^cTrace quantities only.

^dNI = not identified in the records as having been deposited.

^eIn A-South only.

^fIn A-North only.

^gIn northern Walk-In Pit only.

Source: McCauley 1985a.

Approximately 50,000 individual disposals have been tabulated from records and entered into computer data bases. Hundreds of pages of historical text, many of them handwritten, have been typed. Roughly 15,000 Bear Creek Burial Grounds disposals previously computerized by the Maintenance Division have been accessed. Despite this work, however, the resulting picture is incomplete. Many records have been routinely discarded under approved procedures, and many items of information were never recorded at all. Additional documents may yet be uncovered. Nevertheless, the information presented here gives a fairly complete overview of the wastes contained in the Bear Creek Burial Grounds.

Disposals at the Burial Grounds between the 1950s and 1980 can be categorized according to the following three distinct disposal practices:

- **Blanket Disposals** (which account for approximately two-thirds of all wastes delivered). Operating areas in the Y-12 Plant requested that routine pickups be made of their waste by initiating a "blanket request" to the Maintenance Division. Each blanket was given a number and was for a particular kind of waste. These blankets were valid only for the period of time (months or years) that the particular kind of waste was being generated. There have been ~350 blankets active at one time or another since the early 1960s, and they accounted for about two-thirds of the traffic to the Bear Creek Burial Grounds.
- **Hazardous Disposals** (which account for a small proportion of the total wastes). Several times per day, nonroutine disposals were made of materials that were hazardous in nature—toxic, explosive, reactive, etc. These disposals were made by submitting

Table 2.13. Solvents discharged into the isolation area (in gal)

Solvent	1970	1971	1972	1973	1974	1975	1976	1977	1978	1979	1980	1981	Totals
Vythene	410		50	850	1575	210		555	500	505	1,865	1,785	8,305
Varsol	245	410	1,000	720	490	650	615	1,000	900	560	810	661	8,061
Fuel and water	450	500	700	110	15	680	1,165	2,500	400	250	790	205	7,765
Perchloroethylene	100			540	180	475	1,150	930	1,160	710	1,000	1,390	7,635
Freon	20	85	520		265	345					1,390	1,470	4,095
Alcohol	25	25	550	490	280	210	20	290	600	200	410	189	3,289
Solvents from Bldg. 9208 (benzene, toluene, xylene, alcohol)	40	265	210	125	50	50		110	125		190	5	1,170
Trichloroethylene		75	50		220				110	225		55	735
MEK			160		120				75	175			530
Miscellaneous	6,910	10,640	6,760	4,165	1,605	1,880	2,900	4,615	5,130	2,850	7,240	2,730	57,425
Gallons discharged	8,200	12,000	10,000	7,000	4,800	4,500	5,850	10,000	9,000	5,475	13,695	8,490	99,010

Source: McCauley 1984b.

UCN-2109 forms (i.e., Request for Hazardous Materials Disposal forms) to the Maintenance Division. These forms describe the hazards and handling requirements involved in the disposal, as determined by the requester. A significant portion of these hazardous disposals were made in the Bear Creek Burial Grounds.

- **Other Disposals** (which account for approximately one-third of all wastes delivered). Many one-time disposals were made of materials that pose no particular hazard. These include such things as cans, bottles, boxes, glass, discarded equipment, and many other miscellaneous items that are generated routinely during industrial operations. This type of material accounts for the bulk of nonblanket material taken to the Bear Creek Burial Grounds.

When material was taken to the Bear Creek Burial Grounds, it was segregated and disposed of in particular areas. Initially, there were only two areas: Burial Ground 1 and Burial Ground 2 (BG-1 and BG-2). As time passed, these areas became subdivided, with BG-1 becoming Burial Ground A (BG-A) and BG-2 being divided into a variety of areas, primarily B, C, D, E, and J (Fig. 2.6).

Trenches were dug in each of these areas as needed, with each trench receiving its own identification number. Thus, BG-A3E, for example, referred to the east end of trench #3 in the A area. Some areas were set aside for particular materials; these included carbon foam (BG-A8FOAM), aerosol cans (BG-A11AERCAN), asbestos (BG-A12ASBP), crucible oxides (BG-BCRUCOX and BG-CCRUCOX), depleted uranium alloys (BG-BDEPUCPD), depleted uranium (BG-BD38 and BG-CD38), and thorium (BG-BTHOR).

Disposals of waste material in the Bear Creek Burial Grounds came, not only from the Y-12 weapons program but also from ORNL, ORGDP, and outside sources as well. In addition, Y-12 wastes were disposed of in numerous places other than the Bear Creek Burial Grounds, including disposal sites at other installations.

Y-12 Plant Maintenance records contain a number of sources of data for creating the Burial Grounds waste inventory. The following sources of data were used:

- **Activity Reports** (Energy Systems 1983, Parts 2 and 3). The records that cover the longest time span are the Maintenance Division's Monthly Activity Reports. These are available from January 1957 to the present, with roughly 1 month per year missing from the files. The waste disposal information in these reports is in two parts: text (Energy Systems 1983, Part 2) and tables (Energy Systems 1983, Part 3).

The text provides a general overview of each month's operations. Prior to 1970, the text was handwritten. All available text has been entered into the Y-12 work processing system.

The tables record how many pounds and loads were delivered to specific burial trenches, with some brief information on total pounds of alloys and contaminated materials disposed of during the month. The format of the tables changed over the years, and sites other than Bear Creek were included. Some entries are illegible; care was taken to ensure copies of the tables reflect all information that could be discerned.

- **Requests for Hazardous/Classified Disposals (2109s/6538s)** (Energy Systems 1983, Parts 4 and 5). For nonroutine disposals of hazardous and/or classified materials, a Request for Hazardous Disposal form (UCN-2109) or Request for Classified Disposal form

(UCN-6538) was filled out listing special safety and security considerations for the disposal. Approximately 11,300 UCN-2109s and 1,600 UCN-6538s were found on file at various locations around Y-12, with most dating from 1960 to the present. One file folder with information dating from late 1954 to early 1955 was also found. Information from each of these forms was transcribed and entered into a computer data base. Some entries have been deduced from other available sources. Such entries are denoted by an asterisk, which appears by the entries on the data base printout. Asterisks by the printout observation number indicated a deduced disposal area. This information is presented in Energy Systems 1983, Part 4.

In order to provide a concise tabulation of what materials were disposed, an index of the 2109/6538 information has also been prepared. This index presents only the material description field of the data base sorted in alphabetical order by individual materials along with an item number that can be used to find the complete entry in the detailed report. The 2109/6538 index is presented in Energy Systems 1983, Part 5.

- **Kardex Blanket File** (Energy Systems 1983, Part 6). An early 1963 Activity Report text entry mentions the initiation of a Kardex file in which blanket disposals were recorded. Around 600 cards from this file were located, containing over 27,000 entries from the 1964–1980 time period on over 170 different blankets. Each entry consists of a blanket number, a date, and a quantity. No information on disposal sites was included on the cards, but it is known that many entries did not go to the Bear Creek Burial Grounds. For example, plutonium-contaminated materials were gathered under blanket requests and sent to ORNL for disposal. Definitions of the blanket numbers were obtained from other sources. Data from the Kardex Blanket File are in Energy Systems 1983, Part 6.
- **Tanker Logs** (Energy Systems 1983, Part 7). These records document roughly 1,800 liquid disposals on ~25 different blankets during the 1971–1980 time frame. Little, if any, overlap exists between information on the Kardex Blanket File and the Tanker Logs. Disposal site information for each disposal is included in the Tanker Log data. The Tanker Logs are computerized and presented, sorted by burial site, in Energy Systems 1983, Part 7.
- **Trip Logs** (Energy Systems 1983, Parts 8 and 9). Beginning in the 1960s, the drivers of Maintenance Division waste disposal vehicles kept logs of what materials were picked up, where they were picked up, and where they were taken. These logs record blanket, hazardous, and other disposals, not only to the Bear Creek Burial Grounds, but to many other areas as well. Approximately three of these Trip Log sheets were generated each working day. These sheets are on file for the years 1965–1979. A stratified sample of 290 days of Trip Log sheets, 8% of the total, was transcribed and entered into a computer data base. The disposals from these sheets that were delivered to the Bear Creek Burial Grounds are listed in Energy Systems 1983, Part 8.

There are several reasons why only 8% of the available Trip Logs were tabulated. The lack of standardized format made interpretation of the log entries difficult, and complete tabulation in the time available would have been effectively impossible. Most of the Bear Creek Burial Grounds disposals were already on 2109s, 6538s, and the Kardex Blanket File, all of which were to be completely tabulated. The most significant reason, however, was that the 8% sample was deemed adequate for statistically estimating what would have been found if all Trip Log sheets had been examined. Verification of this technique was provided when the statistical estimates for total uranium disposed of at the Bear Creek Burial Grounds between 1965 and 1979 equalled the amount recorded by the

ongoing SWIMS program (8,700 metric tons). (The equivalent SWIMS figure for 1954–1964, before Trip Logs were kept, was 3,400 metric tons.) Other verification efforts were also very successful. The results of this mathematical exercise are presented in Energy Systems 1983, Part 9.

- **Waste Management Inventory—Manifest System** (Energy Systems 1983, Part 10). Since 1980, the Maintenance Division has placed all disposal information in a computer data base known as the Waste Management Inventory—Manifest System. Tanker Logs, Trip Logs, and the Kardex Blanket File were discontinued at that time, although 2109s, 6538s, and Activity Reports continue to be generated and filed. The majority of this data base refers to material taken to the Y-12 Sanitary Landfill. Over 13,000 disposals to the Bear Creek Burial Grounds have been recorded. These entries, sorted in burial trench order, are detailed in Energy Systems 1983, Part 10.

From the use of these data sources an overall inventory of the Bear Creek Burial Grounds was prepared. There exists a certain amount of overlap between the various sources, and this overlap has been used to allow confirmation of some of the values. Although the inventory gives a fairly complete picture of wastes delivered to the Burial Grounds, there are two exceptions.

- No records have been found detailing the composition of disposals made in the initial trenches of Burial Ground 1. (These trenches later became Burial Ground A, trenches #1 through #5.)
- An inventory of materials placed in the B and C areas during the early 1960s comes from an engineering drawing (E-CV-62373). Data from 1960–1964 are not accounted for in any other data source. Data after 1964 probably overlap with other information reported below, but the extent of this overlap is impossible to determine.

The Maintenance Division's Activity Reports began reporting, in 1963, the number of blanket disposals made during a given month. This information has been tabulated, and a total of ~50,000 blanket disposals over the past 20 years was obtained. The blanket disposal inventory accounts for 40,000 of these blanket disposals. Thus, ~80% of the blanket disposals serviced by Maintenance over the past 20 years have been directly tabulated. Since blanket disposals account for only two-thirds of the total Bear Creek Burial Grounds disposals, the tabulated data accounts for approximately half (80% of two-thirds) of all disposals made at the Bear Creek Burial Grounds since 1963. The remaining 20% of blanket disposals serviced by Maintenance since 1963 would be on Kardex Blanket File cards that have not been found. Even though the amounts remain unknown, however, the types of waste materials these cards would document *are* known. In some cases, these blankets were found in the Trip Log Sample and are so noted.

Hazardous disposals make up a very small amount of the total material sent to the Bear Creek Burial Grounds, but this small amount is of great potential interest. The great variety of materials found through tabulation of 2109s and 6538s does not lend itself to a summary inventory such as that presented for blanket disposals. The number of 2109s found corresponds closely with the number of individual hazardous disposals mentioned in the Maintenance Division's Activity Reports, so the records presented in this area are essentially complete for the time period covered.

The estimated one-third of all Bear Creek Burial Grounds traffic that is neither blanket nor hazardous disposals is roughly equivalent to the "DEBRIS" material type in the statistical estimates presented in Part 9 of Energy Systems 1983 for the 1965-1979 time period.

Another major source of information examined was Nuclear Materials Control and Accountability waybill records. These records document the movement of accountable nuclear materials such as uranium. In general, disposals documented in this system are also recorded in the various Maintenance Division records. However, a type of waybill known as a "741" allows separate identification of material that was shipped to Y-12 from other installations for disposal since 1977.

2.3.5.2 Burial Grounds closure activities

When RCRA Part A permit applications were submitted, trenches and disposal areas in the Burial Grounds that received hazardous waste after November 19, 1980, were designated and given interim status.

A variety of wastes have been disposed of in the interim-status units in the Burial Grounds. They include the following:

- Walk-in Pits—uranium metal saw fines, organics, and unspecified laboratory wastes.
- Isolation Area—solvents including alcohol, perchloroethylene, trichloroethylene, freon, and others.
- Trenches A-7 through A-12—mostly depleted uranium-contaminated solid waste, including wood, metal, paper, glass, debris, and others; smaller quantities of liquids including oils, machine coolants, and solvents.
- Rock Pits and Standpipes—some solvents and machine coolants.
- Trenches C-18, C-20, and C-21—various types of solid wastes contaminated with beryllium, beryllium oxide, thorium, and enriched uranium, some contaminated with solvents; small quantities of acid.

BG-A and BG-C West were closed in place as a landfill and covered with an engineered cap under RCRA in 1989. At the time of this writing, a project is under way to cap BG-C East, BG-B, and the Walk-in Pits with a RCRA-style cap.

The Burial Grounds were granted interim status in December 1984 and therefore are subject to the groundwater monitoring requirements described in TN Rule 1200-1-11-.05. Prior to this date, groundwater monitoring at the site was conducted in accordance with two DOE monitoring programs. Historical monitoring at the site was conducted from 1975 until 1987 as part of a program to identify groundwater contamination sources at the Y-12 Plant. Characterization monitoring at the Burial Grounds began in 1983 in response to the MOU signed by DOE, EPA, and TDEC. The monitoring continued until 1986. Characterization at the site was performed to evaluate the extent of contamination in soils, stream sediments, groundwater, and surface water.

Groundwater quality data collected during historical and characterization monitoring verified the presence of contaminants in groundwater at the Burial Grounds. Therefore, interim status assessment monitoring was implemented at the site in lieu of detection

monitoring. TDEC granted approval in November 1990 for the final closure of BG-A. The site was closed as an HWDU and is subject to postclosure care.

2.3.5.3 ORPs

The ORPs were constructed to prevent downstream transport of oil in Tributaries NT-6 and NT-7 (Fig. 2.6). Although overflow of mop water from standpipes had occurred previously, oil was first observed seeping from the west ends of trenches in BG-A into Tributary NT-7 and from the soil surface above the oil disposal pits, between 1969 and 1971. At this time, oil addition to the BG-A South trenches was discontinued (Bailey 1979). To prevent oil transport down Tributary NT-7 to Bear Creek, ORP 1, with maximum dimensions of 55 m (180 ft) in length, 34 m (110 ft) in width, and 2 m (6 ft) in depth, was constructed at the southwest corner of BG-A in May 1971. The pond was equipped with underflow pipes to permit water discharge while retaining floating oils in the impoundment. Simultaneously, a drainage ditch was constructed about 18 to 30 m (60 to 100 ft) west of and parallel to the tributary, which rejoined Tributary NT-7 about 15 m (50 ft) south of the pond, to divert uncontaminated surface runoff around the pond. A sump was later dug at the western end of one of the trenches east of ORP 1 to collect oil, which was pumped into the pond; a pipe was also installed between the end of a second trench and the pond to permit direct oil transfer to the pond (Bailey 1979).

During the mid-1970s, oil seepage was observed from the west-facing slope west of trench A-14 into Tributary NT-7 (Bailey 1979). In 1974, ~57,000 L (15,000 gal) of oil were removed from the surface of ORP 1; an additional 19,000 L (5000 gal) were removed in early 1975. In 1979, an additional 68,000 L (18,000 gal) of oil were removed (Battelle 1986).

Oil seepage was observed on a steep slope above Tributary NT-6 at the eastern end of trenches A-14 and A-15 before 1972 (Bailey 1979), and in May 1972 a smaller impoundment, ORP 2, was constructed on Tributary NT-6 in the northeast corner of BG-A. No significant oil accumulation was reported on the surface of ORP 2 from 1975 to closure of the pond in 1989 (Battelle 1986).

Both ponds were closed and capped under an approved RCRA closure plan in 1989. During closure operations a total of 980 m³ (3,842 yd³) of sludge, sediment, and soil was excavated from ORP 1, ORP 2, and NT-7 (Collins 1990a). A total of 980 screening samples (rapid turnaround) was taken during excavation to guide removal activities. A total of 287 certification samples of remaining soils was taken, and all of the contaminated areas in ORP 1, ORP 2, and NT-7 were certified to have less than 25 ppb PCBs with the exception of the seepage area (Seep 1, Fig. 2.6) to the west of BG-A North. In this area, higher-than-anticipated concentrations of PCBs were encountered during excavation (1,200 ppb). This area was closed without achieving the 25 ppb acceptance criteria. The ponds and the section of NT-7 north of ORP 1 were then covered with an engineered multilayered cap. The portion of NT-7 below ORP 1 was capped with a clay cap at this time. A new channel was constructed for NT-7 and this tributary was rerouted ~ 15 m (50 ft) to the west of its original course. As part of this effort, a seepage collection system was installed northeast of ORP 1 to intercept seepage from BG-A. The collection system consists of a gravity drain to a currently operational pump station.

The Burial Grounds are drained by Tributaries NT-6, NT-7, and NT-8. Seepage zones from adjacent burial areas have been observed on all three tributaries (Fig. 2.6); thus, all

three may serve as sources for contaminant entry into Bear Creek and BCV groundwater. Seepage collection systems have been installed in NT-7 and NT-8 catchments to collect leachate from the Burial Grounds that seeps into surface water. Collection system A is located in Tributary NT-7 and collects an average of 2 L/min (0.5 gpm) from Seep 1 (Fig. 2.6). Collection system B, known as the Leachate Collection System (seeps 2, 3, and 4), is located in NT-8. Flow in system B is variable and may reach 76 L/min (20 gpm) in rainy weather, whereas flow in system A shows little variation.

2.4 RELEASES

Releases from the Bear Creek source areas, as discussed in preceding sections, were the result of routine operations at land disposal units. Descriptions of releases to soil, sediments, surface water, and groundwater are provided in Chap. 3.

3. CHARACTERIZATION OF ENVIRONMENTAL SETTING AND SITE CONCEPTUAL MODEL

This chapter provides detailed descriptions of the source areas that make up BCV OU 1: S-3 Ponds, SL 1, Boneyard/Burnyard, Oil Landfarm, and Burial Grounds (including ORPs 1 and 2). All of these sites are within the Bear Creek Hydrologic Regime (BCHR); a description of the hydrology of Bear Creek is also included in this chapter.

A substantial amount of sampling/analytical data (surface water, groundwater, and soil/sediment) exists for the BCHR and for the source areas in BCV OU 1. Although the bulk of the environmental monitoring data has been collected since 1982, some historical data from before 1982 are available. This chapter summarizes the environmental monitoring data relevant to the preparation of this document. Most of the data presented have been published or were summarized in previous reports (e.g., Turner et al. 1991).

Where surface water, groundwater, and soil data are discussed in subsequent sections, references are made to health and environmental assessment criteria. These include regulatory as well as guidance and advisory criteria. For example, maximum contaminant levels (MCLs) promulgated under the Safe Drinking Water Act (SDWA) represent one type of assessment criteria for use when evaluating monitoring data for surface water and groundwater. Concentrations of contaminants that are acutely or chronically toxic to aquatic life are another type of assessment criteria. Background concentrations of contaminants in an area of study represent yet another type of criteria.

Sampling and analytical methods for many environmental contaminants have been evolving rapidly over the past decade. Sampling and analytical QA/QC protocols also have become more rigorous during this period, and in some cases substantial amounts of QA/QC data exist to accompany sampling/analytical data for units in BCV. These data include results for trip blanks, field blanks, equipment rinsate blanks, process blanks, field duplicates, blind splits, matrix spikes, and standard reference materials.

Evaluation of the quality of the environmental monitoring data for BCV revealed the concordance of results among the various sampling and analytical activities that have been performed in the Bear Creek watershed. Both the nature and concentrations of COCs were similar regardless of which group performed the sampling and analyses. For example, uranium concentrations in Bear Creek surface water are reported consistently by Bechtel National, Inc. (BNI); Roy F. Weston, Inc.; the U.S. Geologic Service (USGS); the Y-12 Plant; and ORNL to occur at levels up to ~1.0 mg/L in the headwaters and to decrease to ~0.05 mg/L downstream. This discussion should not be taken to imply that no sampling or analytical problems have occurred during previous monitoring of Bear Creek and its floodplain. The formal and informal QA/QC protocols used by each investigative group have detected problems that were either corrected or the data were flagged as suspicious.

3.1 GEOGRAPHY

Descriptions of the geographic location of each unit in BCV OU 1 were presented in Chap. 2. BCV OU 1 is located between Bear Creek to the south and Pine Ridge to the

north (plant grid directions, plant grid is $\sim 34^\circ$ W of true north). Each unit is located on the south-facing slope of Pine Ridge, north of the Bear Creek main stem. Topographic relief varies from shallow near the valley floor, to steep in the highest elevations of the Burial Grounds (Fig. 3.1). Elevations within the study area range from 222 to 390 m (740 to 1300 ft) above mean sea level. Tributaries of Bear Creek have eroded V-shaped valleys in the south slope of the Pine Ridge. These valleys trend almost uniformly north-south and have generated localized relief with slopes toward the southeast and northwest.

Installation of caps on some waste sites has modified the local topography of individual units, details of which can be found in the Architect-Engineer Title III Reports for individual sites. In general, the caps have resulted in smooth topographic features: in the valley floors almost level areas of ground raised above the valley floor and on the sides of ridges, smooth mounds.

The S-3 Ponds are located at the head of Bear Creek watershed, close to the drainage divide between the UEFPC Hydrologic Regime and BCHR. The site has been capped and is now an almost level asphalt-covered parking lot located at the western end of the Y-12 Plant (Fig. 2.5).

The Oil Landfarm, SL 1, and Boneyard/Burnyard are all located on the lower slopes of Pine Ridge, near the valley floor west of the S-3 Ponds. A RCRA cap has been installed over the Oil Landfarm. A RCRA-style cap covers the HCDA, and a 0.6-m (2-ft)-thick clay cap covers SL 1. These caps are large, open, grass-covered areas that are almost level. The remainder of the Boneyard/Burnyard that is not capped is gently sloping, uneven ground.

The Burial Grounds occupy the upper and lower slopes of Pine Ridge, extending to the valley floor. RCRA caps have been installed on most of BG-A, on BG-C West, and on ORPs 1 and 2. The topographic relief at these sites is generally low, and the contours are gentle. BG-B and the Walk-in Pits, which have not been capped, occupy an elevated area between BG-A and BG-C. BG-D, BG-E, and BG-J have not been capped and occupy the steeper upper slopes of Pine Ridge above BG-C.

3.2 CLIMATE

Detailed discussions of climate can be found in Sect. 4.4 of Vol. I of the *RCRA Facility Investigation Plan General Document for the Y-12 Plant* (Welch 1989) and in Sect. 3.1 of the *Clinch River RCRA Facility Investigation Plan* (Energy Systems 1990). Oak Ridge, Tennessee, is part of the southeast climatological region of the United States. This region is characterized by a moderate continental forest climate with mild winters [coldest month above 0°C (32°F) but below 18°C (64.4°F)]. The Oak Ridge area has a temperate climate that is moderated by the Blue Ridge Mountains to the east and the Cumberland Plateau to the west.

Weather patterns in Oak Ridge are generally temperate, with warm, humid summers and cool winters. The annual mean temperature is about 15°C (58°F), with a January mean of about 3.5°C (38°F) and a July mean of about 25°C (77°F). Relative humidity in mid-afternoon averages about 55%. At night humidity is higher, averaging 85% at dawn.

Fig. 3.1. Location of sites in BC

LEGEND:

— 900 — TOPOGRAPHIC
CONTOURS

CONTOUR INTERVAL = 20 FT

SCALE: 1" = 1200'

BEAR CREEK VALLEY OPERABLE UNIT 1 RI WORK PLAN

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93044/XREF/BASEMAP	93044/PLOT/342BC1-8
93044/XREF/MANGAREA	
93044/XREF/CONTOUR1	
93044/XREF/CONTOUR2	

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Precipitation in this portion of the Tennessee Valley is seasonally distributed (McMaster 1967). The mean annual rainfall is ~136 cm (54.4 in.) (Fig. 3.2), but during the drought of 1981-88, yearly precipitation fell below this average value. Winter months usually have the most rain, with another peak in July, when thunderstorms are common. Autumn is usually the season of lowest rainfall.

Precipitation data from the Walker Branch watershed, which is located ~0.8 km (0.5 mile) south of BCV, were used to estimate rainfall in BCV during 1984. In January 1985, Energy Systems installed the Burial Grounds rain gauge to provide daily precipitation data. Comparisons between these gauges have generated confidence concerning the validity of transferring precipitation data between the two catchments.

Average loss of water to the atmosphere by evapotranspiration is about 76 cm (30.4 in.) annually in the Oak Ridge area, or about 55% of the total annual precipitation. Evapotranspiration is at a maximum from July to September, during the vegetative growing season. Runoff is greatest in the winter when evapotranspiration is at a minimum.

Prevailing winds in the area follow the general topography of the surrounding ridges (Fig. 3.3). The down-valley draft, coming from the northeast and identified with gravitational flow down local slopes and the Tennessee Valley, prevails during inversion conditions of late evening through mid-morning. Daytime up-valley flow is from the southwest when regional or synoptic flows aloft become strong enough to dominate the opposing local gravitational winds.

Oak Ridge is one of the country's calmest wind areas. Average wind speed for the Oak Ridge area is 7.1 kph (4.4 mph). The Cumberland Plateau and the Smoky Mountains divert severe storms; local irregular ridges further minimize the air movement and wind impact.

3.3 DEMOGRAPHY

3.3.1 Population Information

The Y-12 Plant is located at the northeast boundary of the ORR across Pine Ridge from the commercial center of the city of Oak Ridge, which has a population of 27,310 (U.S. Department of Commerce 1991). As of June 1993, the Y-12 Plant employed 6190 full-time Energy Systems employees and ~1000 on-site employees of M-K Ferguson, DOE, and ORNL. A pumping station located on the north bank of Melton Hill Lake at Clinch River supplies water to the Y-12 Plant water treatment facility and the city of Oak Ridge water treatment plant. A pumping station for domestic water is also located near mile 14.5 on the Clinch River, also within the ORR boundary.

3.3.2 Land Use

The East Tennessee Development District, which includes the city of Oak Ridge, is rural in character. Knoxville is the district's regional center and the only city in the region with a population in excess of 50,000. Oak Ridge and other cities in the district are within the 7500 to 50,000 population range, are incorporated, and have a definite central core that provides major employment and trade opportunities. Of the district's 1,700,010 ha (4,200,800 acres), ~80% is in agricultural and forest land use (Table 3.1).

Fig. 3.2. Annual precipitation record for the Oak Ridge area (*Source: Energy Systems 1991*).

Fig. 3.3. 1990 wind rose for 10-m level at the west end of the Y-12 Plant (Source: Energy Systems 1991).

Table 3.1. Land use data for the Eastern Tennessee Development District

Land use category	Hectares	Acres	Percent
Residential	35,074	86,670	2.1
Commercial	1,950	4,820	0.1
Industrial	20,700	51,150	1.2
Recreational	269,129	665,030	15.8
Agricultural	685,904	1,694,900	40.3
Public and quasi-public	28,430	70,250	1.7
Forest	658,823	1,627,980	38.8
Total	1,700,010	4,200,800	100.0

Source: DOE 1982.

In contrast to the district's general land use patterns, the city of Oak Ridge is an urban center with minimal agricultural activities. Land not owned by the federal government consists of 5510 ha (13,615 acres) and is divided into more than 9500 parcels. Table 3.2 categorizes urban land use data for the city of Oak Ridge. Most of this land is either residential or vacant, and more than 1903 ha (4700 acres) of the vacant land is suitable for housing (DOE 1982).

Table 3.2. Urban land use data for the city of Oak Ridge, Tennessee^a

Land use category	Hectares	Acres	Percent
Residential	1,640	4,050	28.9
Vacant	2,247	5,550	39.6
Recreational	749	1,850	13.2
Commercial	104	256	1.8
Industrial	47	115	0.8
Public transportation, parking	480	1,185	8.5
Private transportation	33	82	0.6
Utilities and communications	52	129	0.9
Services	314	775	5.5
Total	5,666	13,992	99.9

^aExcludes federally owned lands within the corporate city limits.

Source: DOE 1982.

Several categories of land use are in close proximity to the ORR. The nearest privately owned residential properties are in the Poplar Springs, Sugar Grove Valley, Oak Ridge, and Scarboro communities. Oak Ridge and Scarboro are located on the northern boundary of the ORR. Poplar Springs is located 3.2 km (2 miles) west across the Clinch River, and Sugar Grove Valley is located 2.4 km (1.5 miles) north. Other residential areas include Bradbury, Edgewood, and Lawnville. Bradbury is located across the Clinch River ~8 km (5 miles) south. Edgewood and Lawnville are located immediately west-northwest of the Poplar Springs community. Agricultural use includes limited-scale private gardening; raising of tobacco, corn, wheat, and soybeans as cash crops; raising of beef cattle; and dairy farming. Some areas are also used for commercial logging.

Industrial land use includes Phyton Technologies, Inc.; Scientific Ecology Group (SEG); and IT Corporation's Bear Creek Radiological Laboratory. SEG, Phyton, and the IT Corporation laboratory are within 3.2 km (2 miles) of the Y-12 Plant. TVA facilities, including the Melton Hill Dam, the Bull Run Steam Plant, and the Kingston Steam Plant, are all more than 9.6 km (6 miles) from the Y-12 Plant.

Recreational areas near the ORR include the city of Oak Ridge parks (LaSalle, Cedar Hill, Al Bissell, Elm Grove, Milt Dickens, and Scarboro parks); the city of Oak Ridge swimming pools and sporting facilities; South Hills Golf Course; Watts Bar Lake Embayment/Clinch River waterway, which is used as a recreational area by both pleasure boaters and fishermen; a number of small camping areas and boat launching ramps; a small dirt-surface racetrack located ~6.4 km (4 miles) south, which attracts several thousand spectators during the racing season; and a public swimming area 11.3 km (7 miles) southeast at Melton Hill Dam. There are no public recreational facilities, except the previously mentioned boat launching ramp, within 3.2 km (2 miles) of the ORR. Sport hunting of game birds and game animals occurs seasonally in the region, and deer hunting is authorized on some parts of the ORR as a conservation measure.

The ORR lies within the corporate limits of the city of Oak Ridge. A buffer zone surrounds each of the facilities to provide security, space for expansion, and isolation from the general public. Tracts totaling ~5666 ha (14,000 acres) are allocated around operating reactors, waste disposal areas, and streams that receive routine waste releases and burial ground seepages. Acreage used for high-voltage transmission lines, pipelines, transportation corridors, and security fences amounts to ~2023 ha (5000 acres) (DOE 1982).

About 80% of the ORR is part of a comprehensive forest management program that divides the reservation into 27 compartments. These compartments range in size from 148 to 486 ha (365 to 1200 acres). Vegetational features unique to the area are excluded from timber harvest operations, and 40 individual sites are designated for environmental research.

The Y-12 Plant is situated at the eastern end of the ORR, adjacent to the commercial center of Oak Ridge and contains ~324 ha (800 acres). Land uses associated with the Y-12 Plant facility include several waste storage or disposal areas, of which a parcel of ~61 ha (150 acres) serves as a sanitary waste landfill for solid wastes; ~26 ha (65 acres) located west of the main plant within the Bear Creek watershed is used as a burial ground for low-level radioactively contaminated solid wastes generated by the Y-12 Plant; and ~1.6 ha (4 acres) is an abandoned quarry outside the Bear Creek watershed that was formerly used as a disposal and dilution basin for selected nonradioactive chemical wastes, primarily sodium, lithium, and potassium from Y-12 Plant operations.

3.4 GEOLOGY AND SOILS

3.4.1 General Geology

The DOE ORR, which includes BCV OU 1, is located in East Tennessee within the Valley and Ridge Physiographic Province, part of the Appalachian fold and thrust belt (Fig. 3.4). The area is characterized by a succession of east-west trending ridges and valleys, which formed as a result of differential erosion of the southeast dipping (locally 45° to vertical) clastic and carbonate lithologies that make up the individual thrust sheets.

BCV is bounded to the north and south (plant grid directions, plant grid is ~34° W of true north) by Pine Ridge and Chestnut Ridge, respectively. Maximum relief of the ridges is ~91 m (300 ft). These ridges and valleys are part of the White Oak Mountain (WOM) thrust sheet, which exposes Lower Cambrian to Ordovician rock units in this area (Fig. 3.4). The stratigraphy of the area, from oldest to youngest (and exposed from grid-north to grid-south), includes the Lower Cambrian Rome Formation (exposed on Pine Ridge), the Cambrian Conasauga Group (exposed in BCV), and the Knox Group (exposed on Chestnut Ridge). The Conasauga Group is further divided into six formations of alternating shale and carbonate-rich lithologies. From oldest to youngest these are the Pumpkin Valley Shale, the Friendship Formation, the Rogersville Shale, the Dismal Gap Formation, the Nolichucky Shale, and the Maynardville Limestone (Table 3.3).

BCV OU 1 is underlain by Cambrian limestones, shales, and siltstones of the Conasauga Group. Sandy shales and siltstones of the Rome Formation underlie Pine Ridge. Chestnut Ridge is formed on siliceous dolostones of the Knox Group. Bedrock throughout the valley is overlain by unconsolidated deposits of varying thickness and composition. Geologic exposure is generally poor, and the locations of geologic contacts in BCV have been inferred for the most part from borehole data (Fig. 3.5) (HSW 1992). The S-3 Site and the Oil Landfarm are located entirely on the Nolichucky Shale; the Boneyard/Burnyard and the SL 1 are located partially on the Nolichucky Shale and partially on the Maynardville Limestone. The southern section of the Burial Grounds is located on Nolichucky Shale, and the remainder of the Burial Grounds overlies the Maryville Limestone, the Rogersville Shale, the Friendship Formation, and the Pumpkin Valley Shale.

At the immediate headwaters of Bear Creek south of the S-3 Ponds, the creek bed lies in the upper Nolichucky Shale. Within 122 m (400 ft) to the southwest, the creek crosses the contact of the Nolichucky Shale–Maynardville Limestone and remains in the Maynardville Limestone along much of BCV, flowing roughly parallel to regional geologic strike (N55E).

3.4.1.1 Stratigraphy

The strata of greatest interest with respect to contaminant migration in groundwater are the formations of the Conasauga Group (Fig. 3.5), in particular the Maynardville Limestone Formation, which is considered to be part of the Knox aquifer (Solomon et al. 1992). The overlying unconsolidated material, resting on the weathered bedrock surface, is in continuity with this aquifer. For general descriptions of the stratigraphy of other units in the ORR, refer to Hatcher et al. (1992) for an overview of the ORR geology; to Haase, Walls, and Farmer (1985) for the Conasauga Group and Rome Formation; to Lee and Ketelle (1987) for the Knox Group; and to Lee and Ketelle (1988) for the Chickamauga Group.

LEGEND

- Rockwood Formation Sr
- Sequatchie Formation Os
- Reedsville Shale Or
- Chickamauga Limestone Och
- Knox Group OCk
- Fort Payne Chert Mfp

- Conasauga Group Cc
- Rome Formation Cr
- Chattanooga Shale MDc
- Fault Zone
- Syncline

02-082593-099 3.4

Fig. 3.4. Bedrock geology for DOE ORR.

Table 3.3. Stratigraphy of the Y-12 Plant area

Age	Group or formation	Thickness ^a (m)	Lithology and outcrop	Aquifer/ aquitard
Upper Cambrian- Lower Ordovician	Knox Group	530-880	Massive dolomite, siliceous dolomite, bedded chert, limestone, some clastics; outcrop on Chestnut Ridge	Knox Aquifer
Middle to Upper Cambrian	Conasauga Group (outcrops in the center of BCV and on the South slope of Pine Ridge)			
	Maynardville Limestone	125-145	Dolomitic limestone, limestone	Part of the Knox Aquifer
	Nolichucky Shale	100-150	Shale, limestone	Aquitard
	Dismal Gap Formation ^b	95-120	Limestone, shale	Aquitard
	Rogersville Shale	20-35	Mudstone, shale	Aquitard
	Friendship Formation ^c	30-40	Calcareous shale, limestone	Aquitard
	Pumpkin Valley Shale	90-100	Shale, siltstone, occasional sandstone	Aquitard
Lower Cambrian	Rome Formation	90-125	Shale, siltstone, sandstone, local dolomite lenses; outcrops on Pine Ridge	Aquitard

^aOnly thicknesses of formations at the Y-12 Plant are given.

^bFormerly the Maryville Limestone.

^cFormerly the Rutledge Limestone.

Source: Hatcher et. al. 1992.

The Conasauga Group conformably overlies the Rome Formation, a sequence of interbedded sandstones, siltstones, and shales of variegated olive, maroon, and drab colors that form Pine Ridge. On the ORR, the Conasauga Group is characterized by an interfingering of carbonate and clastic facies. In the vicinity of BCV OU 1, the Conasauga Group varies in thickness between 523 and 576 m (1717 and 1890 ft). The following discussion provides descriptions for each of the six formations present in the Conasauga Group in BCV (Fig. 3.5).

Pumpkin Valley Shale. Oldest of the Conasauga Group formations, the Pumpkin Valley Shale crops out along the southeastern edges of ridges formed on the Rome Formation. Predominantly an olive and maroon, massive clay shale with thin beds of siltstone and sandstone, the Pumpkin Valley Shale is gradational with the underlying Rome Formation but lacks the sandstone that is characteristic of the Rome Formation. Two members of the Pumpkin Valley Shale have been identified (Law 1975), the upper one being more shale rich than the lower one. The lower member exhibits units that are commonly bioturbated or have wavy parallel stratification. The Pumpkin Valley Shale is a very tight formation that is considered one of the poorest aquifers in East Tennessee (DeBuchananne and Richardson

Fig. 3.5. Bedrock geol.

LEGEND:

- APPROXIMATE GEOLOGIC CONTACT
- O&k KNOX GROUP
- €mn MAYNARDVILLE LIMESTONE
- €n NOLICHUCKY SHALE
- €m MARYVILLE LIMESTONE
- €rg ROGERSVILLE SHALE

**BEAR CREEK VALLEY
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1	D. EVANS		8-27-93

XREFERENCES	PLOT FILES
93044/XREF/BASEMAP	93044/PLOT/342BC1-4
93044/XREF/MAINGAREA	
93044/XREF/GEOLGY	

FIGURE # 93044\DWGS\342BC1-4 CAD FILE # 0 ROT. ANGLE

1956). In the vicinity of BCV OU 1, the formation is ~ 79 to 101 m (260 to 330 ft) thick and weathers to form a clay-rich soil with an abundance of shale chips. Soil color varies from reds and violets to green (Hatcher et al. 1992).

Friendship Formation. (Formerly referred to as the Rutledge Limestone.) In the vicinity of BCV OU 1, the Friendship Formation ranges in thickness from 30.5 to 35 m (100 to 115 ft). The formation consists of light grey to white, medium to thinly bedded limestones and micritic limestones, which are interbedded with mudstones and shales that become more calcareous higher in the succession. A persistent, 1.5- to 3-m (5- to 10-ft)-thick bed of maroon to maroon-grey mudstone that occurs near the base of this formation serves as a marker bed within the lower Conasauga Group throughout BCV. Increased amounts of limestone distinguish the Friendship Formation from the overlying and underlying formations. Shale locally represents 75% of the Friendship Formation, which weathers to form a yellow-brown to yellow-orange soil.

Rogersville Shale. In the vicinity of BCV OU 1, the Rogersville Shale is ~ 31 m (100 ft) thick. This formation is comprised of massive to medium-bedded, noncalcareous mudstones, interbedded with brown shales. The shales and mudstones contain subordinate units of glauconitic and/or calcareous siltstones. The rocks of the Rogersville Shale weather to a thin light grey to light red, clayey soil containing shale chips (Hatcher et al. 1992).

Dismal Gap Formation. (Formerly referred to as the Maryville Limestone.) In the vicinity of BCV OU 1, the Dismal Gap Formation ranges in thickness from 116 to 136 m (382 to 445 ft). The formation consists of light to dark grey, fine to coarsely crystalline limestone interbedded with calcareous shales and shaley-siltstones. The Dismal Gap Formation is informally divided into two members (Haase et al. 1985). The lower member consists of medium to thinly bedded calcareous mudstone interbedded with peloidal or oolitic calcareous siltstones. Limestones appear as discrete beds in coarsening upward cycles.

The upper member of the Dismal Gap Formation is characterized by the presence of distinctive flat limestone-pebble conglomerates. In places, the intraclastic limestones are glauconitic or oolite-rich. Interbedded with the clastic limestones are siltstone, mudstone, and shale.

Weathering of the Dismal Gap Formation produces a dark red to reddish-brown clayey soil.

Nolichucky Shale. (Foreman et al. 1991; Rothschild et al. 1984.) In the vicinity of BCV OU 1, the Nolichucky Shale ranges in thickness from 161 to 168 m (528 to 550 ft) (King and Haase 1987). Shale interbedded with intraclastic limestone, thick fossiliferous limestone, and oolitic limestone are characteristic of the Nolichucky Shale in the WOM thrust sheet. Shale is the dominant lithology with an approximate shale to limestone ratio of 1:1.75. The Nolichucky Shale Formation can be subdivided into four zones based on dominant lithologies: (1) thrombotic limestone overlying oolitic limestone and lime mudstone in the uppermost Nolichucky; (2) laminated peloidal packstone, lime mudstone, and shale as the most abundant lithologies toward the top of the Nolichucky Shale; (3) allochthonous oolitic and skeletal packstone and grainstone interbedded with shale as the dominant lithologies in the middle Nolichucky Shale; and (4) intraclastic limestone interbedded with shale and calcareous siltstone as the dominant lithologies in the lower Nolichucky Shale (Foreman et al. 1991). The formation is massive to very thinly bedded and is characterized by a maroon-brown color. The

interbedded limestones typically contain limestone-pebble conglomerates and oolite-rich beds similar to those occurring in the underlying Dismal Gap Formation. Throughout much of the Nolichucky Shale, mudstone/shale and limestone lithologies alternate on a scale of 0.3 to 0.9 m (1 to 3 ft), giving the formation a thickly bedded appearance.

During drilling at the Y-12 Plant, solution cavities, opened fractures, and iron precipitate were commonly observed at downhole depths of up to 15 m (50 ft) in the Nolichucky Shale (Rothschild et al. 1984). A comparison of Maynardville Limestone and Nolichucky Shale descriptions suggests the solution cavity development is not as extensive in the Nolichucky Shale as it is in the Maynardville Limestone.

Upon weathering, the Nolichucky Shale develops a brown to yellow color, clay-rich soil containing shale chips.

Maynardville Limestone. (Shevenell et al. 1992) The Maynardville Limestone is the youngest formation of the Conasauga Group and consists primarily of light-gray to tan, massive to thinly bedded limestone with subordinate amounts of dolostone. Data obtained from the ongoing Maynardville Exit Pathway Monitoring Program demonstrate that on the ORR, the Maynardville is more uniformly dolomitic and exhibits more subtle vertical lithologic differentiation than that described in the type section. In BCV, the Maynardville Formation varies in stratigraphic thickness from 70 to 136 m (230 to 445 ft). The Exit Pathway Monitoring Program subdivided the Maynardville into six different informal lithologic zones and identified a seventh zone to mark the transition between the Maynardville Limestone and the overlying Copper Ridge Dolomite. The zones were divided based on the gamma log signatures and associated lithological characteristics of the individual zones.

The Maynardville Limestone occurs on the northern flanks of Chestnut Ridge and extends to the BCV axis. Much of the groundwater flow through the Maynardville Limestone is controlled by solutionally enlarged fractures and joint sets. Three major joint sets have been identified within the Maynardville Limestone: (1) approximately parallel to geologic strike and along bedding planes; (2) approximately perpendicular to geologic strike N10W and dips steeply; and (3) parallel to strike and perpendicular to bedding planes (Rothschild et al. 1984; Dreier et al. 1987). Dissolution along joints, fractures, and bedding planes in the Maynardville Limestone has formed solutionally enlarged zones, and in some cases, relatively large cavities that have greatly increased the overall permeability of the unit. The larger solution cavities generally occur in the upper part of the saturated zone. A combination of factors, such as structure, depth, and lithology, affect the fracture and water level distribution in the Maynardville Limestone. In general, however, the greatest likelihood of intersecting a cavity in the Maynardville Limestone occurs within Zone 6, whereas Zone 3 appears to be the least likely to contain cavities. Most of the cavities in Zones 6 through 2 were located at depths of <30.5 m (100 ft), suggesting that the active karst system may be relatively shallow in the Maynardville Limestone.

3.4.1.2 Unconsolidated materials

Unconsolidated material overlying bedrock in the BCV consists of weathered bedrock (referred to as residuum), man-made fill, alluvium, and colluvium (Petrich et al. 1984). Residuum comprises a majority of the unconsolidated material in this area. The depths to unweathered bedrock differ throughout the valley because of the different thicknesses of fill and alluvium and the particular weathering characteristics of the bedrock units (Geraghty and

Miller 1985). The total thickness of these materials in the BCV typically ranges from 3 to 15 m (10 to 50 ft) (Hoos and Bailey 1986). Figure 3.6 shows the distribution of different unconsolidated material types in BCV. Table 3.4 contains a summary of current soil background data taken from Phase I of the ORR Background Soil Characterization Project annual report (Energy Systems 1993b). Data from Phase II of this study will be available for the RI.

Much of the fill material consists of asphalt, cement blocks, wire, wood, plastic, excavated weathered shale, limestone, and chert intermixed with clay. In many places, a relict soil horizon has been noted below the fill. Rothschild et al. (1984) gives the following description of fill material in the subsurface at the Y-12 Plant: "Color ranges from grey to red brown to yellow brown and dark brown. Generally, it contains angular gravel-sized clasts of dolomite and weathered shale up to 2 in. across mixed with clay. In some cores, pieces of asphalt, wood roots, and wire are found in the fill. The percentage of dolomite and shale clasts varies greatly and ranges up to ~90% in thin (less than 4 in.) zones. The shale clasts are often stained by manganese oxide and exhibit bedding structure. In some cores, the upper part of the fill is composed dominantly of dolomite clasts, and is highly porous and very unconsolidated. Compactness and consolidation of the fill increases with depth."

Alluvium is present in Bear Creek and its tributaries along present and former stream courses. The alluvium typically is an organic clayey silt with minor amounts of sandy clay, chert, and limestone pebbles. A thin layer of colluvium has been observed at the base of Pine Ridge and Chestnut Ridge in some areas. The colluvium is predominantly a clayey silt containing weathered rock fragments.

Residuum overlies bedrock throughout the valley except in scattered outcrop areas and ranges from silty to sandy clay overlying shale units to a slightly sandy clay overlying limestone units; predominant colors are shades of brown, orange, and grey. With increasing depth, the colors darken and the clay grades to weathered rock that has retained its structural characteristics (saprolite). Bedding planes and joint surfaces in the weathered bedrock commonly are marked by dark reddish-brown and yellow-brown oxide coloration, indicative of a high degree of weathering by circulating groundwater (Rothschild et al. 1984).

Soils of particular importance to BCV OU 1 are those that characteristically develop on the Conasauga Group. These soils include, but are not limited to, the Armuchee, Fullerton, Greendale, Hamblen, Jefferson, Leadvale, Newark, and Sequoia series. Depth to bedrock in the vicinity of BCV OU 1 typically varies from <3 m to 9 m (10 to 30 ft), with the Montevallo series being the most shallow soil type (Lietzke, Lee, and Lambert 1988). In general, these soils are classed as poor for crop growth or pasture. With the exception of a few areas that are suitable for pasture, the general classification of Y-12 Plant soils is for forestry. In the vicinity of BCV OU 1, the Armuchee and Sequoia soil series are most prevalent. The Armuchee is a shaley, silty clay loam becoming more shaley with depth, and the Sequoia is a yellowish-brown silt loam overlying a residuum of acid shale. Chemical and physical properties of these soils are given in Tables 3.5 and 3.6 (Welch 1989).

Fig. 3.6. Soils map for

LEGEND:

- UPPER SLOPE SOILS
- LOWER SLOPE SOILS
- FLOODPLAIN SOILS
- FILL MATERIAL

SCALE: 1" = 1100'

**BEAR CREEK VALLEY
OPERABLE UNIT 1
RI WORK PLAN**

Science Applications
International Corporation

REVISION	DRAWN BY:	CHKD BY:	DATE
0	D. EVANS		5-20-93
1	D. EVANS		8-27-93

XREFERENCES	PLOT FILES
93044/XREF/BASEMAP 93044/XREF/MANGAREA 93044/XREF/FLPLSOIL	93044/PLOT/354CH3-6

FIGURE # 93044\DWGS\354CH3-6 CAD FILE # 0 ROT ANGLE

Table 3.4. Oak Ridge Reservation background soil analytes evaluated quantitatively and qualitatively^a

Analyte	Frequency of detection	Minimum detected concentration	Maximum detected concentration	Median detected concentration	Upper 95% confidence bound on the median
<i>Dismal Gap—quantitatively</i>					
Inorganics (mg/kg)					
Arsenic	4/4	5.300	7.300	6.23910	10.3490
Barium	4/4	77.20	212.0	99.0600	124.690
Beryllium	4/4	0.550	2.200	0.78063	0.99871
Boron	1/3	16.40	21.10	13.5410	23.3130
Chromium VI	4/4	19.40	32.10	24.6530	54.3050
Cyanide	1/3	0.440	0.440	0.13003	0.28146
Manganese	4/4	768.0	2220	997.070	1347.50
Mercury	4/4	0.230	0.400	0.31591	0.34211
Mercury (salts)	4/4	0.230	0.400	0.31591	0.34211
Nickel	4/4	19.50	56.70	23.4810	27.9970
Nickel (salts)	4/4	19.50	56.70	23.4810	27.9970
Strontium	3/3	6.100	16.80	7.93480	10.2180
Vanadium	4/4	27.90	54.00	34.1560	36.8490
Zinc	4/4	42.30	108.0	50.6480	58.1260
Radionuclides (pCi/g)					
Cesium-137	4/4	0.0210	0.900	0.59825	2.47050
Europium-155	4/4	0.0840	0.110	0.09480	0.10530
Potassium-40	4/4	14.000	22.00	16.3465	20.3483
Radium-226	4/4	0.7000	0.860	0.78646	0.96091
Thorium-228	4/4	0.5000	0.940	0.71322	0.86202
Thorium-230	4/4	0.3100	0.830	0.56516	0.71725
Thorium-232	4/4	0.4100	0.970	0.68310	0.81335
Thorium-234	4/4	1.5000	1.900	1.63470	1.77100
Uranium-233/234	4/4	0.6100	1.400	0.93725	1.08890
Uranium-235	4/4	0.0569	0.120	0.07919	0.09499
Uranium-238	4/4	0.7500	1.700	1.02470	1.1549

Table 3.4 (continued)

Analyte	Frequency of detection	Minimum detected concentration	Maximum detected concentration	Median detected concentration	Upper 95% confidence bound on the median
<i>Dismal Gap—qualitatively</i>					
Inorganics (mg/kg)					
Aluminum	4/4	16900	44300	20656	23311
Calcium	2/4	991	1860	907.04	1320.6
Chromium	4/4	19.4	32.10	24.653	54.305
Cobalt	4/4	11.3	36.7	14.523	19.363
Copper	4/4	12.4	30.1	16.137	18.339
Iron	4/4	23800	49000	29438	32946
Lead	4/4	14.6	35.4	20.306	47.993
Lithium	3/3	12.7	27.0	16.165	18.753
Magnesium	4/4	2090	7430	2845.9	3505.9
Potassium	4/4	1890	5390	2300.3	2815.1
Silicon	4/4	461	697	505.90	563.01
Sulfate	3/3	28.0	163	86.736	110.92
Thallium	1/4	0.790	0.790	0.16355	0.60786
<i>Nolichucky—quantitatively</i>					
Inorganics (mg/kg)					
Antimony	1/4	0.490	0.490	0.46320	0.48482
Arsenic	3/4	5.800	6.400	3.08630	5.14460
Barium	4/4	59.70	106.0	75.3620	94.8610
Beryllium	4/4	0.730	0.850	0.78616	1.00580
Chromium VI	3/4	26.40	29.90	9.37800	20.8120
Manganese	4/4	405.0	935.0	653.480	883.140
Mercury	4/4	0.180	0.190	0.18493	0.20027
Mercury (salts)	4/4	0.180	0.190	0.18493	0.20027
Nickel	4/4	15.20	20.00	17.2860	20.6100
Nickel (salts)	4/4	15.20	20.00	17.2860	20.6100
Selenium	3/4	0.560	0.740	0.56611	0.71345
Strontium	4/4	3.200	6.100	4.5520	5.66690
Vanadium	4/4	29.40	35.20	32.4070	34.9620
Zinc	4/4	33.90	40.70	37.8920	43.4870

Table 3.4 (continued)

Analyte	Frequency of detection	Minimum detected concentration	Maximum detected concentration	Median detected concentration	Upper 95% confidence bound on the median
Radionuclides (pCi/g)					
Cesium-137	4/4	0.38000	0.7100	0.52689	2.17570
Curium-247	2/4	0.00530	0.0070	0.00552	0.00649
Europium-155	4/4	0.08400	0.0980	0.08955	0.09947
Potassium-40	4/4	14.0000	17.000	15.1952	18.9151
Radium-226	4/4	0.39000	1.4000	0.74007	0.90422
Thorium-228	4/4	1.20000	2.2000	1.50820	1.82290
Thorium-230	4/4	0.85000	1.2000	0.96648	1.22660
Thorium-232	4/4	1.20000	2.0000	1.49480	1.77980
Thorium-234	4/4	1.30000	1.5000	1.42250	1.54120
Uranium-235	4/4	0.04320	0.0969	0.07127	0.08549
<i>Nolichucky—qualitatively</i>					
Inorganics (mg/kg)					
Aluminum	4/4	20800	25100	22189	25041
Calcium	2/4	498	952	490.00	734.28
Chromium	3/4	26.40	29.90	9.3780	20.812
Cobalt	4/4	11.1	17.5	14.446	19.262
Copper	4/4	11.0	12.7	11.701	13.298
Iron	4/4	23000	32100	27901	31226
Lead	3/4	15.3	20.4	5.5086	13.131
Lithium	4/4	7.60	15.5	10.926	12.426
Magnesium	4/4	1730	2410	2008.0	2473.7
Potassium	4/4	2640	3230	2949.8	3610.1
Silicon	4/4	185	328	244.69	272.31
Sulfate	4/4	14.1	25.4	18.657	23.085

^aSource: Energy Systems 1993b.

Table 3.5. Soil chemical properties^a

Soil group	Corrosivity		pH range	Estimated cation exchange capacity (CEC) ^b	
	Steel	Concrete		Surface	Subsoil
Residuum					
Rome	low	mod-high	4.5-6.0	5-15	10-20
Conasauga	mod	mod	4.5-6.0	10-15	15-20
Knox	high	mod-high	4.5-5.5	5-10	8-20
Colluvium					
Rome/Conasauga	low-mod	mod-high	4.5-6.0	5-10	5-10
Knox	mod	mod	4.5-5.5	5-10	5-10
Alluvium					
Holocene/Modern	mod-high	low-mod	4.5-7.8	5-10	5-10
Pleistocene	mod	mod	4.5-5.5	5-10	5-10

^aData from Lietzke et al. 1988.^bMilliequivalents per 100 g of soil.

3.4.1.3 Structural geology

The Oak Ridge area is underlain by two major northeast/southwest trending thrust faults that dip to southeast and define thrust sheets (Hatcher et. al. 1992) (Fig. 3.7). Pine Ridge, BCV, and Chestnut Ridge all are part of the WOM thrust sheet, which is soled by the WOM thrust fault. The WOM fault is a regional thrust fault of the Valley and Ridge, which demonstrates at least several kilometers of translation. The fault formed during the Permian-Pennsylvanian age Alleghanian Orogeny and has not been historically active. At the ORR, the fault trends parallel to regional strike (N55E) and dips steeply (45°) to the southeast (King and Haase 1987). Bedding plane dip values measured in outcrops cluster around 45° but may steepen to vertical as a result of localized small-scale folding or faulting. Elevations within the study area range from 222 to 390 m (740 to 1300 ft) above mean sea levels.

The WOM thrust fault north of Pine Ridge is very complex and is characterized by a sequence of cross-cutting imbricate fault splays that repeatedly stack the Rome Formation. In the vicinity of Bear Creek, this fault superposes the Cambrian Rome Formation over the younger Cambrian Copper Ridge Dolomite member of the Knox Group. In addition, north of the WOM thrust fault, in the underlying Kingston thrust sheet, Bear Creek flows over slices of the Knox Group and Chickamauga Group that have been complexly stacked and rotated as a result of footwall deformation associated with displacement along the WOM

Table 3.6. Soil physical and engineering properties of the predominant soil groups in the Y-12 Plant area^a
(Revised March 10, 1989)

Surface						Subsoil				
Group	Shrink-swell	LL ^b (%)	PI ^c	Unified class	K factor ^d	Shrink-swell	LL (%)	PI	Unified class	K factor
Residuum										
Rome	low	20-30	NP-8	ML,CL	0.24	low	15-40	NP-13	GM, SM	0.28
Conasauga	low	20-40	5-15	ML,CL	0.32	mod	45-75	17-40	MH,ML,CL,CH	0.37
Knox	low	15-30	3-15	ML,CL	0.28	mod	40-70	20-40	MH,ML	0.24
Colluvium										
Rome/ Conasauga	low	20-35	2-10	SM,SC	0.28	low	20-35	2-10	GM,SM,ML	0.28
Knox	low	20-30	NP-10	ML,CL	0.34	low	20-30	5-15	CL,GC	0.28
Alluvium										
Holocene/ Modern	low	<30	NP-10	ML,CL	0.32	low	20-40	4-20	ML,CL	0.43
Pleistocene	low	20-30	3-10	ML,CL	0.37	low	40-70	10-25	CL,ML	0.32

^aAdapted from Lietzke et al. 1988.

^bLL = liquid point.

^cPI = plastic index.

^dK = soil erodibility factor as defined by the Universal Soil Loss equation (see Sect. 3.4.3).

Source: Welch 1989.

Fig. 3.7. Geologic cross-section based on seismic line on Tennessee 95 (Source: Hatcher et al. 1992).

fault. Because of the complex deformation associated with the WOM thrust fault and the immediately underlying Kingston sheet, it is anticipated that this region contains numerous fault-related fracture zones forming areas of enhanced permeability.

Bear Creek is characterized by numerous cross-strike streams that have a consistent north/south trend and are fairly evenly spaced along the valley, usually between 366 to 610 m (1200 to 2000 ft) apart. Such a consistent orientation suggests that the location of these creek beds is controlled by a geological structural feature, such as a minor tear fault or a prominent fracture trend. Both features are possible, and in Melton Valley, the White Oak Creek Tear Fault shows a similar north trend. If these structures are faults, they show minor displacement but may have produced an associated fracture zone.

Fractures. Because of the large-scale faulting, all geologic units in the ORR are highly fractured. Detailed investigations of Conasauga Group core by Lutz and Dreier (1988) show that five fracture sets occur consistently throughout the unit. One set is parallel to bedding, but the other four are generally perpendicular to bedding. The parallel bedding fractures are mainly release joints. Recent studies elsewhere in the Appalachians suggest that release joints can form at depths up to a kilometer (Engelder 1985). Assuming a regional strike of N55E, the strikes of the high angle sets are approximately N55E, N75W, N15E, and N20W.

Fracture density varies throughout the cores. The two variables that have the greatest effect on density are lithology and bedding thickness. Density is inversely proportional to bedding thickness. High fracture densities are found in shales and interbedded limestone and shale, whereas limestones exhibit lower densities. Mineralization of fracture planes is almost entirely confined to limestone beds. All fracture sets exhibit vein development; however, the N55E direction is the most commonly mineralized.

3.4.2 Soil and Stream Sediment Contamination

There is close physical and chemical association between stream sediments, especially those in tributaries to Bear Creek, and soils in BCV. As a result, these materials have been considered together in many of the studies of contaminant distribution in Bear Creek. These materials, therefore, are considered in the same section of this report. Contaminant background data for soils on the Nolichucky and Dismal Gap formations are available in the Phase I annual report of the ORR Background Soils Characterization Project (Energy Systems 1993b).

3.4.2.1 Bear Creek stream sediments

Soils and sediments in Bear Creek have been sampled and analyzed as part of investigations and studies of contaminant distributions at the various waste sites in BCV. Five locations (BCK 1.30, BCK 4.70, BCK 9.91, BCK 11.09 and BCK 12.10) were selected for screening of contaminants in the Bear Creek floodplain (Fig. 3.8) (Turner et al. 1991). Two or three soil cores up to 67.5 cm (27 in.) long were collected at each site using a vibracorer. Each core was cut into intervals ~7.5 to 20 cm (3 to 8 in.) long. Core sections were subsampled for analysis for volatile compounds; the remainder were air dried, homogenized, and crushed. The ORNL Analytical Chemistry Division performed all of the chemical analyses except the gamma spectroscopy, which was performed in the Environmental Sciences Division's Low-Level Counting Facility.

Fig. 3.8. Location of Bear Creek floodplain soil sampling sites.

Tables 3.7 and 3.8 summarize the results of analyses carried out on these cores. The combined summaries indicate that PCBs in some samples were elevated above the TSCA PCB spill cleanup criterion (25 mg/kg) for soil on the bank of a stream. The summaries also indicate that some samples exceeded the Nuclear Regulatory Commission (NRC)/DOE guideline of 30 to 35 pCi/g for uranium in soil with unrestricted access. Certain metals, especially cadmium, also slightly exceeded expected background concentrations. In most cases, the highest concentrations of contaminants were in the surface soil samples.

3.4.2.2 Burial Grounds soils and stream sediments

In the early 1980s, soils and stream sediments were analyzed as part of a comprehensive sampling and analysis program carried out to evaluate the extent of contamination at the Burial Grounds (McCauley 1985c). More recent studies of soils and stream sediments at the Burial Grounds have been carried out for NT-7 and ORP 1 (Geraghty and Miller 1985; 1988), for ORP 1 (Herbes 1988), and for NT-8 (Bogle et al. 1991) (Fig. 3.9 and Appendix C).

Volatile organic priority pollutant compounds have been detected in the soils near the waste sources on the Burial Grounds and in stream sediments from tributaries to Bear Creek that drain the Burial Grounds. PCBs and metals have been found in soils and sediment near the ORPs and near BG-C West.

Appendix G in Turner et al. (1991) summarizes local background concentrations of a variety of contaminants in surface water and sediment. These background data were collected at Tributary NT-2 upstream of all waste facilities and at two sites (GCK-1.4 and GCK-2.4) in the Grassy Creek watershed, which represents the westward continuation of BCV after Bear Creek turns north through Pine Ridge. A summary of background soil chemical data is in Table 3.4.

PCBs. Several hundred soil and sediment samples for chemical analysis for PCBs have been collected from the ORPs and Tributaries NT-6, NT-7, and NT-8 (Fig. 3.9 and Appendix C). The results of these analyses confirmed the presence of PCBs at concentrations above the action level of 25 mg/kg throughout the sediments of ORPs 1 and 2 and much of Tributary NT-7.

In 1980 and 1982, a series of soil cores was collected from transects on the perimeter of ORP 1 and analyzed for PCBs. Concentrations of PCBs in six samples collected between 0 and 3.7 m (0 and 12 ft) from the pond perimeter ranged from 13 to 1448 $\mu\text{g/g}$. Several samples collected at a depth of 0.6 m (2 ft) contained between 2.8 and 14 $\mu\text{g/g}$. Samples collected from the ORP 1 perimeter in July 1985 showed similar concentrations and distributions. Samples collected from ORP 2 during this same period revealed concentrations as high as 174 $\mu\text{g/g}$ in surface sediments from the perimeter of the pond, with concentrations decreasing with depth and distance from the pond.

In October 1985, five cores were collected from the bottom of ORP 1 and analyzed for PCBs. Concentration of PCBs in surface sediments from these areas ranged as high as 710 $\mu\text{g/g}$ but decreased to 3 $\mu\text{g/g}$ or less at a depth of 67.5 cm (27 in.) (Herbes 1988).

Extensive sediment sampling was conducted throughout the tributary system between 1985 and 1988 and has shown that PCBs are present at concentrations of several hundred micrograms per gram in surface sediments of Tributary NT-7 between ORP 1 and the visible

Table 3.7. Statistical summary of contaminant levels at the five floodplain subsurface soil-sampling localities in Bear Creek sampled April 1985^a

Parameter ^b	N	Number times < d.l. ^c	Mean	Maximum	Minimum
Σ Alpha (pCi/g)	62	0	22.0	103	2.4
Σ Beta (pCi/g)	62	0	22.5	119	6.5
¹³⁷ Cs (pCi/g)	63	23	0.35	2.3	0.01
⁴⁰ K (pCi/g)	63	0	11.0	18.2	4.9
²³⁵ U (pCi/g)	63	51	0.27	2.8	0.07
²³⁸ U (pCi/g)	63	36	9.9	93	0.80
Σ Uranium	62	0	28.0	241	0.51
TCE	61	52	0.03	0.45	0.01
PCE	61	54	0.01	0.06	0.01
PCBs	48	0	3.87	132	0.02
Ag	63	63	1.49	1.70	1.10
Al	63	0	16,540	39,000	10,000
As	63	63	4.1	4.8	2.5
Ba	63	0	120	270	59
Be	63	0	0.96	2.2	0.66
Ca	63	0	1,735	14,000	670
Cd	63	1	2.4	12	0.12
Co	63	0	16.6	37	9.8
Cr	63	0	24	73	13
Cu	63	0	13.0	51	4
Fe	63	0	17,338	39,000	1,300
Ga	63	63	10	12	7
Hg	50	0	8	15	3.6
Li	63	1	19.8	58	5.0
Mg	62	0	1,402	4,100	600
Mn	63	0	988	2,100	220
Mo	62	61	0.57	2.10	0.32
Na	62	0	74	180	33

Table 3.7 (continued)

Parameter ^b	N	Number times < d.l. ^c	Mean	Maximum	Minimum
Ni	62	0	21	63	10
P	63	0	273	490	140
Pb	62	0	29	90	15
Sb	62	48	8.1	17	4.8
Se	63	55	9.8	33	5
Sr	62	0	6.4	18	3.7
Ti	62	0	65	100	41
V	62	0	23	48	13
Zn	63	0	39	130	20
Zr	63	1	4.3	10	1.1

^aUnits are $\mu\text{g/g}$, dry weight, except as noted.

^bPreliminary remediation goals for these parameters are discussed in Chap. 5.

^cd.l. = detection limit.

Source: Turner et al. 1991.

Table 3.8. Statistical summary of contaminant levels in floodplain surface soils
at five sites in Bear Creek sampled April 1985^a

Parameter ^b	N	Number times < d.L. ^c	Mean	Maximum	Minimum
Σ Alpha (pCi/g)	14	0	39.7	103.	9.4
Σ Beta (pCi/g)	14	0	37.2	116.	19.7
¹³⁷ Cs (pCi/g)	14	0	1.04	2.30	0.14
⁴⁰ K (pCi/g)	14	0	11.95	18.10	5.2
²³⁵ U (pCi/g)	14	8	0.55	2.80	0.07
²³⁸ U (pCi/g)	14	3	23.2	92.7	1.5
Σ Uranium	14	0	3.4	241.0	11.0
TCE	14	11	0.02	0.19	0.01
PCE	14	13	0.01	0.03	0.01
PCBs	14	0	1.59	11.30	0.02
Ag	14	14	1.50	1.70	1.10
Al	14	0	18,286.	39,000.	12,000.
As	14	14	4.04	4.85	2.50
Ba	14	0	120.36	150.00	95.00
Be	14	0	1.05	2.20	0.69
Ca	14	0	3,721.	14,000.	1,300.
Cd	14	0	3.96	12.00	0.90
Co	14	0	17.3	37.0	13.
Cr	14	0	22.9	36.0	16.
Cu	14	0	21.7	51.0	9.4
Fe	14	0	17,214.	35,000.	1,100.
Ga	14	14	10.4	12.0	7.5
Hf	11	0	8.2	11.0	5.8
Li	14	0	28.4	58.0	16.0
Mg	14	0	1,685.	4,100.	890.
Mn	14	0	929.	1,900.	610.
Mo	14	13	0.64	2.10	0.32
Na	14	0	80.	180.	35.
Ni	14	0	30.	63.	18.
P	14	0	362.	490.	220.
Pb	14	0	41.	90.	15.
Sb	14	11	8.3	17.	4.8
Se	14	13	8.8	19.	5.
Sr	14	0	9.6	18.	5.9

Table 3.8 (continued)

Parameter ^b	N	Number times < d.l. ^c	Mean	Maximum	Minimum
Ti	14	0	63.	90.	44.
V	14	0	23.	41.	16.
Zn	14	0	62.	130.	31.
Zr	14	0	5.0	6.9	3.4
Depth	14	0	1.50	1.50	1.50

^aUnits are $\mu\text{g/g}$, dry weight, except as noted.

^bPreliminary remediation goals for these parameters are discussed in Chap. 5.

^cd.l. = detection limit.

Source: Turner et al. 1991.

area of oil seepage at the west end of BG-A South (Figs. 3.10 and 3.11). However, concentrations decrease with depth to below $25 \mu\text{g/g}$ at 0.6 m (2 ft) (Figs. 3.10 and 3.11). Borings in this study also recorded free-phase light nonaqueous phase liquids (LNAPL) on groundwater (Geraghty and Miller 1988).

PCBs have been found in areas of the Tributary NT-7 floodplain soils south of ORP 1 at depths of concentrations as high as $574 \mu\text{g/g}$. Because PCBs adhere strongly to soil particles, elevated concentrations in the Tributary NT-7 floodplain soils appear to be related to sediment transport from upstream sources (Herbes 1988). Concentrations in Tributary NT-7 sediments at the confluence with Bear Creek were reported to range from 5.2 to $28.2 \mu\text{g/g}$ in 1983 and 1984 and from 2.1 to $14.1 \mu\text{g/g}$ in 1986 (Herbes 1988).

Sediment from the drainage ditch north of ORP 2 had a maximum PCB concentration of $33 \mu\text{g/g}$. PCBs at concentrations between 2.3 and $13.8 \mu\text{g/g}$ are present in Tributary NT-6 sediments throughout its length between the confluence with the intermittent stream draining ORP 2 and the confluence with Bear Creek. Concentrations of PCBs in Tributary NT-9 ranged from $<0.1 \mu\text{g/g}$ on the western branch, which does not flow through the Burial Grounds, to $50 \mu\text{g/g}$ collected near the confluence with Bear Creek.

Oil seeping from BG-C into NT-8 was investigated in 1990. Soils were sampled in transects across visible sites of seepage and had concentrations of PCBs that ranged from below detection limits to $80 \mu\text{g/g}$. Vertical distribution of PCBs generally decreased with depth, and concentrations of PCBs exceeding $25 \mu\text{g/g}$ were limited to the upper 30 cm (12 in.) of sediment. Based on this sampling event, the total volume of PCB-contaminated sediment in the upper east fork of Tributary NT-8 that exceeds the proposed action limit is estimated to be 34 m^3 (1200 ft^3) (Bogle et al. 1991).

During closure operations at the ORPs and BG-A, all soils and sediments with PCBs $>25 \mu\text{g/g}$ were excavated and removed from ORPs 1 and 2, NT-7, and visible oil stain areas near BG-A North. The channel of NT-7 was rerouted $\sim 15 \text{ m}$ (50 ft) to the east of its original course.

LEGEND:

- GW-046 SOIL CORE
- T8-S3 SURFACE SAMPLE
- * 8-2 STREAM SEDIMENT
- SOIL SAMPLE TRANSECT
- SEEP
- CLAY CAPPED DISPOSAL UNIT
- RCRA CAPPED DISPOSAL UNIT
- UNCAPPED DISPOSAL UNIT

SCALE: 1" = 400'

BEAR CREEK VALLEY OPERABLE UNIT 1 RI WORK PLAN

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93044/XREF/BCSAMP	

FIG. 2 93044\DWG\471BC1-1 0
FIGURE # CAD FILE # ROT ANGLE

Fig. 3.11. Oil contamination in soil near ORP 1, west of BG-A South (Source: Geraghty and Miller 1988).

Uranium. Because large quantities of uranium are known to have been deposited in the Burial Grounds, soil and sediment sampling campaigns conducted in Tributaries NT-6 and NT-7 since 1985 have included total uranium analyses. Concentrations measured in sediments north of the Burial Grounds generally range from 1 to 10 $\mu\text{g/g}$. Unlike PCB concentrations, which are present at far higher levels in Tributary NT-7 sediments than in Tributary NT-6 sediments, uranium concentrations are similar in both tributaries and are as high as 225 $\mu\text{g/g}$ in the lower reaches of the tributaries. The arithmetic mean of nine samples collected from three locations in the stream channel and adjacent banks from Tributary NT-7 south of ORP 1 was 84 $\mu\text{g/g}$, while the mean of an equal number of samples collected from three similar locations in Tributary NT-6 south of ORP 2 was 98 $\mu\text{g/g}$ (Herbes 1988). Total uranium in the surface sediment collected in January 1990 from the upper reaches of Tributary NT-8 ranged from 212.2 to 345 $\mu\text{g/g}$. In sediment cores from NT-8, total uranium values ranged from background concentration to 7250 $\mu\text{g/g}$ (Bogle et al. 1991). No correlation was observed between PCB and uranium concentrations.

Metals, cyanide, and semivolatiles. When compared to a control site on Tributary NT-7 upstream from the Burial Grounds, samples collected in September 1983 did not exhibit elevated levels of toxic metals, cyanide, or phenols in ORP 1 or in Tributary NT-7 sediments downstream from the pond (McCauley 1985c, Radian 1993a). Contaminants found in the sediments were arsenic, beryllium, cadmium, chromium, cobalt, copper, lead, nickel, silver, zinc, cyanide, and total phenols. A sample collected at this same location in December 1987 revealed similar results; no semivolatile contaminants were detected in Tributary NT-7.

VOCs. Sediment samples collected along the upper reaches of Tributary NT-8 in January 1990 indicated that elevated levels of VOCs exist in the sediments. Generally, the distribution of VOCs at the sediment sampling locations mirrors their distribution in surface water samples. The highest summed concentration of VOCs in a sediment core was 34,599 $\mu\text{g/g}$. The major VOCs detected in the sediments along the upper reaches of Tributary NT-8 were trichloroethylene, tetrachloroethylene, 1,1-trichloroethane, 1,1-dichloroethane, cis-1,2-dichloroethene, chloroethane, methylene chloride, vinyl chloride, benzene, toluene, and Freon 113 (Bogle et al. 1991).

Summary. Numerous analyses for PCBs and uranium have been performed in soil and sediments from the ORPs and tributaries NT-6, NT-7, and NT-8. The horizontal and vertical extent of PCB contamination was well defined at the ORPs and along tributaries NT-6, NT-7, and NT-8 prior to RCRA remediation activities. Uranium contamination has been defined to a more limited extent; the horizontal and vertical extent of soil contamination within the unsaturated zone has not been fully defined.

3.4.2.3 Oil Landfarm soils and stream sediments

Because of concerns about levels of PCBs that may have been present in the Oil Landfarm from disposal of contaminated oils, several rounds of sampling for PCBs in Oil Landfarm soils have occurred since 1981. The highest level detected in Oil Landfarm soil during this period was 110 $\mu\text{g/g}$ (BNI 1983). A total of 200 samples were collected in 1985 in a randomized block design with the objective of delineating areas containing PCB levels $>25 \mu\text{g/g}$. Several portions of plots were identified over this action level; the highest concentration measured was 60.5 $\mu\text{g/g}$. PCB contamination was principally confined to the upper 15 cm (6 in.) of the soil; concentrations declined by more than 90% between the surface [0 to 15 cm (0 to 6 in.) interval] and the 30- to 45-cm (12- to 18-in.) interval in

10 depth profiles. Highest concentrations were found in the east and west plots (Figs. 3.12 and 3.13) (Herbes 1988) (Appendix A).

Soil samples collected in the 1985 study (Herbes 1988) were also analyzed for uranium. Concentrations of uranium were similarly highest in the east and west plots. The highest mean uranium concentration observed was 204 $\mu\text{g/g}$ (in Plot 7/8). Lowest concentrations were between 13 and 27 $\mu\text{g/g}$ in the north plot group. As was observed for PCBs, uranium concentrations declined rapidly with soil depth (Fig. 3.13).

There were 48 soil samples collected and analyzed by the Y-12 Plant Analytical Laboratory for total carbon, PCBs, and a series of metals. The results are summarized in the BNI (1983) report. Maximum levels (in $\mu\text{g/g}$) for the metals analyzed were arsenic (73), barium (340), beryllium (290), cadmium (6), chromium (88), lead (550), mercury (55), selenium (0.4), silver (8), thorium (20), and uranium (273). As would be expected from the historical pattern of oil application, contaminant concentrations were uniformly lowest in the north plots. Concentrations of beryllium were highest in Plots 1/2 (mean of four analyses: 213 $\mu\text{g/g}$), 5/6 (112), and 9/10 (75). Beryllium levels averaged 32 $\mu\text{g/g}$ in the west plots and 8 $\mu\text{g/g}$ in the north plots. With the exception of one sample from Plot 1/2, all mercury and lead concentrations were less than 4 $\mu\text{g/g}$ and 75 $\mu\text{g/g}$, respectively (Appendix A).

Composite samples were obtained in September 1983 from "light" and "dark" areas (differentiated visually) of Plots 1-12, 13-18, and 19-24 (the east, west, and north areas, respectively). The five samples (no "dark" area was identified in the north plots) were analyzed for VOCs, acid- and base-neutral extractables, ICP-analyzable metals, and pesticides/PCBs. The results are summarized in the BNI (1983) report. In general, the Plots 1-12 (east) "dark" sample contained the highest concentrations of contaminants. Major organic contaminants found, and maximum concentrations (in micrograms per gram), were tetrachloroethylene (8.4), trans-1,2-dichloroethylene (2.7), trichloroethylene (0.75), 1,1,1-trichloroethane (0.50), bis(2-ethylhexyl) phthalate (520), fluoranthene (140), phenanthrene (13), phenol (29), and 2,4-dimethylphenol (31).

A number of cores from monitoring wells installed around the perimeter of the Oil Landfarm in 1983 by BNI were analyzed for the organic contaminants just listed (Fig. 3.12, Appendix C). Principal organic contaminants identified, and the highest levels quantified (in micrograms per gram), included toluene (1.6), methylene chloride (0.96), tetrachloroethylene (1.1), and trans-1,2-dichloroethylene (0.40). PCBs were found in three core segments from well GW-5, located east of the Oil Landfarm, at total concentrations of 1.6 $\mu\text{g/g}$ to 2.8 $\mu\text{g/g}$ at depths of 0 to 1.5 m (0 to 5 ft) (BNI 1983).

Several water and sediment samples were collected by BNI near the confluences of tributaries NT-3 and NT-4 with Bear Creek during 1983 and 1984 (Turner et al. 1991). Levels of VOCs, phenols, and beryllium were all below detection limits (0.01, 0.001, and 0.001 mg/L, respectively) in these samples. Both dissolved and particulate PCBs were also below the detection limit (0.0003 mg/L) in single samples collected in 1983.

PCBs were undetectable ($<0.05 \mu\text{g/g}$) both in tributary and in Bear Creek sediments. Concentrations of beryllium; heavy metals (cadmium, chromium, cobalt, copper, lead, nickel, uranium, and mercury); and phenols showed no increases in the tributary sediments above upstream Bear Creek levels (Turner et al. 1991).

Fig. 3.12. Map showing the location of soil and sediment samples.

LEGEND:

- GW-046 SOIL CORE
- * 8-2 STREAM SEDIMENT
- CLAY CAPPED DISPOSAL UNIT
- RCRA CAPPED DISPOSAL UNIT
- UNCAPPED DISPOSAL UNIT

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FIGURE # 93044\DWGS\4718C1-2 0
CAD FILE # ROT. ANGLE

in the Oil Landfarm (see Appendix C for sources of data and results).

Fig. 3.13. Mean concentrations of PCBs and oil/grease with depth at the Oil Landfarm (*Source: Herbes 1989*).

3.4.2.4 Boneyard/Burnyard soils and stream sediments

There is little information available concerning the Boneyard/Burnyard. A number of areas containing radiologically contaminated materials were defined: one on the extreme west perimeter of Boneyard/Burnyard at a depth of 7.5 to 10 cm (3 to 4 in.) that extended across NT-3 for at least 11 m (35 ft) toward the SL 1 (maximum beta-gamma of 1200 cpm) and showed no direct source for contamination. The other area of radiological contamination was located near radiologically contaminated barrels that may have contained mop water. One such area was located in the northwest corner of the Boneyard/Burnyard (maximum beta-gamma of 300 cpm in the soil and 60,000 cpm in the barrel); the second barrel was located in the southeast corner and had beta-gamma readings of 1000 cpm. Elevated levels were also indicated in the main stem of Bear Creek during this survey (300 cpm).

Data from an investigation carried out by BNI for several grab samples of sediment collected downstream of the Boneyard area show elevated levels of mercury (6.7 to 140 $\mu\text{g/g}$), uranium (18 to 200 pCi/g), several polynuclear aromatic hydrocarbons, and phthalates (Fig. 3.12, Appendix C). Analytical results for groundwater downgradient of the area also have indicated the possibility of migration of a number of contaminants from the HCDA (Boneyard/Burnyard), including organic compounds and uranium (Turner et al. 1991).

3.4.2.5 S-3 Site soils and stream sediments

Minimal information is available concerning the extent of soil and sediment contamination in the vicinity of the S-3 Site. Past sampling efforts have primarily involved the characterization of waste materials and of stream sediments and precipitates from Bear Creek. The extent of contamination in unconsolidated materials surrounding the S-3 Site has not been accurately delineated (Fig. 3.14, Appendix C).

Chemical data for stream sediments, residuum, and soils near the S-3 Site waste management area are presented by McCauley (1985a). TVA reported data on additional sediment samples collected from Bear Creek and EFPC. These data indicate that the Bear Creek stream sediments downstream from the S-3 Ponds waste management area have a higher metal content than the Bear Creek tributary sediments (background). Few organics were detected, and those detected were present in trace quantities.

Analyses of residuum samples collected during drilling of monitoring wells GW-100, -101, -102, and -103 revealed that arsenic, barium, chromium, lead, and mercury were present in high concentrations relative to background values in Table 3.4 (Fig. 3.14). The total concentration of these metals in each sample exceeded several hundred parts per million (Geraghty and Miller 1985). Only trace levels of organic contaminants were detected in the residuum samples. No trends associated with depth or area were determined (Appendix C).

Turner et al. (1988) sampled a spoils pile composed of soils excavated while constructing the tank farm west of the S-3 Site during 1985. While no tested samples exhibited the characteristic of extraction procedure (EP) toxicity, gross alpha activity exceeded the background value of 20 pCi/g in all samples; gross beta activity exceeded the background value of 55 pCi/g in 8 of 31 samples. Laboratory analysis indicated that mercury was also present at levels above background. There has been no documented attempt to assess the horizontal or vertical extent of contamination in the unconsolidated materials surrounding the S-3 Ponds.

Fig. 3.14. Location of soil and stream sediment samples

LEGEND:

- GW-046 SOIL CORE
- * 8-2 STREAM SEDIMENT
- RCRA CAPPED DISPOSAL UNIT

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93044\DWGS\471BC1-3 0
 FIGURE # CAD FILE # ROT ANGLE

sites at the S-3 Site (see Appendix C for sources and results).

Treatment of the contents of the S-3 Ponds produced a sludge of basic pH, consisting primarily of calcium carbonate, metal hydroxides, and by-products from biodegradation. Compositions of S-3 sludges are described in Chap. 2 and in Appendix A. Metals and radionuclides were not found in leachate from either weak acid (EP Toxicity) or groundwater solutions (Energy Systems 1988). In the fall of 1984, 32 core samples were taken from each of the four ponds for constant head permeability tests. Generally, the sludges in each pond were characterized by the following permeability values: 3.5×10^{-4} cm/sec for the southeast pond, 5.5×10^{-5} cm/sec for the northeast pond, 1.7×10^{-4} cm/sec for the southwest pond, and 4.2×10^{-5} cm/sec for the northwest pond (Energy Systems 1988).

3.4.3 Soil Erosion

The erosion potential of soil at a site can be estimated by the use of the Universal Soil Loss Equation (USLE) (USDA 1978). The USLE can be used to design soil conservation practices and procedures to minimize accelerated soil loss. Potential erosion hazards can be evaluated for larger tracts of land by evaluating slope gradient, length of slope, amount of very fine sand and silt in the upper soil, and the thickness of soil that can retain rainfall before the onset of overland flow.

A soil erodibility factor [K] estimates the probability that a soil particle can be detached from the surface and transported downslope by rainfall (USDA 1983). The K factor can be estimated by using a nomograph that relates K factor to five soil properties: (1) percentage of silt plus very fine sand, (2) percentage of sand >0.10 mm, (3) organic matter content of the soil layer that is exposed to rainfall, (4) soil structure, and (5) limiting permeability. Lietzke et al. (1988) calculated K factors for soils at the Bear Creek Valley Low-Level Waste Disposal Development and Demonstration (LLWDDD) Program Site, and these are presented in Table 3.9.

Lietzke et al. (1988) identified four areas in BCV where soil erosion potential is high: (1) land with slopes $>25\%$, (2) all areas that have been seriously eroded in the past, (3) shallow soils underlain by bedrock, and (4) soils that have high silt plus clay content. Most soils in BCV have the potential for accelerated erodibility if they are disturbed or left unprotected, with the exception of soils in the floodplains where sediments generally collect.

3.5 SURFACE WATER

The surface water system in BCHR is comprised of Bear Creek and its tributaries. Headwaters of Bear Creek are located at the west end of the Y-12 Plant near the S-3 Ponds. In its upper reaches, Bear Creek follows a relatively straight course along geologic strike that lies very close to the contact between the Conasauga and Knox Groups. The original channel, located on the west side of the S-3 Ponds site, was filled with rubble during pond construction and rerouted to its present location (Law 1983). Approximately 7.2 km (4.5 miles) downstream of its headwaters, Bear Creek turns northward through a gap in Pine Ridge and empties into EFPC ~9.7 km (6 miles) upstream of the Clinch River. The parallel alignment of the tributaries located in the shale units in BCV is believed to represent the surface expression of joint and fracture patterns in the bedrock (Geraghty and Miller 1985) and/or other structural features.

Table 3.9. Range of estimated [K] values for major soils on the Low-Level Waste Disposal Development and Demonstration Program site^a

Geology and soil number	In-place fine earth	In-place whole soil	Disturbed whole soil^b
Nolichucky residuum Soil No. 51			
Solum	0.26–0.34	0.24–0.31	0.28–0.34
Saprolite	0.41–0.49	0.03–0.04	0.01–0.03
Maryville residuum Soil No. 42			
Solum	0.19–0.28	0.10–0.12	0.13–0.18
Saprolite	0.30–0.36	0.02–0.03	0.01–0.05
Rogersville residuum Soil No. 36			
Solum	0.07–0.34	0.07–0.19	0.06–0.17
Saprolite	0.26–0.33	0.03–0.23	0.01–0.04
Pumpkin Valley residuum Soil No. 24			
Solum	0.26–0.44	0.16–0.20	0.17–0.24
Saprolite	0.30–0.44	0.03–0.23	0.01–0.17
Rogersville-Maryville-Nolichucky colluvium Soil No. 49			
Solum	0.19–0.26	0.12–0.23	0.20–0.28
Pumpkin Valley residuum Soil No. 26			
Solum	0.32–0.42	0.21–0.39	0.21–0.28
Old alluvium Soil No. 951			
Solum	0.36–0.52	0.36–0.52	0.38–0.56

^aData from Lietzke et al. 1988.

^bEstimated range of [K] values for bare soil between disturbance and 1 year afterward.

As discussed in Sect. 1.5, the scope of BCV OU 1 RI with respect to surface water is to address the extent to which units of BCV OU 1 may be contributing contaminants to surface water. In some cases, contaminants may be transported from source areas to Bear Creek via groundwater flow paths. In these cases, the scope of this work plan addresses the extent to which the source area may be contributing contaminants to groundwater. Remediation of surface water and detailed determination of the interactions between surface water and groundwater will be addressed in the BCV OU 4 RI.

3.5.1 Bear Creek

The channel of Bear Creek along much of its course through BCV is formed in the shallow alluvium overlying the Maynardville Limestone. Drilling activities in areas proximal to Bear Creek have indicated that a zone of potentially interconnected solution cavities is present in the upper part of the saturated zone in this formation. Because of the hydraulic interaction between the creek and this solution cavity system, the Bear Creek channel contains losing reaches where surface water discharges from the creek to the groundwater system and gaining reaches where groundwater discharges to the creek.

A dye tracer test was performed in 1989 to evaluate the hydrologic interaction between Bear Creek and the Maynardville Limestone (Geraghty and Miller 1989). During this study, a review of historic stream-flow data for Bear Creek was performed. Results of this review indicated that Bear Creek loses flow to the groundwater system during extreme low-flow conditions over much of its length from near the S-3 Ponds to a spring located at BCK 4.56. During normal flow conditions, Bear Creek loses flow to the groundwater system in its upper reaches and gains flow in its lower reaches downstream of BCK 9.41 (Geraghty and Miller 1989). It was concluded from the dye-tracer test that there was strong evidence indicating that surface water lost in losing reaches of Bear Creek between BCK 11.64 and BCK 9.41 is substantially discharged back into Bear Creek from spring SS-5 (Geraghty and Miller 1989).

3.5.2 Stream Flow Data Collection

Discharge data have been collected in the upper part of Bear Creek at many locations over different periods from 1986 through 1990. Monitoring stations are shown in Fig. 3.15, and their locations are listed according to BCK.

Periodic flow monitoring in Bear Creek has been carried out by the USGS since 1986. Flow and limited chemical data have been monitored by staff in the Environmental Sciences Division at ORNL. Details of past surface water sampling activities along Bear Creek are given in Turner et al. (1991).

Flow data in the main stem, tributaries, and springs allow the different sections of the stream to be classified as losing or gaining reaches. The data also indicate flow in solution cavities in the Maynardville Limestone underlying Bear Creek. This flow apparently conducts water from the upper reaches between BCK 9.53 and BCK 12.46 to lower reaches of Bear Creek.

Fig. 3.15. Location of previously sampled surface monitoring sites near Bear Creek.

7-070693-096

3.5.3 Effects of the Solution Cavity System

The solution cavity system within the Maynardville Limestone, which conducts flow westward under Bear Creek, plays an important role in the surface water hydrology of BCV. The cavity system is thought to be an important avenue for contaminant transport because contaminants in the surface water can be lost to the system, which extends outside the BCHR. Although the precise outlets of the cavity system are unknown, there is evidence that at least some flow from the cavity system emerges in springs discharging from solution openings along bedding planes and joints in the rock units adjacent to Bear Creek.

Both springs SS4 and SS5 are perennial, flowing from the base of Chestnut Ridge on the southern side of the valley. Both streams also show higher average specific conductivity than the main-stem water. The high specific conductivity is an indication that the water from the headwaters (BCK 12.46) or Tributary NT-1 (BCK 12.39), or from a similar source, has been transported to these springs.

3.5.4 Stormflow System

Detailed water budgets for the ORR indicate that ~90% of active subsurface flow occurs through a 1- to 2-m-deep (3.3- to 6.6-ft) stormflow zone (Solomon et al. 1992). Natural areas of the ORR are heavily vegetated, and the stormflow zone approximately corresponds to the root zone. Infiltration tests indicate that this zone is as much as 1000 times more permeable than the underlying vadose zone. During rain events, the stormflow zone partially or completely saturates and then transmits water laterally to the surface water system (Fig. 3.16). When the stormflow zone becomes completely saturated, overland flow occurs.

Where such excavations as waste trenches penetrate the stormflow zone, a commonly observed condition known as bathtubbing can occur in which the excavation fills with water (Ashwood and Spalding 1991) (Fig. 3.16). Between rain events, as the stormflow zone drains, flow rates decrease dramatically, and flow becomes nearly vertical toward the underlying vadose zone. The transmissive capability of the stormflow zone is created primarily by root channels, worm holes, clay aggregation, and fractures, collectively referred to as macro pores. Although highly transmissive, large pores comprise only ~0.2% of the total void volume of the stormflow zone. Because most of the water mass resides within less transmissive small pores, advective-diffusive exchange between large pores and small pores reduces contaminant migration rates relative to fluid velocities in large pores.

The effect on stormwater flow of excavation and surface modifications at the various waste sites is not known. Seeps have been identified in NT-7 that apparently drain from the ends of trenches in BG-A, and this implies that the man-made structures in the Burial Grounds have a strong effect on stormflow hydrology. The extent of this effect must be identified as an important data gap because the stormflow zone is the most likely pathway for contaminant movement away from sites in BCV OU 1 to Bear Creek and its tributaries.

3.5.5 Surface Water Contamination

Surface water in the Upper Bear Creek watershed is affected by surface and subsurface drainage from the waste Burial Grounds, the Oil Landfarm, and the S-3 Ponds. These sources have contributed organic and inorganic chemical contaminants to Bear Creek. In recent years,

Fig. 3.16. Conceptual model for the stormflow and shallow groundwater hydrogeologic regime at BCV OU 1.

Table 3.10. Summed concentrations ($\mu\text{g/L}$) of VOCs in Bear Creek surface water at various sites

Site	Date						
	Sept. 1983	Mar. 1984	Aug. 1984	Aug. 1985	Jan. 1986	May 1986	Sept. 1986
BCK 12.39	NA ^a	NA	<10	<10	<10	<10	<10
BCK 11.83	NA	NA	<10	<10	10	<10	<10
BCK 11.49	ND ^b	<10	ND	<10	<10	<10	NA
BCK 11.09	ND	ND	NA	NA	10	NA	<10
BCK 10.32	NA	ND	NA	<10	<10	NA	<10
BCK 9.91	1291	201	225	147	874	210	210
BCK 9.43	52	97	11	46	225	11	23
BCK 9.40	NA	NA	NA	24	34	<10	12
BCK 7.87	NA	NA	<10	<10	<10	<10	<10
BCK 5.15	NA	NA	NA	<10	<10	<10	<10
BCK 3.25	NA	NA	NA	65 ^c	<10	<10	<10

^aNot applicable.

^bNot detected.

^cCaused by methylene chloride, a likely laboratory contaminant.

Source: McCauley 1985c.

actions have been taken to reduce the input of contaminants to Bear Creek from several of these sources, and several of these sources have also been closed and capped. Intensive monitoring of stream chemistry was initiated at about the same time many of these actions were being taken.

3.5.5.1 Organic contaminants

Surface waters from Bear Creek and tributaries were analyzed for a broad spectrum of organic priority pollutants during 1983 and 1984 (McCauley 1985c). Further sampling conducted during 1985 and 1986 was restricted to the Bear Creek stem. Surface water sampling is also included in the annual groundwater quality assessment reports (GWQARs) (HSW 1992). Only VOCs were detected in the aqueous phase in the stem of Bear Creek. These compounds consist primarily of halogenated aliphatics and low molecular weight aromatics. Trace levels of tetrachloroethene, toluene, and trichloroethene were detected ($\sim 10 \mu\text{g/L}$ each) in the uppermost reaches of Bear Creek (BCK 12.36), while much higher levels of organics were detected farther downstream in the vicinity of the Burial Grounds. Trans-1,2-dichloroethene, tetrachloroethene, trichloroethylene, and vinyl chloride predominated; traces of other halogenated aliphatics were detected. Table 3.10 illustrates the summed concentrations ($\mu\text{g/L}$) of VOCs in Bear Creek surface water at various sites. The

results show the highest concentration of VOCs were detected at BCK 9.91 immediately downstream of the Burial Grounds. VOC concentrations decreased rapidly downstream from this site, undoubtedly caused by volatilization (Callahan et al. 1979). Concentrations generally dropped to about 5% of the level at BCK 9.91 within 0.5 km (0.3 mile) and were always less than 10 $\mu\text{g/L}$ at BCK 7.87. Sources of the VOCs in this area at that time were BG-A through BG-D, which are underlain by groundwater that is highly contaminated by VOCs and drained by tributaries (NT-7 and NT-8) that are contaminated with VOCs to levels 10 to 20 times those observed in Bear Creek.

Levels of summed VOCs in sections of tributaries NT-7 and NT-8 exceeded 2–3 mg/L in March and August 1984. High levels of VOCs were detected in Bear Creek at BCK 9.91 in September 1983 and January 1986. Although tributaries were not sampled at this time, detection of VOCs in the main stem indicates that levels of VOCs in NT-7 and/or NT-8 were also high. Recent measurements (Bogle et al. 1991) of VOCs in the headwater seeps of NT-8 exceeded 3.5 mg/L (summed concentrations). The observed levels approach those known to produce acute toxicity in sensitive species, and could be chronically toxic to stream biota.

3.5.5.2 Inorganic contaminants

Before the discontinuing waste inputs to the S-3 Ponds at the headwaters of Bear Creek, the upper reaches of the stream were acidic and highly enriched with many inorganic constituents (ERDA 1975; Turner and Kamp 1984; Energy Systems 1984a). During 1974 and 1975, upper Bear Creek (BCK 11.1, 11.9) exhibited pH values ranging from 3.5 to 6.8 over a 5-month period (ERDA 1975), while from 1981 through 1983 a pH range of 3.9 to 7.5 was observed in the reach closer to the S-3 Ponds (BCK 12.1, 12.5) (Table 3.11).

Before neutralization of the S-3 Ponds in summer 1983, upper Bear Creek (BCK 12.55) pH ranged from 4.0 to 4.5 during July–September (Energy Systems 1984a). High levels of aluminum (more than 100 ppm) were noted in the acidified waters. The levels of acidity and aluminum observed before September 1983 are toxic to fish and sensitive aquatic invertebrates (Altshuler and Linthurst 1983) and would adversely affect the biotic community of Bear Creek. The acidity in the upper reaches of Bear Creek was neutralized as it was carried farther downstream. It is safe to assume that toxic conditions existed in upper Bear Creek before neutralization of the S-3 Ponds and that this toxicity should have diminished within several kilometers downstream.

Profound changes in the chemistry of upper Bear Creek occurred following neutralization of the S-3 Site in 1983. Within several months, pH rose to more than 7.0, aluminum dropped from 97 mg/L to 4 mg/L, nitrate decreased from 1900 to 300 mg/L, uranium decreased from 21 to 1 mg/L, and other metals also declined significantly (Energy Systems 1984a,b). Intensive chemical and biological monitoring of the Bear Creek watershed was initiated after these changes took place.

Chemical analyses of Bear Creek surface waters, since the S-3 Ponds were neutralized in 1983, have been conducted by BNI (McCauley 1985c); Roy F. Weston, Inc. (Appendix C, Turner et al. 1991 and Southworth et al. 1992); Tennessee Valley Authority (TVA) 1986; USGS (Pulliam 1985); and ORNL (Southworth et al. 1992). These analyses show high concentrations of dissolved inorganic salts in the upper reaches of Bear Creek, typified by electrical conductivities of 2000 to 4000 $\mu\text{mho/cm}$ and total dissolved solids concentrations of 2000 to 5000 mg/L. The contribution of solutes from the S-3 Ponds groundwater plume

dominates the major ion inorganic chemistry of Bear Creek. The total salt content, as indexed by conductivity, closely follows the pattern expected if Bear Creek water were diluted only with uncontaminated groundwater downstream from BCK 12.4. Solute inputs and dilution vary as a result of variations in precipitation runoff and infiltration, with the highest solute concentrations generally occurring during periods of low flow. Stream monitoring data from 1992 are summarized in Table 3.12.

Table 3.11. Mean pH (range in parentheses) of Bear Creek in three surveys conducted over a 10-year period

(Hydrogen ion concentrations were used in the calculation of mean pH)

Sample sites (BCK)	Sampling period/number of measurements	
	October–March 1974–75 ^a / 5 measurements	July–January 1983–84 ^b / 24 measurements
BCK 12.55	NS ^c	4.4 (4.0–4.6) ^d 6.7 (6.1–7.3) ^e
BCK 12.5	NS	NS
BCK 12.0	NS	NS
BCK 11.9	4.68 (3.50–5.09)	NS
BCK 11.1	6.32 (3.90–6.80)	NS
BCK 6.3	NS	NS
BCK 4.3	7.21 (6.95–7.58)	NS
BCK 1.8	7.12 (6.88–7.30)	NS

^aERDA 1975.

^bHSEAD 1984a,b.

^cNS = not sampled.

^dBefore September 29, 1983.

^eAfter September 29, 1983.

Bear Creek surface water is highly enriched [relative to Grassy Creek reference sites (Fig. 3.15)] in many inorganic ions. The downstream variation in major ion chemistry is depicted in Table 3.12. Aluminum, barium, calcium, chloride, magnesium, manganese, nitrate, potassium, sodium, sulfate, strontium, and uranium are conspicuously high in the headwaters of Bear Creek and decline gradually in concentration downstream. Lithium and boron are slightly elevated in the Bear Creek headwaters relative to Grassy Creek reference sites, but the highest levels occur below the Burial Grounds. Ammonia, beryllium, cadmium, cobalt, copper, lead, nickel, silver, and zinc are elevated in the uppermost reaches of Bear Creek, but decline to approximately control levels or below detection limits within a short distance downstream. A comparison of maximum concentrations in Bear and Grassy Creeks, reported toxicity values, and EPA water quality criteria for the protection of freshwater biota are presented in Table 3.13.

Table 3.12. Concentrations (mg/L) of major ions in unfiltered Bear Creek water at various sites downstream from the S-3 Site in 1992^a

Parameter	BCK 11.97	BCK 9.40	BCK 4.55	BCK 0.63	NT 13
Calcium	209.75	60.75	45.25	27.25	18.05
Magnesium	25.95	10.60	11.93	8.35	3.18
Aluminum	0.50	0.64	0.49	0.26	0.57
Sodium	32.00	8.33	4.28	4.13	4.15
Potassium	4.8	2.65	1.78	1.7	1.58
Manganese	0.86	0.08	0.07	0.10	0.77
Strontium	0.57	0.15	0.09	0.08	0.06
Barium	0.38	0.09	0.07	0.07	0.06
Uranium	0.18	0.09	0.03	0.04	<0.001
Nitrate (NO ₃ -N)	107.94	13.03	5.23	5.12	<0.2
Sulfate	47.75	19.63	12.83	10.73	10.80
Chloride	30.98	11.4	6.03	6.38	1.03
Alkalinity (as HCO ₃)	122.50	133.25	129	118	59.50

^aMean values for four quarterly samples taken on March 10, June 1, September 8, and December 16, 1992.

Hardness-dependent screening criteria for surface water will be calculated by using site-specific data for contaminant concentrations and hardness. The 5% lower-bound concentration of each contaminant will be used as a conservative screen. The equation to be used to calculate the screening criteria is presented in U.S. EPA Region IV guidance (EPA 1992e):

$$\text{Screening Value} = \exp\{[a \times \ln(\text{Hardness})] + b\},$$

where hardness is expressed as mg CaCO₃/L. In the absence of hardness data, hardness will be estimated as

$$(\text{mg Ca/L}) \times 2.497 + (\text{mg Mg/L}) \times 4.116$$

(Standard Methods for the Examination of Water and Wastewater, Method 309A). The parameters *a* and *b* are specific for each element and are different for acute and chronic exposures. They are listed in Table 3.14.

Table 3.13. Maximum concentrations (mg/L), acute toxicity ranges, and water quality criteria for inorganic ions found at elevated levels in Bear Creek

Ion	Grassy Creek	Bear Creek	Acute toxicity (LC ₅₀)	Water quality criterion	Chronic screening criteria ^a
Aluminum	0.6	5.8	0.2–38.0	0.75	0.087
Ammonia	0.04	0.39		3.5 ^b	
Barium	0.044	1.2	13.5–105.0		
Boron	0.01	1.1	900.0		0.75
Cadmium	0.004	0.16	0.09–7.2	0.0011 ^c	
Calcium	42.0	600.0			
Chloride	7.4	230.0		860	230
Copper	0.02	0.04	0.014–1.0	0.012 ^c	
Fluoride	0.3	2.0			
Lead	0.1	0.4	1–482	0.0032 ^c	
Lithium	0.1	0.8			
Magnesium	17.0	82.0			
Manganese	0.087	8.8	1.5–1000.0		
Nickel	0.01	0.08	1.8–188.0	0.160 ^c	
Nitrate	0.6	440.0			
Potassium	0.9	14.0			
Silver	0.03	0.12	0.02–1.0	0.120 ^c	
Sodium	8.2	67.0			
Strontium	<0.5	1.5	86–10,000		
Sulfate	2.0	109.0			
Uranium	2.0	2.0	2.8–5.0		
Zinc	0.014	0.022	0.78–14.3	0.110 ^c	

^aEPA Region IV.

^bpH 7, 25°C.

^cHardness dependent, to be calculated from site-specific data.

Sources: Southworth et al. 1992, EPA 1976, EPA 1992e.

Most of these inorganics are relatively nontoxic; many, such as calcium, magnesium, sodium, and sulfate, are typical dissolved constituents of fresh waters. The pH of Bear Creek was generally between six and seven from late 1983 to the present. The high acidity noted before 1983 was not observed.

3.5.6 Conclusions

Surface water in Bear Creek and those tributaries located in the Y-12 Plant waste disposal areas has been sampled extensively, and in some cases, intensively, since 1983. Virtually all sampling has been conducted under non-stormflow conditions. The analyses performed have included both inorganic and organic contaminants and have typically been comprehensive, including in one case EPA Appendix IX contaminants. Results of these analyses suggest that nitrate, uranium, aluminum, and cadmium are the chief contaminants warranting human health and ecological concern. The concentrations of all these contaminants decrease downstream from the headwaters near the S-3 Ponds suggesting that the primary release to surface water is related to this source. Since neutralization of the wastes in the S-3 Ponds and capping of this site, concentrations of contaminants in Bear Creek have diminished.

Table 3.14. Parameters for hardness-dependent aquatic screening criteria

Element	Acute screening value		Chronic screening value	
	a	b	a	b
Cadmium	1.128	-3.826	0.7852	-3.49
Chromium III	0.019	3.698	0.819	1.561
Copper	0.9422	-1.464	0.8545	-1.463
Lead	1.273	-1.48	1.273	-4.703
Nickel	0.846	3.3612	0.746	1.1645
Silver	1.72	-6.52	<i>b</i>	<i>b</i>
Zinc	0.8473	0.9604	0.8473	0.7614

^aCalculated by the following equation:

Screening Value = $\exp\{[a \times \ln(\text{Hardness})] + b\}$, where hardness is expressed as mg CaCO₃/L.

^bNo criterion set.

Given the extensive previous characterization of Bear Creek surface water by several agencies (Y-12 Plant; USGS; Roy F. Weston, Inc.; BNI; and TVA) and the general concordance of results, it seems likely that all contaminants of concern for surface waters have been identified. However, previous work has not addressed either the locations of high contaminant input to surface waters (e.g., active springs) or stormflow transport of contaminants from the source areas to the main stem of Bear Creek.

3.6 GROUNDWATER

As discussed in Sect. 1.5, the scope of the BCV OU 1 RI with respect to groundwater is to address the extent to which units of the BCV OU 1 may be contributing contaminants to groundwater. Migration of contaminants from the individual source areas is being addressed by the BCV OU 4 RI. It is the objective of this RI work plan to identify data gaps and completely characterize contaminant releases and potential releases at each site.

3.6.1 Existing Well Information and Descriptions

Information on location and construction details for the wells discussed in this document is available in *Updated Subsurface Data Base for Bear Creek Valley, Chestnut Ridge, and Parts of Bethel Valley on the U.S. Department of Energy Oak Ridge Reservation* (Jones et al. 1992). The locations of current groundwater monitoring wells in BCV are shown in Fig. 3.17.

3.6.2 Bear Creek Hydrologic Regime

Water level data are collected, at a minimum, each time monitoring wells are sampled as part of the Y-12 Plant compliance program and semiannual water level measurements independent of sampling events were initiated in August 1992 on over 600 wells in BCV (Fig. 3.17). The results of water level measurements throughout BCV are summarized in the annual GWQAR (HSW 1992). Recent results reveal a groundwater piezometric surface (Fig. 3.18) sloping in from recharge areas on ridges to discharge along Bear Creek. The piezometric surface suggests a groundwater divide just east of the S-3 Ponds area as the eastern boundary of the BCHR.

The groundwater system underlying the BCHR generally consists of three parts: the stormflow zone (as discussed in Sect. 3.5.3), vadose zone, and groundwater zone. The groundwater zone consists of an upper saturated zone comprised of unconsolidated materials and shallow bedrock and a lower saturated zone consisting of the six formations of the Conasauga Group. For practical purposes, the upper and lower zones can be considered a single aquifer of relatively low water transmitting capacity, with the exception of transmission in conduits. The occurrence and movement of groundwater in the bedrock are closely related to the presence of bedding planes, joints, fractures, and solution cavities. In the shale formations, bedding planes and bedding plane fractures have resulted in preferential flow paths along strike. Flow perpendicular to bedding is retarded by the low permeability of the shale and is largely limited to areas where joints or fractures cut across bedding.

Depth to groundwater in the BCHR varies both spatially and temporally. The configuration of the water table generally is a subdued replica of the surface topography, being at high elevations beneath hills and at lower elevations in the valley bottoms. In general, elevations of the water table are lowest from October to December and highest from January to March.

An upward hydraulic gradient generally is present in the shale formations of the Conasauga Group throughout the BCHR at depths ranging from 14 to 152 m (45 to 500 ft). Upward gradients range from 0.003 to 0.129; gradients at individual sites in BCV OU 1 are described in the following sections. At shallow depths, the vertical gradients may be reversed at times in response to recharge events (Geraghty and Miller 1986). Recharge to the shallow groundwater system occurs through the surface veneer of unconsolidated materials.

LEGEND:

- BEDROCK WELLS < 200'
- ABANDONED BEDROCK WELLS < 200'
- UNCONSOLIDATED WELLS
- ABANDONED UNCONSOLIDATED WELLS
- ▲ BEDROCK WELLS > 200'
- * 1993 COMPLIANCE SAMPLING LOCATIONS

0 500 1000 2000
SCALE: 1" = 1000'

BEAR CREEK VALLEY OPERABLE UNIT 1 RI WORK PLAN

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XREFERENCES	PLOT FILES
93044/XREF/BASEMAP	93044/PLOT/342BC1-7
93044/XREF/MANGAREA	
93044/XREF/OU4WELLS	

93044\DWGS\342BC1-7 0
FIGURE # CAD FILE # ROT ANGLE

the active flow system in Upper BCV (Source: HSW 1992).

Fig. 3.18. Groundwater elevations in the un

LEGEND:

SCALE: 1" = 1100'

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RI WORK PLAN

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93044/XREF/BASEMAP	93044/PLOT/342BC1-2
93044/XREF/MANGAREA	
93044/XREF/GWCONT	
93044/XREF/OJ4WELLS	

93044\DWGS\342BC1-2

0

FIGURE 14

CAD FILE #

ROT ANGLE

consolidated aquifer in BCV (Source: HSW 1992).

3.6.2.1 Flow direction and rate

Groundwater movement in the BCHR is influenced locally by the surface drainage system and its associated solution cavities. Figure 3.19 illustrates the general pattern of groundwater flow.

The direction of groundwater movement generally is controlled by the potentiometric gradient and the permeability of the medium. In a completely isotropic aquifer, the flow direction of groundwater should be normal to the equipotential contours of the aquifer. However, in an anisotropic aquifer such as the Conasauga, the local flow directions are strongly controlled by the orientation of the bedding planes, fractures, and joints as well as by the distribution of solution cavities. Thus, a map of potentiometric contours based on observed head differences in wells in BCV indicates only the general direction of groundwater flow.

Figure 3.18 is a contour map of the water table at the BCHR based on water levels recorded in wells in 1991 (HSW 1992). In the Maynardville Limestone underlying Bear Creek, the direction of groundwater flow is southwest along the valley axis. The water table map is representative of the general water table configuration observed over the past several years (Geraghty and Miller 1986, 1987, 1988, 1989; HSW 1991, 1992). Although the water table fluctuates throughout the year, the pattern of fluctuation is similar in all wells, so that the water table slope remains essentially unchanged. No reversals of horizontal hydraulic gradient have been observed within the BCHR.

The average rate of groundwater flow in the Nolichucky Shale and the Maryville Limestone ranges from ~43 to 762 m/year (140 to 2500 ft/year) (Geraghty and Miller 1989). An average flow rate in the Maynardville Limestone was calculated at 4 m/d (12.8 ft/d); however, a tracer test in the Maynardville Limestone in the Bear Creek drainage basin yielded a rate of travel of ~61 m/d (200 ft/d) through structural/solution features to a spring (Geraghty and Miller 1989). Poor estimates of groundwater flow rate are to be expected when the Darcy equation for groundwater flow is applied to an anisotropic aquifer like this one. The assumptions made for using this equation, such as that groundwater flow paths are parallel to hydraulic gradient, may not apply and, therefore, may give misleading results.

S-3 Ponds. The groundwater divide between BCHR and UEFPC Hydrologic Regime is located just east of the S-3 Site (Fig. 3.20). North of the S-3 Ponds, groundwater movement is toward the south and southeast. South of the ponds, groundwater movement is to the southwest and northeast along the valley axis, defining the groundwater divide. The groundwater contours in Fig. 3.20 are representative of the general groundwater table observed over the past few years (HSW 1991, 1992; Geraghty and Miller 1986, 1987, 1988, 1989). A water table mound that was evident below the S-3 Ponds while the ponds were accepting liquid wastes has diminished in size and is now almost indistinguishable. During the time this groundwater mound was present, the groundwater divide passed under the S-3 Ponds; this is evidenced by the nitrate plume in groundwater below the S-3 Ponds that is in groundwater in both hydrological regimes (Turner et al. 1991).

The range of groundwater levels recorded in wells at this site is shown in Fig. 3.21 in relation to the waste units at the S-3 Site. From these data it is apparent that some of the waste materials in the former S-3 Ponds are located below the groundwater table.

Fig. 3.19. Generalized cross-section of BCV showing inferred groundwater flow paths.

Fig. 3.20. Groundwater elevations in the unco

olidated aquifer at the S-3 Site (Source: HSW 1992).

Fig. 3.21. Cross-section A-A' at the S-3 Site showing groundwater elevations in relation to the waste unit.

Data obtained from monitoring well clusters at the S-3 Ponds indicate upward hydraulic gradients in the shale formations of the Conasauga Group of between 0.003 and 0.129 (Geraghty and Miller 1986). Water level data from the Maynardville Limestone indicate downward flow in this formation, with gradients ranging from 0.008 to 0.124.

Oil Landfarm, SL 1, and Boneyard/Burnyard. The water table contours generally indicate southerly and southeasterly directions of groundwater movement at the Oil Landfarm (Fig. 3.22). Data obtained in October 1986 from monitor well clusters at the Oil Landfarm indicated upward gradients ranging from 0.01 to 0.25. Water level data from monitor well clusters in the Maynardville Limestone indicate downward flow components to depths as great as 61 m (200 ft). The calculated downward gradients range from 0.008 to 0.124 (Geraghty and Miller 1989). The downward flow may indicate that the solution cavities common to this unit function as an effective sink for the aquifer system. Horizontal gradients at the Oil Landfarm range from 0.03 to 0.07.

The range of groundwater levels recorded in wells at this site is shown in Fig. 3.23 in relation to the waste unit and the cap at the Oil Landfarm.

Burial Grounds. The groundwater table contour map for the Burial Grounds (Fig. 3.24) indicates southerly and southeasterly directions of groundwater movement northwest of Bear Creek. In the zone underlying Bear Creek, the direction of groundwater flow is southwest along the valley axis. Data obtained in July 1986 from monitor well clusters at the Burial Grounds indicated upward hydraulic gradients in the Nolichucky Shale ranging from 0.06 to 0.14. Upward gradients were also recorded in the Maynardville Limestone at the Burial Grounds ranging from 0.14 to 0.24. Horizontal hydraulic gradients in the Burial Grounds range from 0.03 to 0.1. In general, hydraulic gradients were steeper in areas with steep topographic slopes and also near Bear Creek.

The range of groundwater levels recorded in wells at this site is shown in Figs. 3.25 and 3.26 in relation to the waste units at the Burial Grounds. From these data it is apparent that the bottom of some of the trenches in BG-A and BG-C is located below the groundwater table.

3.6.2.2 Aquifer characteristics

The hydraulic properties of the geologic units that exert the greatest influence on contaminant transport at BCV OU 1 have been quantified by field tests. Hydraulic conductivity tests have been conducted in the unconsolidated zone and bedrock zone wells at several sites in BCV. In the unconsolidated zone, these values commonly range from 3.4 to 91.4 m/year (11 to 300 ft/year) (BNI 1983, 1984). In shallow bedrock, the values range from 0.3 to 12 m/year (1 to 40 ft/year) in the Nolichucky Shale and 274 to 1280 m/year (900 to 4200 ft/year) in the Maynardville Limestone. At the S-3 Site, hydraulic conductivity values range from 30 to 232 m/year (98 to 760 ft/year) in the unconsolidated zone and 3 to 12 m/year (10 to 40 ft/year) in the shallow bedrock zone (Geraghty and Miller 1987). Results of tests conducted in deep bedrock well were far lower, with the lowest being 0.01 m/year (0.02 ft/year), indicating a significant decrease in conductivity with depth (Geraghty and Miller 1987).

Mean values for hydraulic conductivity in the Nolichucky Shale and the Maryville Limestone, estimated using the USGS MODFLOW groundwater flow model, range from 0.1

to 0.5 m/d (0.4 to 1.7 ft/d) in the Nolichucky Shale and from 0.06 to 0.2 m/d (0.2 to 0.7 ft/d) in the Maryville Limestone (Geraghty and Miller 1989). Effective porosities ranging from ~ 1 to 2% were determined in all samples from the Pumpkin Valley Shale in nearby Melton Valley (deLaguna et al. 1963) and are assumed to be representative of the remainder of the formations in the Conasauga Group, except the Maynardville Limestone.

Hydraulic characteristics of the Conasauga Group, including transmissivity and storage coefficient, have been calculated from data obtained from aquifer pumping tests. The anisotropy of the aquifer was confirmed by these tests, each being characterized by elliptical water level cones of depression, elongate in the direction of the strike of strata, which is indicative of a higher degree of transmissivity in that direction.

The transmissivity of the Nolichucky Shale has been determined by pumping tests performed near the Burial Grounds and at the S-3 Site. At the Burial Grounds test site, transmissivity values determined by Law (1983) from wells open to unconsolidated material or bedrock averaged 990 L per day per m (Lpd/m) [260 gal per day per ft (gpd/ft)]. Storage coefficients determined from the test data were about 3×10^{-3} , indicative of semiconfined conditions. The magnitude of the anisotropy was calculated by Smith and Vaughn (1985) as approximately three times more transmissive along strike. An aquifer pump test with tracers was conducted at the same time which agreed with these results (Lozier et al. 1987). Test results also indicated a longitudinal/transverse transmissivity ratio of 20 to 30, a longitudinal dispersivity of ~ 3 m (10 ft), and an effective transport porosity on the order of 4×10^{-3} .

Pumping tests in the Nolichucky Shale at the S-3 Site produced similar estimates of aquifer properties. Transmissivity values determined in wells open only to bedrock ranged from 380 to 440 Lpd/m (100 to 115 gpd/ft), and storage coefficients ranged from 1×10^{-4} to 4×10^{-4} (Geraghty and Miller 1986). Each of the tests in the Nolichucky Shale was characterized by a low discharge rate, 19 L/min (5 gpm), with considerable water level drawdown in the pumped well.

3.6.2.3 Recharge/discharge relationships

Groundwater recharge in the BCV area is from local precipitation, which averages 137.5 cm/year (55 in./year). Recharge to the bedrock zone is most effective in places where the unconsolidated zone is thin or very permeable. This generally is the case on Pine Ridge, which is believed to be a major recharge area for the formations of the Conasauga Group in the BCV. Short losing sections of Bear Creek also may act locally to recharge the aquifer.

Natural groundwater discharge in the BCV is through streams, springs, and evapotranspiration. Bear Creek, its tributaries, and the solution cavity system along Bear Creek are the major discharge areas for shallow and intermediate depth groundwater moving through the interbedded strata of the BCV. Groundwater discharges sustain the base flow of Bear Creek (Geraghty and Miller 1985).

3.6.3 Groundwater Contamination

Groundwater in the BCHR has been affected by subsurface contamination transported from the Burial Grounds, the Oil Landfarm, and the S-3 Site. These sources have contributed

Fig. 3.22. Groundwater elevations in the uncon

LEGEND:

- UNCONSOLIDATED WELLS
 ■ ABANDONED UNCONSOLIDATED WELLS

— 880 — GROUND WATER CONTOURS

GROUND WATER CONTOUR DATA
 FROM APRIL 29 TO MAY 3, 1991
 COUNTOUR INTERVAL = 10 FT

CLAY CAPPED DISPOSAL UNIT

RCRA CAPPED DISPOSAL UNIT

UNCAPPED DISPOSAL UNIT

SCALE: 1" = 400'

**BEAR CREEK VALLEY
 OPERABLE UNIT 1
 RI WORK PLAN**

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93044/XREF/ASEMAP	93044/PLOT/3428C1-6
93044/XREF/MANGAREA	
93044/XREF/OU4WELLS	
93044/XREF/GWCONT	

FIGURE # 93044\DWG\3428C1-6 0
 CAD FILE # ROT ANGLE

Fig. 3.23. Cross-section B-B' at the Oil Landfarm showing groundwater elevations in relation to the waste unit.

Fig. 3.24. Groundwater elevations in the upper portion of the site.

LEGEND:

- UNCONSOLIDATED WELLS
 ■ ABANDONED UNCONSOLIDATED WELLS

— 880 — GROUND WATER CONTOURS

GROUND WATER CONTOUR DATA FROM APRIL 29 TO MAY 3, 1991
 CONTOUR INTERVAL = 10 FT

CLAY CAPPED DISPOSAL UNIT

RCRA CAPPED DISPOSAL UNIT

UNCAPPED DISPOSAL UNIT

SCALE: 1" = 400'

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93044/XREF/ASEMAP	93044/PLOT/342BC1-1
93044/XREF/MANGAREA	
93044/XREF/GWCONT	
93044/XREF/GW4WELLS	

FIG. 3.23 93044\DWG\342BC1-1 0
 FIGURE # CAD FILE # ROT. ANGLE

consolidated aquifer at the Burial Grounds (Source: HSW 1992).

Fig. 3.25. Cross-section C-C' at the Burial Grounds showing groundwater elevations in relation to the waste unit.

Fig. 3.26. Cross-section D-D' at the Burial Grounds showing groundwater elevations in relation to the waste unit.

organic and inorganic contaminants to groundwater; however, the relative contribution from each unit is not well known. Since actions were taken to isolate wastes from groundwater and surface water, extensive groundwater sampling has been carried out. Numerous reports characterizing groundwater in BCHR have been prepared as part of the GWQARs for the Upper Bear Creek Hydrogeologic Regime at the Y-12 Plant (Geraghty and Miller 1986, 1987, 1988, 1989, 1990; HSW 1991, 1992). These reports specify the locations and construction details for a number of monitoring wells in Bear Creek watershed. Extensive discussions and presentations of groundwater quality data are contained in the BCV OU 4 RI work plan. The following sections contain a summary of groundwater contamination in BCHR. For detailed information concerning groundwater quality, refer to the GWQARs just mentioned.

3.6.3.1 Organic contamination

VOCs are the primary organic contaminants of concern detected in groundwater in the BCHR. Dissolved VOCs are present in groundwater at all sites in BCV OU 1 in the shallow, intermediate, and deep bedrock zones of the underlying Nolichucky Shale. Total VOC concentrations in groundwater underlying the Burial Grounds are among the highest reported in BCHR. At this site they are most widespread in the unconsolidated shallow aquifer, underlying the whole Burial Grounds waste area (Fig. 3.27). The concentrations of VOCs are highest below BG-A, with the maximum value in 1990 of 7556 $\mu\text{g/L}$ (HSW 1991). The VOC plume in the bedrock aquifer below the Burial Grounds is restricted to an area below the BG-A South, and the aerial extent of contamination decreases with increasing depth in the aquifer to a minimum of 152 m (500 ft). VOCs commonly detected in groundwater are tetrachloroethene, trichloroethene, trans-1,2-dichloroethene, and 1,1-dichloroethene. Vinyl chloride, 1,1-dichloroethylene, and 1,1,1-trichloroethane have also been detected in samples from wells containing the highest summed VOC concentrations.

A VOC plume is present in the unconsolidated aquifer underlying most of the Oil Landfarm. Concentrations of total VOCs in groundwater below the Oil Landfarm range up to 1000 $\mu\text{g/L}$ in shallow bedrock. At shallow and intermediate bedrock depths an apparently continuous dissolved VOC plume extends west from the S-3 Site for about 1829 m (6000 ft) to just west of the Oil Landfarm. The distribution of this plume correlates with the lowermost members of the Maynardville Limestone (HSW 1992). Inputs to this plume have probably come from the S-3 Site, the Oil Landfarm, other sites in the former S-3 Waste Management Area, and sites in BCV OU 2.

Accumulations of dense nonaqueous phase liquids (DNAPLs) were discovered below the southern border of BG-A. Subsequent to this discovery, an investigation was initiated to obtain information on the mode of occurrence and distributions of DNAPLs in groundwater below the Burial Grounds (Haase and King 1990). The DNAPLs consist of free-phase tetrachloroethane, trichloroethene, and 1,1,1-trichloroethane with concentrations of PCBs ranging from 14,000 to 27,000 $\mu\text{g/L}$.

Due to the upward hydraulic gradient that exists in the Nolichucky Shale (and some sections of the Maynardville Limestone) it is unlikely that dissolved VOCs have migrated downward through this formation. VOCs in groundwater can migrate via two pathways: as free-phase DNAPLs or as dissolved VOCs. Free-phase DNAPLs will tend to migrate downward and are thought to be relatively immobile in the horizontal plane. The presence

Fig. 3.27. Distribution of VOCs in groundwater in the BCHR (Source: HSW 1992).

of DNAPLs has been confirmed in the bedrock aquifer below the Burial Grounds and DNAPLs may be present below the Oil Landfarm; however monitoring activities here have not encountered any free-phase product.

DNAPLs that migrate through a geologic medium will leave behind a trail of residual DNAPL (i.e., nonmoving free-phase hydrocarbons). Consequently, the DNAPL column acts as a long-term source of dissolved VOCs to groundwater passing through the column. A trail of residual DNAPL that may be leached to groundwater in the future probably remains in the vadose zone below the Burial Grounds, the Oil Landfarm, and the S-3 Ponds.

3.6.3.2 Inorganic contamination

The primary source of inorganic groundwater contamination appears to be the S-3 Site. Based on discussions of the surface water data presented in Sect. 3.5.2, it is apparent that groundwater discharge to the headwaters reach of Bear Creek is an important part of the water balance and contaminant input for the stream in this region. Monitoring results published in GWQARs (HSW 1992) have demonstrated that shallow groundwater in this region contains significant contaminants whose source was the S-3 Ponds (Table 3.15). Dye tracer tests (Geraghty and Miller 1989) have documented the interaction between the surface water and active groundwater flow system.

Of the range of constituents found in the S-3 Ponds, the nitrate ion is the most mobile component and should be transported conservatively in both surface and subsurface environments. Consequently, it constitutes an important tracer with which to monitor contaminant migration from the source region. Figure 3.28 is a strike parallel section that illustrates the down-valley transport of contaminants (represented by nitrate) from the S-3 Ponds source area. The nitrate plume appears to be concentrated <61 m (200 ft) below the ground surface throughout most of the valley. Contaminants may have penetrated deeper near the source due to the dense (high salinity) nature of effluent discharged to the ponds.

Geraghty and Miller (1987) report that discharge of contaminated groundwater to the headwaters reach of Bear Creek has resulted in observed nitrate concentrations ranging between 443 and 1950 mg/L nitrate-N for samples collected between 1984 and 1986 at a stream sampling station ~610 m (2000 ft) downstream from the S-3 Ponds. Recent dye tracer studies have revealed that water from losing reaches of Bear Creek between BCK 11.64 and 9.41 reappears in spring SS-5 (Geraghty and Miller 1989). Recent results from sampling at springs in BCV (HSW 1992) suggest that spring SS-5 may have nitrate concentrations elevated above background (NT-11; see Table 3.12).

Total concentrations of arsenic, barium, cadmium, chromium, lead, mercury, and selenium exceed their respective MCLs or proposed MCLs in groundwater samples from one or more wells in the BCHR. With the exception of elevated concentrations of barium, cadmium, chromium, lead, and mercury in groundwater near the S-3 Ponds, the occurrence of elevated metal concentration is sporadic and could be the result of typical problems associated with sampling low-permeability formations (i.e., large amounts of suspended materials resulting from purging the well). The S-3 Ponds have been identified as the most likely source for metals in groundwater at the headwaters of Bear Creek (HSW 1992).

Table 3.15. Nitrate, annual mean gross alpha activity, and gross beta activity exceeding MCLs, 1990 (After HSW 1992)

Well number	Aquifer zone ^a	Nitrate (as N) ^b (mg/L)	Gross alpha activity ^c (pCi/L)	Gross beta activity ^c (pCi/L)
<i>S-3 Site</i>				
GW-100	U	388	<MCL ^d	<TL ^e
GW-101	U	2,200	96 ± 49	<TL
GW-122	I	770	<MCL	<TL
GW-124	I	160	<MCL	556 ± 15
GW-125	D	28	<MCL	<TL
GW-127	U	<MCL	267 ± 35	308 ± 46
GW-236	U	421	19 ± 3	203 ± 9
GW-243	S	6,230	8,205 ± 386	60,750 ± 1,345
GW-244	S	NA ^f	67 ± 75	3,240 ± 290
GW-245	S	4,545	79 ± 102 ^g	2,765 ± 270
GW-246	S	3,530	143 ± 86	13,900 ± 540
GW-247	S	7,380	56 ± 64	4,240 ± 370
GW-276	U	454	971 ± 160	4,110 ± 320
GW-277	S	1,690	<MCL	2,900 ± 290
GW-345	U	398	<MCL	244 ± 6
GW-346	S	811	<MCL	<TL
GW-348	S	10	<MCL	<TL
GW-526	I	845	21 ± 34	<TL
GW-615	I	9,690	307 ± 79	554 ± 173
GW-616	I	274	22 ± 41	<TL
<i>Spoil Area I</i>				
GW-314	S	10	<MCL	<TL
GW-315	S	16	<MCL	55 ± 3
<i>Rust Spoil Area</i>				
GW-306	S	11	<MCL	<TL
GW-307	U	18	<MCL	<TL
GW-308	U	15	<MCL	<TL
GW-309	S	18	<MCL	50 ± 4
GW-310	U	23	<MCL	62 ± 3
<i>Oil Landfarm</i>				

Well number	Aquifer zone ^a	Nitrate (as N) ^b (mg/L)	Gross alpha activity ^c (pCi/L)	Gross beta activity ^c (pCi/L)
GW-005	U	BDL ^h	31 ± 2	<TL
GW-064	S	10	<MCL	<TL
GW-085	S	185	<MCL	57 ± 3
GW-087	U	BDL	54 ± 5	<TL
GW-225	I	97	<MCL	<TL
GW-226	S	<MCL	20 ± 2	<TL
GW-227	S	17	25 ± 2	69 ± 3
GW-228	I	37	<MCL	<TL
GW-369	I	11	<MCL	<TL
GW-537	U	819	40 ± 16	354 ± 48
GW-601	D	29	<MCL	<TL

Well number	Aquifer zone ^a	Nitrate (as N) ^b (mg/L)	Gross alpha activity ^c (pCi/L)	Gross beta activity ^c (pCi/L)
<i>Bear Creek Burial Grounds</i>				
GW-052	U	28	40 ± 4	120 ± 8
GW-061	S	11	<MCL	<TL
GW-621	S	12	<MCL	<TL

^aU = unconsolidated zone; I = intermediate depth bedrock zone (85–300 ft); D = deep bedrock zone (>300 ft); and S = shallow bedrock zone (Unc–85 ft).

^bmg/L = milligrams per liter.

^cpCi/L = picocuries per liter ± counting error.

^dMCL = maximum contaminant level.

^e<TL = results less than 50 pCi/L threshold limit.

^fNA = not analyzed.

^gCalculated mean activity including 1989 data.

^hBDL = below detection limit.

Moore and Toran (1992) investigated the possibility that, following the closure of the S-3 Ponds, the shallow groundwater and surface water system near the Bear Creek headwaters would be naturally flushed of contaminants and show a temporal decrease in contaminant concentrations. They used the groundwater monitoring data from two water table wells near the S-3 Ponds to illustrate an exponential decrease in nitrate concentrations beginning in mid-1987 (Fig. 3.29). Data from other monitoring wells do not support the natural flushing hypothesis. However, Moore and Toran suggest that purging associated with the sampling process in Nolichucky wells may obscure any trends in data because the water sampled represents a mix of water from fractions that are influenced differently by the matrix diffusion process.

The natural flushing hypothesis is supported by historical data from a headwater stream location (BCK 11.97) where Moore and Toran (1992) have recorded decreasing concentrations of major elements (e.g., aluminum, calcium, magnesium) (Fig. 3.30) that may be associated with acid leaching of the rocks as well as decreasing concentrations of constituents associated with the waste disposal (nitrate, uranium).

Additional support for the hypothesis is found in the historical resurgence of fish populations within Bear Creek (Southworth et al. 1992). Their analysis of data from 1984 through 1988 showed that fish populations were returning to the headwaters of Bear Creek after a long absence. Their data for macroinvertebrate populations in the stream bed did not show a similar trend through 1988.

3.6.3.3 Radionuclides

The radiochemical quality of groundwater was evaluated with respect to the MCL for gross alpha activity (15 pCi/L) and the threshold value (50 pCi/L) for gross beta activity. A plume of elevated gross alpha and beta exists in groundwater below the S-3 Ponds (HSW 1992). The plumes are located in both the shallow unconsolidated and the bedrock groundwater aquifers; however, the plumes are more extensive in the unconsolidated aquifer.

Fig. 3.28. Nitrate in groundwater in the Maynardville Limestone along the axis of BCV (Source: HSW 1992).

Fig. 3.29. Average normalized concentrations for nitrate as nitrogen in GW-106 and GW-276 (Source: Moore and Toran 1992).

Fig. 3.30. Average normalized concentrations for major constituents in Bear Creek (Source: Moore and Toran 1992).

Gross alpha activity is most extensive in the unconsolidated aquifer zone near the S-3 Site (19–971 pCi/L), and isolated pockets of elevated alpha activity are found in the unconsolidated zone at the Oil Landfarm (31–54 pCi/L). A maximum gross alpha value of 8205 ± 386 pCi/L was detected in 1990 in shallow bedrock well GW-243 located adjacent to the S-3 Site. Gross alpha activity in the shallow bedrock zone below the S-3 Ponds was extensive and most wells within 305 m (1000 ft) downstream of the ponds contain elevated gross alpha counts (10 to 8205 pCi/L). A second plume of elevated gross alpha activity in the shallow bedrock was detected in the east central portion of the Oil Landfarm that possibly extends further east (11 to 18 pCi/L). A plume of elevated gross alpha activity has been detected in the intermediate bedrock aquifer but is less extensive than that in the shallow aquifer. Chemical data from 1989 suggest that the primary alpha emitting radionuclide is likely to be uranium; however, isotopes of radium neptunium and americium may also contribute to the gross alpha activity (HSW 1992).

The gross beta activity in groundwater from 19 wells exceeded the 50 pCi/L threshold value and is usually associated with ^{99}Tc . Contamination was detected in plumes below the S-3 Ponds and the Oil Landfarm. The highest mean value for beta activity was reported from a shallow bedrock well located near the S-3 Ponds ($60,750 \pm 1345$ pCi/L). Similarly to the gross alpha activities, plumes of elevated gross beta activity exhibited the highest counts and were most extensive in the shallow bedrock zone below the S-3 Ponds (2900 to 60,750 pCi/L). Gross beta activities were elevated in the unconsolidated zone and intermediate bedrock aquifers below the S-3 Ponds. Elevated beta activities were also detected in three wells at the Oil Landfarm.

3.6.4 Conclusions

Extensive groundwater sampling throughout BCV has revealed contaminant plumes of VOCs emanating from each of the source units in BCV OU 1 and a plume of inorganic contaminants, characterized by nitrate, that emanates from the S-3 Ponds near the headwaters of Bear Creek and underlies Bear Creek for considerable distance downstream.

Free-phase DNAPLs have been detected in the intermediate and shallow groundwater bedrock aquifers. Dissolution of residual DNAPL in the bedrock and unconsolidated zone aquifers is suspected as the major source of dissolved VOCs in groundwater in the BCHR. Dissolved VOCs were detected in high concentrations in groundwater from the shallow and intermediate aquifers. Particularly high concentrations of dissolved VOCs were found in water from the unconsolidated aquifer below the Burial Grounds and the Oil Landfarm.

Detectable concentrations (suggestive of contamination) of nitrate, a relatively mobile anion, are found down valley and generally define the maximum horizontal extent of a contamination plume that extends westward from the S-3 Ponds. Radionuclide contamination associated with this nitrate plume is restricted to within 305 m (1000 ft) of the S-3 Ponds, reflecting the retardation effect of chemical sorption reactions between radionuclides and geologic media. Radionuclide groundwater contamination is also associated with the Oil Landfarm but is not as extensive as that found at the S-3 Ponds. Most contaminant source areas overlie Nolichucky soils and bedrock. Migration of some contaminants (e.g., metals) within the Nolichucky appears to be retarded. Down-valley transport of contaminant plumes is apparently controlled by more active conduit flow within the Maynardville and interactions between Maynardville groundwater and Bear Creek surface water.

Preliminary analyses of historical groundwater and surface water quality data suggest that source area controls (e.g., RCRA caps) may be producing measurable decreases in contaminant concentrations. These source area controls contribute to the decrease in contaminant concentrations by the halting of waste disposal and neutralizing of the sludges. Verification of these results will require additional analysis of groundwater and surface water quality data from monitoring points that are demonstrably connected to the actively flowing portion of the groundwater system.

3.7 ECOLOGY

BCV OU 1 lies outside the main Y-12 Plant complex and as such is in close proximity to areas of potential ecological sensitivity. Ecological studies of the Bear Creek watershed were initiated in May 1984 and are continuing at the present time. A summary of the results of the sampling conducted from May 1984 through early 1990 is presented in Southworth et al. (1992). These studies consisted of an initial detailed characterization of the benthic invertebrate and fish communities in Bear Creek, followed by an ongoing monitoring phase (Fig. 3.31) with reduced sampling intensities. The characterization phase involved two approaches: (1) in-stream sampling of benthic invertebrate and fish communities in Bear Creek to identify spatial and temporal patterns in distribution and abundance and (2) laboratory bioassays of water samples from Bear Creek and selected tributaries to identify potential sources of toxicity to biota. The monitoring phase of the ecological program relates to the long-term goal of identifying and prioritizing contaminant sources and assessing the effectiveness of remedial actions. Activities of the characterization phase are continuing at less frequent intervals.

Before 1940, most of BCV was cleared and used for agriculture (Southworth et al. 1992). Most of the agricultural land has been planted in pines, with a strip of pines and mixed hardwoods bordering much of the stream. About 65% of the BCV watershed is wooded (McMaster 1967).

3.7.1 Flora

Vegetation in the Bear Creek watershed is predominantly oak, oak-hickory associations on the upper slopes and ridgetops, and planted pine along the creek and floodplain area. Pine Ridge is generally an oak, oak-hickory association with chestnut oak, red oak, tulip poplar, white oak, and hickory as the dominant species. The lower areas along Bear Creek and Bear Creek Road are mainly planted loblolly pine with some scattered areas of planted shortleaf, natural pine, and oak-hickory communities. Also of interest is a large area of big bluestem grass that grows along Bear Creek Road. Large stands of this species are uncommon in east Tennessee. Plant communities in Bear Creek are depicted in Fig. 3.32 as components of the general plant communities found on the ORR (Welch 1989). This site is being considered as a National Environmental Research Park (NERP) Reference Area. More detailed information is available for portions of the Bear Creek watershed area on forest compartment maps and in various documents (Bradburn and Rosenbalm 1984; Parr and Pounds 1987). Table 3.16 lists threatened and endangered flora and fauna found within the ORR.

Fig. 3.31. Locations of existing biological monitoring sites.

General Vegetative Cover

LEGEND

PINE/PINE-HARDWOOD FORESTS	BOTTOMLAND HARDWOOD FOREST
CEDAR, CEDAR-PINE, AND CEDAR-HARDWOOD FOREST	NORTHERN HARDWOOD FOREST
OAK-HICKORY FOREST	GRASSLAND, DEVEGETATED AREAS, AND CULTURAL FEATURES

Fig. 3.32. General vegetation cover types on the ORR (Source: DOE 1980).

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**Table 3.16. Threatened and endangered plant and animal species on the ORR
(Tennessee Specific)**

Common name	Scientific name
<i>Flora</i>	
Spreading false-foxglove	<i>Aureolaria patula</i>
Appalachian bugbane	<i>Cimicifuga rubifolia</i>
Pink lady-slipper	<i>Cypripedium acuale</i>
Tall larkspur	<i>Delphinium exaltatum</i>
Northern bush-honeysuckle	<i>Diervilla lonicera</i>
Mountain witch-alder	<i>Fothergilla major</i>
Golden seal	<i>Hydrastis canadensis</i>
Butternut	<i>Juglans cinerea</i>
Canada lily	<i>Lilium canadense</i>
Fen orchid	<i>Liparis loeselii</i>
Ginseng	<i>Panax quinquefolius</i>
Tuberculed rein-orchid	<i>Platanthera flava</i> v. <i>herbiola</i>
Purple fringeless orchid	<i>Platanthera peramoena</i>
<i>Mammals</i>	
River otter	<i>Lutra canadensis</i>
Gray bat	<i>Myotis grisescens</i>
Indiana bat	<i>Myotis sodalis</i>
<i>Birds</i>	
Cooper's hawk	<i>Accipiter cooperii</i>
Sharp-shinned hawk	<i>Accipiter striatus</i>
Bachman's sparrow	<i>Aimophila aestivalis</i>
Grasshopper sparrow	<i>Ammodramus savannarum</i>
Northern harrier	<i>Circus cyaneus</i>
Peregrine falcon	<i>Falco peregrinus</i>
Bald eagle	<i>Haliaeetus leucocephalus</i>
Osprey	<i>Pandion haliaetus</i>
Red-cockaded woodpecker	<i>Picoides borealis</i>
Bewick's wren	<i>Thryomanes bewickii</i>
<i>Fish</i>	
Tennessee dace	<i>Phoxinus tennesseensis</i>

Source: Kroodsmma 1993.

3.7.2 Rare Plant Species

Two areas contain rare plant populations, and these have been designated NERP Natural Areas. The Canada lily (*Lilium canadense*) occurs in NERP Natural Area 13 at the base of Pine Ridge. It grows along the forested edges of the powerline right-of-way in the Bear Creek drainage area. This plant species is state listed as threatened. The other site, NERP Natural Area #4, contains the southern rein orchid (*Platanthera flava*). This is forested area with numerous wetlands around seeps and contains old stream or flood channels. An uncommon aquatic plant, *Orontium aquaticum*, occurs here also. The southern rein orchid is state listed as a species of special concern. The site has also been registered as a State Natural Area in an agreement between DOE and the TDEC (Howell 1986).

3.7.3 Terrestrial Fauna

BCV contains elements of the majority of wildlife habitat types found on the ORR (Fig. 3.33), and, therefore, the expected terrestrial fauna is that described for the entire reservation (Welch 1989).

Hardwood and mixed hardwood/conifer habitats are the most abundant habitat types in Bear Creek watershed, followed by pine plantation and grassland habitats, with considerable riparian habitat along the length of Bear Creek. Species commonly found in these habitats are listed below.

Hardwood and mixed hardwood/conifer habitats. In wooded areas the eastern gray and southern flying squirrels, southeastern shrew, eastern mole, short-tailed shrew, white-footed mouse, white-tailed deer, and eastern chipmunk may be found. Predators such as the bobcat and weasel occur here. Birds commonly found in forest areas include the yellow-shafted flicker, red-bellied woodpecker, hairy woodpecker, downy woodpecker, blue jay, Kentucky warbler, pine warbler, ovenbird, Carolina chickadee, tufted titmouse, and scarlet tanager. Hawks (red-shouldered, red-tailed, and broad-winged) are commonly found on the ORR, as are wild turkeys, which have been reintroduced to the area in recent years. Amphibians and reptiles found in the forest habitat include the dusky salamander, American toad, eastern box turtle, ground skunk, worm snake, black racer, rat snake, black king snake, milk snake, and copperhead.

Pine plantation habitat. Very early stages of planted pine areas contain species similar to those in old-field habitats, and the small mammals present are much the same as in the late stages of old fields. The populations tend to be smaller, though, because of less tree diversity. Pine plantations with a dense canopy and no understory are essentially barren of both small and large mammals except around the edges. As plantations are thinned and canopies opened, undergrowth develops and provides habitat for species similar to those found in early- to mid-stage hardwood-mixed hardwood/conifer forests. Avian species have a low preference for the pure pine areas bordering the transmission line corridors. Pine warblers and white-throated sparrows are common, but few other species are evident. These habitats are little used by reptiles or amphibians (Johnson 1964).

General Wildlife Habitats

LEGEND

HARDWOOD/MIXED HARDWOOD HABITAT	AQUATIC HABITAT
PINE PLANTATION HABITAT	CAVE HABITAT
OLD FIELD AND GRASSLAND HABITAT	BUILDING HABITAT

Fig. 3.33. General wildlife habitats on the ORR (Source: DOE 1980).

Old-field and grassland habitat. In the grassland/forb stage of vegetation, the principal species of small mammals are the southeastern shrew, least shrew, short-tailed shrew, eastern harvest mouse, hispid cotton rat, pine mouse, and the eastern cottontail rabbit. The eastern mole occurs in areas of loose soil. Closely mowed or grazed areas and dense kudzu growth is good habitat for the groundhog. Also found there are the striped skunk, coyote, red fox, and white-tailed deer. In more brushy vegetation, the white-footed mouse, golden mouse, and opossum may be found. Bird species found in this habitat include bobwhite, red-tailed hawk, field sparrow, towhee, blue grosbeak, meadowlark, and red-winged blackbird. The eastern bluebird population has increased with the establishment of bluebird nesting boxes. The yellow-breasted chat is found in old fields. Numerous frog, toad, lizard, and snake species are found in the old-field areas.

Aquatic and riparian habitats. Many reptiles and amphibians occur in the various aquatic and wetland areas, including turtles, queen snake, water snake, salamanders, and frogs. The muskrat and beaver are bound closely to aquatic habitats. Rice rats, mink, and raccoons are also found in these areas. Many large mammals come frequently to this habitat to drink, and various small species are present at the water's edge. The American bald eagle occurs occasionally as a transient. The Canada goose, great blue herons, and green-backed herons nest on the ORR.

Threatened and endangered species. Three mammal species on the federal list as endangered may occur on the ORR but have not been verified. These are the gray bat (*Myotis grisescens*), Indiana bat (*Myotis sodalis*), and eastern cougar (*Felis concolor*). Only one mammal species listed by the Tennessee Wildlife Resources Agency (TWRA) as in need of management, the southeastern shrew (*Sorex longirostris*), is known to occur on the reservation (Kroodsma 1987). The pine snake (*Pituophis melanoleucas*) is listed as threatened by the TWRA, and its presence in the Bear Creek watershed has been verified. The Cooper's hawk (*Accipiter striatus*) occurs in forests throughout the ORR and is listed by the state as threatened (Kroodsma 1987). Rare plants occurring on the Y-12 Plant site have been characterized by Hardy et al. (1992).

3.7.4 Aquatic Fauna

There have been 19 species of fish found in Bear Creek in recent quantitative monitoring efforts conducted at seven sites along virtually the entire length of Bear Creek (Southworth et al. 1992). Minnows (blacknose dace, *Rhinichthys atratulus*; Tennessee dace, *Phoxinotus tennesseensis*; stoneroller, *Camptostoma anomalum*; and creek chub, *Semotilus atromaculatus*) were the predominant fish species found upstream from the weir at BCK 4.55. Below the weir, larger species (northern hogsucker, *Hypentelium nigricans*; white sucker, *Catostomus commersoni*; and rockbass, *Ambloplites rupestris*) were more common, the diversity of minnow species increased, and darters were found. Conclusions of the recent fish-monitoring studies were that much of Bear Creek had limited fish fauna (low species richness) characterized by robust population parameters (high densities and biomass). The uppermost site (BCK 12.36) did not have a stable, resident fish population. Water from this site was commonly toxic to fathead minnow larvae in toxicity testing and contained high levels of dissolved salts as a result of input of contaminated groundwater from the S-3 Site. The next two monitoring sites downstream, BCK 11.83 and BCK 11.09, had low fish density and biomass in 1984 and 1985 but showed recovery in later sampling. No impacts on the fish of Bear Creek were evident in the vicinity of inputs from the Burial Grounds (BCK 9.91 and BCK 9.40) despite the fact that qualitative surveys found no fish in the tributaries (NT-6-8) draining that site.

The yellowfish madtom (*Noturus flavipinnis*), listed by the state as endangered, has been found in Bear Creek. The Tennessee dace, a major constituent of the fish population above the weir at BCK 4.55, is listed as a species in need of management. Its habitat is protected by the state of Tennessee (Starnes and Etnier 1980). In Bear Creek, it occurs at every site above the weir and in at least four tributaries (NT-13, NT-14, NT-18, and ST-7).

Quantitative sampling of benthic invertebrates was conducted monthly at nine sites from BCK 12.36 to BCK 3.25 during June 1984–May 1985 as part of Phase I of the ecological monitoring program for Bear Creek and at quarterly intervals thereafter (Southworth et al. 1992). A total of 126 distinguishable taxa were collected in Bear Creek, including crustaceans (Isopoda, Amphipoda, and Decapoda); aquatic worms (Oligochaeta); snails (Gastropoda); mussels (Pelecypoda); and insects (Insecta). There were 11 orders of insects collected in Bear Creek, including springtails (Collembola), mayflies (Ephemeroptera), dragonflies and damselflies (Odonata), crickets and grasshoppers (Orthoptera), stoneflies (Plecoptera), true bugs (Hemiptera), alderflies and fishflies (Megaloptera), caddisflies (Trichoptera), butterflies and moths (Lepidoptera), beetles (Coleoptera), and true flies (Diptera).

The invertebrate fauna of Bear Creek showed a pattern of increasing density, biomass, and taxonomic diversity and richness with increasing distance downstream from the uppermost sampling site (BCK 12.36) (Southworth et al. 1992). The paucity of benthic invertebrates found in the upper reaches of Bear Creek contrasted sharply with reference sites (unimpacted streams of similar size), which had relatively diverse and abundant assemblages of macroinvertebrates. Water from BCK 12.36 and BCK 11.83 was toxic to *Ceriodaphnia*, an invertebrate test species, but not to fathead minnow larvae, in toxicity tests run in March 1988. At the time of these tests, streamflow in Bear Creek was higher than normal, and contaminants from the S-3 Ponds groundwater plume were diluted more than would be the case under baseflow conditions. Under low-flow conditions, it is likely that Bear Creek water would be toxic to *Ceriodaphnia* as far downstream as BCK 9.4. While evidence of adverse effects on the fish communities of Bear Creek was not noted at sites downstream from BCK 11.83, the benthic fauna appeared to be more sensitive, with clear differences in faunal composition from unimpacted reference sites at all sites except BCK 3.25, where complete recovery appears to have occurred. Species intolerant of pollution (mayflies, stoneflies, and caddisflies) were absent in the upper reaches and became more common downstream. Mayflies, which are particularly sensitive to toxic metals, were virtually absent at all sites except BCK 3.25. Unlike the fish data, which provide evidence of ecological recovery in Bear Creek since 1984, the benthic macroinvertebrate fauna do not appear to have changed in a manner indicative of either improving or degrading water quality since 1984 (Southworth et al. 1992).

No threatened or endangered species of aquatic macroinvertebrates have been collected in Bear Creek.

A limited amount of data have been collected on the bioaccumulation of contaminants by aquatic biota in Bear Creek. Fish collected from lower Bear Creek contained elevated levels of mercury (0.2 to 0.5 $\mu\text{g/g}$, wet weight) in 1982 (Van Winkle et al. 1984) and in 1984 (TVA 1986). While well below the U.S. Food and Drug Administration (FDA) action level, these data do indicate the presence of biologically available mercury in the Bear Creek system. Fish from lower Bear Creek were also analyzed for cadmium, chromium, silver, arsenic, and nickel in 1984 (TVA 1986), and levels were found to be typical of fish from uncontaminated environments.

PCB contamination in fish from Bear Creek was evident in 1982, when rockbass collected from the lower reaches of the creek were found to contain 0.65 ± 0.29 ppm (wet weight) PCBs (McElhaney 1982). TVA analyzed fish from lower Bear Creek for organic priority pollutants, including PCBs in 1984; levels of all substances were below detection limits. Fish were collected from lower Bear Creek in 1987 as part of an effort to evaluate the importance (relative to other sources) of the UEFC discharge at New Hope Pond as a source of PCB contamination to lower Poplar Creek and the Clinch River. In this collection, PCBs in rockbass averaged 0.28 ± 0.12 ppm ($n=8$), a value comparable to the level of contamination found in sunfish in the lower third of EFPC and well above the level typical of fish from uncontaminated sites (0.02 ± 0.01 ppm). Clams held in Bear Creek (BCK 4.55) for 1 month as part of the same study accumulated 1.01 ppm PCBs (versus 0.05 ppm in controls), indicating that Bear Creek now contains a source of biologically available PCBs.

PCB contamination has been observed in fish in EFPC, Poplar Creek, and the Clinch River/Watts Bar Reservoir downstream from Bear Creek (TVA 1986). It is unlikely that a substantial fraction of this contamination is attributable to Bear Creek because sources of PCBs and similar levels of contamination are found in fish from these systems far upstream in EFPC from the mouth of Bear Creek (TVA 1986). A similar situation exists for mercury (TVA 1986).

The Biological Monitoring and Abatement Program (BMAP) studies have contributed a great deal of information on ecological impacts to Lower EFPC of ongoing releases from the Y-12 Plant, especially impacts on the aquatic biota. However, they were not designed to reveal effects of deposited sediments on either aquatic or terrestrial biota. In particular, the BMAP studies do not address the following points.

- Contribution of in-stream sediments and sediment pore water to contaminant burden in aquatic biota. BMAP sample locations were intentionally chosen in riffle areas to avoid the contributions of heavily contaminated soils and sediments. Although it is clear from BMAP results that Y-12 Plant releases are a significant ongoing source of contaminants to the fish and other biota, studies on the effects of in-stream sediments were less definitive.
- Contaminants in fish were determined using fillets rather than whole fish. This approach reflects a concern for human health. However, piscivorous birds and animals consume internal organs, skin, fat, and in some cases bone, as well as muscle tissue. These tissues may selectively accumulate lipophilic contaminants or bone-seeking inorganics and radionuclides. Therefore, to be relevant to an ecological risk assessment, analyses of fish and other animals sampled should be done on whole organisms.
- BMAP was designed to assess impacts of waterborne releases to Lower EFPC. It does not include studies of terrestrial biota, which are more likely to be impacted by contaminants in the floodplain soils and groundwater.

To address these data gaps, the following approaches are proposed. Details of the proposed studies are provided elsewhere in this report:

- Fish and other aquatic biota will be sampled at locations at which in-stream sediments are present in significant quantities to provide a potential to contribute to contaminant loading of the biota.

- Aquatic and terrestrial biota will be analyzed on a whole-body basis rather than using muscle or organ tissue only.
- Terrestrial biota will be sampled as necessary to provide a comprehensive description of the distribution of contaminants through the terrestrial ecosystem.

3.8 RADIOLOGICAL SURVEY

An aerial radiological survey was conducted in 1989 encompassing 440 km² (176 square miles) that included the entire ORR and adjacent areas. Typical background exposure rates were found to vary from 7 to 11 microroentgens per hour. Preliminary data collected in a 1992 aerial radiological survey corroborate the 1989 data and are currently being evaluated. Data from this and a more recent survey that included several geophysical measurements will be available for use during the RI.

3.9 SITE CONCEPTUAL MODEL

Site conceptual models have been developed to illustrate the potential pathways of contaminant migrations, or flows, among various abiotic media and, ultimately, to receptors. Chapter 1 outlines the scope of the risk assessment to be carried out in the BCV OU 1 RI. Ecological risk will be assessed for BCV OU 1 by determining the potential for lower members of the food chain (vegetation, earthworms, tadpoles, shrews) to uptake contaminants and by modeling the potential effects on the ecosystem. Before a site-wide ROD is issued for Bear Creek Valley, an integrated baseline risk assessment for Bear Creek Valley will be carried out, and at that time high members of the food chain will be addressed in more detail. The generalized pathways for a human receptor and for ecological receptors are shown in Figs. 3.34, 3.35, 3.36, and 3.37. Figure 3.38 provides the hydrogeological submodel for the BCV OU 1 conceptual model.

3.9.1 Soils and Sediments

Uranium, PCBs, VOCs and various metals have been identified in soils and tributary sediments at each of the units within BCV OU 1. In general, wastes were either placed in unlined pits or spread over the surface. Soil contamination resulted from mixing of wastes with soil, by leaching of contaminants from wastes into surrounding soils, or by deposition of contaminants at seeps where contaminated groundwater discharges. High levels of soil contamination were identified at the Burial Grounds and the Oil Landfarm; however, the aerial extent and depth of contamination at these sites are not well known. Distribution and level of contamination in soils at the Boneyard/Burnyard and the S-3 Ponds is not well defined.

Stream sediments in tributaries that drain the units in BCV OU 1 are derived from erosion of soils within each tributary's catchment area. The nature and extent of contaminants in stream sediments reflect that of the soils in the area and the fact that contaminants may precipitate or sorb to stream sediments because of changes in water chemistry at the point of discharge.

Fig. 3.34. Bear Creek Valley conceptual site model for human receptors.

Fig. 3.35. Bear Creek Valley OU 1 conceptual site model for human receptors.

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Fig. 3.36. Simplified food web for Bear Creek Valley OU 1 ecosystem.

Fig. 3.37. Bear Creek Valley ecological conceptual model.

Fig. 3.38. Site conceptual model for transport of contaminants within BCHR.

Some areas of high soil and sediment contamination have been excavated and are now capped with RCRA caps; these areas are BG-A, BG-B, and BG-C; ORP 1 and ORP 2; and the Oil Landfarm. The HCDA has been capped with a RCRA-style cap, and sediments in Tributary NT-7 downstream of ORP 1 and SL 1 have been capped with 0.6 m (2 ft) of clay. The Walk-In Pits are currently being capped. The level of contamination in soils surrounding the caps resulting from downgradient migration of contaminants in groundwater or stormflow has not been defined.

3.9.2 Groundwater

Each of the major hydrologic units can be subdivided into the following subsurface zones: stormflow, vadose, and groundwater (Fig. 3.38). Flow within and between any of these zones is influenced by topography, surface cover, geologic structure, and lithology. Analyses by Moore (1988) suggest that a combination of topographic relief and a marked decrease in permeability as depth increases restrict the majority of flow to the shallow portions of the system.

The stormflow zone approximately corresponds to the root zone in naturally vegetated portions of the ORR. Moore (1988) suggests that ~90% of active subsurface flow occurs through a 1- to 2-m-deep (3.3- to 6.6-ft) stormflow zone in those areas. During storm events, the stormflow zone transmits water laterally toward surface water discharge points. Surface runoff occurs when the stormflow zone becomes completely saturated.

Water flow within the stormflow zone occurs primarily through root channels, worm holes, and fractures. While these large pores transmit most of the water, the majority of water mass (and contaminant storage) within the zone resides within the less transmissive surrounding soil matrix wherein hydroscopic forces immobilize the water. Diffusion of contaminants between the matrix and large pores, during times of active stormflow, effectively reduces the rates of contaminant flushing relative to fluid flow through the large pores.

The effect that construction of the Burial Grounds waste pits and RCRA caps has on the hydrology of the stormflow system is uncertain. Caps were not installed with drains or grout curtains designed to prevent stormflow from travelling below the caps. Upgradient burial trenches that are oriented perpendicular to the topographic slope (such as trenches in BG-A) may collect water, whereas trenches oriented parallel to slope (such as BG-D, -E, and -J) may enhance transmission of water through the stormflow zone.

The vadose zone underlies the stormflow zone and varies in thickness from a maximum near ridge tops to a minimum where the water table approaches the ground surface (e.g., near perennial streams). The vadose zone is comprised of a regolith of clays and silts derived from the weathering of underlying bedrock and, in some cases, extends into the underlying bedrock. The regolith often retains relict structures (e.g., fractures) from the parent bedrock. These structures appear to serve as discrete, oriented, permeable flow paths. Between recharge events, flow within the vadose zone decreases with water moving toward the water table.

The major contaminant source areas in BCV tend to be located in the Nolichucky strike belt. During the operational history of the source areas the contaminants were either shallowly buried within the vadose zone or were spread on the ground surface. As a result,

most of the contaminants that are migrating within BCV were mobilized by water moving through the stormflow zone, the vadose zone, and the water table interval of the groundwater zone.

Migration of DNAPLs downward below the Burial Grounds and the Oil Landfarm is evidenced by the high concentrations of organics in groundwater below these sites, and in one case to direct sampling of free-phase product in groundwater. DNAPLs may have migrated to a minimum of 152 m (500 ft) below the surface below the Burial Grounds. The extent to which DNAPLs are being leached from waste pits and the vadose zone at these sites is not known. However, DNAPLs that migrate through geologic media will leave behind a trail of residual DNAPL (i.e., nonmoving free-phase hydrocarbons). In this way there probably remains a trail of residual DNAPL in the vadose zone below the Burial Grounds that may be leached to groundwater in the future. The extent of organic contamination of soil and the vadose zone within BCV OU 1 has not been defined.

3.9.3 Surface Water

At any one time, surface water in streams draining the units of BCV OU 1 is comprised of three possible components: water that has entered the stream from springs where shallow groundwater discharges, water from seeps and springs that are the discharge point of the stormflow system, and overland flow. The relative contribution to the total flow of each of these components depends on recent and present precipitation conditions.

3.9.4 Site Conceptual Model for Human Receptors

Site conceptual model flow diagrams for human receptors, for the entire Bear Creek Valley (i.e., for both the integrator OU and each source OU) and for BCV OU 1 (a source OU), are illustrated in Figs. 3.34 and 3.35, respectively. The Bear Creek Valley site conceptual model is shown in order to illustrate the relationship between the various source OUs (BCV OU 1, OU 2, and OU 3) and the integrator OU (BCV OU 4).

3.9.4.1 Site conceptual model for Bear Creek Valley

For Bear Creek Valley (Fig. 3.34), seven potentially contaminated media are considered and include soil/waste, sediment, surface water, shallow groundwater (active flow), groundwater (deep), air, and the food chain. The primary sources of the contaminants are the S-3 Ponds, Oil Landfarm, Burial Grounds, Bone Yard/Burn Yard, Sanitary Landfill 1, Rust Spoil Area, Spoil Area 1, and SY 200. The primary and secondary release mechanisms are chemical releases (including leaks and spills), runoff, infiltration/percolation, groundwater discharge, soil resuspension (wind generated dust), plant and animal uptake, and volatilization.

The soil is a potential pathway of migration via runoff to surface water, percolation/infiltration into groundwater, resuspension of particles (wind generated dust) into the air pathway, uptake into plants and animals, and volatilization of contaminants into the air. The soil exposure medium is divided into soil/waste (associated with each source OU) and soil/sediment (the BCV flood plain soil). Soil/waste and sediments will be evaluated, in terms of human health risk, (1) using the current on-source OU industrial scenario and (2) using the future hypothetical on-source OU residential scenario in a baseline risk assessment (BRA). The results of the source OU BRAs will also be included in the final Bear Creek Valley BRA [i.e., BRA for BCV OU 4 Phase II, refer to Figure 3.34]. For

soil/waste (BCV OU 1 and OU 2) and sediment (BCV OU 4), exposure pathways/routes which will be evaluated include ingestion of soil, inhalation of resuspended soil (dust), dermal contact with soil, inhalation of volatiles from the soil, and external exposure to radionuclides in the soil. For soil/sediment (BCV floodplain soils), the risk assessment includes evaluation of these same pathways for exposure to floodplain soils or soils that have been irrigated (for agricultural/gardening purposes) with contaminated surface water.

Contaminant releases into the air exposure medium may occur as a result of resuspension of dust (from soil) and volatilization of contaminants from soil, surface water, and groundwater; residential uses of surface water and groundwater (e.g. showering) can result in volatilization of contaminants. Inhalation is the primary exposure route for contaminants in air and this media will be evaluated in the current and future on-source OU and integrator-OU risk evaluations.

The surface water exposure medium is believed to have the greatest potential for off-source OU exposure. The potential for contaminant movement into groundwater is increased due to surface water infiltration to groundwater. In addition, groundwater can discharge to surface water; in Fig. 3.34, this secondary release mechanism is shown only as a dotted line. Surface water associated with a source OU will be evaluated in an integration point (IP) assessment (refer to Sec. 6.4) in Phase I of BCV OU 4 and in a BRA assessment in Phase II of BCV OU 4. Exposure pathways that will be evaluated for the surface water are inhalation of volatiles, dermal contact, and ingestion of the water.

Groundwater, both deep and active flow, is another potential exposure medium and is a mechanism for off-source OU transport of contaminants (Fig. 3.34). Again, groundwater associated with a source OU will be evaluated in an IP assessment in Phase I of BCV OU 4 and in a BRA assessment in Phase II of BCV OU 4. The exposure pathways which will be evaluated include ingestion, inhalation of volatiles, and dermal contact with the groundwater.

Biota (i.e., vegetables and fish) may accumulate contaminants that are present in the soil, irrigation water, and surface water. These exposure media will be evaluated (in BCV OU 4 Phase II BRA) using (1) the hypothetical residential (gardener) on-source OU scenario BRA, for soils, and (2) using the future residential (gardener and fisherman) integrator OU BRA, for the flood plain soils and surface water.

In summary, the site conceptual model for the entire Bear Creek Valley (Fig. 3.34) is shown to illustrate the relationships between the source OUs and the integrator OU. In addition, the different types of risk assessments which will be performed (i.e., industrial and residential BRAs, and IP assessments). In the short term, for Phase I of BCV OU 4 (integrator OU), the surface water integrator and the active flow groundwater will be evaluated in an IP risk assessment. The soil/waste in the source OUs (BCV OU 1 and OU 2) will be evaluated using both an industrial and a residential scenario in BRAs. The IP assessment will be used to evaluate the off-source OU exposure. In the future (BCV OU 4 Phase II), all the BRAs for each source OU (i.e., BRAs for soil/waste) and the integrator-OU (which includes all surface water and groundwater associated with all of BCV) will be evaluated and integrated into a Bear Creek Valley BRA (i.e., BRA for BCV OU 4 Phase II).

3.9.4.2 Site conceptual model for Bear Creek Valley OU 1

Illustrated in Fig. 3.35 is the site conceptual model for use in human health risk assessment for BCV OU 1; also included in this figure are media associated with BCV OU 1 that will be evaluated in the integrator OU (i.e., BCV OU 4). There are surface water, groundwater, sediment, soil, and wastes associated with this source OU; the groundwater, sediment, and surface water will be evaluated in BCV OU 4. Only exposure to the soils and the inventoried wastes will be evaluated, in terms of risk to human health, for this source OU (i.e., for BCV OU 1).

The soil associated with BCV OU 1 will be evaluated, in terms of risk to human health, using both the current industrial land-use and the future hypothetical residential land-use scenarios. Exposure pathways/routes which will be evaluated include ingestion of soil, inhalation of resuspended soil (dust), dermal contact with soil, inhalation of volatiles from the soil, and external exposure to radionuclides in the soil.

Human receptors (in a future hypothetical residential land-use scenario) could also be at risk from direct exposure to the buried wastes (inorganics, organics and radionuclides) placed in the Burial Grounds; however, under the current land-use conditions, human receptors are not exposed to these buried wastes. A statistical sampling and analysis of classified and unclassified inventory lists for the BCV OU 1 Burial Grounds is available. A conservative estimate of the risk to a human receptor will be determined using the inventory list of the wastes deposited in the Burial Grounds (best *estimates* for the total volume of each contaminant deposited in the pits will be used). Exposure pathways to be considered include ingestion, dermal contact, inhalation, and external exposure. A bounded risk estimate will be determined in which the lower bound will be estimated using a reasonable maximum exposure (RME) assessment and an upper bound will be estimated by evaluating an acute exposure scenario.

3.9.4.3 Ecological conceptual model

Because they may be mobile and may be consumed by mobile species and because they may concentrate contaminants from exposure media, biota may serve as significant transporters of contaminants. The uptake and transport of contaminants by OU 1 aquatic and terrestrial biota are partially addressed as part of BMAP, a requirement of the Y-12 Plant NPDES permit. Figure 3.37 shows the BCV ecological conceptual model and outlines the sections of this model that will be addressed by the BCV OU 1 RI. The ecological site conceptual model is based on the simplified food web for BCV (Figs. 3.36 and 3.37). BCV OU 1 as a source unit will address ecological receptors with a range that is limited to the source units. These biota are vegetation, soil invertebrates, and on-source herbivorous and vermivorous wildlife. Where the range of a species extends beyond the source OU, risk will be evaluated in the integrator OU (BCV OU 4). Results to date indicate elevated tissue concentrations of mercury and PCBs in Bear Creek biota. Further study of the terrestrial and wetland flora and fauna will be carried out to assess ecological impacts to on-site receptors.

4. POTENTIAL RECEPTORS AND EXPOSURE PATHWAYS

This chapter summarizes potential current and future human and ecological receptors that may be exposed directly or indirectly to contaminants in the BCV OU 1 waste sources. Potential receptors to be considered in the BCV OU 1 risk assessments were presented in Chap. 3 as part of the site conceptual model. This chapter discusses in more detail the potential receptors that will be considered while developing the sampling and analysis plan for the site.

Potential receptors to the waste units that compose the BCV OU 1 must be defined for both current and future land use scenarios. Current receptors are identified in order to make intermediate decisions in the CERCLA process, such as determining the need for an expedited removal action. Future receptors must be identified in order to make the decisions involving long-term management of the waste unit.

An exposure pathway is the course that a chemical agent takes as it moves from source to receptor. An exposure pathway may include the following:

- A source of potential contamination
- A mechanism for release from the source [e.g., chemical partitioning from solid to liquid (waste to rainwater) that infiltrates into and through the unit] and a route of migration from the source to a receiving environmental media (e.g., vertical transport through the unsaturated zone to the groundwater)
- A point of exposure (e.g., soils)
- A route of exposure (e.g., ingestion)
- A receptor (i.e., human or ecological)

The integration of all these factors results in a complete exposure pathway. Complete exposure pathways for BCV OU 1 are presented in Chap. 3 as part of the site conceptual model.

At BCV OU 1, the sources of potential contamination are buried waste and any soils surrounding the waste that have become contaminated by means of leaching from the unit. Some exposures may occur through direct contact with the waste. Other exposures may occur through indirect contact after the waste has migrated away from the source. The primary release and transport mechanisms away from the source to the point of exposure in surrounding media include the following:

- Vertical leaching to the groundwater
- Vertical/horizontal leaching through the stormflow system to surface water
- Runoff to surface water and sediments
- Volatilization into the air
- Particulate emissions into the air

Once the exposure point has been identified for each medium, an exposure route must be identified. Exposure routes are the mechanisms by which receptors contact the chemicals.

The potential routes of direct exposure associated with the BCV OU 1 receptors include the following (these direct exposure routes will be evaluated in the BRA for BCV OU 1):

- Soil/waste ingestion
- Dermal contact with soil/waste
- Inhalation
 - of volatiles in soil
 - of fugitive dusts
- Direct exposure to contaminants
 - acute exposure
 - penetrating radiation

Indirect exposure to the waste may occur through the following pathways and will be evaluated in BCV OU 4:

- Groundwater pathways
 - ingestion of groundwater
 - inhalation while showering
 - uptake by plants
 - ingestion of lower members of the food chain that are contaminated
 - dermal contact with groundwater used for domestic purposes
 - dermal contact while showering
- Surface water pathways
 - ingestion of surface water
 - dermal contact with surface water
 - inhalation of volatiles from surface water
 - uptake into food chain (aquatic flora and fauna)
- Air pathways
 - inhalation of volatiles in water

Additional secondary pathways may be identified through the course of the RI. Exposure to soil and wastes associated with BCV OU 1 will be evaluated in this RI; groundwater and surface water (for BCV OU 1) will be evaluated in BCV OU 4 Phase I and II.

4.1 HUMAN POPULATIONS

Two general types of land use categories will be considered during this risk assessment process: residential and industrial. Current land use of the BCV OU 1 site is light industrial. Although no industrial activity actually takes place in the OU, the waste units in BCV OU 1 support or have supported the Y-12 Plant production operations, and thus light traffic continues around some of the waste units.

It is too early in the CERCLA process to attempt to identify future land use of the ORR. The strategic planning for future use of the government-owned property at the Y-12 Plant is a continuing topic of discussion. It is likely that the results of several source OU risk

assessments and cost-benefit analyses performed during FSs will help make the decisions concerning future land use. Because future land use for the ORR has not been decided, the BCV OU 1 preliminary hazard evaluation and (this RIWP) baseline risk assessment (BCV OU 4 Phase II) will consider at a minimum the following land use scenarios:

- Government controls remain in place, and activity at the site is **industrial**.
- Government control is lost, and the land is used for **residential** purposes.

Additional future land use scenarios, such as recreational, may be evaluated in the site-wide baseline risk assessment (BCV OU 4 Phase II). For purposes of this BRA, it is assumed that the BCV OU 1 land will be industrial and/or residential. This BRA's data collection activities will provide data to evaluate any (residential, industrial, and recreational) of the potential land use scenarios.

4.1.1 On-Source OU Receptors

4.1.1.1 Current land use

Two factors play a role in limiting the number of potential current receptors to the BCV OU 1 area. First, as with the entire Y-12 Plant, BCV OU 1 is a controlled access area, surrounded by fences and guarded by active patrolling. Second, due to the nature of the waste disposal activities at the site and the fact that the waste unit lies outside of the production area, relatively few personnel are routinely in the area. On-source OU human receptors are workers that may be Energy Systems employees or contractor personnel. Transient visitors may enter the site, likely associated with ongoing remediation efforts; however, the frequency and duration of these visitors is likely to be minimal. There is little potential for a trespasser to enter the area because of the emphasis on security at the Y-12 Plant. No sensitive subpopulations are assumed to exist on site.

4.1.1.2 Future land use and sensitive populations

Based on the assumption that the government loses control of the land and that future land use is residential, future potential receptors would include both children and adults who reside and garden on site. Gardening includes the vegetable/fruit ingestion pathway. In addition, these receptors could hunt and fish on the land. However, based on the assumption that the government retains control of the land, on-source human receptors would likely be industrial/remediation workers. Children would be a sensitive population under this future residential land use assumption.

4.1.2 Off-Source OU Receptors

Because BCV OU 1 is a source term OU, and off-source OU receptors are affected by means of migration through media of the integrator OU, protection of off-site receptors will be handled indirectly. Groundwater and surface water flow is being addressed extensively as part of the BCV OU 4 integrator OU. Current risk to off-source receptors exposed to contaminants in these media will be addressed in the IP assessment of the integrator OU RI. However, because the ultimate goal of a source control measure is to decrease or eliminate releases and migration to any receptors, the risk assessment and remedial action selection for the source OU must address migration to the off-source OU receptors.

To distinguish the scope of the BCV OU 1 risk assessment from that of the BCV OU 4 integrator OU, it is assumed that the groundwater and surface water immediately adjacent to the waste sources are receptors and that it is the goal of the source OU to eliminate contaminant migration to these environmental receptors. To do this, the integrator unit will provide groundwater and surface water PRGs at the points where the two OUs interface. These PRGs will be designed to protect the downstream receptors. Additional information for evaluating exposures to off-source OU receptors will likely be generated by the integrator OU (e.g., the dilution attenuation factor from the water adjacent to the source to the off-site receptor). Distinguishing the responsibilities of the BCV OU 1 risk support from those of the BCV OU 4 risk support will continue as the RI/FS for each progresses.

Off-source OU receptors could be members of the general public who live downstream and downgradient of the site and may be exposed indirectly to waste from BCV OU 1; off-source OU receptors also include future potential receptors located adjacent to the source OU.

Currently, nine public water supply systems serve ~91,500 people within a 32-km (20-mile) radius of the ORR. The intake for Kingston is located on the Tennessee River at ~0.6 river kilometers (0.37 miles) above the confluence of the Clinch and Tennessee rivers. Sixteen public groundwater supplies are located within a 32-km (20-mile) radius of the Y-12 Plant. Of these sources, the closest to the Y-12 Plant is the Allen Fine Spring supplying the Dixie-Lee Utility District in Loudon County, which serves 6700 people. Private single-family wells are common in adjacent rural areas not served by public water supply. Most of the residential wells, other than those adjacent to the city of Oak Ridge, are south of the Clinch River.

4.1.2.1 Regional demography

The demography of East Tennessee and the Oak Ridge region is discussed in Sect. 3.3. Current and future off-source OU human receptors are members of the general public who live downstream and downgradient of the site and may be exposed indirectly to waste from BCV OU 1. This off-source OU population can be considered a receptor only if it is shown (in BCV OU 4) that the BCV OU 1 wastes migrate into the surface water (Bear Creek) and groundwater (and into the creek and/or the Clinch River) surrounding the sources. Potential receptors include members of the public who may come into contact with waste in the stream outside the plant boundary, users of water in the Clinch River, and recreational users of the Clinch River. Since no groundwater wells exist between BCV and the points at which groundwater mixes with surface water, it was assumed that groundwater well users are not current off-source OU receptors; however, they will be considered as potential future receptors in the BCV OU 4 BRA and IP assessments.

4.1.2.2 Sensitive populations

Children who live off-source OU may be considered as sensitive subpopulations in the BCV OU 4 risk assessment. There are no hospitals, nursing homes, or day-care centers located within a reasonable distance downgradient of BCV.

4.2 ECOLOGICAL POPULATIONS

As with the human health assessment, distinguishing the responsibilities of the BCV OU 1 ecological risk support from those of the BCV OU 4 ecological risk support will be an ongoing process. Identifying ecological receptors differs from identifying human receptors in that land use/access control does not drive the identification. Future ecological receptors are assumed to be the same as current receptors in the absence of potential remediation and changes of land use.

To distinguish the scope of the BCV OU 1 risk assessment from the scope of the BCV OU 4 integrator OU for this work plan, it is assumed that aquatic receptors in the surface water immediately adjacent to the waste sources are receptors of concern, and that it is the goal for the source OU to eliminate contaminant migration to these environmental receptors as well as on-site terrestrial receptors. To do this, the integrator unit should provide ecologically-based surface water PRGs at the points where the two OUs interface. These PRGs will account for protection of the aquatic receptors. It will then be the goal for the source OU to prevent contaminant migration from the unit to the surface water at levels exceeding the PRGs.

Although the waste units that make up BCV OU 1 are considered light industrial, there are few buildings or concrete areas (with the exception of the asphalt parking lot near the S-3 Site). The BCV OU 1 area is primarily grass-covered fields surrounded by woodlands. Some of the units are covered with caps, and thus contact with the waste by ecological receptors may only occur for burrowing animals. However, other units do not have caps, and terrestrial and aquatic receptors may contact the waste directly. In addition, seeps and springs in several areas of the site may be used as water sources by terrestrial receptors. These seeps and springs may also provide habitat for aquatic biota such as diptera, amphipoda, and salamanders; these biota would be directly exposed to contaminants in these seeps and springs. Thus, on-site terrestrial receptors and associated food chain receptors are of concern for the BCV OU 1 RI. Terrestrial receptors that can be associated specifically with BCV OU 1 may include mammals, insects, worms, reptiles, and plants. Risks to browsing animals and predators, most of which have ranges larger than the source units, will be evaluated in the site-wide baseline ecological risk assessment to be completed as part of the RI for OU 4.

4.2.1 Flora: Terrestrial and Aquatic

Flora on the ORR are described in Sect. 3.7. Flora, including wetland and pond flora, will be addressed as receptors potentially at risk from contaminants and also as a pathway to herbivorous fauna in BCV. Grasses are the food source of some herbivorous animals and are known to accumulate metal and radionuclide contaminants. They will be sampled as monitors of exposure via food to herbivorous mammals and birds. Vegetation characteristic of contaminated areas identified by Phase I sampling will be sampled later in Phase I to determine the nature of contamination in flora. If significant uptake of contaminants is found, statistically based sampling will be done to determine the distribution of contaminated flora.

4.2.2 Fauna: Terrestrial and Aquatic

The abundance and productivity of bird and mammal populations are the primary assessment endpoints for ecological risk assessments on the ORR. Fauna on the ORR are described in Sect. 3.7. All of these fauna are considered likely receptors in the ecological

conceptual model. However, risks to birds and mammals with ranges larger than the source units and to aquatic biota in tributaries to Bear Creek will be evaluated as part of the RI for BCV OU 4.

Complete exposure pathways may exist for terrestrial organisms found on and near BCV OU 1. Small mammals, insects, and worms may be directly exposed to wastes and contaminated soils, especially as a result of burrowing in the subsurface soils. Inhalation of volatile organics in burrows or evaporating from soils and seeps may also expose terrestrial biota. Terrestrial mammals may be exposed both topically and through soil ingestion. Birds and mammals may feed on contaminated food and may ingest soils directly. Contaminated soils carried by wind may be inhaled by birds and animals.

Aquatic biota may be exposed to contaminants in surface water by ingestion, direct contact, and gill absorption to contaminants in sediments by ingestion and direct contact, and to contaminants in both water and sediment by ingestion of contaminated food. Exposure of aquatic biota to contaminants in the main stem of Bear Creek will be evaluated by the Bear Creek integrator OU RI. However, aquatic biota in tributaries draining the source OUs and in wetlands, seeps, and springs within the OU boundaries will be addressed by this OU RI.

Radionuclides may cause exposure either internally or by direct irradiation. Terrestrial animals may be exposed by direct irradiation from contaminated soil in which they burrow or on which they forage; the degree of exposure will depend on their behavior.

Perhaps the most significant pathway for exposure to contaminants at the BCV OU 1 area is the food web. Feeding habits and other aspects of lifestyle determine the significance of these exposures. For example, shrews prey on earthworms, which may have exterior contamination from adhering soils and interior contamination from ingesting soils. Therefore, shrews may ingest more soils than voles, whose diet is mainly vegetarian. Some insects of terrestrial origin spend their entire life on or in the soil. Earthworms are highly exposed to contaminants in soil by both dermal contact and ingestion. They are likely sources of contamination for shrews, some birds, and other small animals. Plants are highly exposed to contaminants in soils by root uptake and by surface deposition of airborne dust. Animals may, in turn, feed on the vegetation.

Wetlands constitute areas of special ecological concern because of their legally protected status and because of their unusual ecology. Wetlands are those areas that are inundated or saturated by surface water or groundwater at a frequency and duration sufficient to support, and under normal circumstances do support, a prevalence of vegetation typically adapted for life in saturated soil conditions (40 *CFR* 20.3; 33 *CFR* 328.3). Wetlands are defined by three essential attributes: wetland hydrology, hydrophytic vegetation, and hydric soils.

Contaminant releases in wetland soils may be different from those in drier soils because water-saturated soils may experience increased or diminished leaching of contaminants and because pH, oxidation level, and other characteristics of soil chemistry are likely to be different from drier sites. Wetlands also typically have different populations of flora and fauna from drier habitats, so transfers into the food web are different and require separate evaluation.

A few indicator species will be used as representatives of their trophic levels in the BCV OU 1 ecosystem. These species were chosen because they are expected to be abundant

and easily surveyed or sampled and because they represent a key step in the transfer of contaminants from soil, sediment, or water to predators. Contaminant burden will be measured for aquatic insects and grasses and browse; both population surveys and body burden analyses will be conducted for earthworms, amphibians, and terrestrial mammals.

Wetland areas in BCV OU 1 are likely to act as sinks for contaminants in surface runoff. Although these areas are not significant habitats for fish, they may provide good habitat for amphibians. Amphibian larvae (tadpoles) are exposed to contaminants via water, sediment, and vegetation and are, therefore, good indicators of contamination of the aquatic habitats. They subsequently leave the aquatic habitats in which they are initially exposed and enter the food chain of terrestrial predators, providing a link from surface runoff and sediment transport of contaminants to terrestrial predators at some distance from the areas of contaminant accumulation. Ponds and wetland areas containing amphibian larvae will be identified, and the development of larvae will be monitored visually. When they have neared maturity, they will be sampled for contaminant body burden.

Earthworms are continuously exposed to soil contaminants by direct contact and by ingestion. They are a food source to some birds and small mammals. Population densities and contaminant body burdens, both total and depurated, will be determined by sampling earthworms. If contaminant levels are elevated and population densities are reduced, earthworm toxicity tests will be performed on the soil (see Sect. 6.1.1.4).

Small mammals can be particularly effective biomonitors of soil contaminant effects in the food web if the species used are abundant, easily caught, do not migrate long distances, have a widespread distribution, and have generalized food habits. Their relatively short life span (generally less than 1 year) facilitates long-term monitoring of contaminants because each generation reflects the present type and amount of contaminants in the environment; thus increases, decreases, or the addition of new contaminants can be detected. The small mammals likely to be captured in BCV OU 1 include white-footed mouse (*Peromyscus leucopus*), short-tail shrew (*Blarina brevicauda*), masked shrew (*Sorex cinereus*), and field vole (*Microtus pennsylvanicus*). Each of these species meets the requirements of good environmental indicators mentioned above. In particular, the insectivorous short-tail shrew is well suited because of its relatively high food intake and dietary needs. It bioaccumulates contaminants such as metals and PCBs if they are present in invertebrate populations, and much of its food source lives in the soil, close to BCV OU 1 contamination, for some part of its life cycle. The white-footed mouse is abundant and is expected to provide a potential pathway to higher trophic levels (i.e., predators); it will be analyzed for contaminant levels. Voles represent exposures via a primarily herbivorous pathway.

Woodchucks (*Marmota monax*) are exposed to subsurface contaminants by soil ingestion, direct dermal contact, and external irradiation. They will be trapped and analyzed for radionuclide uptake. Although their larger home range makes it difficult to identify the locations of the contaminant sources, they are good indicators of exposures to subsurface soil near and in the waste areas.

4.2.3 Future Land Use

Unmanaged terrestrial habitats typically undergo a succession of changes as grasses are replaced by forbs and trees, with accompanying changes in animal populations. In the absence of institutional controls in BCV OU 1, areas that are currently maintained as grass would

revert to forest, with changes in the distribution and abundance of terrestrial biota. In addition, engineered caps may break down or be penetrated by burrowing animals, allowing contaminants to be released to soil and surface water. Therefore, potential future receptors near waste areas include the current potential receptors, as well as birds and other forest-dwelling animals.

Residential, industrial, and agricultural are the potential future land uses that will be evaluated. In addition, a potential future land use in which waste sites are allowed to revert to forests would provide a mixture of exposures. Potential receptors under this scenario would be the same in the unmanaged habitat scenario.

5. IDENTIFICATION OF INVESTIGATION REQUIREMENTS

Many of the requirements for performing an effective site investigation are based on the need to evaluate the site in terms of the two CERCLA "threshold" criteria: compliance with ARARs and protection of human health and the environment (EPA 1988b). This chapter provides specific considerations of these requirements. Both compliance with ARARs and evaluation of potential health and ecological risks are dependent on understanding the potential COCs at the site. Table 5.1 lists potential COCs for BCV OU 1 by waste site. These lists were compiled from studies referenced in Chap. 3 of this work plan. To some extent, the COC list is a reflection of the types and amounts of data available on a site and thus may not be comprehensive. Data with which to develop a COC list are not available for SL 1 or the Boneyard/Burnyard. Thus, one of the goals of the Stage 1 sampling efforts for these sites will be to identify COCs.

5.1 IDENTIFICATION OF ARARS

CERCLA was passed by Congress and signed into law on December 11, 1980 (Public Law 96-510). This act was intended to provide for "liability, compensation, cleanup, and emergency response for hazardous substances released into the environment and the cleanup of inactive waste disposal sites." SARA, adopted on October 17, 1986 (Public Law 99-499), did not substantially alter the original structure of CERCLA but provided extensive amendments to it.

Even before CERCLA, DOE issued Order 5400.1 to establish environmental protection program requirements, authorities, and responsibilities for DOE operations. This order seeks to ensure compliance with applicable federal, state, and local environmental protection laws and regulations, executive orders, and internal DOE policies. DOE Order 5400.1 has been modified to meet new requirements imposed by amendments to the federal environmental laws.

Section 121 of CERCLA specifies that remedial actions for cleanup of hazardous substances must comply with requirements or standards under federal or more stringent state environmental laws that are applicable or relevant and appropriate to the hazardous substances or particular circumstances at a site. Inherent in the interpretation of ARARs is the assumption that protection of human health and the environment is ensured.

The purpose of this report is to supply a preliminary list of available federal and state chemical- and location-specific ARARs that might be considered for the cleanup of soil at BCV OU 1. In addition, tables of groundwater and surface water ARARs are presented for comparison with analytical detection limits for contaminants in these media. If the RI determines that leachate and runoff from the contaminated soils at BCV OU 1 are causing contamination in surface water and groundwater at levels that surpass ARARs for the contaminants, remedial alternatives to isolate and control the sources to prevent further contamination of surface and groundwater will be addressed during the FS phase. Action-specific ARARs will be identified and analyzed when remedial alternatives are selected during the FS phase. The process of ARAR identification is an iterative one that is continually changing as the RI/FS progresses. Therefore, this list of ARARs represents a compilation of

Table 5.1. Chemicals of potential concern at BCV OU 1

	Groundwater/surface water contaminants of concern			Soil/waste contaminants of concern		
	Radionuclides	Inorganics	Organics	Radionuclides	Inorganics	Organics
Burial Grounds		Cadmium Chromium	1,1-dichloroethane 1,1-dichloroethylene 1,2-dichloroethylene Tetrachloroethylene 1,1,1-trichloroethane Trichloroethylene PCBs	Thorium-228 Thorium-230 Thorium-232 Uranium-234 Uranium-235 Uranium-238	Comprehensive Target Analyte List (TAL)	Comprehensive Target Compound List (TCL) analytes, plus additional tentatively identified compounds (TICs)
Oil Landfarm	Gross alpha Gross beta Strontium-90 Technetium-99 Tritium	Barium Cadmium Chromium Lead Mercury	1,1-dichloroethane 1,2-dichloroethylene Tetrachloroethylene Trichloroethylene PCBs	Gross alpha Gross beta Strontium-90 Technetium-99 Tritium	Barium Cadmium Chromium Lead Mercury	Comprehensive TCL analytes, plus additional TICs
S-3 Site	Gross alpha Gross beta Strontium-90 Technetium-99 Thorium-228 Thorium-230 Thorium-232 Tritium Uranium-234 Uranium-235 Uranium-238	Aluminum Barium Beryllium Boron Cadmium Chromium Cobalt Copper Lead Mercury Nickel Nitrate (as N) Uranium	Acetone 2-butanone Carbon tetrachloride Chloroform 1,1-dichloroethylene Methylene chloride Tetrachloroethylene 1,1,1-trichloroethane	Gross alpha Gross beta Plutonium-238 Strontium-90 Technetium-99 Thorium-228 Thorium-230 Thorium-232 Uranium-234 Uranium-235 Uranium-238	Aluminum Barium Beryllium Boron Cadmium Chromium Cobalt Copper Lead Mercury Nickel Nitrate, nitrite Phosphates Selenium Thorium Uranium	Comprehensive TCL analytes, plus additional TICs

potential ARARs, of which subsets will be used or to which additional ARARs will be added as further site characterization is completed.

It is understood that DOE will comply with the requirements of NEPA as specified in DOE Order 5440.1D, "National Environmental Policy Act Compliance Program." Further, DOE Order 5400.4, "Comprehensive Environmental Response, Compensation, and Liability Act Requirements," calls for integration of NEPA and CERCLA requirements for DOE remedial actions at CERCLA sites. This issue has been reaffirmed in FFA §I(A)(3) and §III(A)(2) and Secretary of Energy Notice (SEN) of February 5, 1990 (SEN-15-90), which was issued to ensure that DOE's NEPA activities are carried out in a centralized and uniform manner. Therefore, the regulations found in NEPA will not be addressed in this report as ARARs; however, the federal and state regulations protecting environmental resources which may be identified at a site during a NEPA assessment are discussed in Sect. 5.1.3.

Title I, §111(c)(6) of CERCLA mandated that the Occupational Safety and Health Agency (OSHA) promulgate standards for protection of employee health and safety during hazardous waste operations at RCRA or CERCLA sites and during emergency response to hazardous substance releases. The final regulations for "Hazardous Waste Operations and Emergency Response" (29 *CFR* 1910) have appeared in 54 *FR* 9294 (Final Rule, March 6, 1989). These regulations are designed to protect workers involved in cleanup operations at uncontrolled hazardous waste sites and to provide for worker protection during initial site characterization and analysis, monitoring activities, materials handling activities, training, and emergency response. These regulations do not apply to those workers who would not be exposed.

Federal construction activities involving no potential for hazardous substance exposure are covered by the OSHA standards in 29 *CFR* 1926 ("Federal Service Contracts" and 29 *CFR* 1910 "General Industry"). DOE also addresses occupational safety in DOE Orders 5480.11 ("Radiation Protection for Occupational Workers"), 5480.4 ("Environmental Protection, Safety, and Health Protection Standards"), 5483.1A ("Occupational Safety and Health Program for Contractors at GOCO Facilities"), 5480.9 ("Construction Safety and Health Program"), and 5480.10 ("Contractor Industrial Hygiene Program"). ARARs apply to those federal and state regulations that are designed to protect the environment and do not generally apply to occupational safety regulations. EPA requires compliance with the OSHA standards in §300.150 of the National Contingency Plan (NCP), not through the ARARs process. Therefore, neither the regulations promulgated by OSHA nor the DOE orders related to occupational safety are addressed as ARARs in this chapter; these regulations are addressed in the site-specific Health and Safety Plan for BCV OU 1. (See Chap. 9.)

The following is a listing of the definitions of terms used throughout this report:

Applicable requirements are "those cleanup standards, standards of control, and other substantive requirements, criteria, or limitations promulgated under federal environmental or state environmental or facility citing law that specifically address a hazardous substance, pollutant, contaminant, remedial action, location, or other circumstance at a CERCLA site" (40 *CFR* 300.5).

Relevant and appropriate requirements are "those cleanup standards, standards of control, and other substantive requirements, criteria, or limitations promulgated under federal

environmental or state environmental or facility citing law that, while not applicable to a hazardous substance, pollutant, contaminant, remedial action, location, or other circumstance at a CERCLA site, address problems or situations sufficiently similar to those encountered at the CERCLA site that their use is well suited to the particular site” (40 *CFR* 300.5).

Requirements under federal or state law may be *either* applicable *or* relevant and appropriate to CERCLA cleanup actions, but not both. However, requirements must be *both* relevant *and* appropriate for compliance to be necessary. In the case where a federal and a state ARAR are available or where there are two potential ARARs addressing the same issue, the more stringent regulation must be selected. However, CERCLA §121(d)(4) provides several ARAR waiver options that may be invoked, providing that the basic premise of protection of human health and the environment is not ignored. A waiver is available for state standards that have not been applied uniformly in similar circumstances across the state. In addition, CERCLA §121(d)(2)(C) forbids state standards that effectively prohibit land disposal of hazardous substances.

CERCLA on-site remedial response actions must only comply with the substantive requirements of a regulation and not the administrative requirements to obtain federal, state, or local permits [CERCLA §121(e) and FFA §XXII]. To ensure that CERCLA response actions proceed as rapidly as possible, EPA has reaffirmed this position in the final NCP (55 *FR* 8756). **Substantive requirements** pertain directly to the actions or conditions at a site, whereas **administrative requirements** facilitate their implementation. EPA recognizes that certain of the administrative requirements (such as consultation with state agencies, reporting, etc.) are accomplished through the state involvement and public participation requirements of the NCP. These administrative requirements should be observed if they are useful in determining cleanup standards at the site (55 *FR* 8757).

In the absence of federal- or state-promulgated regulations, there are many criteria, advisories, guidance values, and proposed standards that are not legally binding but may serve as useful guidance for setting protective cleanup levels. These are not potential ARARs but are “to-be-considered” (TBC) guidance.

5.1.1 Chemical-Specific ARARs

Chemical-specific requirements set health- or risk-based concentration limits or discharge limitations in various environmental media for specific hazardous substances, pollutants, or contaminants (53 *FR* 51437). These requirements generally set protective cleanup levels for the chemicals of concern in the designated media or else indicate a safe level of discharge that may be incorporated when considering a specific remedial activity.

BCV OU 1 consists of five sites: the Burial Grounds (including the ORPs), the Oil Landfarm, the Boneyard/Burnyard, Sanitary Landfill I, and the S-3 Ponds. The sites include unlined ponds and trenches, walk-in pits, burial grounds, and landfills, both capped and uncapped. Some of the caps are RCRA caps. Shallow groundwater and surface water runoff drain into Bear Creek. Historical and current operating conditions at this OU are discussed in detail in Chap. 2. A preliminary list of COCs detected in soils, surface water, and groundwater at BCV OU 1 includes metals, PCBs, organics, inorganics, VOCs, and radionuclides. The sites are contaminated with RCRA hazardous wastes in addition to the radiological contamination. Although limited in number, chemical-specific standards have been

established under several statutes, including the SDWA, the Clean Water Act (CWA), and the Clean Air Act (CAA).

5.1.1.1 Groundwater and surface water

Available chemical-specific ARARs that have been promulgated under federal and state law are listed in Table 5.2 for COCs that have been detected in surface water and groundwater at BCV OU 1. They are presented here for comparison with analytical detection limits for contaminants in these media. If the RI determines that leachate and runoff from the contaminated soils at BCV OU 1 are causing contamination in surface and groundwater at levels that surpass ARARs for the contaminants, remedial alternatives to isolate and control the sources to prevent further contamination of surface water and groundwater will be addressed during the FS. If groundwater contamination from BCV OU 1 is confirmed, cleanup of the groundwater will be addressed as a separate OU, and a detailed analysis of applicable ARARs will be addressed at that time. Shallow groundwater and surface water runoff from BCV OU 1 flows into Bear Creek, which is classified by the Tennessee Water Quality Control Board for industrial water supply, fish and aquatic life, recreation, irrigation, and livestock watering and wildlife uses (TDEC Rules, Chap. 1200-4-4). The state water quality criteria for protection of fish and aquatic life and for the consumption of aquatic organisms (the criteria for protection of recreation uses) are listed in Table 5.2 for comparison with surface water sampling results to determine extent of any contamination. In the absence of a state criterion for a specific chemical, a federal criterion is listed if available. If surface water contamination from BCV OU 1 is identified, one goal of any remedial actions taken will be to control the on-site source of that contamination; surface water itself will not be remediated as part of this OU.

5.1.1.2 Air

Primary National Ambient Air Quality Standards (NAAQS) for six chemicals appear in 40 *CFR* 50. The NAAQSs take into consideration all sources of exposure to a given chemical and establish ceilings that are not to be exceeded in the United States. NAAQSs are established as the criteria that state and local governments must plan to achieve and thus are not directly enforceable. Under §110 of the CAA, states are required to implement regulations to achieve the NAAQSs. The ambient air quality standards established by the TDEC Division of Air Pollution Control appear in Chap. 1200-3-3 of the Rules of the TDEC and are identical to the federal primary NAAQSs. The ambient standards of 1200-3-3 are translated into source-specific emission limitations (Rules of the TDEC, Chaps. 1200-3-4 to 1200-3-21) that must be considered ARARs if they apply to any of the remedial alternatives selected. Tennessee state air emission standards are considered "hybrid" (i.e., both chemical- and action-specific) ARARs and will be analyzed as action-specific ARARs following selection of remedial alternatives. In particular, the regulations governing fugitive dust emissions will be addressed.

National Emission Standards for Hazardous Air Pollutants (NESHAPs) for various industrial sources that emit one of several pollutants are established in 40 *CFR* 61. Most of the NESHAPs are generally neither applicable nor relevant and appropriate to cleanup at CERCLA sites because they regulate specific types of sources that would not be expected to be found at a CERCLA site. However, the NESHAPs will be considered as potential "hybrid" (i.e., both chemical- and action-specific) ARARs on a site-specific basis during the

Table 5.2. Chemical-specific federal and state regulations for protection of groundwater and surface water at BCV OU 1

($\mu\text{g/L}$ except where noted)^a

Chemical	Safe Drinking Water Act MCLs ^{b,c}	Safe Drinking Water Act MCLGs ^{b,d}	TDEC WQC Ingestion of Aquatic organisms ^e	TDEC WQC Fish and Aquatic Life CMC ^f	TDEC WQC Fish and Aquatic Life CCC ^f
Acetone	—	—	—	—	—
Aluminum	0.05-0.2 ^g	—	—	750 ^h	87 ^h
Barium	2000	2000	—	—	—
Beryllium	4 ⁱ	4 ⁱ	1.3	—	—
Boron	600 ^j	—	—	—	750 ^{h,k}
2-Butanone (Methyl ethyl ketone)	—	—	—	—	—
Cadmium	5	5	—	4 ^{l,m}	1 ^{l,m}
Carbon tetrachloride	5	zero	44	—	—
Chloroform	100 ⁿ	—	4700	—	—
Chromium (total)	100	100	—	—	100
Chromium (trivalent)	—	—	670,000	—	—
Chromium (hexavalent)	—	—	—	16	11
Copper	TT(1300) ^o	1300	—	18 ^{l,m}	12 ^{l,m}
Cobalt	—	—	—	—	—
1,1-Dichloroethane	—	—	—	—	—
1,1-Dichloroethylene	7	7	32	—	—
1,2-Dichloroethylene (cis- isomer)	70	70	—	—	—
Lead	TT(15) ^o	zero	—	82 ^{l,m}	3 ^{l,m}
Mercury	2	2	0.15	2.4	0.012
Methylene chloride	5	zero	16,000	—	—
Nickel	100	100	4600	1400 ^m	160 ^m
Nitrate	10,000	10,000	—	—	—
Nitrate and nitrite (as N)	10,000	10,000	—	—	—
PCBs (total)	0.5	zero	0.001	—	0.001
Tetrachloroethylene	5	zero	88	—	—
1,1,1-Trichloroethane	200	200	170,000	—	—
Acetone	—	—	—	—	—
Trichloroethylene	5	zero	807	—	—
Vinyl chloride	2	zero	5250	—	—
Radionuclides					
Gross alpha activity, pCi/L	15 ^{p,q}	zero ^q	—	—	—

Table 5.2 (continued)

Chemical	Safe Drinking Water Act MCLs ^{b,c}	Safe Drinking Water Act MCLGs ^{b,d}	TDEC WQC Ingestion of Aquatic organisms ^e	TDEC WQC Fish and Aquatic Life CMC ^f	TDEC WQC Fish and Aquatic Life CCC ^f
Gross beta activity, mrem/year	4 ^{r,s}	zero ^r	—	—	—
Strontium-90, pCi/L	8 ^t	—	—	—	—
Technetium-99	—	—	—	—	—
Tritium, pCi/L	20,000 ^t	—	—	—	—
Uranium (natural)	20 ^u	—	—	—	—

^aFederally promulgated regulations under the SDWA and Tennessee state regulations under the CWA are included in this table.

^bTennessee state drinking water standards (Chapter 1200-5-1 of the Rules of the TDEC) are identical to the federal.

^cMCL = maximum contaminant level.

^dMCLG = maximum contaminant level goal.

^eRules of the TDEC, Chapter 1200-4-3-.03; criteria for recreational use designation; a 10⁻⁵ risk factor is assumed for carcinogens. WQC = water quality criterion.

^fCMC = criterion maximum concentration; CCC = criterion continuous concentration; Rules of the TDEC, Chapter 1200-4-3-.03.

^gFederal and Tennessee secondary MCL; no primary MCL is available.

^hFederal WQC; Tennessee does not have a fish and aquatic life criterion for this chemical. Value for aluminum is at pH 6.5-9.0. Source: *EPA Region IV Criteria Chart* (December 1992).

ⁱ57 FR 31776 (July 17, 1992); effective January 17, 1994. Tennessee is in the process of revising its MCL to match this federal MCL.

^jEPA Office of Drinking Water Lifetime Health Advisory (November 1992).

^kMinimum standard for long-term irrigation of sensitive crops.

^lWater hardness dependent criteria; based on a water hardness of 100 mg/L as CaCO₃.

^mValues are for the dissolved form of this metal.

ⁿThis number is the MCL for total trihalomethanes which includes the sum of the concentration of chloroform, bromodichloromethane, dibromochloromethane, and bromoform.

^oNumber in parentheses is an "action level" that, when measured in the 90th percentile at the consumer's tap, triggers initiation of corrosion control studies and treatment requirements.

^pThis interim MCL includes radium-226 but excludes radon and uranium.

^qProposed MCL and MCLG of 15 pCi/L and zero, respectively, excludes radon, uranium and radium-226 (56 FR 33050, July 18, 1991).

^rProposed rule (56 FR 33050, July 18, 1991); final rule expected 1994; current interim MCL is 4 mrem/year, the same as the proposed MCL.

^sIf two or more radionuclides are present, the sum of their annual dose equivalent to the total body or to any organ shall not exceed 4 mrem/year.

^tAverage annual concentration assumed to produce a total body or organ dose of 4 mrem/year.

^uApproximately equal to 30 pCi/L.

selection of remedial alternatives; in particular, the NESHAPs for radionuclides will be reviewed.

5.1.1.3 Soils/sediment

Very little legislation or guidance is available governing cleanup criteria for contaminated soils at CERCLA sites. RCRA has addressed land disposal of treated hazardous wastes in its land disposal restrictions (LDRs) (40 *CFR* 268), and these will be addressed as action-specific ARARs during the FS. Since some of the sites under BCV OU 1 have been closed under RCRA, with RCRA caps and RCRA hazardous waste as defined in 40 *CFR* 261 present at the sites, they could be subject to RCRA corrective action regulations. Requirements for corrective action for SWMUs have been proposed as a new Subpart S of 40 *CFR* 264 (55 *FR* 30798, July 27, 1990). EPA has proposed a risk-based approach to establish media cleanup standards for surface water, groundwater, soil, and air. These standards are to be established at concentrations that ensure protection of human health and the environment and are to be set for each medium of concern during the remedy selection process. Target cleanup levels may initially be set at the RCRA action levels but modified as appropriate to reflect site-specific exposure assumptions (55 *FR* 30826). Final promulgation of this rule is projected for December 1993.

A portion of the proposed Subpart S was promulgated as a final rule for corrective action at Corrective Action Management Units (CAMUs) and Temporary Units (TUs). This rule establishes two new sections in Subpart S ("Corrective Action for Solid Waste Management Units"): 40 *CFR* 264.552 (CAMUs) and 40 *CFR* 264.553 (TUs) (58 *FR* 8658, February 16, 1993). Both of these units function solely to manage wastes that are generated at a RCRA facility for the purpose of implementing remedial actions (i.e., remediation wastes). As a result of this rule, regulatory requirements for remediation wastes will differ from those for "as-generated" wastes regulated under Subtitle C, especially the LDRs and minimum technology requirements (MTRs). The substantive requirements of this rule are expected to be ARARs for remediation of CERCLA sites that involve the management of RCRA hazardous waste. In this context, CAMU designations would be incorporated into CERCLA decision documents and will be analyzed further for action-specific ARARs when remedial alternatives are selected during the FS phase.

EPA has determined that the RCRA treatment standards are generally inappropriate or unfeasible when applied to contaminated soil or debris (55 *FR* 8760). Therefore, EPA proposed separate rulemakings to establish treatment standards for disposal of such contaminated soil and debris. The Notice of Proposed Rulemaking (NPRM) for debris appeared January 9, 1992 (57 *FR* 958), and the final rule on August 18, 1992 (57 *FR* 37194). The revised standards (40 *CFR* 268.45) require contaminated debris to be treated prior to land disposal using extraction, destruction, or immobilization technologies.

The NPRM for soil appeared October 24, 1991 (56 *FR* 55160); this rulemaking initiative is presently on hold while a new schedule for action is negotiated (Houlberg et al. 1993). The rule will be analyzed for ARARs or TBC when available. In the interim, EPA has developed guidance for obtaining and complying with a treatability variance for soil and debris that are contaminated with RCRA hazardous wastes for which treatment standards have already been set (OSWER Directive 9347.3-06FS, July 1989). These will be considered as TBC guidance when remedial alternatives are selected and more information becomes available on waste types.

TDEC has proposed a Hazardous Substance Site Remedial Action Rule (TDEC Rules, Chap. 1200-1-13) that is currently undergoing revisions based on public comments; it is unclear when this rule will be promulgated (Binford 1993). This rule states that, where numerical standards are not available for cleanup of soils at hazardous waste sites, soil cleanup levels shall be developed based on methods approved or determined by TDEC or by utilizing a site-specific risk assessment approved by TDEC. When promulgated, this rule will be legally applicable to cleanup at BCV OU 1; in the interim, it will serve as TBC guidance.

The regulations found in TSCA contain storage, disposal, and cleanup requirements for materials contaminated with PCBs. These regulations limit concentrations of PCBs disposed of in soil to 50 ppm and require that any soils containing concentrations of PCBs greater than 50 ppm (40 *CFR* 761.60) be incinerated (40 *CFR* 761.70) or stored in a chemical waste landfill (40 *CFR* 761.75). TSCA also specifies requirements that must be achieved for PCBs disposed of by incineration (40 *CFR* 761.70), in a chemical waste landfill (40 *CFR* 761.75), or by other disposal methods [40 *CFR* 761.60(a)(5)(iii)].

5.1.1.4 Other TBC guidance

EPA has suggested cleanup values for lead in soils based on studies of blood lead levels in exposed children. The EPA OSWER Directive 9355.4-02 (dated September 7, 1989) recommends a cleanup level for soils of 500 to 1000 ppm lead. However, EPA has distributed a draft memorandum (dated June 1992) recommending a cleanup level for lead in soil of 500 ppm, based on a new lead uptake/biokinetic model. This threshold level could be revised based on site-specific information (DOE 1993).

Although not an ARAR, EPA has published a TSCA PCB spill cleanup policy (52 *FR* 10688) which recommends cleanup standards for PCBs of 25 to 50 ppm for sites with restricted access; a 10-ppm cleanup level is recommended for residential and unrestricted access rural areas. In this latter case, a 25-cm (10-in.) cap of clean soil must cover the site. In the EPA guidance report for remedial actions at Superfund sites containing PCBs, preliminary remediation goals are set at 1 ppm for residential land use (a risk of 10^{-5}) and between 10 and 25 ppm for industrial and/or remote areas (a risk of 10^{-4}) (EPA 1990d). Alternatives should reduce concentrations to these levels or limit exposures. These values are considered TBC guidance, not ARARs.

In the absence of federal- or state-promulgated ARARs, or in the case where ARARs are not adequately protective, EPA states a preference for Reference Doses (RfDs) or reference concentrations (RfCs) for systemic toxicants, carcinogen slope factors (CSFs) for carcinogens, or Office of Drinking Water (ODW) Health Advisories (EPA 1988b) for drinking water contaminants. The RfDs, RfCs, and CSFs are available through the Integrated Risk Information System (IRIS) (EPA 1991a) and the EPA *Health Effects Assessment Summary Tables* (EPA 1991b).

The EPA ODW has developed nonregulatory Health Advisories for concentrations of noncarcinogenic contaminants in drinking water at which no adverse health effects would be expected to occur. Lifetime Health Advisories available for those COCs at BCV OU 1 which do not have promulgated MCLs and secondary maximum contaminant levels (SMCLs) are listed in Table 5.2.

5.1.2 Radiation Protection Standards

Very few applicable standards are available for the cleanup of radioactively contaminated CERCLA sites. The Atomic Energy Act (AEA) of 1954 and its amendments delegated authority for control of nuclear energy to DOE, NRC, and EPA. In addition, certain states have regulatory authority and programs for radioactive waste. EPA's regulations are derived from several other statutes as well and cover many types of activities and all types of radioactive materials. The NRC licenses the possession and use of various types of radioactive materials at certain types of facilities. Tennessee is an NRC-Agreement state and, as such, has its own authority and licensing regulations.

DOE is authorized to control all types of nuclear materials at sites under its jurisdiction and is exempt from the NRC licensing and regulatory requirements. Therefore, NRC regulations are not considered to be legally applicable for CERCLA cleanup at DOE facilities. The DOE regulations for handling and cleanup of radioactive materials are outlined in a series of internal DOE orders that are legally binding to DOE contractors but are not considered by EPA to be ARARs. The DOE orders, however, are functionally equivalent to the NRC requirements, and include all "appropriate" requirements from the NRC regulations. For the purposes of development of ARARs, DOE orders will be treated as TBC guidance rather than the NRC regulations. These NRC regulations will be considered as potentially relevant and appropriate requirements during selection of remedial alternatives during the FS.

In addition to RCRA hazardous wastes, BCV OU 1 shows evidence of contamination with radionuclides, including technetium-99, strontium-90, tritium, plutonium, and several isotopes of thorium and uranium, as well as gross alpha and beta activity. The proper definition of "mixed low-level radioactive and hazardous waste" has caused considerable debate with regard to dual jurisdiction by EPA and NRC. However, EPA has published a clarification of the problem (53 *FR* 37045, September 23, 1988), as did DOE previously [52 *FR* 15937, May 1, 1987, and DOE Order 5400.3 ("Hazardous and Radioactive Mixed Waste Program," dated February 22, 1989)]. In effect, mixed wastes are those containing a RCRA hazardous waste as defined in 40 *CFR* 261 and a radioactive waste subject to the AEA. RCRA regulations apply to the hazardous component of the waste, and AEA regulations apply to the radioactive component. When the application of both standards is conflicting or inconsistent, RCRA yields to the AEA. Tennessee received final authorization to regulate radioactive mixed waste on July 3, 1986 (53 *FR* 37045, September 23, 1988); however, the state has not implemented any regulations or guidance related to the handling of mixed waste.

On May 26, 1992, EPA published a notice in the *Federal Register* (57 *FR* 22024) proposing to find that DOE has made all but one of the demonstrations required in its application, under RCRA rules in 40 *CFR* 268.5, for a 1-year, case-by-case extension of the May 8, 1992, effective date of the LDRs applicable to certain mixed wastes generated or stored at the K-25 Site, the Y-12 Plant, and ORNL, as well as 28 other DOE facilities. Comments were due back to EPA by July 27, 1992, and no official final action has been taken on this proposal as yet. In June, 1992, EPA and DOE signed a Federal Facilities Compliance Agreement (mixed-waste FFCA) to bring mixed waste generation and storage facilities on the ORR into compliance with environmental law. The mixed waste-FFCA allows DOE ORR facilities to continue to generate and store mixed wastes while addressing LDR mixed waste compliance issues. The mixed wastes covered under the mixed waste-FFCA include flammable and corrosive liquids, solvents, paint waste, waste oils and organics, and solid mixed wastes.

In addition, the Federal Facility Compliance Act of 1992 (Public Law 102-386, October 6, 1992) amends §6001 of RCRA to waive immunity of the United States with respect to substantive and procedural requirements regarding control, abatement, or management of solid or hazardous waste. This waiver of immunity includes injunctive relief, administrative order, or civil or administrative penalties or fines, and subjects the federal government to the full range of available enforcement tools to penalize isolated, intermittent, or continuing violations. A delayed effective date of 3 years from enactment of the Federal Facility Compliance Act of 1992 is applied to the waiver of sovereign immunity for mixed waste regulated under RCRA §3004(j) so long as that waste is managed in accordance with other applicable requirements. Further, DOE is expressly exempt from the 3-year effective date for mixed waste in violation of RCRA §3004(j) so long as DOE has in effect a plan that has been submitted and approved pursuant to RCRA §3021(b) and an order requiring compliance with such plan has been issued pursuant to RCRA §3021(b). DOE has published its strategy for development of a national compliance plan for DOE mixed waste (57 *FR* 57170, December 3, 1992), and plans to publish a draft Compliance Plan in October 1993. On October 21, 1993, EPA published an Advance Notice of Proposed Rulemaking 40 *CFR* 195, "Radiation Site Cleanup Regulations." Although the proposed rule is not expected until the end of 1994, the rule is expected to be finalized before the FS timeframe. This may warrant consideration for ARARs in the RI stage.

5.1.2.1 EPA regulations

EPA has promulgated MCLs for radionuclides in community water systems. These MCLs appear in two forms—concentration limits for certain alpha-emitting radionuclides (40 *CFR* 141.15) and an annual dose limit for the ingestion of certain beta- and gamma-emitting radionuclides (40 *CFR* 141.16). MCLs and maximum contaminant level goals (MCLGs) were proposed for radon and uranium and repropoed for ²²⁶Ra and ²²⁸Ra, beta, and photon emitters on July 18, 1991. Final promulgation of the concentration limits is not expected until 1994. As with the chemical-specific MCLs, these may be used to compare to detection limits for contaminants in groundwater at BCV OU 1; they are listed in Table 5.2.

Table 5.3 lists EPA and DOE radiation protection standards that are described below. These regulations and standards generally address specific types of activities and will be analyzed in detail for action-specific ARARs during the FS phase. General requirements are discussed below.

Subpart H of 40 *CFR* 61 addresses atmospheric radionuclide emissions from DOE facilities and may be applicable to airborne emissions during cleanup of BCV OU 1. EPA has issued a final NESHAP rule (54 *FR* 51654, December 15, 1989) that limits emissions of radionuclides to the ambient air from DOE facilities to amounts that would not cause any member of the public to receive an effective dose equivalent of 10 mrem/year (40 *CFR* 61.92).

Environmental protection standards for the management, storage, and disposal of spent nuclear fuel, high-level wastes, and transuranic (TRU) wastes are found in 40 *CFR* 191. Management and storage shall be conducted so as to provide a reasonable assurance that no member of the public in the general environment shall receive a combined annual dose equivalent of greater than 25 mrem to the whole body and 75 mrem to any organ (40 *CFR* 191.03). 40 *CFR* 191.04 established alternative standards for DOE facilities if EPA determines that such standards prevent any member of the public from receiving a continuous

Table 5.3. Radiation protection standards that may be ARARs for BCV OU 1

Regulation	Applicability	Exposure conditions	Standard
40 CFR 61	National Emission Standards for Hazardous Air Pollutants for DOE facilities	Public exposure, airborne emissions	10 mrem/year
40 CFR 141	Drinking water maximum contaminant levels	Community water systems, gross beta	4 mrem/year
40 CFR 191	Spent nuclear fuel, high-level and transuranic wastes	Public exposure, all sources	25 mrem/year (total body); 75 mrem/year (thyroid)
DOE Order 5400.5 ^a	Radiation Protection of the Public and the Environment	Public exposure, all sources	100 mrem/year
		Temporary maximum exemption	500 mrem/year
		Aquatic organism exposure, absorbed dose	1 rad/d
DOE Order 5820.2A	Radioactive Waste Management	Public exposure, all sources, excluding air	25 mrem/year
		Public exposure, atmospheric releases	10 mrem/year

^aProposed as 10 CFR 834 (58 FR 16268, March 25, 1993).

exposure of more than 100 mrem/year effective dose equivalent or an infrequent exposure of more than 500 mrem/year effective dose equivalent from all sources, excluding natural background and medical exposures.

5.1.2.2 DOE orders

DOE orders are not promulgated regulations and thus are not considered to be ARARs by EPA. They are, however, binding between DOE and Energy Systems because of contractual agreements. The radiation exposure limits for the general public defined in DOE Order 5400.5 ("Radiation Protection of the Public and the Environment," February 8, 1990) are an effective dose equivalent of 100 mrem/year from all exposure pathways and all DOE sources of radiation and a dose of less than 500 mrem/year as a temporary maximum exemption under specially permitted and DOE-approved circumstances (see Table 5.3). In addition, effluent releases to surface water must not result in exposures to aquatic organisms exceeding an absorbed dose of 1 rad/d. The overriding principle of the DOE order is that all releases of radioactive material shall be as low as reasonably achievable (ALARA).

DOE Order 5400.5 lists Derived Concentration Guides (DCGs) for radionuclide isotopes which are based on a committed effective dose equivalent of 100 mrem/year for ingestion of air or water. For liquid wastes containing radionuclides that are discharged to surface waters, the best available technology must be used if the receiving water, at the point of discharge, would receive radioactive material at a concentration greater than the DCG. Implementation of the best available technology process is not required if annual releases to surface water are below the DCG. In the case of releases of multiple radionuclides, the sum of the fractional DCGs must not exceed unity. In addition, effluent releases to surface water must not result in exposures to aquatic organisms which exceed an absorbed dose of 1 rad/d. These DCGs will be addressed if any remedial alternatives result in effluent releases to air or water.

DOE Order 5400.5, Chap. IV, presents guidelines for cleanup of residual radioactive material and management of sites with residual radioactivity above the specified guidelines, and these guidelines will be analyzed for action-specific ARARs during the FS.

DOE has proposed the 5400.5 radiation protection standards for the public and the environment for codification at 10 *CFR* 834 (58 *FR* 16268, March 25, 1993). When final, these standards will be legally applicable for cleanup at DOE sites.

DOE Order 5820.2A ("Radioactive Waste Management," September 9, 1988) states that the management of low-level radioactive waste must ensure that external exposure to the waste and concentrations of radioactive material that may be released into surface water and soil does not exceed 25 mrem/year to any member of the public. Releases to the atmosphere shall not exceed 10 mrem/year (Table 5.3). Reasonable effort should be made to maintain releases to the environment to ALARA levels. Chapter II of the order addresses management of TRU wastes. The order pertains to the management of radioactive waste and contains closure and postclosure care requirements; these will be addressed for action-specific ARARs during the selection of remedial alternatives in the FS.

5.1.2.3 Other TBC guidance for radiological risk assessment

The EPA Office of Radiation Programs has derived slope and unit risk factors for radionuclides of concern at remedial sites for each of three major exposure pathways (inhalation, ingestion, and external exposure to contaminated soil). These are available in the EPA *Health Effects Assessment Summary Tables* (EPA 1991b).

5.1.3 Location-Specific ARARs

Location-specific requirements "set restrictions upon the concentration of hazardous substances or the conduct of activities solely because they are in special locations" (53 *FR* 51394). Federal and state location-specific ARARs that might be pertinent to remedial actions at BCV OU 1 are identified in the remainder of this section and listed in Table 5.4.

Geologic characteristics. There are no indications of salt-dome formations, salt-bed formations, or underground mines on or near BCV OU 1. However, if any caves are located in areas where remedial activities may occur, the regulations found in RCRA [40 *CFR* 264.18(c)] which prohibit the placement of noncontainerized or bulk liquid hazardous waste in caves might be ARARs. The faults on the ORR are ancient (pre-Holocene) and stable, and the possibility of fault movement is considered extremely unlikely (Chance 1986); therefore, seismic regulations would not be applicable.

Table 5.4. Tentative location-specific ARARs for BCV OU 1

Location characteristic(s)	Requirement(s)	Operating condition(s)	Citation(s)
Wetlands			
Presence of wetlands as defined in Executive Order 11990 §7(c) and 40 CFR 6, Appendix A §4(j)	Whenever possible, actions must avoid or minimize adverse impacts on wetlands and act to preserve and enhance their natural and beneficial values. New construction in wetlands areas should be particularly avoided unless there are no practicable alternatives Wetlands protection considerations shall be incorporated into planning, regulating, and decision-making processes	Agency action which involves: — acquiring, managing, and disposing of lands and facilities — providing federally undertaken, financed, or assisted construction and improvements — conducting federal activities and programs affecting land use	Executive Order 11990 40 CFR 6.302(a) 40 CFR 6, Appendix A 10 CFR 1022
Presence of wetlands as defined in 40 CFR 230.3(t) and 33 CFR 328.3(b)	Action to avoid degradation or destruction of wetlands must be taken to the extent possible. Discharges for which there is a practicable alternative with less adverse impacts or those which would cause or contribute to significant degradation are prohibited. If adverse impacts are unavoidable, action must be taken to enhance, restore, or create alternative wetlands	Action involving discharge of dredge or fill material into wetlands	Clean Water Act §404 40 CFR 230 33 CFR 323

Table 5.4 (continued)

Location characteristic(s)	Requirement(s)	Operating condition(s)	Citation(s)
Floodplains			
Within 100-year floodplain	Facility must be designed, constructed, operated, and maintained to prevent washout of any hazardous waste by 100-year flood	RCRA ^a -defined listed or characteristic hazardous waste (40 CFR 261) -or- RCRA-permitted treatment, storage, or disposal facility	40 CFR 264.18(b)
Within "lowland and relatively flat areas adjoining inland and coastal waters and other floodprone areas such as offshore islands, including at a minimum, that area subject to a one percent or greater chance of flooding in any given year." [Executive Order 11988 §6(c) and 40 CFR 6, Appendix A §4(d)]	Action shall be taken to reduce the risk of flood loss, minimize the impact of floods on human safety, health and welfare, and restore and preserve the natural and beneficial values of floodplains. The potential effects of actions in floodplains shall be evaluated and consideration of flood hazards and floodplain management ensured. If action is taken in floodplains, alternatives that avoid adverse effects and incompatible development and minimize potential harms shall be considered	Action which involves: — acquiring, managing, and disposing of lands and facilities — providing federally undertaken, financed, or assisted construction and improvements — conducting federal activities and programs affecting land use	Executive Order (EO) 11988 40 CFR 6.302(b) 40 CFR 6 (Appendix A) 10 CFR 1022
Aquatic resources			
Within area encompassing or affecting waters of the state of Tennessee as defined in TCA ^b 69-3-103(32) and the presence of wildlife or aquatic life	Discharge of "substances" which "will result or will likely result in harm, potential harm or detriment to the health of animals, birds, fish, or aquatic life" is prohibited	Action involving the discharge of any pollutants into the waters of the state [see TCA 69-3-103(18) and (21) for noninclusive list]	Tennessee Water Quality Control Act of 1977 (TCA 69-3-101 <i>et seq.</i>) Stream Use Classifications (TDEC ^c Rules — Chap. 1200-4-4)

Table 5.4 (continued)

Location characteristic(s)	Requirement(s)	Operating condition(s)	Citation(s)
Floodplains			
Within 100-year floodplain	Facility must be designed, constructed, operated, and maintained to prevent washout of any hazardous waste by 100-year flood	RCRA ^a -defined listed or characteristic hazardous waste (40 CFR 261) -or- RCRA-permitted treatment, storage, or disposal facility	40 CFR 264.18(b)
Within "lowland and relatively flat areas adjoining inland and coastal waters and other floodprone areas such as offshore islands, including at a minimum, that area subject to a one percent or greater chance of flooding in any given year." [Executive Order 11988 §6(c) and 40 CFR 6, Appendix A §4(d)]	Action shall be taken to reduce the risk of flood loss, minimize the impact of floods on human safety, health and welfare, and restore and preserve the natural and beneficial values of floodplains. The potential effects of actions in floodplains shall be evaluated and consideration of flood hazards and floodplain management ensured. If action is taken in floodplains, alternatives that avoid adverse effects and incompatible development and minimize potential harms shall be considered	Action which involves: <ul style="list-style-type: none"> — acquiring, managing, and disposing of lands and facilities — providing federally undertaken, financed, or assisted construction and improvements — conducting federal activities and programs affecting land use 	Executive Order (EO) 11988 40 CFR 6.302(b) 40 CFR 6 (Appendix A) 10 CFR 1022
Aquatic resources			
Within area encompassing or affecting waters of the state of Tennessee as defined in TCA ^b 69-3-103(32) and the presence of wildlife or aquatic life	Discharge of "substances" which "will result or will likely result in harm, potential harm or detriment to the health of animals, birds, fish, or aquatic life" is prohibited	Action involving the discharge of any pollutants into the waters of the state [see TCA 69-3-103(18) and (21) for noninclusive list]	Tennessee Water Quality Control Act of 1977 (TCA 69-3-101 <i>et seq.</i>) Stream Use Classifications (TDEC ^c Rules — Chap. 1200-4-4)

Table 5.4 (continued)

Location characteristic(s)	Requirement(s)	Operating condition(s)	Citation(s)
Floodplains			
Within 100-year floodplain	Facility must be designed, constructed, operated, and maintained to prevent washout of any hazardous waste by 100-year flood	RCRA ^a -defined listed or characteristic hazardous waste (40 CFR 261) -or- RCRA-permitted treatment, storage, or disposal facility	40 CFR 264.18(b)
Within "lowland and relatively flat areas adjoining inland and coastal waters and other floodprone areas such as offshore islands, including at a minimum, that area subject to a one percent or greater chance of flooding in any given year." [Executive Order 11988 §6(c) and 40 CFR 6, Appendix A §4(d)]	Action shall be taken to reduce the risk of flood loss, minimize the impact of floods on human safety, health and welfare, and restore and preserve the natural and beneficial values of floodplains. The potential effects of actions in floodplains shall be evaluated and consideration of flood hazards and floodplain management ensured. If action is taken in floodplains, alternatives that avoid adverse effects and incompatible development and minimize potential harms shall be considered	Action which involves: — acquiring, managing, and disposing of lands and facilities — providing federally undertaken, financed, or assisted construction and improvements — conducting federal activities and programs affecting land use	Executive Order (EO) 11988 40 CFR 6.302(b) 40 CFR 6 (Appendix A) 10 CFR 1022
Aquatic resources			
Within area encompassing or affecting waters of the state of Tennessee as defined in TCA ^b 69-3-103(32) and the presence of wildlife or aquatic life	Discharge of "substances" which "will result or will likely result in harm, potential harm or detriment to the health of animals, birds, fish, or aquatic life" is prohibited	Action involving the discharge of any pollutants into the waters of the state [see TCA 69-3-103(18) and (21) for noninclusive list]	Tennessee Water Quality Control Act of 1977 (TCA 69-3-101 <i>et seq.</i>) Stream Use Classifications (TDEC ^c Rules — Chap. 1200-4-4)

Aquatic resources. There are no known designated wilderness areas, wildlife refuges, or scenic rivers on the ORR or within range of the reservation such that remedial action would likely impact these resources. However, if any remedial action is taken which affects wildlife or alters Bear Creek, which is designated by the state of Tennessee for protection of fish and aquatic life, recreation, irrigation, and wildlife and livestock watering, then the Fish and Wildlife Coordination Act (16 USC 661 *et seq.*), the Tennessee Water Quality Control Act of 1977 (TCA 69-3-101 *et seq.*), the Clean Water Act §404, 40 CFR 230, 40 CFR 323, and 40 CFR 6.302(g) may be applicable. If any action involves the discharge of dredge or fill material into an aquatic system, the Clean Water Act §404 and 40 CFR 230 may also be implicated.

Slightly over 2023 ha (5000 acres) of the ORR have been designated as a DOE NERP, which also includes research areas both within and without the NERP itself. In addition, DOE has also designated areas on the ORR as DOE-NERP Reference Areas and DOE-NERP Natural Areas. However, none of the designated research areas, reference areas, or natural areas are on or near BCV OU 1 (Parr and Pounds 1987); therefore, none of the uses and restrictions applicable to these resources would be ARARs for BCV OU 1.

Wetlands and floodplains. It appears that parts of BCV OU 1 may be located in the 100- and 500-year floodplains of Bear Creek. If any remedial alternatives are selected that would impact these floodplains, the requirements found in Executive Order 11988, 40 CFR 264.18(b), 40 CFR 6.302(b), 40 CFR 6 (Appendix A), and 10 CFR 1022 would provide ARARs for the BCV OU 1 site. A preliminary survey of the wetlands on the ORR revealed the presence of several wetlands areas near the BCV OU 1 sites, specifically, just north of the Burial Grounds (Parr 1992). If wetlands will be affected by any remedial alternative chosen for the BCV OU 1 site, consideration should be given to Executive Order 11990, 40 CFR 6.302(a), 40 CFR 6 (Appendix A), 10 CFR 1022, CWA §404, and 40 CFR 230 for applicable requirements.

Rare, threatened, or endangered species. There are a number of federal- and state-listed endangered or threatened plant and animal species on the ORR. An in-depth discussion of these species appears in Volume No. 24 of *The Resource Management Plan of the Oak Ridge Reservation* (Kroodsmma 1987). The state of Tennessee has also designated approximately 80 species of wildlife as "in need of management," and some of those occur in the area including the ORR (Tennessee Wildlife Resources Commission Proclamation 86-29). Pursuant to TCA 70-8-104(b), it is unlawful for any person to knowingly destroy the habitat of "wildlife in need of management." Should any remedial actions at the BCV OU 1 site impact any federally listed endangered or threatened species, the provisions found in the Endangered Species Act of 1973 (16 USC 1531 *et seq.*), 50 CFR 492, 40 CFR 6.302(h), and the Fish and Wildlife Coordination Act (16 USC 661 *et seq.*) may be implicated as ARARs. If any proposed actions impact state-listed endangered or threatened animal species, the Tennessee Non-Game and Endangered or Threatened Wildlife Species Conservation Act of 1974 (TCA 70-8-101 *et seq.*) may provide ARARs. The prohibitions of the Tennessee Rare Plant Protection and Conservation Act of 1985 (TCA 11-26-201 *et seq.*) do not apply to a landowner, lessee, or other person entitled to possession of the land on which the species is located (TCA 11-26-209). This also includes managers in the case of publicly owned land and those with written permission of the landowner or manager (TCA 11-26-209). These exclusions would apparently apply to ORR. However, the purpose of the statute, to protect and preserve rare plants, should be considered TBC guidance for any remedial actions at BCV OU 1 if rare plants are identified.

Historic sites and archaeological findings. The region surrounding ORR is rich in both archaeological and historic resources, and a number of studies have indicated the presence of

abundant resources on the reservation. These surveys are summarized in Volume 3, Appendix B, of the *Resource Management Plan for the U.S. Department of Energy Oak Ridge Reservation* (Sanders 1984). No archaeological sites were identified at the Y-12 Plant; in addition, there are no historic buildings at BCV OU 1. In the event a NEPA survey indicates the presence of archeological resources at the BCV OU 1 site, the Archaeological Resources Recovery Act (16 USC 470aa-ll), the Archaeological and Historic Preservation Act (16 USC 469a-c), 43 CFR 7, and 40 CFR 6.301 may provide ARARs.

5.1.4 Action-Specific ARARs

Performance, design, or other action-specific requirements set controls or restrictions on particular kinds of activities related to the management of hazardous waste (52 FR 32496). Selection of a particular remedial action at a site will invoke the appropriate action-specific ARARs that may specify particular performance standards or technologies as well as specific environmental levels for discharged or residual chemicals. Action-specific ARARs are established under RCRA, CAA, CWA, SDWA, and TSCA. Action-specific ARARs for BCV OU 1 will be developed during selection of remedial alternatives.

5.2 IDENTIFICATION OF PRELIMINARY REMEDIATION GOALS

The first step in the FS scoping process is to develop remedial action objectives (RAOs). RAOs are site-specific, qualitative goals that define the extent of cleanup required to achieve a CERCLA response action (EPA 1988a). RAOs address COCs, media of concern, potential exposure pathways and remediation goals (EPA 1991c). General RAOs for BCV OU 1 follow:

- Prevent current and future direct exposures to contaminants resulting in an incremental lifetime cancer risk greater than 10^{-6} or a hazard index greater than 1.
- Prevent transport of contaminants from the sources to the groundwater or surface water that would result in unacceptable risks to the humans and aquatic biota in the receiving water bodies (i.e., those media addressed as part of BCV OU 4).
- Prevent current and future exposures of ecological receptors to contaminants resulting in body burdens harmful to the receptors or their predators.

Preliminary remediation goals (PRGs) are a subset of RAOs. PRGs are chemical-specific, medium-specific numerical concentration limits that address all contaminants and all pathways found to be of concern at the site. Guidance is available from the EPA for developing risk-based PRGs (EPA 1991c). PRGs are developed early in the RI/FS process. They are dependent on the identification of ARARs as well as on the knowledge of the risk assessment process (EPA 1991c). PRGs should be based on readily available environmental or health-based ARARs, ambient water quality criteria, and other criteria, advisories, or guidance (EPA 1991c). Many identified ARARs have been derived from risk levels that would not meet the CERCLA objectives of "protectiveness of human health." In other words, PRGs based on ARARs could be less stringent than criteria based on the 10^{-4} to 10^{-6} risk level.

PRGs for BCV OU 1 must be developed that address both RAOs. Because few ARAR exist that address direct contact with soils, risk-based PRGs must be developed to address the first objective. PRGs for the second objective can be addressed using a simple fate and transport

model developed specifically to establish soil cleanup levels based on potential migration to groundwater. Both of these methods are discussed in detail below.

5.2.1 Soil PRGs Based on Direct Exposures

Soil risk-based PRGs are developed using guidance from EPA, "Development of Risk-Based Preliminary Remediation Goals" (EPA 1991c). Development of initial risk-based PRGs requires the following information:

- Chemicals of potential concern.
- Environmental media of potential concern (soil).
- Probable future land use and exposures.
- Chemical-specific toxicity information.
- Target risk levels.

A preliminary list of chemicals of potential concern have been identified in previous studies and are listed in Table 5.1.

As discussed in Section 4.1.1, it is premature to designate a probable future land use to develop PRGs. In lieu of actual knowledge of future land use, it would be premature to make screening decisions without considering EPA's default residential scenario, the same scenario on which RCRA soil action levels are based. Thus, PRGs will be presented for both industrial and residential land uses. Pathways evaluated for the residential scenario include incidental soil ingestion, inhalation of volatiles and particulates, and food chain uptake. The sole pathway evaluated for the industrial scenario is incidental soil ingestion (Fig. 3.33).

Toxicity data used to develop PRGs are cancer slope factors and reference doses from the IRIS database (EPA 1992f) and Health Effects Assessment Summary Tables (HEAST) (EPA 1991b; 1992e).

Target risk levels (TR) must be established for carcinogens and a target hazard quotient (THQ) and target hazard index (THI) must be established for noncarcinogens. Once these levels are established, they can be used in conjunction with toxicity data and exposure equations to calculate PRGs. The NCP and EPA (1991c) suggest a default target risk of 10^{-6} and a cumulative site HI < 1 (EPA 1991c). Site-specific data can be collected on several parameters listed in Table 5.5. This will be done prior to using PRGs for Stage 1 risk assessment work. (See Sect. 6.4.1 in Chap. 6.)

The equations used to calculate industrial and residential exposures are derived from EPA (1991c). The parameters used are listed in Table 5.5. Some parameters used to determine the chemical-specific volatilization factors for volatile organics can be found in EPA (1991c).

Table 5.5. Parameters used to calculate preliminary remediation goals

Parameter	Units	Industrial land use		Residential land use	
		Cancer effects	Noncancer effects	Cancer effects	Noncancer effects
Target risk (TR)	unitless	1×10^{-6}		1×10^{-6}	
Target HI (THI)	unitless		0.2		0.2
Body weight (BW)	kg	70	70		
Averaging time (AT)	year	70	25	70	30
Exposure frequency (EF)	days/year	250	250	350	350
Exposure duration (ED)	year	25	25		
Soil ingestion rate (IR_s)	mg/d	50	50		
Age-adjusted soil ingestion factor (IF)	mg-year/kg-d			114	114
Inhalation rate (IR_a)	m ³ /d	20	20		
Particulate emission factor (PEF)	m ³ /kg	4.63×10^9	4.63×10^9		
Volatilization factor (VR)	m ³ /kg	Chemical-specific ^a	Chemical-specific ^a		

^aEPA 1991c (page 29, equation 8).

PRGs for soil ingestion by biota can be calculated similarly. However, because RfDs for humans include larger safety factors than are necessary for biota, soil PRGs for biota are typically higher than those for humans. Therefore, the more conservative human values would take precedence.

5.2.2 Soil PRGs Based on Migration to Groundwater

The second RAO for BCV OU 1 is to prevent contaminants from leaching from the source to groundwater or surface water at levels that would be unacceptable in those water bodies. EPA has developed several models for use in determining soil cleanup levels based on potential contaminant migration to the groundwater (EPA 1989c). No such model is known to exist for migration to surface water at this time, but several models such as the Modified Universal Soil Loss Equation are being reviewed for use in this capacity. Soil PRGs for aquatic biota could be calculated from soil transport modeling results along with equilibrium partitioning data on COCs and the corresponding AWQC or State of Tennessee criterion continuous concentration.

One model for addressing migration to the groundwater is the revised Summers Model (Summers et al. 1980; EPA 1989c). The Summers Model is a simple one-dimensional analytical

model that requires minimal data. For this reason, it is appropriate for developing PRGs. For actual modeling for the RI and BRA, more sophisticated models, such as SESOIL or MULTIMED, will be used. Parameter values used in the Summers Model for developing PRGs are listed in Table 5.6. In some cases, default parameters were used, which will be updated as site-specific data are gathered during Phase 1 sampling. These parameter values reflect conservative estimates of what is currently known about the hydrologic conditions at the site. Although the values can be used to estimate PRGs, additional data will need to be obtained for the RI report. The revised Summers Model makes the following assumptions:

- The soil/water system is at equilibrium.
- No contaminant degradation is occurring.
- The unsaturated soil zone is homogeneous down to the aquifer.
- Contaminants are mixed throughout the depth of the aquifer beneath the contaminant source.

The model does not account for any contaminant dilution or attenuation due to horizontal transport within the aquifer. Acceptable soil concentrations are therefore determined based on the assumption that groundwater must meet acceptable levels within the groundwater directly beneath the source. Groundwater PRGs are being developed as part of BCV OU 4. In lieu of risk-based groundwater PRGs, MCLs will be used as the acceptable groundwater level.

The equations presented below have been incorporated into a Microsoft® EXCEL spreadsheet model to calculate PRGs. Because ranges or distributions of data exist for several of the above parameters, Crystal Ball® software can be used to provide decision makers with the distribution of PRG values that reflect the uncertainty associated with the input parameters. Decision makers can then determine the confidence level on which they will select the PRG. Most decisions made in the early stages of CERCLA reflect at least an upper 95% confidence on the mean result. This capability may be extremely useful at the point in the FS-Proposed Plan process when PRGs must be revised into final remediation levels. The Summers Model is described below.

The concentration of a chemical in groundwater is a function of the amount of the chemical infiltrating through the soil column to the groundwater and the amount of the chemical already present in the groundwater. The chemical concentration is also determined by the volume of water into which the leachate is dissolved. The equation for the Summers Model is

$$C_w = \frac{(Q_p C_p) + (Q_a C_a)}{Q_p + Q_a},$$

where

C_w = PRG concentration in water (mg/L),

Table 5.6. Parameters used in the Summers Model

Parameter	Symbol	Value	Units	Reference
Volumetric flow rate of infiltration	Q_p	6176.5	ft ³ /day	Calculated by $Q_p = V_{dz} \times A_p$ (EPA 1989c)
Darcy velocity in downward direction	V_{dz}	0.28	ft/day	Hydraulic conductivity for unsaturated zone
Horizontal area of contamination	A_p	21789	ft ²	Default value from EPA, 1991a (2025 m ²); Also used to calculate VF and PEF
Volumetric flow rate of groundwater	Q_a	37.7	ft ³ /day	Calculated by $Q_a = V_d \times h \times w$ (EPA 1989c)
Hydraulic conductivity	K	0.28	ft/day	Average value for Y-12 area (Solomon et al. 1992)
Hydraulic gradient	I	0.03	unitless	Estimated from potentiometric maps of the Y-12 area
Darcy velocity in aquifer	V_d	0.0085	ft/day	Calculate by $V_d = K \times I$
Aquifer thickness	h	30	ft	Average thickness of shallow groundwater zone at Y-12
Perpendicular spill width	w	147.6	ft	Default value from Risk Assessment Guidance for Superfund Part B (EPA 1991c)

Q_p = volumetric flow rate of infiltration into the aquifer (ft³/day), where $Q_p = V_{dz} \cdot A_p$, V_{dz} = Darcy velocity in downward direction (ft/day) and A_p = horizontal area of spill (ft²),
 C_p = concentrations of pollutant in the infiltration at the unsaturated-saturated zone interface (μg/L),

Q_a = volumetric flow rate of groundwater (ft³/day), where $Q_a = V_d \cdot h \cdot w$ and V_d = Darcy velocity in aquifer (ft/day), h = aquifer thickness (ft), and w = width of spill perpendicular to flow direction in aquifer (ft),

C_a = initial or backward concentration of pollutant in aquifer (mg/L).

V_{dz} is estimated as the average annual precipitation minus surface runoff and evapotranspiration for the area, assuming all precipitation infiltrated through the soil.

The Darcy velocity in the aquifer (V_d) is estimated by

$$V_d = KI,$$

where K is the hydraulic conductivity (ft/day) and I is the hydraulic gradient (unitless).

It was assumed that the background concentrations of the chemicals (C_a) were equal to zero, and equations were rearranged to solve for C_s , the PRG concentration in soil:

$$C_s = \frac{C_{gw}(Q_p + Q_d)K_d}{Q_p},$$

where K_d is the soil/water equilibrium partitioning coefficient (mL/g).

5.2.3 Soil PRGs for BCV OU 1

Table 5.7 presents soil PRGs for use in performing the BCV OU 1 RI. In keeping with the spirit of the guidance that PRGs should reflect readily available information *and* be health protective the table presents risk-based PRGs, ARARs/TBC, and other considerations for establishing PRGs, such as background and reasonable cost-effective detection limits.

Values are presented for each of the three methods described above: direct exposures, both industrial and residential land use, and exposures via potential migration to groundwater. Note that all PRGs are above CLP detection limits; thus these detection limits should provide data that is useable in the risk assessment process. Also of note is that PRGs for chemicals that are relatively soluble (i.e., tetrachloroethylene and trichloroethylene) are lower than PRGs based on direct contact. On the other hand, PRGs for PCBs (Aroclors) and several metals are more stringent via direct contact pathways because of the low migration potential of these relatively insoluble compounds. Information used to develop the PRGs based on the leaching potential is preliminary and thus these values are subject to change.

In using initial PRGs in the early stages of the alternative screening process, engineers should understand that PRGs may be modified and should make the design of alternatives flexible. Chemicals may be added or deleted from the list of chemicals of concern, or PRGs may need to be modified based on the identification or deletion of additional limiting pathways. Modified PRGs will be presented in the operable unit feasibility study reports. Modified PRGs are refined into final remediation levels and presented in the Record of Decision. Final remediation levels must meet threshold criteria of "protection of human health and the environment" and "compliance with ARARs" (EPA 1991c):

However, the NCP also allows for modification of PRGs during final remedy selection based on the "balancing" and "modifying" criteria and factors relating to uncertainty, exposure and technical feasibility. [EPA 1991c]

Note should be taken that, with the exception of recommending the inclusion of environmental ARARs in the selection of PRGs, the Risk Assessment Guidance for Superfund RAGS Part B (EPA 1991c) addresses mainly human health effects, especially for soil PRGs. The potential need for ecological-based soil PRGs is being evaluated.

Table 5.7. Soil preliminary remediation goals

Chemical	Risk-based PRGs					ARAR/ TBC- based PRGs (mg/kg)	Other considerations	
	Direct contact with soil				Protection of groundwater			
	With Institutional Controls - industrial worker ^a		Without Institutional Controls - EPA default residential ^a		Leaching to groundwater - Summer's Model ^c			
	Cancer effects ^d (mg/kg)	Toxic effects ^e (mg/kg)	Cancer effects (mg/kg)	Toxic effects (mg/kg)	MCL-based ^f (mg/kg)		Background ^g (mg/kg)	CLP detection limits ^h (mg/kg)
Arsenic	3.3E+00	6.1E+02	3.7E-01	8.2E+01	1.0E+01	--	7.4E+00	
Barium	NC	1.4E+05	NC	1.9E+04	2.4E+03			
Beryllium	1.3E+00	1.0E+04	1.5E-01	1.4E+03	1.3E+00	--	8.5E-01	
Boron	NC	1.8E+05	NC	2.5E+04	NA	--	--	
Cadmium (food)	1.1E+04	2.0E+03	NC-O	2.7E+02	2.5E+00	--	--	
Chromium (VI)	1.6E+03	1.0E+04	NC-O	1.4E+03	1.5E+02	--	5.2E+01	
Cobalt	NC	5.3E+03	NC	7.1E+02	NA	--	9.2E+00	
Copper	NC	7.6E+04	NC	1.0E+04	NA	--	2.2E+01	
Lead	NC	NE ^k	NC	NE	1.5E+01	500-1000 ^k	1.7E+01	
Mercury	NC	6.1E+02	NC	8.2E+01	2.0E-02	--	1.2E-01	
Nickel	NC	4.1E+04	NC	5.5E+03	6.5E+01	--	1.8E+01	
Nitrate	NC	3.3E+06	NC	4.4E+05	NA	--	--	
Nitrite	NC	2.0E+05	NC	2.7E+04	NA	--	--	

Table 5.7 (continued)

Chemical	Risk-based PRGs					ARAR/ TBC- based PRGs (mg/kg)	Other considerations	
	Direct contact with soil				Protection of groundwater			
	With Institutional Controls - industrial worker ^a		Without Institutional Controls - EPA default residential ^b		Leaching to groundwater - Summer's Model ^c			
	Cancer effects ^d (mg/kg)	Toxic effects ^e (mg/kg)	Cancer effects (mg/kg)	Toxic effects (mg/kg)	MCL-based ^f (mg/kg)		Background ^g (mg/kg)	CLP detection limits ^h (mg/kg)
Selenium	NC	1.0E+04	NC	1.4E+03	3.9E+01			
Thallium	NC	1.4E+02	NC	1.9E+01	3.0E+00	--	8.6E+00	
Uranium ⁱ	NC	6.1E+03	NC	8.2E+02	3.4E+01	--	1.0E+00	
ORGANICS								
Acetone	NC	2.0E+5	NC	2.7E+4	NA	--	--	
Aroclor-1248	7.4E-01	NT	8.3E-02	NT	5.4E+00	10-50 ^m	--	
Aroclor-1254	7.4E-01	NT	8.3E-02	NT	1.0E+01	10-50	--	
Aroclor-1260	7.4E-01	NT	8.3E-02	NT	1.2E+01	10-50	--	
2-Butanone	NC	1.0E+05	NC	1.4E+04	NA	---	--	
Carbon Tetrachloride	4.4E+01	1.4E+03	4.9E+00	1.9E+02	5.2E-02	--	--	
Chloroform	9.4E+02	2.0E+04	1.0E+02	2.7E+03	1.8E-01	--	--	
1,1-Dichloroethane	NC	2.0E+05	NC	2.7E+04	6.3E-03			
1,1-Dichloroethylene	9.5E+00	1.8E+04	1.1E+00	2.5E+03	4.3E-03			

Table 5.7 (continued)

Chemical	Risk-based PRGs					ARAR/ TBC- based PRGs (mg/kg)	Other considerations	
	Direct contact with soil				Protection of groundwater			
	With Institutional Controls - industrial worker ^a		Without Institutional Controls - EPA default residential ^b		Leaching to groundwater - Summer's Model ^c			
	Cancer effects ^d (mg/kg)	Toxic effects ^e (mg/kg)	Cancer effects (mg/kg)	Toxic effects (mg/kg)	MCL-based ^f (mg/kg)		Background ^g (mg/kg)	CLP detection limits ^h (mg/kg)
1,2-Dichloroethylene	NC	1.8E+04	NC	2.5E+03	1.8E-01	--	--	
Methylene Chloride	7.6E+02	1.2E+05	8.5E+01	1.6E+04	1.7E-03	--	--	
Tetrachloroethylene	1.1E+02	2.0E+04	1.3E+01	2.7E+03	3.3E-02	--	--	
Trichloroethylene	5.7E+02	NT	6.4E+01	NT	3.3E-02	--	--	
1,1,1-Trichloroethane	NC	1.8E+04	NC	2.5E+04	NA	--	--	
RADIONUCLIDES (pCi/g)^a								
Plutonium 238	3.1E+00	NT	5.0E-02	NT	NA			
Strontium 90	9.7E+01	NT	2.4E+01	NT	NA			
Technetium 99	1.7E+02	NT	2.3E+00	NT	NA			
Tritium	5.9E+04	NT	1.5E+04	NT	NA			
Thorium 228	2.0E-01	NT	2.6E-03	NT	NA			
Thorium 230	2.1E+00	NT	2.6E-02	NT	NA			
Thorium 232	4.2E+00	NT	5.4E-02	NT	NA			

Table 5.7 (continued)

Chemical	Risk-based PRGs					ARAR/ TBC- based PRGs (mg/kg)	Other considerations	
	Direct contact with soil				Protection of groundwater			
	With Institutional Controls - industrial worker ^a		Without Institutional Controls - EPA default residential ^b		Leaching to groundwater - Summer's Model ^c			
	Cancer effects ^d (mg/kg)	Toxic effects ^e (mg/kg)	Cancer effects (mg/kg)	Toxic effects (mg/kg)	MCL-based ^f (mg/kg)		Background ^g (mg/kg)	CLP detection limits ^h (mg/kg)
Uranium 234	3.7E+00	NT	4.7E-02	NT	NA			
Uranium 235	4.7E-04	NT	5.9E-06	NT	NA			
Uranium 238	5.2E+00	NT	6.7E-02	NT	NA			

^aPathways evaluated for the worker exposure include incidental ingestion of soil and inhalation of particulate and volatile emissions (EPA 1991a).

^bPathway evaluated for the residential scenario is soil ingestion (EPA 1991c). Parameters used are listed in Table 5.1.

^cPRGs based on leaching to groundwater are calculated using the Summer's Model (EPA 1989c). Data required for the model are listed in Table 5.1. K_d s used are for clayey soils (Energy Systems 1992a).

^dDeveloped using above mentioned pathways and cancer slope factors from the IRIS database (EPA 1992f), the Health Effects Assessment Summary Tables (HEAST) (EPA 1991b, 1992e).

^eDeveloped using above mentioned pathways and reference doses from the IRIS database (EPA 1992f), the Health Effects Assessment Summary Tables (HEAST) (EPA 1991b, 1992e).

^fSoil cleanup level is based on preventing concentrations in the groundwater from reaching MCL levels. If MCL was not available, no PRG soil value was estimated.

^gUnited States Geological Survey (USGS) data are reported for comparison purposes. Data represent mean value for the Eastern United States (Shacklette and Boerngen 1984). Site-specific data will be used as they become available.

^hContract required detection limits from EPA CLP Statement of Work: Inorganics - ILM 02.0, Organics - OLM 01.8. A conversion of 1 gram to mL volume was used to convert water limit to soil limits.

ⁱ"-" = no ARAR or TBC available. Few soil ARARs exist.

^jNC = Not a carcinogen; NC-O = Not an oral carcinogen; NT = Not a chemical toxicant; NA = No data available.

^kNE = Not estimated. EPA has requested that lead toxicity be evaluated using the EPA Lead Biokinetic/Uptake Model. Using this model, EPA estimated that acceptable lead concentrations are between 500-1000 ppm (Clay 1991).

^lPRG estimated based on uranium toxicity.

^mFrom TSCA PCB spill cleanup policy (52 FR 10688).

5.3 IDENTIFICATION OF POTENTIAL CLASSES OF REMEDIAL TECHNOLOGIES

The two types of response actions allowed by CERCLA regulations at a site are removal actions and remedial actions. A removal action is designed to respond to immediate threats to public health or welfare or to the environment and can be undertaken immediately upon the discovery of the threat. On the other hand, remedial actions are designed to reduce risk and must be preceded by the RI/FS process. A removal response can be followed by, or be a part of, remedial response.

5.3.1 Removal Actions

EPA-established requirements for removal actions taken at CERCLA sites are listed in Sects. 300.410 and 300.415 of the NCP. Classes of removal actions include the following:

- Cleaning up or removing released hazardous substances.
- Acting in the event of the threat of release of hazardous substances into the environment.
- Monitoring, assessing, and evaluating the release or threat of release of hazardous substances.
- Preventing, minimizing, or mitigating damage to the public health or welfare or to the environment.

Removal actions must, to the extent practical, contribute to the efficient performance of any long-term remedial action for the release [55 FR 8843 and CERCLA Sect. 104(b)].

The following is a list of potential removal actions that may be necessary at BCV OU 1:

- Removal or overpacking of drums and other containers.
- Excavation of highly contaminated soils, especially those that are radioactive.
- Drainage controls (i.e., capture and treatment) of contaminated shallow groundwater.
- Drainage controls (i.e., capture and treatment) of contaminated seeps.
- Fences, warning signs, or other security measures.

5.3.2 Remedial Actions

EPA-established requirements pertaining to the remediation of CERCLA sites are listed in Sect. 300.430(a)(1)(iii) of the NCP. These requirements have been considered in developing the PRGs, which help to identify appropriate remedial action alternatives. The classes of remedial technologies that may be appropriate to BCV OU 1 are summarized below.

- **Institutional controls** such as fences and deed restrictions may be used alone or as a supplement for engineering controls, as appropriate, to prevent exposure to hazardous substances.
- **Containment technologies** such as capping, vertical barriers, horizontal barriers, and dust suppression may be used, as appropriate, to control contaminated areas that pose a

relatively low long-term threat and for sites where treatment is impractical (e.g., in a pyrophoric setting). Hydraulic controls are considered containment technologies.

- **Removal technologies** such as excavation may be used, as appropriate, alone or in combination with treatment to remove contamination or to abate the further migration of contaminants.
- **Treatment technologies** such as solidification/stabilization or physicochemical, biological, thermal, or electrokinetic technologies may be used, as appropriate, alone or in combination with any of the above technologies to (1) abate principal threats such as highly mobile or toxic substances, (2) prevent further migration of contaminants, or (3) remove contamination. Where possible, treated material that no longer poses a threat to human health or the environment will be returned to the excavation site or treatment will occur in situ.
- **Innovative technologies** may be considered when they offer the potential for superior treatment performance or lower costs for performance similar to that of demonstrated technologies.

During the FS, alternatives will be developed to protect human health and the environment by eliminating, reducing, and/or controlling risks posed through each exposure pathway. The development and evaluation of the remedial action alternatives will reflect the scope and complexity of the site problems being addressed.

Table 5.8 presents a preliminary range of general response actions and associated technologies that may be used to develop alternatives to achieve the RAOs discussed in Sect. 5.2. This table is an initial determination as to which general response actions might be applied to the units. Identifying potential technologies at this stage will help identify data needed to evaluate them. Early identification of technologies will also determine whether treatability studies are needed.

The general response actions shown in Table 5.8 are (1) no action, (2) institutional actions, (3) source containment, and (4) removal/treatment actions. The "no-action" response required by CERCLA serves as a baseline comparison against which the risk association with all other actions can be measured. The potential principal contaminants of interest affecting the soil were listed earlier in Table 5.1. They consist mainly of metals, nitrates, organics, and radionuclides. As additional information becomes available, contaminants may be added or deleted.

Table 5.8. Identification of preliminary general response actions, and preliminary remedial technologies for soil and waste

General response actions	Remedial technology
No action:	No action
Institutional controls:	Institutional technologies <ol style="list-style-type: none"> 1. Access barriers/fences 2. Deed restrictions 3. Monitoring
Containment actions:	Containment technologies: <ol style="list-style-type: none"> 1. Capping 2. Vertical barriers 3. Horizontal barriers 4. Hydraulic isolation
Removal actions:	Removal technologies: <ol style="list-style-type: none"> 1. Excavation
Treatment:	Treatment technologies: <ol style="list-style-type: none"> 1. Physicochemical treatment 2. Biological treatment 3. Thermal treatment 4. Electrokinetic treatment 5. In situ treatment
Disposal:	Disposal technologies: <ol style="list-style-type: none"> 1. On-site disposal 2. Off-site disposal

6. SITE DATA NEEDS

The purpose of this chapter is to identify the data necessary to meet the objectives of the RI that were discussed in Sects. 1.7 and 1.8. As outlined in Chap. 1, the staged RI/FS process will be carried out in a three-stage approach:

- **Stage 1** involves a preliminary hazard evaluation for each site, and a decision will be made whether a removal action, interim action, no action, or an additional RI is necessary.
- **Stage 2** involves a baseline risk assessment and risk-based remedial goals, screening of remedial action alternatives for each site, and issuance of a ROD.
- **Stage 3** involves bringing the source OUs and the integrator OU in BCV to a valley-wide ROD.

The objective of the RI is to provide legally defensible data that will allow the necessary decisions to be made at each stage of the RI/FS process. The field work described in this work plan is designed to obtain data that will be adequate to complete Stage 2 and will provide most of the data relating to these source areas for Stage 3.

The general data needs for this RI include physical property parameters for soils, contaminant concentration data for soils and shallow groundwater, data for waste characterization, and data concerning the hydrological and geochemical properties of soils and the shallow groundwater system. Much of these data are available in the literature from previous studies and, where possible, will be incorporated into the RI to reduce the amount of sampling required to meet the DQOs. Data used for the RI must be supported by sufficient documentation to ensure that the data are valid for decisions made using these data to be legally defensible. At Stage 1, decisions for removal actions or RI continuance can be made based on data for which documentation of QA is below the equivalent of EPA Level III. Therefore, Stage 1 of the staged RI/FS process has been completed without further sampling at most of the sites in BCV OU 1, for which extensive historical data are available, and Stage 2 is beginning (Fig. 6.1). Where it is possible that a no-further-investigation decision will be made at a site, a goal of EPA Level III (or equivalent documentation at Stage 1) will be set to support this decision.

The sites in BCV OU 1 are at the following stages in the staged RI/FS process:

- **Stage 1:** Boneyard/Burnyard, HCDA, and SL 1.
- **Stage 2:** S-3 Site, Oil Landfarm, and the Burial Grounds.

The goal for Stage 2 is to develop a single, consistent database for soils, groundwater, and surface water—a database that has undergone sufficient scrutiny to be defensible. To achieve this goal all data will undergo validation at a modified Level III, and 10% at Level IV, as described in Sect. 8.9. For historical data to reach this goal, sampling may be required in order to confirm values where documentation of historical QA data is insufficient. The general data needs for the BCV OU 1 sites are described in Table 6.1. Data needs for BCV OU 1 were developed using assumptions made with the site background information described in Chap. 2 and the historical data described in Chap. 3.

CERCLA process for source sites not closed under RCRA

CERCLA process for source sites closed under RCRA

Fig. 6.1. The staged RI/FS process.

Table 6.1. BCV OU 1 RI work plan site data needs

Data need	Data type	Sample type or method	Media	Data use	Data evaluation and interpretation	Quality level
Physical properties	Soil grain size, permeability, void ratio, soil density, bulk density, total organic carbon content, cation exchange capacity, soil pH ^a	Undisturbed samples (Shelby tubes)	Unconsolidated earth materials (includes soil, fill, saprolite)	<ul style="list-style-type: none"> Evaluate release mechanisms for transport from the source to groundwater, surface water, and air pathways Screen remediation technologies Remediation design 	These data will be illustrated on maps and cross-sections. Data will be used in conjunction with analytical data to estimate the mass flux of contaminants transported from the source to the saturated groundwater environment	
	Hydraulic head measurements	Electronic probe graduated to 0.01 ft	Groundwater	<ul style="list-style-type: none"> Determine hydraulic gradients; establish thickness of unsaturated zone; establish seasonal fluctuation in water table Screen remediation technologies Remediation design 	Data will be illustrated as potentiometric maps, hydraulic gradients will be input into mass transport calculations. Hydrographs will also be made to establish seasonal variation in the water table	
	Standard compaction test	Undisturbed sample (Shelby tube)	Soil	<ul style="list-style-type: none"> Determine compactibility of soil, plastic and liquid limits, and compressibility Screen remediation technologies Remediation design 	<p>Volume reduction of compacted soil</p> <p>Soil behavior on wetting</p> <p>Determine load bearing capacity of soil</p> <p>Data will be illustrated on tables</p>	
	Topographic survey	Survey	Soil	Assess surface contours	Determine volume estimates Grading requirements	

Table 6.1 (continued)

Data need	Data type	Sample type or method	Media	Data use	Data evaluation and interpretation	Quality level
	Visually inspected soil properties, and split barrel penetration blow counts	Soil/split barrel penetration tests	Unconsolidated earth materials (includes soil, fill, and saprolite)	<ul style="list-style-type: none"> Evaluate release mechanisms for transport from the source to groundwater and surface water pathways; establish thickness and volume of unconsolidated zone with the site. These data will be used to qualitatively interpret the distribution of physical properties determined from laboratory analysis Screen remediation alternatives Remediation design 	Data will be plotted on boring logs, maps, and cross-sections and will be used to establish the relative degree of heterogeneity within the unconsolidated zone	
Contaminant distribution	Analytical parameters to be determined for each site to be investigated	Soil samples will be collected by hand auger or split-spoon. CLP methods will be used for laboratory analysis	Soil (source)	<ul style="list-style-type: none"> Define the vertical and horizontal extent of contamination within the source area. Define the boundary and volume of soil to be remediated. Data will also be input into mass transport calculations to assess the future threat of release to groundwater and surface water environments. Data will be used as input for the human health and ecological risk assessments 	Data will be presented on maps and cross-sections along with physical property data; qualitative correlations will be made between soil physical properties and contaminant distribution and occurrence; statistical analysis will be performed and the data presented in table format	90% delivered at EPA Level III and validated at modified Level III as described in Sect. 8.9. 10% delivered at EPA Level IV and validated at Level IV
Leachability of contaminants	Metals including Hg, U, and organics; analytes detailed in 40 CFR 260	EPA SW 846		Define the relative leachability of the contaminants. Data will be used as input in the ecological risk assessment	Data will be evaluated in conjunction with leaching models	

Table 6.1 (continued)

Data need	Data type	Sample type or method	Media	Data use	Data evaluation and interpretation	Quality level
	Analytical parameters to be determined for each site to be investigated	Groundwater samples will be taken twice: once during the wet season and once during the dry season. Dissolved and unfiltered samples will be collected. CLP methods will be used	Groundwater (pathway)	<ul style="list-style-type: none"> Assess the influence of seasonal changes on contaminant concentrations in groundwater. Data will be used as input for the Baseline Risk Assessment for BCV OU 4 and the final ROD for BCV 	Data will be presented on maps and cross-sections. Mass transport rates will be calculated	90% delivered at EPA Level III and validated at modified Level III as described in Sect. 8.9. 10% delivered at EPA Level IV and validated at Level IV
	Analytical parameters to be determined for each site investigated	Filtered and unfiltered samples will be collected during storm events. CLP methods will be used for analysis. Flow rates will also be measured	Surface water (pathway)	<ul style="list-style-type: none"> Determine if contaminants are being transported by surface water to BCV Use in the ecological risk assessment 	Data will be presented on tables	90% delivered at EPA Level III and validated at modified Level III as described in Sect. 8.9. 10% delivered at EPA Level IV and validated at Level IV
	Inorganics, PCBs, radionuclides	Samples collected from areas of contaminated soil adjacent to source areas	Terrestrial vegetation Terrestrial animals	Determine if contaminants are being transported from contaminated soil to the terrestrial food web. Use data as input to the ecological risk assessment	Data will be presented in tables	CLP Level V (SAS ^b)
	Inorganics, PCBs, radionuclides	Samples collected downslope from seeps	Terrestrial vegetation Terrestrial animals	Determine if contaminants are being transported from source areas to the food web via groundwater. Use data as input to the ecological risk assessment	Data will be presented in tables	CLP Level V (SAS)
		Sampling with scoop. EPA-approved toxicity tests	Surface water Soil	Determine if contaminants in physical media are toxic to biota. Use data as input to the ecological risk assessment	Data will be presented in tables	CLP Level V (SAS)

6.1 PHYSICAL CHARACTERIZATION ACTIVITIES

6.1.1 General Site Physical Characterization Activities

6.1.1.1 Soils and solid wastes

Data are needed to characterize the nature and extent of contamination in soils and wastes and to evaluate current and future on-site and off-site contaminant migration pathways. Data indicating the nature and extent of contamination are needed not only to determine the extent of contamination at each site but also to allow evaluation of future fluxes of contaminants along migration pathways.

In order to determine the magnitude of the contamination and the relative location of the contaminant sources (trenches, pits, ponds) for these various contaminants, a two-phase sampling design is being proposed based on the DQOs discussed in Sect. 1. During the first phase of the RI, the nature and magnitude of contaminant concentrations will be determined in OU 1. During the second phase, the distribution and extent of the major COCs (as identified during Phase 1) will be determined. The two phases of field work will be carried out in one operation with concurrent evaluation of data to optimize phase 2 design.

Phase 1 Sampling Design. A stratified and systematic random sampling design has been used during Phase 1 of the RI. As will be discussed in Chap. 7, BCV OU 1 is divided into a number of different strata (waste areas): the S-3 Site, the Oil Landfarm, the Boneyard/Burnyard, The Sanitary Landfill, and the various waste areas (trench groupings) within the Bear Creek Burial Grounds. Within each of the BCV OU 1 strata, a number of potential sampling points have been selected that are located immediately downgradient and along strike from each of the capped areas or located directly on capped areas discussed in Chap. 7. A systematic random sample will then be selected for each area with the number of samples selected being proportional to the relative size of each area.

After the samples for the Phase 1 RI have been collected and analyzed, a five-part decision rule will be employed, using the EPA's *Risk Assessment Guidance for Superfund, Vol: Human Health Evaluation Manual, Part A* (EPA 1989d) to determine whether the concentration in the samples collected in BCV OU 1 exceeds established detection limits, background concentrations (if known), established action levels, or preliminary remediation goals. Specifically, COPCs will be considered for further evaluation if any one of the following applies:

- The analyte was positively detected in at least one sample.
- The analyte concentration was measured at levels significantly higher than naturally occurring levels (background) of the same analyte.
- The analyte concentration exceeded established action levels.
- The analyte was identified but the exact concentration was unknown (i.e., J-qualified data) or was detected at levels significantly higher than the same chemicals detected in associated blank samples.
- The analyte concentration was higher than established preliminary remediation goals.

A systematic process, following the decision rule just discussed, will be used to narrow the list of COPCs. COPCs will be associated, where possible, with specific media (soil, groundwater, surface water) and specific sites in OU 1. This will allow considerable streamlining of sampling and analytical techniques.

Phase 2 Sampling Design. The specific sampling design to be used in Phase 2 of the RI will largely depend on the results obtained in Phase 1. As with Phase 1, a statistically based sampling design will be used to characterize the distribution and extent of the specific COCs determined in Phase 1. Design of phase 2 sampling will be based on guidance in the *Soil Sampling Quality Assurance Users Guide* (Barth et al. 1989).

Since the list of contaminants to be investigated in Phase 2 should be considerably smaller than the list of contaminants analyzed in Phase 1, many more samples can be collected to characterize the extent of contamination. A series of rapid field screening techniques that involve rapid sample collection and analysis techniques (including hydropunch sampling, field GCs, field XRF, and NAA) will be used to quickly identify the extent of contamination. The results from these sample collections and surveys will be verified by subsequent analytical laboratory analysis for statistically selected subsets of samples.

Phase 2 sampling will be designed and carried out as soon as adequate Phase 1 information is available. It is proposed that the field operation will be conducted as a single effort with concurrent preliminary evaluation of data to allow phase 2 sampling to get under way.

Because of the inherent difficulty, potentially high risks, and high costs of evaluating the nature and extent of waste materials, an inventory of waste in the Burial Grounds and historical data concerning wastes in the Oil Landfarm and the S-3 Site will be used.

As was detailed in Chap. 2, ~50,000 individual disposals to the Burial Grounds have been tabulated from the Y-12 Plant records. The disposal records cover three main types of disposal practices: (1) blanket disposals that account for about two-thirds of the total waste in the Burial Grounds, (2) nonroutine hazardous disposals that account for a small proportion of the total wastes, and (3) other disposals that account for one-third of the total waste. Volumes or masses of wastes have been determined using, in order of completeness and usability of the data, the following sources:

- **Y-12 Maintenance Division Monthly Activity Reports.** These reports are available from January 1957 to the present (with roughly 1 month missing from the files per year).
- **Requests for Hazardous/Classified Disposals (UCN-2109s and UCN-6538s).** Approximately 11,300 UCN-2109s and 1,600 UCN-6538s have been entered into a computer data base.
- **Kardex Blanket File.** Approximately 600 cards with 27,000 entries from the 1964–1980 time period covering waste disposals under 170 blankets have been analyzed.
- **Tanker Logs.** These records cover ~1,800 liquid disposals on ~25 different blankets during the 1971–1980 time frame. These logs have been transcribed into a computer data base.

- **Vehicle Trip Logs.** These logs cover blanket, hazardous/classified, and other deliveries from the 1960 to 1980 time frame. Most of the information in vehicle trip logs is already recorded by the previous files and records, and the logs themselves are not standardized so they are difficult to evaluate. For these reasons, roughly 8% of vehicle trip logs (covering 290 days) were sampled and examined to statistically estimate the total deliveries for this time period. This analysis has been verified by comparison to the ongoing SWIMS program.
- **Waste Management Inventory—Manifest System.** Since 1980, all disposal information has been placed on a computerized data base. Over 13,000 disposals to the Bear Creek Burial Grounds have been recorded by this system. Since 1980, disposal records are complete and the quantities and compositions of wastes disposed of in the Burial Grounds since that time can be determined precisely.

The quantity and composition of ~80% of the blanket disposals mentioned in the Monthly Activity Reports have been tabulated. The blanket disposals account for approximately two-thirds of the total waste in the Burial Grounds; thus 53% of the waste materials in the Burial Grounds has been tabulated using these records. It has been assumed that the waste documented by the 20% of blanket records that have not been tabulated are similar to those recorded in the 80% of records that have been tabulated. Therefore, using these blanket disposal records, wastes in the Burial Grounds can be characterized.

The total number of UCN-2109s for hazardous/classified waste disposals mentioned in the Monthly Activity Reports corresponds closely to the number of records tabulated, so the records of this type are essentially complete. The hazardous/classified deliveries account for a small fraction of total deliveries to the Burial Grounds, but the composition of many of these deliveries is of greater potential interest.

For the estimated one-third of all Bear Creek Burial Grounds deliveries that were neither hazardous nor blanket the Vehicle Trip Logs have been used to determine the quantity and makeup of the wastes. For these materials a random sample of working days with a statistical analysis of the sample was carried out and is detailed in the waste inventory report (Energy Systems 1983). Through the use of this random sample, the volume and composition of this remaining one-third of the wastes have been determined with ranges at specified levels of uncertainty.

In summary, the existing analysis of the waste delivery records and document activities at the Burial Grounds provides an inventory of the waste materials with a measurable degree of uncertainty. This provides a better characterization of waste materials than can be obtained by sampling all trenches in the Burial Grounds and with less risk posed to RI personnel.

Soil samples are needed to determine the nature and extent of secondary contamination that has migrated from the original location of the waste materials and has been captured by surrounding soils. Contaminants may have migrated either as nonaqueous-phase liquids and been trapped in pores or sorbed to soil particles, or as dissolved phase and been precipitated or sorbed at the surfaces of soil materials.

On the whole, past sampling of soils in BCV has not followed and/or documented QA/QC procedures that meet current requirements. The historical soil database for each site is also relatively limited in extent; in many cases, the nature of the soils has changed since the

last sampling event (e.g., a site has been capped or soils have been excavated). Where historical data are inadequate, soil sampling will be conducted to verify the level and extent of soil contamination.

The Summers Model for soil leaching (Summers et al. 1980; EPA 1989c) will be used to evaluate PRGs (see Chapt. 5 of this work plan). More sophisticated models such as HELP, MULTIMED, or SESOIL will be used to address the leaching of contaminants from the soil to the groundwater. Using Monte Carlo Simulations will allow statistical ranges for parameter values to be input and will provide output in the form of a likely range of values. Thus, allowances will be made for natural soil variability and waste heterogeneity.

Physical properties of selected soils will be determined to provide the information required for selecting engineering alternatives in the FS (Table 6.2).

6.1.1.2 Groundwater and surface water

A substantial amount of groundwater monitoring data concerning the BCHR are available (HSW 1991, 1992); the Y-12 Plant groundwater data base contains information on 827 wells. Changes in groundwater quality that are the result of changes to a single source OU are more likely to be identified in shallow groundwater close to the source than in the deep aquifer where contributions from various sources conflow and become indeterminable. The groundwater data base was screened using the following criteria to obtain a list of wells the historical data from which can be used to analyze shallow groundwater quality at the sites in BCV OU 1 and additionally support the BCV OU 4 RI.

- **Well location**—Wells within the boundaries and in the immediate vicinity of each site are used to monitor shallow groundwater.
- **Aquifer and depth**—Only wells that are screened in the unconsolidated deposits or that are less than 9.1 m (30 ft) deep are used to monitor shallow groundwater.

Wells listed in Table 6.3 that are identified with a “W” in the OU 1 column meet the criteria for location, depth, and aquifer. Data from these wells will be used in the RI to analyze historical groundwater quality trends. An “S” in the OU 1 column indicates that historical data concerning soil analysis will be used from this boring, and reference is made to where these data are available.

The sample plan will not call for additional sampling of these wells because some will be sampled as part of the continued groundwater monitoring program at the Y-12 Plant, some will be sampled for the BCV OU 4 RI, and for others analytical data show little or no change over the existing monitoring period. Wells that will be sampled by this RI are listed in Chap. 7.

Data from deeper wells in the BCHR have been referred to in the site characterization sections of this RI work plan.

Validation of data concerning groundwater chemical analysis must be accomplished for the data set that is used in the RI. The repeatability of analyte concentrations displayed in the historical data partially validates these data, and 100% of these data are assumed to be validated to the equivalent of EPA Level III. Continued monitoring of all the wells listed in Table 6.3 is not warranted because sufficient historical data are available for each well to

Table 6.2. Geotechnical and physical parameters for each soil sample

Analytical parameter	Analytical method
Permeability	SW 846 Method 9100
Hydraulic conductivity	SW 846 Method 9100
Porosity	EM 1110-2-1906
Grain size (sieve)	ASTM D422
Grain size (hydrometer)	ASTM D422
Mineral composition	X-Ray Diffraction/electron spectroscopy
Bulk density	ASTM D4531-86
Cation exchange capacity	SW 846 Method 9081
pH	SW 846 Method 9045
Total organic carbon (bulk)	SW846 Method 9060 ^a
Total organic carbon (<75 μm)	SW 846 Method 9060 ^a
Distribution coefficient (Kd)	ASTM Method 4646
Moisture content	ASTM D4959-89
Aterberg limits	
Shear strength	

^aAs modified for solid media per laboratory-specific methodology. Laboratory method will be submitted for review and approval.

Table 6.3. BCV OU 1 groundwater monitoring wells

Well	Depth (ft)	Formation ^a	Aquifer ^b	Core (yes or no)	Screened (yes or no)	Status ^c	OU 1 ^d	First year data	Last year data
<i>S-3 Site</i>									
GW-100	20.7	€mn	U/wB	N	Y	F	S ^e	NA	NA
GW-101	17.5	€n	U/wB	N	Y	F	W,S ^e	1986	1991
GW-102	24.0	€n	U	N		P	S ^e	NA	NA
GW-103	25.0	€n	U/wB	N		P	S ^e	NA	NA
GW-105	17.0	€n	U/wB	N	Y	F	W	1986	1990
GW-107	14.2	€n	U	N	Y	F	W	1986	1990
GW-127	24.0	€mn	U/wB	N	Y	F	W	1986	1990
GW-269	30.0	€m	U	N	Y	F	W	1986	1990
GW-276	18.5	€n	wB	N	Y	F	W	1986	1990
GW-325	17.9	€n	wB	N	Y	F	W	1987	1991
GW-617	18.0	€mn	U	N	Y	F	W	1990	1992
<i>Bear Creek Burial Grounds</i>									
GW-014	13.2	€n	U/wB	N	Y	F	W,S ^f	1987	1990
GW-015	7.9	€n	U/wB	N	Y	F	S ^f	NA	NA
GW-016	17.1	€n	U/wB	N	Y	F	S ^f	NA	NA
GW-018	20.8	€n	wB	N	Y	F	W,S ^f	1991	1991
GW-020	66.1	€n	wB	Y		P	S ^f	NA	NA
GW-021	15.0	€n	wB	N		P	S ^f	1987 _g	1987 _g
GW-024	75.7	€m	wB	Y		P	S ^e	NA	NA
GW-026	20.2	€m	wB	N		P	S ^e	NA	NA
GW-027	30.5	€m	wB	N		P	S ^e	1987 _h	1988 _h
GW-034	45.5	€m	wB	N	Y	F	S ^e	NA	NA
GW-035	62.0	€m	wB	Y		D	S ^e	NA	NA
GW-036	39.1	€rg	wB	N	Y	F	S ^e	NA	NA
GW-037	70.0	€m	wB	Y	Y	F	S ^e	?	?
GW-042	30.0	€pv	wB	N	Y	F	W	1987	1992
GW-045	15.2	€mn	U/wB	N	Y	F	W	1987	1990
GW-046	20.5	€n	U/wB	N	Y	F	W,S ^f	1987	1990

Table 6.3 (continued)

Well	Depth (ft)	Forma- tion ^a	Aquifer ^b	Core (yes or no)	Screened (yes or no)	Status ^c	OU 1 ^d	First year data	Last year data
GW-047	25.5	En	wB	N	Y	F	W,S ^f	1987	1990
GW-052	19.5	Emn	U	N	Y	F	W	1990	1992
GW-053	39.7	Emn	U/wB	Y	Y	F	W,S ^f	1990	1992
GW-057	25.0	Emn	wB	Y	Y	F	W	1991	1992
GW-061	25.0	Emn	wB	Y	Y	F	W	1989	1992
GW-078	21.1	En	wB	N	Y	F	W	1990	1991
GW-080	30.0	Eng	wB	N	Y	F	W	1990	1992
GW-083	30.0	En	wB	Y	Y	F	W	1990	1992
GW-089	25.0	Er	U	N	Y	F	W	1990	1992
GW-237	13.7	Emn	U/wB	N	Y	F	W	1988	1988
GW-242	17.0	E _p v	wB	N	Y	F	W	1987	1990
GW-287	12.5	En	wB/B	N	Y	F	W	1987	1992
GW-291	16.98	Em	wB/B	N	Y	F	W	1988	1991
GW-622	20.5	En	wB	N	Y	F	W	1990	1992
GW-624	27.2	En	wB	N	Y	F	W	1990	1991
GW-641	24.3	En	wB	N	Y	F	W	1990	1992
GW-643	29.3	Em	wB	N	Y	F	W	1991	1992
GW-654	15.5	Em	wB	N	Y	F	W	1991	1992
<i>Oil Landfarm Area</i>									
GW-001	45.0	Em	U	Y	Y	F	S ⁱ	NA	NA
GW-002	60.0	Em	wB	Y	Y	F	S ⁱ	NA	NA
GW-003	35.2	En	wB	Y		P	S ⁱ	NA	NA
GW-004	58.4	En		Y		P	S ⁱ	NA	NA
GW-005	12.5	En	U	N	Y	F	W,S ^{ij}	1990	1991
GW-007	16.5	En	wB	N	Y	F	W,S ⁱ	1987	1990
GW-008	25.5	En	wB	Y	Y	F	S ⁱ	NA	NA
GW-009	63.8	En		Y		P	S ⁱ	NA	NA
GW-010	15.0	En	U/wB	N	Y	F	W,S ⁱ	1987	1990
GW-011	60.8	En	wB	Y	Y	F	S ⁱ	NA	NA
GW-013	33.8	En	U/wB	Y	Y	F	S ⁱ	NA	NA

Table 6.3 (continued)

Well	Depth (ft)	Formation ^e	Aquifer ^b	Core (yes or no)	Screened (yes or no)	Status ^c	OU 1 ^d	First year data	Last year data
GW-044	70.0	€m	wB	Y	Y	F	S ⁱ	NA	NA
GW-067	16.5	€mn	U	N	Y	F	W	1987	1990
GW-084	34.0	€m	wB	N	Y	F	W	1987	1992
GW-086	33.5	€n	wB	N	Y	F	W	1990	1991
GW-087	19.0	€n	U	N	Y	F	W	1987	1990
GW-097	19.2	€n	wB	N	Y	F	W	1987	1990
GW-637	27.5	€n	wB	N	Y	F	W	1991	1992
GW-638	12.0	€n	U/wB	N	Y	F	W	1991	1992
GW-644	23.4	€n	wB/B	N	Y	F	W	1991	1992
GW-649	20.8	€n	wB	N	Y	F	W	1991	1992

^e€mn = Maynardville; €n = Nolichucky; €m = Maryville; €rg = Rogersville; €pv = Pumpkin Valley; €r = Rutledge.

^bU = unconsolidated overburden; wB = weathered bedrock; B = unweathered bedrock.

^cF = finished; P = plugged and abandoned; D = destroyed.

^dS = soil core data; W = groundwater monitoring well data.

^eGeraghty and Miller 1985.

^fHerbes 1988.

^gWell sampled in 1987 and then plugged.

^hWell sampled in 1987 and 1988 and then plugged.

ⁱMcCauley 1984b.

^jHerbes 1989.

identify trends in groundwater quality for each of those wells. Sampling of selected wells will be carried out, with 10% of data reported with EPA Level IV data packages and 90% with modified EPA Level III (see Sect. 8.9).

Little information is available regarding the dynamics of groundwater recharge and the storm flow system in BCV. Studies performed on Solid Waste Storage Area (SWSA) 6 by ORNL (Solomon et al. 1991) indicate that a significant flux of contaminants can be transported by storm flow and that the presence of protective caps over waste trenches does not necessarily preclude recharge to groundwater through waste pits. The phenomenon of "bathtubbing," where water collects in a waste pit and slowly recharges to groundwater, has been observed in uncapped and capped pits at SWSA 6 (Ashwood and Spalding 1991). Lack of understanding of the storm flow system and aquifer recharge is a data gap at each waste site in this OU. Information will be gathered to identify recharge areas and flow paths of storm waters. In general, this will require storm flow monitoring devices located upgradient and downgradient from sites that are likely to be affected by stormflow (Moore and Toran 1992). Monitoring of the storm flow system at sites in BCV that have not been affected by waste sites was proposed in the draft RI work plan for the integrator OU (BCV OU 4). These data will be available to the BCV OU 1 RI and will be evaluated as being representative of undisturbed storm water flow patterns and background storm water flow quality in BCV.

Surface water is an exit pathway for contaminants from BCV OU 1 and may be used to monitor contaminant migration from the sites. Stream sediments are also an exit pathway for soils eroded from sites in BCV OU 1. Surface water and stream sediments will be sampled to evaluate this pathway as part of the RI for BCV OU 4.

6.1.1.3 Wetlands

Wetlands are defined by three essential attributes: wetland hydrology, hydrophytic vegetation, and hydric soils. They generally include swamps, marshes, bogs, and similar areas. Wetlands at BCV OU 1 must be identified and evaluated because wetland function is legally protected. Wetlands may occur at BCV OU 1, especially near seeps or along Bear Creek or its tributaries, although the presence of wetlands at the S-3 Site is unlikely.

Wetlands at each source area will be identified from historical records, aerial photographs, and walk-over surveys. Soil and sediment samples within wetlands will be used to identify areas of contaminant deposition.

6.1.1.4 Biological media

The baseline risk assessment includes consideration of both risk to humans and risk to natural resources (ecological risk). The process to be followed for ecological risk assessment (ERA) is outlined in *Risk Assessment Guidance for Superfund, Volume II: Environmental Evaluation Manual* (EPA 1989a), *Ecological Assessments of Hazardous Waste Sites: A Field and Laboratory Reference Document* (EPA 1989b), and *Framework for Ecological Risk Assessment* (EPA 1992c). The ERA process differs from the human health risk assessment in that the ERA focuses on populations and communities rather than on individuals. In the ERA process, risks to individuals are addressed only if the individuals are protected under the Endangered Species Act, are unique, or have sufficient other ecological value that individual consideration is merited. Therefore, the ERA will determine if and where there is imminent

and substantial danger to the health of various animal and plant populations and habitats and if and where site remediation may be needed for long-term protection of the environment.

ERA for the source units will be based on entry of contaminants into the food web. Therefore, body burden analyses will be limited to organisms that have limited mobility, are exposed to contaminated physical media, and are important food items (i.e., grasses and browse, earthworms, terrestrial small mammals and burrowing mammals, and amphibian larvae).

Although limited data are available to model biological effects of soil and surface water contaminants on biological receptors, modeled exposures may be much more conservative than field data support, therefore requiring more extensive remediation than actually needed. For this reason, it is prudent to gather direct data on biological exposures to ensure that realistic risk estimates are made. Biota sampling will be designed to reveal movement of contaminants from BCV OU 1 sources into biotic receptors, including the food web as well as individual taxa; to evaluate ecological risks in source areas; and to provide the necessary information for ecological risk assessment of the integrator OU. Toxicity tests will reveal potential toxicological impacts to organisms exposed to physical media.

Specific objectives of the sampling and analysis of biota and associated physical media are as follows:

- To determine relative population levels of selected indicator species at various locations at BCV OU 1 and at nonimpacted reference sites.
- To determine whether species designated as threatened and endangered or in need of management are present in BCV OU 1 or whether habitat essential for their survival and reproduction is present.
- To determine concentrations of COPCs in aquatic biotic systems.
- To determine concentrations of COPCs in terrestrial ecosystems.
- To determine concentrations of COPCs in BCV OU 1 surface water at the time of aquatic biota sampling.
- To determine concentrations of COPCs in BCV OU 1 wetland soils and pond sediments at the time of aquatic biota sampling.
- To determine concentrations of COPCs in natural vegetation used for food.
- To determine concentrations of COPCs in the soils in which these plants were grown.
- To select exposure parameters and relevant scenarios for realistic exposures to higher trophic levels by ingestion of biota.

Analytical data will be used to characterize exposures of aquatic biota and their predators and of terrestrial plants and small mammals and their predators to BCV OU 1 site contaminants. This information will be used along with population survey results to determine whether remedial actions are necessary for the protection of ecological populations at BCV OU 1. Chemical analyses of biota and the physical media by which they are exposed to site contaminants provide a link to demonstrate the extent to which contaminants move from the physical media to the exposed organisms. This link facilitates the identification of areas whose contaminants pose a threat to the biota present and thus helps identify areas in need of

remediation. Consideration of uptake and biotransfer factors can help to set remedial goals for cleanup activities.

The primary sources of contaminants to aquatic and terrestrial biota in BCV OU 1 are releases of chemical and radiological contaminants by leaching from landfills. Surface runoff may serve as a primary release mechanism, transporting contaminants to soil. In addition, contaminants transported by surface runoff to tributaries within BCV OU 1 may expose aquatic biota, terrestrial animals, and waterfowl by ingestion or dermal contact. (Exposures in the tributaries to Bear Creek will be evaluated as part of the RI for BCV OU 4.) Contaminated soil is a secondary source of contaminants released to other environmental receptors and an exposure source to the biotic receptors. Vegetation may be exposed by root uptake or direct contact. Terrestrial animals and birds may be exposed to contaminants in soil directly by ingestion and dermal contact or indirectly by ingestion of vegetation contaminated by uptake of contaminants from soil. Air and groundwater are assumed not to provide significant direct exposures for biota.

Ecological communities will be characterized by historical and field data, using surveys designed to determine the number and distribution of indicator species chosen to reveal significant exposures of biota to site contaminants. Data from each sample site will be compared with data from the appropriate reference site and from each of the other BCV OU 1 sites to determine whether there are differences from an unaffected site or differences among sites related to their location on BCV OU 1. Historical survey information will also be used to determine whether there are threatened and endangered species or critical habitats in the BCV OU 1 area.

Earthworms will be sampled as indicators of biotransfers from soil into the food web. Terrestrial small mammals will be trapped, identified, and counted; and selected individuals will be analyzed for whole-body burden of COCs. Groundhogs will be sampled as integrators of contaminants potentially removed from subsurface soils by burrowing. Aquatic insects and amphibia will be sampled as indicators of contaminant uptake from surface water streams and seeps. Grasses and browse will be analyzed as sources of contaminants to herbivores by ingestion. Communities of terrestrial vegetation will be characterized by walk-over surveys. The sampling locations will be chosen to maximize the comparability of physical media sampling data with biological sample data. All populations and habitat characterizations will be compared with similar characterizations performed on reference areas of BCV not affected by BCV OU 1 site contaminants.

Wetland vegetation will be sampled as an indicator of biotransfers of wetland contaminants into the food web. In wetlands whose surface waters are sufficiently ponded to support aquatic biota, sediment samples will be taken for sediment toxicity tests, and amphibian larvae will be sampled for contaminant body burdens. So that the exposure assessment can be limited to wetland surface water and sediment as contaminant sources, larvae must be sampled rather than adults, which are exposed via terrestrial food sources. Therefore, sampling of amphibians must be done during the short time period when tadpoles are present in surface water. Filtered surface water will be sampled to determine the direct exposure concentrations, as particulate contaminants are typically not available to aquatic biota. These samples may be scheduled to coincide with sampling for the integrator OU.

Based on historical data, contaminants detected above contract reporting limits in surface water, soils, and sediments at BCV OU 1 include metals, PCBs, polycyclic aromatic

hydrocarbons (PAHs), dioxins, furans, and radionuclides. PCBs, PAHs, dioxins, furans, and some inorganics will be of concern because of their bioaccumulative properties. PCBs are toxic and persistent in the environment and will be analyzed in biota samples. Standard EPA SW-846 methods, with revisions as necessary, will be used to analyze for PCBs, dioxins, furans, and PAHs. Volatile organics will not be analyzed in biological tissues. Neutron activation analyses for metals will be performed on all biota samples.

Toxicity testing of surface water and soil will reveal cumulative toxic effects of contaminants in physical media to indicator animals and plants and, by extension, potential toxic effects to biota exposed on site to the contaminated soils. Toxicity testing will be done by standardized EPA-approved procedures.

The approach to sampling of biota will maximize the amount of useful information gained within the sampling budget. Small mammals are exposed to contaminants in surficial soils both directly and via the food web. The prey of small mammals (earthworms and other invertebrates) may bioconcentrate some contaminants (for example, PCBs) that may not be detected in high concentrations in surface soil. Because contaminant concentrations in the soil may not be indicative of the exposure experienced by small mammals, earthworms, grasses and browse will be sample. Burrowing animals may be exposed to contaminants at deeper levels, but they would need to be sampled only if contaminants are found in subsurface soils. Biota sampling will be carried out in a single phase. The optimum season for most biota sampling will be spring and early summer, and, based on the project schedule in Sect. 1, this will be carried out during the first six months of 1995. Therefore, ecological sampling will be carried out when the results of the Phase 1 soil and wetland sediment sampling are reviewed and the sample locations can be optimized using Phase 1 data.

The process for determining what ecological sampling will be done is outlined in Fig. 6.2. Background contaminant concentrations will be determined at uncontaminated reference sites that possess environmental characteristics (e.g., soils, vegetation, etc.) similar to those present at BCV OU 1. Areas where soil contaminants exceed twice background concentrations will be identified, and a vegetation sample and an earthworm sample will be taken within each of these hot spots. If contaminant concentrations in the biota indicate bioconcentration relative to the soil, additional samples will be taken according to a statistical design that will reveal the area of contamination. If vegetation appears unhealthy in the hot-spot area, soil samples will be tested for phytotoxicity by a seed germination assay. If the population density of earthworms is below the 5% lower bound of the density at the reference site, the population will be tested to indicate the boundaries of the toxic zone.

When amphibian larvae are found in ponds, seeps, springs, and wetlands, they will be observed visually to evaluate normal development and will be sampled as they reach maturity. One surface water and one sediment sample will be taken from each area. If contaminant concentrations are elevated, additional samples will be taken to delineate the area of contamination. If amphibian larvae are not found in contaminated ponds or wetlands when they are present in reference areas, toxicity of the surface water will be evaluated by a *Ceriodaphnia* toxicity test.

Fig. 6.2. Decision process for Phase 1 biota sampling and analysis.

6.1.1.5 Ecological endpoints

An ecological endpoint is a characteristic of an ecological component (e.g., increased mortality in fish) that may be affected by exposure to a stressor (Suter 1990). Two types of endpoints are discussed in Suter's report: assessment and measurement endpoints. Assessment endpoints are explicit expressions of actual environmental value that is to be protected. Measurement endpoints are measurable responses to a stressor that are related to the valued characteristics chosen as the assessment endpoints (Suter 1990). When an assessment endpoint can be measured directly, the measurement and assessment endpoint are the same. In most cases, however, the assessment endpoint cannot be measured directly, so a measured endpoint is selected that can be related, either qualitatively or quantitatively to the assessment endpoint. Assessment policy goals and measurement endpoints are described in Table 6.4 and were developed within the context of the guidance in EPA (1992c) and Suter (1993).

6.1.2 Site-Specific Physical Characterization Activities

Site-specific data needs are summarized in Table 6.5. Figure 6.1 locates each site in the staged RI process that is described in Chap. 1.

6.1.2.1 S-3 Site

Historical data are sufficient to show that the S-3 Site constitutes no immediate hazard to human health or the environment; however, hazardous wastes are buried at this site and may pose a long-term risk to human health or to the environment. Therefore, using the staged RI process outlined in Chap. 1, Stage 1 was completed, and the decision was made to continue the RI. This section outlines the data needed to complete Stage 2 of the staged RI/FS process for the S-3 Site.

S-3 Site soils and solid wastes. Prior to installation of the RCRA cap, sludges in the ponds were analyzed for a range of contaminants and then stabilized using a coarse aggregate that was probably limestone. Outside the physical boundaries of the S-3 Ponds, little information has been gathered concerning the nature and extent of contamination in the soils that may have occurred as the result of reprecipitation or sorption of contaminants on particle surfaces in soils and that may contaminate groundwater by matrix diffusion. Data exist from four soil cores (GW-100, GW-101, GW-103, and GW-102) that indicate the presence of PCBs surrounding the S-3 Site both in shallow soils [<0.6 m (2 ft)] and at depth [3.7 m (12 ft)]. At GW-102, chlordane was found between depths of 4–12 ft. Traces of VOCs were also found in the soil samples from GW-102. Radioactivity (chiefly as uranium) and concentrations of metals above background levels (especially mercury) were detected in stream sediments and chemical precipitates in the main branch of Bear Creek adjacent to the S-3 Site and in NT-1 (Turner and Kamp 1984). This indicates that, during operation of the S-3 Ponds, radionuclides and metals migrated via groundwater to surface water and precipitated in the stream bed.

To address the nature and extent of contamination in waste materials, documentation to support validation of the existing data set concerning waste composition is needed. The nature and extent of contamination that has migrated to soils surrounding the S-3 Site by way of groundwater need to be determined.

Table 6.4. Assessment policy goals and measurement endpoints

Policy goal	Assessment endpoint	Measurement endpoint
The conservation of threatened, endangered and rare species and their critical habitats	No adverse impact on survival or reproduction of any state or federally designated threatened or endangered species and their critical habitat on the BCV	<ul style="list-style-type: none"> a. Surveys of the presence or absence of species and their habitats in the BCV b. Maintenance of plant community composition and/or structure required for rare plant and other support species c. The abundance and distribution of plant and animal species that would support threatened, rare, or endangered species
The protection of migratory birds	No killing of or adverse impact to migratory birds as a result of exposure to site-specific stressors	Contaminant concentrations in selected BCV food sources for migratory birds (to be evaluated as part of the BCV OU 4 RI)
The preservation of wetlands	The presence and structure/function of wetlands in relation to contaminants	Survey of wetlands and associated contamination levels, if any, on the Bear Creek floodplain
No adverse effects from contaminants to aquatic indicator organisms and/or predators that feed on them	<ul style="list-style-type: none"> a. The ratio of concentrations of contaminants in ponds and wetland surface water to water quality criteria for protection of aquatic life should be ≤ 1 b. The ratio of body burdens of contaminants in indicator organisms to levels expected to cause toxicological effects should be ≤ 1 c. The ratio of body burdens of contaminants in aquatic biota to levels expected to be protective of predators should be ≤ 1 d. Surface water in ponds and wetlands should not be toxic to aquatic biota 	<ul style="list-style-type: none"> a. Contaminants in water b. COC concentrations in whole body samples of indicator species c. COC concentrations in whole body samples of aquatic indicator species d. <i>Ceriodaphnia</i> toxicity tests
The existence of a terrestrial animal community indicative of undegraded conditions	Diversity, abundance, and distribution of terrestrial animals should not result in a $\geq 20\%$ decrease compared with background conditions	Population abundance and distribution by habitat

Table 6.4 (continued)

Policy goal	Assessment endpoint	Measurement endpoint
No adverse effect from contaminants to terrestrial indicator organisms	<ul style="list-style-type: none">a. The ratio of body burdens of contaminants in indicator organisms to levels expected to cause toxicological effects should be ≤ 1b. Contaminated soils should not be toxic to plants or soil invertebrates	<ul style="list-style-type: none">a. COC concentrations in whole body samples of terrestrial indicator speciesb. Seed germination and earthworm toxicity tests on contaminated soils

Sources: EPA 1992c; Suter 1993; Suter 1990.

Table 6.5. Rationale for sampling OU 1 sites

Site name	Sampling type	Rationale
S-3 Ponds	Radiological survey Soil borings	Confirm nature and extent of waste materials currently in place. Determine nature and extent of contaminants that have migrated away from the ponds and now reside in the surrounding soils. Determine hydrogeologic and chemical properties of contaminated soils for transport models. Define the stormflow system. Gather soil physical properties data for FS.
Oil Landfarm	Soil borings	Define nature and extent of soil contamination downgradient of and along strike from the Oil Landfarm cap. Gather soil physical properties data for FS.
Boneyard/Burnyard and Hazardous Chemical Disposal Area	Walk-over, radiological, and geophysical survey Surface soils Soil borings Groundwater	Perform geophysical and radiological screening of the whole site to identify hot spots. Sample soil to define the nature and extent of soil contamination in hot spot areas and to determine if soil contamination exists outside of hot spot areas. Determine quality of shallow groundwater in contact with, immediately downgradient of, and along strike from hot spot areas. Define the surface water and soil erosion pathways. Gather soil physical properties data for FS.
SL 1	Radiological survey Soil borings Groundwater	Determine if groundwater immediately downgradient of SL 1 is impacted by this site
Burial Grounds	Radiological survey Soil borings Surface soils Groundwater Surface water	Define the nature and extent of waste materials in various waste trenches. Define nature and extent of soil contamination in the vicinity of waste trenches. Identify radiological hot spots outside the waste trenches. Determine surface water quality during the stormflow stage. Determine water quality in shallow groundwater immediately downgradient of and along strike from trenches. Define the stormflow system and the likelihood of storm water impacting contaminated soil beneath the cap. Gather soil physical properties data for FS

S-3 Site groundwater and surface water. Eight groundwater monitoring wells in the vicinity of the S-3 site meet the criteria for screen depth and location to be used by the RI (Table 6.3). For shallow groundwater in the unconsolidated deposits, no well exists upgradient of the S-3 Site; however, the S-3 Site is close to the groundwater divide between BCV and UEFPC hydrogeologic regimes, and it is unlikely that any sources for contamination exist upgradient. The present groundwater monitoring wells are adequate to monitor groundwater off site.

The effect of the S-3 Site cap on the hydrology of the stormflow system is not known. However, because the S-3 Site is almost completely surrounded by asphalt roads and other construction and is located on a topographic high, it is likely that the site is isolated from the stormflow recharge area; therefore, it is unlikely that the wastes below the cap at the S-3 Site will be affected by stormflow.

Historical data indicate that past contamination of surface water draining the S-3 Site has come mainly from contaminants migrating through the groundwater to seeps and springs in Bear Creek. The presence of the asphalt cap over the ponds precludes any direct contamination of surface water by contaminated soils and waste at the site. The contaminant pathways to surface water by way of groundwater will be addressed by the current groundwater monitoring network and surface water sampling proposed for BCV OU 4. Surface water and groundwater sampling is, therefore, not warranted at this site.

Biota sampling and toxicity testing. Because there is very little natural habitat at the S-3 Site, biota sampling and toxicity testing will not be done at this site.

6.1.2.2 Oil Landfarm

Historical data are sufficient to show that the Oil Landfarm constitutes no immediate hazard to human health or the environment; however, soil contamination exists at this site and may pose a long-term risk to human health or the environment. Therefore, using the staged RI process outlined in Chap. 1, Stage 1 was completed and the decision was made to continue the RI (Fig. 6.1). This section outlines the data needed to complete Stage 2 of the staged RI/FS process for the Oil Landfarm.

Oil Landfarm soils and solid waste. Before the Oil Landfarm was capped in 1990, extensive soil sampling was carried out in the various plots within the Landfarm. More than 200 samples were taken in a two-phase study and analyzed for PCBs, uranium, oil, and grease (Herbes 1988). Soils exceeding 25 mg/kg PCBs were excavated and removed from the site. The remaining soils were left in place.

Nineteen stream sediment samples and nine soil borings were taken in the vicinity of the Oil Landfarm, the Boneyard/Burnyard, and SL 1 during four studies between 1983 and 1991 (McCauley 1984a; Herbes 1989; Geraghty and Miller 1985; Turner et al. 1991). PCB contamination has been recorded in soils outside the Oil Landfarm cap at a depth of 11 m (6 ft) (Chap. 8); however, no information concerning soil contamination is available for surface soils in this area. Tributary sediments show little or no contamination; this indicates that there is little contamination of shallow soils outside the Oil Landfarm cap.

To address the nature and extent of contamination in soils beneath the cap, documentation is needed to support QA procedures.

A soil leaching model will be used to evaluate the potential for off-site migration of contaminants from soil beneath the cap. Use of a Monte Carlo Simulation allows statistical ranges for parameter values to be input, and this model provides output in the form of a likely range of values. Data required for input to this model were listed in Table 6.2. These data can be obtained by testing native soils surrounding the Oil Landfarm cap. Large quantities of oils were disposed of at this site. The extent of soil contamination resulting from migration of LNAPL away from the site on groundwater needs to be determined.

Oil Landfarm groundwater and surface water. Eighteen groundwater monitoring wells in the vicinity of the Oil Landfarm meet the criteria for screen depth and location to be used by the RI (Table 6.1). Upgradient wells that were screened in shallow groundwater in the unconsolidated deposits are wells GW-13 and GW-43. The remainder of the wells in the unconsolidated deposits are located downgradient to the south and west of the Oil Landfarm. The present shallow groundwater monitoring wells are adequate to monitor changes in shallow groundwater downgradient of the Oil Landfarm cap.

The recharge area for the stormflow upgradient of the Oil Landfarm is isolated from this site by the Burial Grounds access road. Therefore, stormflow will not be evaluated at this site.

Habitats. Habitat delineations in the Oil Landfarm vicinity will reveal the locations of protected habitats and habitats where the ecological indicator receptors are likely to be exposed to site contaminants. During Phase 1 sampling, wetlands in the vicinity of the Oil Landfarm will be identified from historical information and walkover surveys. The size and location of each wetland, as well as a description of the associated vegetation, will be recorded. Aquatic and terrestrial habitats will be delineated throughout the Oil Landfarm site. The locations of various habitat types will be entered in the Government Information System mapping system.

Biota sampling and toxicity testing. Organic constituents of waste oil are likely to be toxic to earthworms and other soil organisms. Metal and radionuclide contaminants are also likely to be transferred from soil to the food web. Biota sampling will be focused on areas shown during Phase 1 to contain physical media contaminants. Likely sampling sites are downslope from the cap, where leachate from the waste areas may have contaminated surface soils.

Grasses and browse and earthworms will be sampled in areas of probable surface soil contamination as delineated by the screening radiological survey. If the populations are reduced (as described in Fig. 6.2), then the soil samples will be taken at each location for phytotoxicity and earthworm toxicity tests. Amphibian larvae will be sampled for whole-body contaminant burdens in wetlands that are found to be contaminated and that contain sufficient water to support the development of amphibians. One surface water sample and one sediment sample will be analyzed for COCs at each location.

Terrestrial small mammals will be sampled in appropriate habitat identified by the habitat survey. Statistically designed sampling of grasses and browse and earthworms will be done in areas shown in Phase 1 to contain contaminated biota (Sect. 6.1.1.4). Terrestrial burrowing mammals will be sampled in appropriate habitats identified by the habitat survey. Wetlands shown in Phase 1 containing contaminated amphibian larvae will be sampled further to delineate the extent of contamination in soil, sediment, and water by taking additional soil or paired surface water/sediment samples to be analyzed for COCs. Wetland areas will be divided into subareas, roughly 20 m² (66 ft²), and a sample location will be placed at random within each subarea.

6.1.2.3 Boneyard/Burnyard and HCDA

Historical data are sufficient to indicate that the Boneyard/Burnyard and HCDA do not constitute immediate hazards to off-site receptors. These data are not sufficient to indicate whether these sites constitute immediate hazards to on-site human health or the environment. Uncontrolled soil contamination exists at this site and may pose both immediate and long-term risks to human health or the environment. Therefore, using the staged RI process outlined in Chap. 1 and on pages 6-1 through 6-5, Stage 1 must be completed for the Boneyard/Burnyard before the staged RI/FS process can continue (Fig. 6.1). This section outlines the data needed to complete Stages 1 and 2 of the staged RI/FS process for the Boneyard/Burnyard and HCDA. Data needs for Stage 2 will be refined based on the results of Stage 1. The likely situation that may require a removal action at the completion of Stage 1 is the presence of partially buried uncontrolled hazardous waste that constitutes a hazard to on-site workers. This waste will need to be excavated before the staged RI/FS can continue.

Boneyard/Burnyard soils and solid wastes. A walk-over radiological survey and an electromagnetic conductivity survey (Energy Systems 1984a) have identified a number of areas of likely contamination in and near the Boneyard/Burnyard. Soil samples from soil borings on the northwestern perimeter of the Boneyard/Burnyard show contamination with VOCs, PCBs, and metals. A single stream sediment sample from NT-4, downstream of this point, recorded contamination with VOCs in 1984. The northwestern margin of the Boneyard/Burnyard is the location of a distinct conductivity high, suggesting a possible location for wastes. This area was also the site of high alpha and beta emissions in the walk-over radiological survey. The geophysical survey recorded sporadic conductivity highs throughout the Boneyard/Burnyard; however, the sample spacing was too great to define areas of high conductivity.

The Boneyard/Burnyard is a ~3.2-ha (8-acre) area of land; to address the nature and extent of waste and contamination in soils, a systematic phased approach in which rapid field screening of contaminant distribution followed by stratified random sampling based on the results of the screening will be employed. The data needs for Stage 1 will be satisfied by the field screening, and a decision for removal action will be made on completion of this stage. Data from the screening process will then be used on which to base a stratified sampling plan in Stage 2 of the RI.

The field screening will consist of a systematic observational and radiological walk-over survey followed by a geophysical survey. The walk-over survey will be conducted to identify partially buried waste material or waste material that is close to the surface and that constitutes a hazard. The locations of waste materials and trenches at the Boneyard/Burnyard are not known; geophysical screening methods will be used to determine the locations of the trenches. Electromagnetic conductivity has already proved to be an effective method for locating contamination at this site, and recent improvements to these techniques have resulted in better methods for data acquisition and interpretation (Nyquist and Emery 1993). To locate waste trenches and areas where solid wastes possibly exist, an electromagnetic survey will be carried out using equipment capable of screening the upper 1 to 2 m (3.3 to 6.6 ft) of the soils. Both the walk-over surveys and the geophysical survey will cover adequate ground outside the boundaries of the Boneyard/Burnyard to ensure that background levels are recorded. There are almost no data that relate to the nature and extent of soil contamination. For Phase 1, soil sampling needs to be carried out on a coarse grid over the whole site.

Phase 2 soil sampling will need to be carried out by trenching and/or soil boring at likely targets identified by the geophysical and radiological surveys and the Phase 1 soil sampling as outlined in this work plan. Trenches are preferable to borings because they provide substantially more information where heterogeneous waste are concerned and require little additional effort at this site where groundwater is probably <5 m (16.4 ft) below ground surface. Soil samples will be taken from the trenches and borings and analyzed for all the analytes in the COPC list for this site.

Boneyard/Burnyard groundwater and surface water. Groundwater at the Boneyard/Burnyard is presently monitored in five wells screened in the unconsolidated deposits (GW-5, GW-67, GW-84, GW-86, and GW-87).

Wells GW-84 and GW-86 are screened in the unconsolidated deposits upgradient of the Boneyard/Burnyard. Wells GW-67 and GW-87 are screened in unconsolidated deposits downgradient on the southern margin of the site near Bear Creek. Well GW-5 is located on the western bank of Tributary NT-3 and, therefore, may not capture groundwater from the Boneyard/Burnyard; however, the boundaries of the Boneyard/Burnyard are not well defined and contaminants associated with operations at the Boneyard/Burnyard may not be restricted to the eastern bank of NT-3. Groundwater for Well GW-087 has shown contamination with VOCs (trichloroethene, 1,1,1-trichloroethane, perchloroethene, dichloroethene, and benzene) and radionuclides (gross alpha, gross beta, and thorium), an indication that shallow groundwater at the Boneyard/Burnyard has been contaminated.

The present shallow groundwater monitoring wells are adequate to monitor changes in groundwater moving off site; however, no information is available concerning groundwater quality on site. Shallow groundwater will be sampled in selected soil borings.

Habitats. Habitat delineations at the Boneyard/Burnyard and HCDA will reveal the locations of protected habitats and habitats where the ecological indicator receptors are likely to be exposed to site contaminants. During Phase 1 sampling, wetlands in the vicinity of the Boneyard/Burnyard and HCDA will be identified from historical information and walkover surveys. The size and location of each wetland, as well as a description of the associated vegetation, will be recorded. Aquatic and terrestrial habitats will be delineated throughout the site. The locations of various habitat types will be entered in the Government Information System mapping system.

Biota sampling and toxicity testing. Organic constituents of hazardous wastes are likely to be toxic to earthworms and other soil organisms. Metal and radionuclide contaminants are also likely to be transferred from soil to the food web. Biota sampling will be done during Phase 2 and will be focused on areas shown during Phase 1 to contain physical media contaminants. Likely sampling sites are downslope from the cap, where leachate from the waste areas may have contaminated surface soils.

Grasses and browse and earthworms will be sampled in areas of probable surface soil contamination as delineated by the screening radiological survey. If the populations are reduced (as described in Fig. 6.2), then the soil samples will be taken at each location for phytotoxicity and earthworm toxicity tests. Amphibian larvae will be sampled for whole-body contaminant burdens in wetlands that are found to be contaminated and that contain sufficient water to support the development of amphibians. One surface water sample and one sediment sample will be analyzed for COCs at each location.

Terrestrial small mammals will be sampled in appropriate habitat identified by the habitat survey. Statistically designed sampling of grasses and browse and earthworms will be done in areas shown in Phase 1 to contain contaminated biota (Sect. 6.1.1.4). Terrestrial burrowing mammals will be sampled in appropriate habitats identified by the habitat survey. Wetlands shown in Phase 1 containing contaminated amphibian larvae will be sampled further to delineate the extent of contamination in soil, sediment, and water by taking additional soil or paired surface water/sediment samples to be analyzed for COCs. Wetland areas will be divided into subareas, roughly 20 m² (66 ft²), and a sample location will be placed at random within each subarea.

6.1.2.4 SL 1

Historical data are not sufficient to indicate whether the SL 1 constitutes an immediate hazard to human health or the environment. Hazardous wastes may not exist at this site; therefore, it is likely that this site poses no immediate or long-term risk to human health or the environment. Using the staged RI process outlined in Chap. 1, Stage 1 needs to be completed for SL 1 and may result in a recommendation for no further investigation, unless hazardous wastes are discovered (Fig. 6.1). Y-12 Maintenance records will be evaluated to determine whether hazardous wastes were put into SL 1. Soil borings will be installed downgradient from the SL 1 site. In addition, soil borings and monitoring wells installed at the OLF and BYBY sites will serve as upgradient wells so SL 1. Sampling of existing downgradient wells in Bear Creek will also be carried out. If hazardous wastes are not being released from SL 1 such that the net effect on groundwater is that SL 1 alone would not cause contaminant concentrations to be above ARARS, then no further investigation will be recommended.

6.1.2.5 Bear Creek Burial Grounds

Historical data for the Bear Creek Burial Grounds indicate that there are substantial volumes of hazardous waste materials at this site that may pose long-term risks to human health or the environment. Activities at the Burial Grounds conducted under RCRA closure operations have substantially reduced the immediate hazards posed by these sites. BG-A North, BG-A South, BG-B, BG-C West, and ORPs 1 and 2 have been capped with RCRA caps. The WIPs and BG-C East are currently being capped. A clay cap covers the original location of Tributary NT-7. Water now drains from this area through a constructed drainage channel. A leachate collection system, consisting of a series of buried porous and double-line pipes, captures leachate from various seeps in the NT-7 and NT-8 watersheds. Following the staged RI process outlined in Chap. 1, minimal data will be required to complete Stage 1. The RI will continue whether removal actions are required or not; therefore, the data needs outlined in this section constitute those needed for completion of Stages 1 and 2 of the staged RI/FS process for the Burial Grounds.

Burial Grounds soils and solid wastes. A total of 132 soil and 29 stream sediment samples have been taken within the boundaries of the Burial Grounds in three separate investigations between 1983 and 1990 (Geraghty and Miller 1985; Herbes 1988; Bogle et al. 1991).

Both liquid and solid wastes were deposited in the Burial Grounds during its operation. It is assumed that liquid wastes exit the trenches by leaching to groundwater or flowing through seeps to the tributaries to Bear Creek that drain the Burial Grounds. This process has been observed by past investigations, and a leachate collection system is currently

operating at this site. Solid wastes remain in place, and leachate formed when water contacts these wastes may migrate by way of the same routes as liquid wastes.

Liquid wastes that migrated from the Burial Grounds were captured by the ORPs located in Tributaries NT-6 and NT-7. Sediments from these ponds that exceeded 25 mg/kg PCBs were excavated and removed; the remainder are in place and covered with RCRA caps.

PCB contamination has been recorded in stream sediments from Tributaries NT-7 and NT-8, which drain the central and western portions of the Burial Grounds; but PCB contamination was not detected in NT-6, which drains the eastern portion of the Burial Grounds. PCB contamination has also been recorded in the main stem of Bear Creek downstream of the Burial Grounds. Uranium was present in Bear Creek sediments and in tributary sediment samples taken from NT-6, NT-7, and NT-8 near their confluence with Bear Creek. Elevated concentrations of metals (nickel and lead) were also present in these samples.

Soil core data indicate PCB contamination of soils to the east of BG-A South in an area that has now been capped; however, PCBs were absent from soil samples taken south of BG-A South. PCB contamination associated with elevated uranium concentrations has been recorded in soil core samples taken northwest of BG-A North in an area that has since been capped.

Cores recovered from within the boundaries of the Walk-In Pits (North and South), BG-C East, and BG-B show that PCB-contaminated soils exist throughout the entire area in varying concentrations. A single soil sample recorded PCB contamination (0.11 ppm) beyond the physical boundaries of BG-B to the east. Butyl benzyl phthalate has been recorded at levels from 4440 ppb to 5300 ppb in soil cores east of Walk-In Pit North and BG-B. Samples from the uppermost 1.8 m (6 ft) of soil in this area were reported to have elevated concentrations of arsenic (17,000 to 20,000 ppm). Analysis of an archived sample in 1985, however, indicated an arsenic concentration of 100 ppm.

No soil sample data exist for the areas to the south or west of BG-B, BG-C, and Walk-In Pits North and South. No soil sample data are available for the areas in the vicinity of the remaining pits. BG-B, the Walk-In Pits, and BG-C East have been or are currently being covered with a RCRA cap; therefore, direct sampling of these areas cannot be carried out. Areas downgradient and along strike from the caps will be sampled.

Solid wastes in the Burial Grounds have been isolated from the environment either by covering with a RCRA cap or by institutional controls; therefore, they do not pose an immediate threat to human health or the environment through direct contact pathways. Data concerning surface water quality are not sufficient to determine if an immediate threat is posed by units in the Burial Grounds. Surface water is not adequately monitored in the vicinity of the Burial Grounds during storm flow events. To complete Stage 1 of the staged RI process outlined in Chap. 1 and the beginning of Chap. 6, data are needed to determine the flux of contaminants from uncontrolled seeps to surface waters. This will require a comprehensive survey of all seepage areas in the Burial Grounds, during both base flow and storm flow conditions. Water samples will be taken from seeps that are outside the capture zone of the two leachate collection systems in the NT-7 and NT-8 drainage basins. If seepage is detected that constitutes an immediate hazard, a removal action will be required to

remediate this problem. Likely removal action could be construction of a leachate collection system or modification of existing collection systems.

Trenches in the Burial Grounds received a variety of wastes; therefore, the most cost-effective remedial alternative will probably address each trench on an individual basis, as well as addressing groups of trenches as a whole. In general, waste trenches that received liquid wastes will have very different remedial alternatives from those that contain only solid wastes. To determine the nature and extent of waste materials at the Burial Grounds in Stage 2 of the staged RI, Y-12 Plant Maintenance records will have been examined to create a preliminary inventory of waste materials known to have been placed in the waste trenches. (See Sect. 2.3.5.1.) Using the preliminary waste inventory and environmental data described in Chap. 3, trenches will be grouped into the following two categories: those that received combined liquid and solid wastes and those that received only solid wastes.

For trenches that received liquid wastes, it is assumed that the majority of the wastes have migrated away from the trench and are not present in their original forms. In this situation, characterization of the nature and extent of residual contamination in soils in the vicinity of the trench (derived from migration of liquid contaminants) will be determined. Where seeps have been identified that can be associated with individual trenches or groups of trenches, soils in the migration pathway from the trench to the seep will be sampled. Aqueous liquid wastes are assumed to have mixed with groundwater and to have migrated as plumes away from the trenches along the groundwater flow path. Contaminant plumes such as this may also have left contamination in soils along the flow path where metal contaminants have adsorbed to soil particles. LNAPL wastes will have migrated laterally away on top of the groundwater table and will have left a trail of residual LNAPL in soils above the groundwater table along the migration pathway.

DNAPLs will have migrated downward to an impermeable barrier, which most likely is the bedrock aquifer. These wastes will have left a trail of residual DNAPL droplets in the aquifer material and, in the case where a trench does not intercept the groundwater table, in vadose zone material below the trench.

Where liquid wastes have migrated away from Burial Ground trenches, residual oils remain in the soil. Soil sampling will be necessary to characterize the nature and extent of soil contamination. Soils will be sampled downgradient of trenches known to have had liquid waste materials deposited in them. Priority will be given to trenches located near observed seeps. For uncapped areas of the burial grounds where soil contamination is likely, surface soils will be sampled.

For trenches that contain solid wastes, the inventory of Y-12 Plant Maintenance records will be used to determine the types of wastes disposed in the trenches. The Y-12 Plant records are extensive and cover more than 30 years of waste deliveries to the Burial Grounds. The inventory of Y-12 Plant Maintenance records provides estimates on types and volumes of wastes disposed in the trenches.

To address the potential for migration of contaminants from the Burial Ground trenches, samples of waste leachate will be obtained, where possible. Leachability modeling of waste materials is not possible; however, measurement of the leachability of contaminated soil and modeling of leaching of contaminants from soils will be carried out. Data required for soil leachate modeling are listed in Table 6.2.

Burial Grounds groundwater and surface water. Twenty-six groundwater monitoring wells in the vicinity of the Burial Grounds meet the criteria for screen depth and location to be used by the RI (Table 6.3); three of these wells have been plugged and abandoned. One upgradient well (GW-042) is screened in unconsolidated deposits; two wells (GW-642 and GW-643) are screened in the shallow bedrock aquifer. The present shallow groundwater monitoring wells are adequate to monitor groundwater downgradient.

The recharge area for the storm water flow system upgradient of the Burial Grounds is not covered. The effect on the stormflow system of roads in the Burial Grounds is not known; however, the waste trenches farthest upgradient are most likely to be affected by stormflow because there are no barriers to stormflow up slope of these trenches. Data are needed to address the potential for storm flow to come into contact with the waste. Stormflow monitoring tubes will be installed upgradient and downgradient of waste trenches likely to be affected by stormflow.

Habitats. Habitat delineations in the vicinity of the Burial Grounds will reveal the locations of protected habitats and habitats where the ecological indicator receptors are likely to be exposed to site contaminants. During phase 1 sampling, wetlands in the vicinity of the Burial Grounds will be identified from historical information and walkover surveys. The size and location of each wetland, as well as a description of the associated vegetation, will be recorded. Aquatic and terrestrial habitats will be delineated throughout the site. The locations of various habitat types will be entered in the Government Information System.

Biota sampling and toxicity testing. Organic constituents of hazardous wastes are likely to be toxic to earthworms and other soil organisms. Metal and radionuclide contaminants are also likely to be transferred from soil to the food web. Biota sampling will be done during Phase 2 and will be focused on areas shown during Phase 1 to contain physical media contaminants. Likely sampling sites are downslope from the cap and at surface seeps, where leachate from the waste areas may have contaminated surface soils.

Grasses and browse and earthworms will be sampled in areas of probable surface soil contamination as delineated by the screening radiological survey. If the populations are reduced (as described in Fig. 6.2), then the soil samples will be taken at each location for phytotoxicity and earthworm toxicity tests. Amphibian larvae will be sampled for whole-body contaminant burdens in wetlands that are found to be contaminated and that contain sufficient water to support the development of amphibians. One surface water sample and one sediment sample will be analyzed for COCs at each location.

Terrestrial small mammals will be sampled in appropriate habitat identified by the habitat survey. Statistically designed sampling of grasses and browse and earthworms will be done in areas shown in Phase 1 to contain contaminated biota (Sect. 6.1.1.4). Terrestrial burrowing mammals will be sampled in appropriate habitats identified by the habitat survey. Wetlands shown in Phase 1 containing contaminated amphibian larvae will be sampled further to delineate the extent of contamination in soil, sediment, and water by taking additional soil or paired surface water/sediment samples to be analyzed for COCs. Wetland areas will be divided into subareas, roughly 20 m² (66 ft²), and a sample location will be placed at random within each subarea.

6.2 SAMPLING AND ANALYTICAL REQUIREMENTS

Sampling and analytical requirements for all sites in BCV OU 1 are detailed in Sect. 7.4.

6.3 DATA EVALUATION AND INTERPRETATION

All data collected as part of this RI will be entered into a database compatible with the Oak Ridge Environmental Information System (OREIS). Ten percent of the data will be validated to EPA Level IV quality standards, and the remaining 90% will be validated to a modified EPA Level III as discussed in Sect. 8.9. The validated data will then be used for statistical analysis and data presentation as needed to accomplish the data use discussed in Sects. 1.8.1 and 1.8.3 of the DQOs and identified in Table 6.1. The 10% to be validated to EPA Level IV will be chosen by field observation and will be taken regularly throughout the field exercise.

Data generated from analysis of soil samples will be summarized statistically in data tables. The tables will contain the minimum, maximum, mean, and standard deviation for each potential COC. These data will be used to calculate human health risk as described in Sect. 6.4 and to aid in defining the nature and extent of contamination.

Contamination by depth profiles will be generated for all data from each borehole. All values above the PRGs established in Sect. 5.2 will be graphed in a different color to aid in establishing preliminary remediation boundaries. Cross-sections will also be prepared from selected profiles to establish the vertical and horizontal extent of contamination. Visually estimated soil characteristics will also be represented on soil logs and cross-sections. Potentiometric maps and contaminant concentration maps will also be prepared. Geotechnical data will be represented on data tables.

Data from the geophysical survey will be interpreted by plotting on maps and contouring to identify likely locations of waste trenches. Conductivity data for soils will be contoured to identify spatial distributions of soil conductivity.

6.4 RISK ASSESSMENT

6.4.1 Human Health Risk Assessment for BCV OU 1

The regulatory agencies involved in Environmental Restoration activities on the ORR have partitioned the sites (in this particular case, the Oak Ridge Y-12 Plant Bear Creek Valley) into two categories of OUs: source Ous and integrator Ous. Source Ous are defined as on-site (property owned/managed by DOE-ORO) waste areas that are currently releasing or have the potential for releasing contaminants into the environment. Integrator Ous can receive the contaminants released from source Ous and transport them to an off-site receptor. Because of the complexity of the hydrogeologic regime and because many different sources could contribute to groundwater and surface water contamination, this classification was necessary. As a result, immediate attention can be given to source Ous while continued monitoring, characterization, and investigations of the integrator OU are being conducted.

The scope of the BCV OU 1 RI work plan entails using the data collected in the RI to accomplish three things: (1) to evaluate the current potential risk posed to human health from exposure to the potentially contaminated media (soil/waste), (2) to evaluate the future potential risk posed to human health from exposure to the potentially contaminated media (soil/waste), and (3) to prioritize how much this source OU is contributing to off-source OU release of contamination (to the integrator OU). These objectives will be accomplished by evaluating on-source OU human exposure to BCV OU 1 media (i.e., soil and inventoried wastes) by means of a BRA. Surface water and groundwater associated with BCV OU 1 will be evaluated in the BCV OU 4 RI in an integration point (IP) assessment (BCV OU 4 Phase 1) and in a valley-wide BRA (BCV OU 4 Phase 2).

The methodology employed to conduct the risk assessments will follow guidelines established in the *Risk Assessment Guidance for Superfund, Volume I: Human Health Evaluation Manual (Part A)* (EPA 1989d) and the annotated outline for an RI/baseline risk assessment report. In addition, recommendations from the Risk Assessment Council and other supporting documentation, including the results of the Background Soil Characterization Project for the ORR, will be followed.

The exposure pathways evaluated in the BCV OU 1 BRA include incidental ingestion of soil, dermal contact with soil, inhalation of wind-generated dust and volatiles from the soil, and external exposure to radionuclide contaminants in the soil. To evaluate the risk posed to human receptors from direct exposure to the inventoried wastes, a method will be used that includes an upper bound (acute exposure) and a lower bound on the risk.

6.4.2 Ecological Risk Assessment for BCV OU 1

The ecological risk assessment process differs from the human health risk assessment process in that it focuses on populations and communities rather than on individuals. Individuals are protected only if they are protected under the Endangered Species Act, are unique, or have such ecological importance that individual consideration is merited. Ecological risk assessment, as described in *Framework for Ecological Risk Assessment* (EPA 1992c), involves basic steps occurring in three primary phases: (1) problem formulation; (2) analysis, which includes exposure characterization and effects characterization; and (3) risk characterization. Problem formulation includes a preliminary characterization of chemical and physical stressors present in the ecosystem (Chap. 3), ecological components likely to be at risk (Sect. 4.2), and assessment and measurement endpoints (Sect. 6.1.1.4). Exposure characterization evaluates the interactions of the stressors with the ecosystem, especially each indicator population or habitat. Effects characterization evaluates the ecological response to the chemical and physical stressors in terms of the assessment and measurement endpoints. Data from both field observations and controlled laboratory studies are used to evaluate ecological effects. Risk characterization integrates the effects of exposure and stressor response on indicator populations or habitat components, using risk quotients as well as weight-of-evidence arguments and then further interprets the ecological significance, including the potential for ecosystem recovery.

6.5 FEASIBILITY STUDY

The following sections describe the tasks necessary to conduct and produce the FS report. The report will follow the format outlined in EPA's *Guidance for Conducting*

Remedial Investigations and Feasibility Studies Under CERCLA (EPA/540/G-88/004) (EPA 1988a) and the annotated outline developed by the DOE-ORO Document Content and Review Committee. The FS report will be prepared to integrate NRDA and NEPA requirements. When EPA and TDEC approve the FS, a Proposed Plan will be submitted presenting the preferred alternative for an interim remedial action, followed by a ROD for interim action.

6.5.1 FS Contractor Responsibilities

The FS contractor will prepare the FS within project quality, cost, and scheduling goals. The FS contractor will also provide engineering and technical resources to support QC efforts, provide project management, and coordinate project activities to ensure that established goals are achieved. Monthly progress reports on the FS process will be issued to DOE. FS contractor activities for BCV OU 1 include scoping activities and technical support of the RI; these activities will be completed on final approval of the ROD for interim action prepared for this OU.

6.5.2 Scope and Assumptions

As stated in Sect. 5.2, the general goal for BCV OU 1 remedial action is to protect human health and the environment from the contaminant sources within the OU. To meet this goal, the CERCLA process will be followed. The process will include characterization of this source control OU to the extent necessary to (1) identify, screen, and develop remedial action alternatives; (2) select a preferred remedial alternative or a set of alternatives; and (3) prepare the required decision documents.

The FS contractor will be responsible for completing the FS process for BCV OU 1. This process will require preparing the following milestone documents: (1) the FS report, which documents the process used to examine the remedial technologies and also incorporates the requirements under the Natural Resource Damage Assessment (NRDA) and NEPA; (2) the Proposed Plan, which is a summary plan for presenting the selected remedial alternative; and (3) the ROD for interim action, which is the legal document that confirms that selected remediation methods meet statutory and regulatory requirements. The FS/Proposed Plan/ROD process may include the following principal Work Breakdown Structure elements:

- Treatability studies.
- FS development.
- NEPA activities.
- FS OU coordination/oversight.
- Proposed Plan development.
- ROD development.

6.5.2.1 Treatability studies

If necessary, treatability studies will be conducted to (1) provide data to allow treatment alternatives to be fully developed and evaluated during the detailed analysis and (2) support the remedial design of a selected alternative.

6.5.2.2 FS development

The FS contractor will use the format presented in the CERCLA RI/FS guidance document and the DOE-ORO Document Content and Review Committee FS annotated outline to develop the FS. As required, exceptions and modifications to the format will be made for site-specific conditions in BCV OU 1. Based on the information obtained during the RI, selected alternatives will be analyzed and compared using the following seven criteria:

- Overall protection of human health and the environment.
- Compliance with ARARs.
- Short-term effectiveness.
- Long-term effectiveness.
- Reduction of contaminant toxicity, mobility, and volume.
- Implementability.
- Cost.

State and community acceptance will be evaluated in the ROD.

The alternatives will be developed in sufficient detail to allow the preferred alternative to be selected in the Proposed Plan. The draft FS report will be issued to EPA and TDEC for review. Submittal of the draft FS report to EPA and TDEC is an FFA milestone.

6.5.2.3 NEPA activities

To comply with the federal facility requirements under NEPA, the remedial alternatives will be evaluated for their impact on the environment. NEPA establishes public policies and goals for protecting environmental quality and mandates procedural requirements to be considered when implementing decisions that may impact the environment. DOE Order 5400.4 requires that NEPA and CERCLA be integrated to the maximum extent practical to avoid duplication of effort. NEPA issues to be addressed include rare and endangered species, archeological studies, wetlands, and floodplains. The environmental impacts of the alternatives will be considered integral to the evaluation and selection process.

6.5.2.4 FS OU coordination/oversight

The objective of the Integrating Contractor for FS oversight activities is to provide review and technical input during the FS. This ensures that the data and information obtained during the RI are available for incorporation into the FS report and that they are adequate to prepare a Proposed Plan and support development of the ROD.

6.5.2.5 Proposed plan development

The Proposed Plan for BCV OU 1 will be prepared to present the preferred remediation alternative to the public. The Proposed Plan documents the investigation process, administrative and regulatory actions, and remedial alternatives. The proposed plan is a primary milestone document made available for public comment and subject to administrative and regulatory reviews.

6.5.2.6 ROD development

The final step in the decision process for any interim remedial action for BCV OU 1 will be preparation of the ROD. This legal document formally describes the preferred remediation alternative and establishes the remediation schedule and monitoring plan for the site. The BCV OU 1 ROD will be written in accordance with the statutory requirements of CERCLA and applicable federal and state requirements. It will contain a decision summary outlining the nature and extent of the contamination and associated risks at BCV OU 1, the evaluation and analysis of the remedial action (RA) alternatives considered, and an explanation of how the selected alternative will meet statutory requirements. The ROD will also contain a responsiveness summary addressing the public comments obtained during public review of the Proposed Plan and public examination of the Administrative Record. The ROD is a primary FFA milestone.

6.6 REMEDIAL DESIGN AND RA PLAN

6.6.1 Remedial Design

This section presents the scope of the Remedial Design (RD) effort required to prepare an RD work plan; perform any required engineering studies; prepare a Title I 30% design package; and prepare Title II 60%, 90%, and 100% final design reports for BCV OU 1 remediation.

6.6.1.1 Scope and assumptions

The scope of the RD work Plan for the remediation of BCV OU 1 is to provide necessary supporting documents to implement each RA activity. These documents require approval from TDEC and EPA. The following sections present details of the scope of work and assumptions associated with remediation of BCV OU 1.

6.6.1.2 RD work plan

The RD work plan will provide the technical and management approach for the RD work. The RD contractor will prepare a draft RD work plan based on the scope and design criteria specified in the Proposed Plan and the ROD for interim action. The RD work plan will include the detailed design process and schedule for the design effort and will be prepared in accordance with CERCLA/NEPA regulations. The RD contractor will incorporate comments from EPA and TDEC and submit the final RD work plan for approval.

6.6.1.3 RD work plan oversight activities

During the preparation of the RD work plan, DOE will provide review and technical input to ensure that the scope of work is adequately defined in accordance with the criteria specified in the ROD for interim action. Review of the draft RD work plan will also ensure that the selected technologies do not violate the intent of the ROD and that the work plan meets all regulatory and administrative requirements.

6.6.1.4 RD report Title I design

Based on the engineering studies and other information available from the RI/FS, the RD contractor will prepare Title I (30%) design/construction drawings for remedial activities showing the extent of remedial activities, site plan, details, and outline of specifications for the work involved.

6.6.1.5 RD report Title II design

On approval of Title I design document, the RD contractor will prepare Title II engineering designs, analyses, and calculations required for all civil, structural, mechanical, and electrical construction, construction drawings, technical specifications, and cost estimates.

All documents will be submitted to Energy Systems for comment at 60% completion. On resolution of comments, the 90% design package will be submitted to Energy Systems and DOE for review and comment. Upon resolution of 90% comments, the final design report (100%) for remedial activities will be submitted to EPA and TDEC for review and approval. Agency comments will then be incorporated to prepare design documents certified for construction.

6.6.1.6 RD oversight activities

During the review of the Title I design documents at 30% and the Title II design documents at 60% and 90%, all participants will provide their review and technical input to help submit the final design document to TDEC and EPA in the given time frame.

6.6.2 Remedial Action

This section presents the scope of the RA activities required to prepare an RA work plan and implement the remediation of BCV OU 1. The risk, uncertainties, and communication issues for the remedial activities are also identified in this section.

6.6.2.1 Scope and assumptions

The scope of the RA includes the RA work plan, construction management, Title III construction services, construction of one remediation design package, construction support, independent certification, and verification.

6.6.2.2 RA work plan

The RA work plan will (1) define the scope and objectives of the RA based on the ROD for interim action and the final RD report; (2) document the specific construction components of the RA; and (3) present the RA schedule, subcontracting strategy, QA plan, health and safety plan, and RA monitoring plan.

After comments from EPA and TDEC are incorporated, the final RA work plan will be prepared for approval and implementation.

6.6.2.3 RA work plan oversight activities

The purpose of this activity is to provide technical input and review during the preparation of the RA work plan. During this time, all the DOE prime contractors will review the work plan. This will ensure that the proposed construction efforts implement the selected remedial action consistent with the ROD for interim action and the final RD report. This oversight activity will also ensure that the bid process and implementation plans comply with administrative and regulatory requirements.

6.6.2.4 RA integration

The objective of this element is to provide construction management, independent certification, Title III services, and construction support as required.

RA integration for BCV OU 1 remedial activities includes the following tasks:

- Ensuring that subcontracted work is performed on schedule, in accordance with all technical requirements, and in compliance with the Environmental Safety and Health Program, the QA Program, the Waste Management Program, and the Security Program.
- Performing field inspections, providing as-built drawings, approving Design Change Notices and Field Change Requests as applicable, and ensuring that construction is accomplished according to final design requirements.

APPROVALS

FIELD SAMPLING PLAN
FOR THE REMEDIAL INVESTIGATION OF THE
BEAR CREEK OPERABLE UNIT 1
AT THE OAK RIDGE Y-12 PLANT

DOE/OR/01-1161/V1&D2

March 1994

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7. FIELD SAMPLING PLAN

7.1 INTRODUCTION AND PURPOSE

BCV OU 1 is one of two source OUs within BCV. The boundaries of the integrator OU for BCV (BCV OU 4) are from the crest of Chestnut Ridge to the crest of Pine Ridge and from the surface water divide near the S-3 Site to the groundwater divide west of State Highway 95 between Bear Creek and Grassy Creek. The source OUs in BCV are described below:

- BCV OU 1 comprises the S-3 Ponds; Oil Landfarm; Boneyard/Burnyard; HCDA; SL 1; Bear Creek Burial Grounds (BG-A North, -A South, -B, -C East, -C West, -D, -E, and -J); ORPs 1 and 2; Walk-In Pits; and Uranium Vaults.
- BCV OU 2 includes the Rust Spoil Area, Spoil Area 1, and the SY-200 Yard.

BCV OU 1 (source OU) is the focus of this RIWP. Groundwater and surface water will not be addressed as media to be remediated in this BCV OU 1 RI. Risk posed to off-source and off-site human receptors by means of groundwater and/or surface water pathways will be addressed in the BCV OU 4 (integrator OU) RIWP. Sediments and soils that constitute floodplain soils will be considered in BCV OU 4 RIWP. For this BCV OU 1 RIWP, risks from exposure to soils and sediments within the BCV OU 1 site boundaries will be evaluated.

The overall RI/FS goal for BCV OU 1 is to select remediation alternatives at reasonable cost that are protective of human health and the environment, that maintain protection over time, and that minimize the volume of untreated waste. The objective of the RI is to provide sufficient information to assess the risks posed to human health and the environment and to support the development, evaluation, and selection of appropriate response alternatives. This work plan provides justifiable rationale and guidance for the collection and analysis of the data required to fulfill the objectives of all stages of the RI.

A staged approach to this RI is outlined in Chap. 1 of this work plan. Though each stage of the RI has different objectives, a goal of the Work Plan is to ensure that data acquired during each stage of the RI are comprehensive and that no repeat sampling of any media at any location is required as a result of changing objectives. The goal of each sample detailed in the sampling plan is to provide sufficient information to support the final remediation of the site. The following objectives will be met to assess the risks contributed by this OU to the overall risk to human health and the environment in BCV and to allow a ROD for interim actions to be issued:

- To determine whether the wastes pose an on-site risk to human health and the environment at the present time
- To determine whether the wastes will pose an on-site risk to human health and the environment in the future
- To determine whether the wastes pose an off-site risk to human health and the environment at the present time
- To determine whether the waste will pose an off-site risk to human health and the environment in the future

- To determine whether the wastes and/or waste derivatives are in compliance with ARARs and/or decision criteria from the integrator OU (OU 4).
- To support the screening of appropriate alternatives during the remedy selection and to provide technically and legally defensible data on which to base credible decisions regarding risk levels and cleanup alternatives

7.2 PROJECT ORGANIZATION AND RESPONSIBILITIES

The BCV OU 1 RI project organization is illustrated in Fig. 7.1. The responsibilities of Energy Systems and the RI subcontractor are presented in Table 7.1.

7.3 PROJECT DESCRIPTION

7.3.1 History and Current Conditions

Six sites at three discrete areas in BCV have been included in BCV OU 1 to be evaluated as contaminant source OUs. Some sites within BCV OU 1 have undergone environmental restoration under RCRA. Part of BG-A North, BG-A South, and BG-C West; the Oil Landfarm; the HCDA; and the S-3 Ponds were closed in place as landfills under RCRA. These actions were based on considerable data concerning contaminants in environmental media at these sites. SL 1 was capped and closed with a single-layer clay cap. No action has been taken at the Boneyard/Burnyard, and there is little or no information concerning contaminants at this site.

The site name, unit number, and unit status of each site in BCV OU 1 are listed in Table 7.2. Detailed descriptions of site operating history and environmental background are presented in Chaps. 2 and 3 of this work plan. A brief description of each site in BCV OU 1 follows.

S-3 Ponds. The S-3 Ponds were constructed in 1951 and consisted of four unlined surface impoundments covering a total area of roughly ~1.5 ha (3.7 acres) with a storage capacity of 38 million L (10 million gal). During operation, up to 21,000 L/d (5500 gal/d) of effluent was pumped to the pond through a pipeline from the Y-12 Plant. The S-3 Site underwent RCRA closure in 1988 and now operates as a parking lot. Preliminary COCs are nitrate, uranium, metals, and organic solvents.

Oil Landfarm. The Oil Landfarm was used for the land application of waste oils and coolants contaminated with beryllium, depleted uranium, PCBs, and chlorinated hydrocarbons. Disposal operations were discontinued in 1982. Final RCRA closure of the site was certified by TDEC in December 1990.

Boneyard/Burnyard. The Boneyard/Burnyard consists of ~3.2 ha (8 acres) that were used from 1943 to 1970 as a disposal site for waste from the Y-12 Plant. Burning and disposal of debris and sanitary, metallic, chemical, and radioactive wastes are known to have occurred. The site has been abandoned and is predominantly covered with grassy vegetation. The southeastern portion of this site is overlain by the HCDA. The HCDA covers ~0.8 ha (2 acres) and was used to burn or neutralize liquid and gaseous wastes from 1975 until 1981.

Fig. 7.1. BCV OU 1 RI functional organizational chart.

**Table 7.1. Energy Systems and RI subcontractor responsibilities
for the Y-12 Plant ER Program assessment**

Role	Subcontractor responsibilities	Energy Systems responsibilities
Project Manager	Coordinate all project plans and activities	Identify project-specific requirements Coordinate project activities, as necessary
Medical Monitoring	Ensure all subcontractor project personnel have updated yearly physicals Ensure all subcontractor project personnel have updated respirator fit tests	Identify for the subcontractor any additional or site-specific medical monitoring requirements
Health Physics	Notify ER Division Project Manager of need for an Energy Systems health physicist (HP), including time and place Provide an HP or trained designee to be on-site during field activities to screen samples, equipment, and personnel, as necessary	Provide an HP, when requested, for screening work areas, samples, equipment, or personnel Prepare tags to approve off-site shipment of samples
Industrial Hygiene	Provide site-specific health and safety plans Provide health and safety audits	Review/approve health and safety plans Provide health and safety audits
Emergency Response	Contact the Plant Shift Supervisor in the event of an emergency at the investigation site Follow Plant Shift Supervisor instructions	Manage all emergency situations (i.e., via the Plant Shift Supervisor)
Utilities	Alert the ER Division Project Manager of the need for a utility check	Provide utility checks
NEPA	Ensure fieldwork is done in accordance with NEPA documentation requirements	Ensure assessment activities are cross checked
Groundwater Protection	Provide location and design of monitoring wells Provide field oversight of well installation	Approval of location and design of monitoring wells Install, plug and abandon, and maintain monitoring wells Sample monitoring wells
Security (access)	Provide necessary information for badges, passes, and access	Ensure badges, passes, and access are requested and obtained
Permits	Identify project-specific permit requirements Provide necessary permit information	Ensure permit is complete Submit permit to correct Y-12 Plant department Coordinate permit approval
Analytical Project Office	Determine project-specific analytical requirements and the number, types, and delivery schedule for samples	Coordinate ER Program analytical requirements

Table 7.1 (continued)

Role	Subcontractor responsibilities	Energy Systems responsibilities
Y-12 Lab (if required)	Provide the number, types, and delivery schedule for samples	Provide lab space and sample analyses Approve project-specific analytical requirements
Quality assurance	Develop site-specific QA plans Provide QA audits	Review QA plans Develop ER programmatic requirements Provide QA audits
Training	Maintain up-to-date training records Ensure project personnel have correct training Identify project-specific training requirements	Assist in identifying project-specific training requirements Audit training records for subcontracted personnel Identify, schedule, and reserve space in required Energy Systems training sessions
Waste Treatment	Develop project-specific waste management plans detailing waste treatment options Provide estimates of types and quantities of waste needing treatment	Review/approve waste management plans Arrange for waste treatment with various Y-12 departments
Waste Transportation	Notify appropriate Energy Systems personnel that waste transportation is needed Correctly package and label waste for transport	Arrange for waste pick-up and disposition of waste
Classification	Identify project-specific activities	Determine the level of classification of project-specific activities
Sample Transportation	Follow Energy Systems-specific requirements	Provide guidance for sample transportation

Table 7.2. Remedial units of BCV OU 1

Remedial Unit	Unit No.	Status
S-3 Ponds	T-004	Closed under RCRA in 1988
Oil Retention Ponds 1 and 2	T-008, T-009	Closed under RCRA in 1990
Oil Landfarm	T-014	Closed under RCRA in 1990
Bear Creek Burial Grounds	D-024	Some sections closed, others undergoing closure
Sanitary Landfill 1	D-101	Closed in 1983
HCDA (Burnyard/Boneyard)	D-024-HC	HCDA covered with a RCRA-style cap

The HCDA is covered with a RCRA cap and was closed with the Oil Landfarm.

SL 1. SL 1 received various types of nonhazardous wastes from the Y-12 Plant. Waste disposal at SL 1 was carried out from 1962 to 1982 at which time the site was graded, capped with clay, and closed in 1983 under a TDEC-approved closure plan in accordance with TDEC rules and regulations for sanitary landfills.

Burial Grounds. The Burial Grounds comprise a series of waste disposal units covering an area of ~140 ha (350 acres). Each waste disposal unit consists of a series of trenches excavated to between 4.3 and 7.6 m (14 and 25 ft) below grade. The trenches received a variety of hazardous and nonhazardous solid and liquid wastes from the Y-12 Plant. All hazardous waste disposal operations were discontinued in 1981. All trenches known to have received RCRA hazardous material have been capped or are in the process of being capped as part of a RCRA closure. The Burial Grounds include ORPs 1 and 2, the Walk-in Pits, and the Uranium Vaults. The ORPs were constructed to intercept seepage from the Burial Grounds trenches. Both ponds were capped and closed under RCRA in 1990. The Walk-in Pits were used to dispose of solid or drummed liquid wastes including pyrophoric saw fines of uranium and thorium and shock-sensitive chemicals. A wide range of contaminants may have been disposed of in the Burial Grounds.

7.3.2 Site Conceptual Model

A site conceptual model that illustrates the potential pathways of contaminant migrations, or flows, among various abiotic media and, ultimately, to receptors is described in Chap. 3 of this RI work plan. The site conceptual model for transport and exposure pathways is presented in Figs. 3.6 and 3.7. The development of the site conceptual hydrologic model for BCV OU 1 relies on a conceptual model for groundwater flow that is applicable throughout the ORR (Solomon et al. 1992). The components of that conceptual model are included in the conceptual model in general and applied to the specific needs of the shallow groundwater/surface water system within BCV. Contributions to the hydrologic system from the soils are by way of contaminant leaching from soils and wastes. Direct contact pathways to human and ecological receptors include direct or acute exposure to soils and waters, ingestion of contaminated soils and/or contaminated surface water or groundwater, food

ingestion, dermal contact with soils, inhalation of volatiles, inhalation of dust, and penetrating radiation.

The site conceptual model is based on the hypothesis that potential contaminant migration pathways exist downgradient from waste trenches. Aqueous contamination may migrate laterally with stormflow or vertically to groundwater with recharge. Ionic contaminants dissolved in aqueous solutions may precipitate or sorb to mineral surfaces in soils along the pathway. LNAPLs may migrate downward until the groundwater table is reached. These chemicals may then migrate laterally away from the waste location, downgradient along the groundwater table. DNAPLs may migrate downward through the groundwater table and into the aquifer. Migration of DNAPL or LNAPL leaves a trail of residual contamination, and vertical changes in the groundwater table may spread this residual contamination in the soil over the range of groundwater table variation.

7.3.3 Identification of Investigation Requirements

The strategy for a source OU in BCV is to stop releases from sources to the surface water and shallow groundwater system and to reduce risks to human health and the environment to acceptable levels based on current and future land uses at the source sites. The sources of contamination at the BCV OU 1 sites to be investigated are soils and waste materials. Because there are few specific ARARs pertaining to soils, the investigation requirements for soils in BCV OU 1 are set by the PRGs. PRGs for this RI were established by two methods. The first set of PRGs was determined by health-based standards based on direct contact pathways (i.e., ingestion and inhalation of contaminated material). The second set of PRGs was calculated using the Summers Model for soil leaching (EPA 1989c). The allowable level of soil contamination can be calculated using this method based on contaminant concentrations in the groundwater that result from soil leaching. To establish PRGs by this method, the Safe Drinking Water Act MCLs and MCLGs were used as an acceptable groundwater concentration, and the acceptable soil concentration was calculated by inverse modeling using the Summers Model. Assumptions used to calculate the PRGs are discussed in Sect. 5.2 of this work plan. Feasible treatment technologies are discussed in Sect. 5.3 of the work plan.

7.4 SITE ACTION PLAN

General site data needs for physical characterization of media in BCV are discussed in Sect. 6.1.1 and summarized in Table 6.1. Site-specific data needs and rationale for sampling at each site are discussed in Sect. 6.1.2 and summarized in Table 6.5.

At each site, data are needed to characterize the nature and extent of contamination in soils and wastes and to evaluate current and future on-site and off-site contaminant migration pathways. The nature and extent of contamination are needed not only to determine the extent of contamination at each site but also to allow evaluation of future fluxes of contaminants along migration pathways.

The site action plan described here uses the staged approach to the RI/FS that is detailed in Chap. 1 of the work plan. BYBY and SL 1 are at Stage 1 of the process; the Burial Grounds, S-3 Site, and the OLF are at Stage 2. This site action plan details the samples to be taken to complete Stage 2 for each site. Soil sampling activities for Stage 2 at the Burial

Grounds and at the BYBY will be carried out in two phases because of the large size of each site [69 and 3.2 ha (170 and 8 acres), respectively] and because of the uncertainty concerning probable locations of contaminants. For each site, the first phase of sampling is detailed here, and the rationale for locating samples in the second phase is described in Sect. 1.8 and in Chap 6. Ecological sampling will be completed for each site as a single phase after the phase 1 soil sampling. This will be carried out during the spring and early summer which is the preferred season for sampling these media.

Historical data from the current groundwater monitoring well system in BCV will be used to analyze historical groundwater quality trends. Sampling of groundwater in a few selected wells will be carried out to ascertain changes in groundwater quality in the vicinity of the source sites.

7.4.1 Technical Procedures for Sample Collection

Section 8.2 discusses the technical procedures that will be used as guidelines to implement the sampling activities at BCV OU 1. All sampling activities, where applicable, will follow the technical procedures established in *Environmental Surveillance Procedures, Quality Control Program* (Kimbrough et al. 1994) and applicable SAIC Field Technical Procedures. In some cases for OU 1, innovative sampling methods are proposed for which Energy Systems has not published procedures (e.g. direct push sampling for soils and groundwater). In such cases procedures will be developed for these methods specifically for their application at sites in OU 1. A procedures manual will be prepared and controlled by SAIC and approved by the Energy Systems Y-12 ER Project Manager and QA Specialist (QAS).

7.4.1.1 Soils

The soil sampling activities will include drilling and sampling soil borings and collecting shallow soil samples using a hand auger. Shallow soil samples (top 12 in.) will be collected in accordance with guidelines specified in method ESP-303-2, Rev. 0. Soil samples may be collected from soil borings by one of two methods: either using a direct push method (such as Geoprobe Systems®) or using conventional mechanical auger systems. The decision as to which sampling method to use will be made on a site-specific basis and will depend upon how successful direct push methods are at penetrating and sampling the formation at each site. When direct push methods are used then a procedure applicable for use of this method at OU 1 will be developed. In addition procedures for handling samples from the direct push systems will be developed. When conventional mechanical auger methods are used then samples will be collected in accordance with guidelines specified in method ESP-303-4, Rev. 0. Soil samples will be put into containers, preserved, and handled in accordance with methods ESP-400, ESP-701, and ESP-800.

Additional field activities during soil boring will be as follows:

- The soil will be visually inspected and described using the Unified Soil Classification System; Munsell Color Charts will be used to describe soil color.
- For auguring, the split-spoon blow count will be recorded per American Society of Testing and Materials Method D 1586-84 on the soil lithology logs.
- For auguring, shelly tube samples will be taken following procedures outlined in ESP 303-5. For direct push methods, shelly tube samples will be taken using a procedure that

will be developed for OU 1. All Shelby tube samples will be screened for radioactivity. The contracted laboratory will possess a current license Special Nuclear Material (SNM) license and will have demonstrated an acceptable method for disposing of laboratory generated waste.

- All sampling and drilling equipment will be decontaminated according to the procedures detailed in methods ESP-801 and ESP-802.
- All sample locations will be surveyed by a licensed surveyor to the nearest 0.03 m (0.1 ft) using NAD 83 State Plane coordinate system. The elevation of the sample location will be surveyed to the nearest 0.003 m (0.01 ft) by the surveyor.
- During the sampling event all soils will be screened for potential organics and radiological contamination following the guidelines established in methods ESP-307-6 and ESP-307-7.

Analytical methods and detection limits used in this study for the analysis of water, soil, and sediment samples are listed in Chap. 8 and Appendix D.

7.4.1.2 Groundwater

Groundwater will be sampled from soil borings using direct push methods at the time of sampling the soil. This activity will be carried out following procedures developed for OU 1. The field sampling plan includes the installation of temporary piezometers that will be used to evaluate the immediate response of groundwater levels to rainfall events and to establish hydraulic gradients. Temporary piezometers will be installed in accordance with SAIC procedures. All piezometers will be screened such that the screen intercepts the groundwater table and the likely range of vertical variation of the groundwater table. This will allow the piezometer to test groundwater at the water table/vadose zone interface and intercept any LNAPL that may be on the groundwater. If recharge to a temporary piezometer is too slow such that a full sample cannot be taken to allow for analysis of all analytes then the sample will be analyzed for a prioritized list of analytes. Prioritization of analytes is as follows: Volatile organics, uranium and thorium, metals, nitrate, radiologicals, other analytes.

Existing monitoring wells that are to be sampled will be purged and sampled using micropurge techniques. A procedure for micropurging will be developed for OU 1. Where soils are fine-grained, and likely to cause difficulty in the purging of wells, wells will be purged at low flow rates until dry. All groundwater samples collected will be put into containers, preserved, and handled in accordance with ESP-400, ESP-701, and ESP-800. Water level measurements recorded from the piezometers will follow the guidance specified in ESP-302-1. Before initiating each groundwater sampling event, all equipment, well materials, and sampling devices will be decontaminated in compliance with ESP-801 and ESP-802. Stormflow tubes will be installed following SAIC procedures. An example of a stormflow tube construction is shown in Fig. 7.2.

7.4.1.3 Surface water/sediments

Surface water samples will be collected, as applicable, in accordance with ESP-301-1; storm event sampling will follow ESP-301-6. Before each surface water sampling event, all sampling equipment will be decontaminated following the guidelines established in methods

Fig. 7.2. Construction diagram of tubes used for monitoring of water levels in the storm flow zone.

ESP-802 and ESP-802. All samples collected will be put into containers, preserved, and handled in accordance with methods ESP-400, ESP-701, and ESP-800.

7.4.1.4 Ecological samples

Terrestrial small mammals will be trapped in snap-traps or bucket traps, identified, and counted; and selected individuals will be analyzed for whole-body burden of COCs following guidelines set forth in TP-309-3. Amphibia will be sampled with dip nets as indicators of contaminant uptake from surface water ponds, seeps, and springs following guidelines in TP-309-5. Earthworms will be sampled according to the requirements of TP-309-8 as indicators of contaminant uptake from soil. A portion of each sample will be weighed, depurated, and reweighed before being analyzed for metals, PAHs, and PCBs, and samples in burn areas, dioxin 2,3,7,8-TCDD. Grasses and browse will be clipped and analyzed as sources of contaminants to herbivores by ingestion following TP-309-2. Toxicity tests will be done by standard EPA-approved laboratory procedures. The sampling locations will be chosen to maximize the comparability of physical media sampling data with biological sample data. All populations and habitat characterizations will be compared to similar characterizations performed on reference areas of BCV not affected by BCV OU 1 contaminants. Strict custody and quality control will be maintained for all of the sampling and analysis for the ERA.

Pond and wetland locations containing sufficient water for the development of amphibian larvae will be identified from the OU 1-wide habitat survey. Three of these locations within the area of contamination and one reference location upslope of contaminants will be chosen at random for evaluation of the potential for transfer of contaminants to amphibian larvae. At the appropriate time of year, the water bodies will be monitored for the development of amphibian larvae. Before they are mature, they will be sampled by dip-netting, individuals will be identified to the species level, and one composite sample of at least 165 g from each water body will be submitted for analysis. A field duplicate will be taken from one location, and a sample of at least 300 g will be taken at one location to allow for the preparation of matrix spike and matrix spike duplicate samples. An equipment rinsate will be performed on applicable equipment. Therefore, it is anticipated that four samples, one duplicate, one rinsate, and one matrix spike/matrix spike duplicate will be submitted.

Target analytes will be metals by neutron activation, PAHs, PCBs, dioxin 2,3,7,9-TCDD, and radiological parameters (gross alpha, gross beta, and gamma spectroscopy). If amphibian larvae are not found in water bodies at a time when other similar water bodies contain them, the surface water will be sampled for *Ceriodaphnia* toxicity tests. If contaminant transfer or toxicity is evident in any of the water bodies, additional water bodies will be sampled during phase 2 to evaluate the extent of exposure to amphibians, and by extension, to other aquatic biota and their predators.

7.4.2 S-3 Site

Operational information and previous sampling at the S-3 Site indicate that soil and groundwater are contaminated with metals, radiological contaminants, PCBs, and possibly VOCs. The sources of contamination at the S-3 Site are the sludge in the former ponds and the soils surrounding those ponds that were contaminated by reprecipitation or sorption of contaminants migrating from the ponds during their operation. A walk-over radiological survey will be conducted along the main stem of Bear Creek above its confluence with NT-1.

This survey will be used to identify likely locations of seeps in Bear Creek adjacent to the S-3 Site.

Soil boring and sampling will be conducted at the S-3 Site and the areas in the immediate vicinity. Ten soil borings will be drilled at locations downgradient of the S-3 Site (Fig. 7.3) to define the nature and extent of secondary contamination in soils. Samples of soils from two of these locations will be analyzed for geotechnical and geochemical parameters. Six borings will be converted to temporary piezometers. Table 7.3 presents the sample apportionment, and Table 7.4 presents a detailed summary of the number of samples and quality control samples to be collected.

7.4.2.1 S-3 Site soils

Soil borings will be sampled to refusal at or near the residuum/bedrock interface. On average this is anticipated to be 6.1 m (20 ft) below grade. Borings will be sampled using 0.6-m (2-ft) split-spoon barrels. Soil borings furthest away from the S-3 Site will be sampled first because they are likely to be the least contaminated borings. All samples will be surveyed in the field with an OVA and scanned for beta and gamma radiation with a GM pancake probe. The OVA reading will be taken immediately after opening the split spoon and will be followed by the GM pancake.

Soil samples will be analyzed at an off-site laboratory for metals, radiologicals (Uranium-234,-235,-238, Thorium-228,-230,-232, Plutonium-238,-239,-240, Americium-241, and Neptunium-237, gamma scan, Strontium-90, Technetium-99, gross alpha, and gross beta), volatiles (VOCs), and PCBs. Table 7.3 presents an estimate of the number of samples to be analyzed per media at the S-3 Site. The sample estimate presented in Table 7.3 is based on ten vertical borings sampled to 6.1 m (20 ft) with continuous sampling at 1.2-m (4-ft) intervals. VOC samples will be collected from the first 0.6 m (2 ft) for each 1.2-m (4-ft) interval sampled for chemical and radiological analysis. The remaining parameters will be composited over the entire 1.2-m (4-ft) interval.

Two shelly tube samples will be collected at the S-3 Site. The samples will be analyzed for the geotechnical and hydrological parameters listed in Table 6.2. The exact location and sample interval of the shelly tubes will be determined in the field by the project geologist after inspection of soil lithology logs prepared for the first soil boring drilled. When a shelly tube sample is taken, chemical and radiological parameters will be analyzed for the first 0.6 m (2 ft) of a 1.2-m (4-ft) sample. The shelly tube sample will be the second 0.6 m (2 ft) of the 1.2-m (4-ft) sample.

If contaminated soil is found then a leaching study will be performed on soils from two sample locations to determine the likelihood of leaching from contaminated soils to groundwater. These data will be used to determine the extent to which contaminated soils contribute to groundwater contamination at this site. These data will also provide a method by which the effectiveness of the leaching models can be evaluated. Standard TCLP methods are not applicable to the goals of the RI at this site. Soil leaching experiments will be carried out in conjunction with the Environmental Sciences Division at ORNL.

Fig. 7.3. Proposed soil sample

LEGEND:

● 3SOIL BORING LOCATION

▨ RADIOLOGICAL WALKOVER ONLY

EXISTING GROUNDWATER WELLS

- SHALLOW - ACTIVE
- SHALLOW - ABANDONED
- INTERMEDIATE - ACTIVE
- INTERMEDIATE - ABANDONED
- ▲ DEEP - ACTIVE
- * RCRA COMPLIANCE MONITORING WELLS

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BEAR CREEK VALLEY OPERABLE UNIT 1 S-3 SITE

REVISION	DRAWN BY:	CHKD. BY:	DATE:
0	P. HOLM	D. MOSS	2/25/94
1	P. HOLM	D. MOSS	3/7/94

XREFERENCES	PLOT FILES
/93044/XREF/BASEMAP /93044/XREF/MANGAREA	/PLOT/93044/593SMPA

93044	/DWGS/593SMPA	N/A
DRAWING #	CAD FILE #	ROT. ANGLE

and monitoring well locations for the S-3 Site.

Table 7.3. Sample apportionment by medium for the S-3 Site

Medium	Source	Location	No. of samples ^a	Frequency	Analysis/procedure ^b	Rationale
Surface soils	Survey	Upper Bear Creek	NA	Once	Scan surface for total activity	Identify seepage points in the channel of Upper Bear Creek
Soils	Soil borings	S-3 1 through S-3 10	50 ^c	Once	Metals, VOCs, radiological parameters, ^d PCBs	Define the nature and extent of contamination in soils
Soils	Soil borings	S-3 3 and S-3 8	2	Once	Physical, hydrological, and geochemical properties	Obtain information necessary to evaluate the potential for contaminant migration through soils and for screening of remedial technologies
Groundwater	Temporary piezometers	S-3 1, 3, 5, 6, 7, & 8	NA	Continuous monitoring	Water table elevation	Determine the elevation of the water table close to the former ponds during varying hydraulic conditions

^aQuality assurance samples are not included.

^bAnalytical methods are discussed in the Quality Assurance Project Plan (Chap. 8 of this RI work plan).

^cThis number assumes samples will be collected at every 4-ft interval to a depth of 20 ft, from ten soil borings.

^dRadiological parameters include gross α , gross β , gamma spectroscopy, ⁹⁰Sr, ⁹⁹Tc, ²²⁸Th, ²³⁰Th, ²³²Th, ²³⁴U, ²³⁵U, ²³⁸U, ²³⁷Np, ²³⁸Pu, ²³⁹Pu, ²⁴⁰Pu, and ²⁴¹Am.

Table 7.4. Number of samples and quality control samples to be collected for the S-3 Site.

Parameter	Number of sites	Sampling frequency	Number of samples	Field duplicate ^a	Trip blanks ^b	Field blanks ^c	Equipment rinsates ^d	Total samples
<i>Soil borings and soil cores</i>								
Metals	10	1	50	3		4	3	60
Radiological parameters ^e	10	1	50	3		4	3	60
Volatile organics	10	1	50	3	10	4	3	70
PCBs	10	1	50	3		4	3	60
Physical and hydrological ^f	2	1	2	1				3

^aField duplicates will be collected at a frequency of 5% of the samples collected per matrix per event (see Sect. 8.7.1.4).

^bOne trip blank is to accompany each cooler containing VOC samples (see Sect. 8.7.1.1).

^cAt a minimum, one sample per event for decontamination source water, and one per lot for ASTM water. At least one set of field blanks will be collected for every three borings completed (see Sect. 8.7.1.3).

^dEquipment rinsates will be collected at a rate of 5% per event (see Sect. 8.7.1.2).

^eRadiological parameters include gross α , gross β , gamma spectroscopy, ^{90}Sr , ^{99}Tc , ^{228}Th , ^{230}Th , ^{232}Th , ^{234}U , ^{235}U , ^{238}U , ^{237}Np , ^{238}Pu , ^{239}Pu , ^{240}Pu , and ^{241}Am .

^fSee Table 6.2 for physical, hydrological, and geochemical parameters.

7.4.2.2 S-3 Site groundwater

Six soil borings (S-3 1, 3, 5, 6, 7, and 8) will be converted to temporary piezometers, and water levels in these wells and existing wells GW-101, GW-127, and GW-276 will be monitored continuously by a data logger over a 2-month period. Continuously monitoring groundwater levels will help to identify wells that are most likely to be connected to the active groundwater flow system in the headwaters of BCHR.

7.4.3 Oil Landfarm

Operational information and previous sampling at the OLF indicate that soil and groundwater are contaminated with metals, PCBs, and VOCs. The source of contamination at the OLF is contaminated soils below the cap and possibly soils surrounding the cap that were contaminated by reprecipitation or sorption of contaminants or residual LNAPL and DNAPL migrating away from the site. Soil boring and sampling will be conducted in areas around the OLF cap. No borings will be made through the cap. Five soil borings will be made outside the OLF cap, four downgradient and one along strike (Fig. 7.4). Soil samples from two boring locations will be used to determine geotechnical data for the FS (Table 6.2). Soil borings will be sampled to refusal at or near the residuum/bedrock interface. Table 7.5 presents the sample apportionment, and Table 7.6 presents a detailed summary of the number of samples and quality control samples to be collected at OLF and SL 1.

7.4.3.1 Oil Landfarm soils

Soil borings will be sampled to refusal at or near the residuum/bedrock interface. On average this is anticipated to be 6.1 m (20 ft) below grade. The borings will be sampled at 1.2-m (4-ft) intervals using 2-ft split spoons. All samples will be surveyed in the field with an OVA meter and scanned for beta and gamma radiation with a GM pancake probe. The OVA reading will be taken immediately after opening the split spoon followed by the GM pancake. The readings from these meters will be recorded for every 0.6-m (2-ft) interval on the soil lithology log.

Soil samples will be analyzed at an off-site laboratory for metals, radiologicals (Uranium-234,-235,-238, Thorium-228,-230,-232, gamma scan, gross alpha, and gross beta), volatiles (VOCs), and PCBs. The results of these borings will be used to determine the nature and extent of subsurface soil contamination that has resulted from contaminant migration away from the former Landfarm. VOC samples will be collected from the first 0.6 m (2 ft) for each 1.2-m (4-ft) interval sampled for chemical and radiological analysis. The remaining parameters will be composited over the entire 1.2-m (4-ft) interval. Table 7.6 presents an estimate of the number of samples to be analyzed per media at the OLF. The sample estimate presented in Table 7.6 is based on five borings advanced to 6.1 m (20 ft) with continuous sampling at 1.2-m (4-ft) intervals. Groundwater is likely to be between 1.5 and 3.4 m (5 and 11 ft) below ground surface. Two Shelby tube samples will be collected from soils surrounding the OLF cap and will be analyzed for geotechnical and hydrological parameters described in Table 6.2. The exact location and sample interval of the Shelby tubes will be determined in the field by the project geologist after inspection of soil lithology logs prepared for the first soil boring drilled.

ARD/BURNYARD

LEGEND:

△ RADIOLOGICAL & GEOPHYSICAL SURVEY GRID

● 3 SOIL BORING LOCATION

▼ SURFACE SOIL SAMPLE

EXISTING GROUNDWATER WELLS

■ SHALLOW - ACTIVE

□ SHALLOW - ABANDONED

● INTERMEDIATE - ACTIVE

○ INTERMEDIATE - ABANDONED

▲ DEEP - ACTIVE

* RCRA COMPLIANCE MONITORING WELLS

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OIL LANDFARM SANITARY LANDFILL 1 BONEYARD/BURNYARD

REVISION	DRAWN BY:	CHKD. BY:	DATE:
0	P. HOLM	D. MOSS	2/25/94
1	P. HOLM	D. MOSS	3/7/94

XREFERENCES	PLOT FILES
/93044/XREF/BASEMAP /93044/XREF/MANGAREA	/PLOT/5933SMPB

93044	/DWGS/5933SMPB	N/A
DRAWING #	CAD FILE #	ROT. ANGLE

ter, and stream sediment sample locations for the Oil Landfarm.

Table 7.5. Sample apportionment by medium for the Oil Landfarm and Sanitary Landfill 1

Medium	Source	Location	No. of samples ^a	Frequency	Analysis/procedure ^b	Rationale
Surface soils	Survey	Adjacent to the SL 1 cap	NA	Once	Scan surface for total activity	Identify seepage points for leachate from SL 1
Surface soils	Hand auger	Adjacent to the SL 1 cap. See map	11 ^c	Once	Metals and PCBs by field screening methods	Define the nature and extent of contamination in surface soils
Surface soils	Hand Auger	TBD	2	Once	Metals, VOCs, radiological parameters ^d , PCBs, and pesticides	Confirm results of field analysis of soils. Provide high quality data for the risk assessment
Soils	Soil borings	OLF 1 through 5 and SL 1 1 and 2	35 ^e	Once	Metals, VOCs, radiological parameters ^d , PCBs, and pesticides. SVOCs in SL 1 1 and 2	Define the nature and extent of contamination in soils adjacent to the caps
Soils	Soil borings	OLF-3, and OLF-4	2	Once	Physical, hydrological, and geochemical properties	Obtain information necessary to evaluate the potential for contaminant migration through soils and for screening of remedial technologies
Groundwater	Soil Borings—direct push grab sample	SL 1 1 and 2	2	Once	Metals, VOCs, radiological parameters ^f , PCBs and pesticides, major ions, and field parameters. Sample by direct push methods.	Assess the contribution to the integrator unit from contaminated soils
Habitat	Survey	Total area	NA	Once	Habitat survey	Evaluate vegetation types and distribution of habitats prior to biota sampling in phase 2

^aQuality assurance samples are not included.

^bAnalytical methods are discussed in the Quality Assurance Project Plan (Chap. 8 of this RI work plan).

^cThis number assumes one surface sample per site at 11 locations with a 250-ft sample separation.

^dRadiological parameters include gross α , gross β , gamma spectroscopy, ²²⁸Th, ²³⁰Th, ²³²Th, ²³⁴U, ²³⁵U, and ²³⁸U

^eThis number assumes seven sample borings will be sampled continuously with 4-ft intervals over 20 ft.

^fRadiological parameters include gross α , gross β , gamma spectroscopy, ²²⁸Th, ²³⁰Th, ²³²Th, ²³⁴U, ²³⁵U, ²³⁸U, ⁹⁹Tc, and ⁹⁰Sr

Table 7.6. Number of samples and quality control samples to be collected for the Oil Landfarm and Sanitary Landfill 1

Parameter	Number of sites	Sampling frequency	Number of samples	Field duplicate ^a	Trip blanks ^b	Field blanks ^c	Equipment rinsates ^d	Total samples
<i>Surface soils</i>								
Metals ^e	11	1	11	1		2	1	15
PCBs ^e	11	1	11	1		2	1	15
Metals	2	1	2	1		2	1	6
Radiochemical parameters ^f	2	1	2	1		2	1	6
Volatile organics	2	1	2	1	1	2	1	7
Semivolatiles organics	2	1	2	1		2	1	6
PCBs and Pesticides	2	1	2	1		2	1	6
<i>Soil borings</i>								
Metals	7	1	35	2		4	2	43
Radiochemical parameters ^f	7	1	35	2		4	2	43
Volatile organics	7	1	35	2	7	4	2	50
Semivolatiles organics	2	1	10	1		4	2	17
PCBs and Pesticides	7	1	35	2		4	2	43
Physical and hydrological parameters ^g	2	1	2	1				3
<i>Groundwater</i>								
Metals ^h	2	1	2	1		2	1	6
Anions ⁱ	2	1	2	1		2	1	6
Radiochemical parameters ^j	2	1	2	1		2	1	6
Volatile organics	2	1	2	1	1	2	1	7

Table 7.6 (continued)

Parameter	Number of sites	Sampling frequency	Number of samples	Field duplicate ^a	Trip blanks ^b	Field blanks ^c	Equipment rinsates ^d	Total samples
PCBs and pesticides	2	1	2	1		2	1	6
Miscellaneous ^k	2	1	2	1		2	1	6

^aField duplicates will be collected at a frequency of 5% of the samples collected per matrix per event (see Sect. 8.7.1.4).

^bOne trip blank is to accompany each cooler containing VOC samples (see Sect. 8.7.1.1).

^cAt a minimum, one sample per event for decontamination source water, and one per lot for ASTM water. At least one set of field blanks will be collected for every three borings completed (see Sect. 8.7.1.3).

^dEquipment rinsates will be collected at a rate of 5% per event (see Sect. 8.7.1.2).

^eBy field screening methods

^fRadiological parameters include gross α , gross β , gamma spectroscopy, ^{228}Th , ^{230}Th , ^{232}Th , ^{234}U , ^{235}U , and ^{238}U

^gSee Table 6.2 for parameters under physical, hydrological, and geochemical.

^hUnfiltered samples only will be collected.

ⁱAnion analyses include sulfate, fluoride, nitrate, chloride, and cyanide.

^jRadiological parameters include gross α , gross β , gamma spectroscopy, ^{228}Th , ^{230}Th , ^{232}Th , ^{234}U , ^{235}U , ^{238}U , ^{99}Tc , and ^{90}Sr

^kMiscellaneous analyses include alkalinity, total suspended solids, total dissolved solids, pH, specific conductance, and turbidity.

7.4.3.2 Oil Landfarm biota

Biota sampling for the OLF, BYBY, and SL 1 will be carried out as one single coordinated sampling event. This biota sampling plan is described in Sect. 7.4.5.3.

7.4.4 Sanitary Landfill 1

There is almost no information for evaluating the extent to which SL 1 may contribute contaminants to groundwater or to surrounding soils. SL 1 is located on the Maynardville Limestone so it is difficult to separate contaminants in groundwater that can be identified as being derived for SL 1 as opposed to the S-3 plume and the BYBY area.

Soil boring and sampling will be conducted at SL 1. Eleven surface soil samples will be taken adjacent to the SL 1 cap and two soil borings will be drilled at the locations downgradient of SL 1 (Fig. 7.4) to define the nature and extent of secondary contamination in soils.

7.4.4.1 Sanitary Landfill soils

Surface soil samples will be taken from the top 12 in. of soil at sites adjacent to the SL 1 cap. Soil samples will be taken using a hand auger. Surface soil samples will be analyzed using rapid screening level analysis for metals and PCBs. Metals will be analyzed using ICP with QC equivalent to Level II, and PCBs will be analyzed colorimetrically after solvent extraction. Two soil samples will be analyzed at an off-site laboratory for metals, radiologicals (Uranium-234,-235,-238, Thorium-228,-230,-232, gamma scan, gross alpha, and gross beta), volatiles (VOCs), semivolatiles (BNAs), PCBs, and pesticides.

Soil borings will be sampled to refusal at or near the residuum/bedrock interface. On average this is anticipated to be 6.1 m (20 ft) below grade. The borings will be sampled at 0.6-m (4-ft) intervals using 2-ft split spoons. All samples will be surveyed in the field with an OVA meter and scanned for beta and gamma radiation with a GM pancake probe. The OVA reading will be taken immediately after opening the split spoon followed by the GM pancake. The readings from these meters will be recorded for every 0.6-m (2-ft) interval on the soil lithology log.

Soil boring samples will be analyzed at an off-site laboratory for metals, radiologicals (Uranium-234,-235,-238, Thorium-228,-230,-232, gamma scan, gross alpha, and gross beta), volatiles (VOCs), semivolatiles (BNAs), PCBs, and pesticides. VOC samples will be collected from the first 0.6 m (2 ft) for each 1.2-m (4-ft) interval sampled for chemical and radiological analysis. The remaining parameters will be composited over the entire 1.2-m (4-ft) interval. Table 7.6 presents an estimate of the number of samples to be analyzed per media at the SL 1 and OLF. The sample estimate presented in Table 7.6 is based on two borings advanced to 6.1 m (20 ft) with continuous sampling at 1.2-m (4-ft) intervals. Groundwater is likely to be between 1.5 and 3.4 m (5 and 11 ft) below ground surface.

7.4.4.2 Sanitary Landfill groundwater

Groundwater will be sampled from each soil boring by one of two methods:

- using direct push methods at the time of sampling the soil, or
- using a temporary monitoring well that will be removed after the sample is taken.

These samples will be a one time grab sample for groundwater. Where the recharge to the temporary screen is low, samples may be taken over an extended period of time using very low flow rates. Groundwater will be analyzed for metals, radiologicals (Uranium-234,-235,-238, Thorium-228,-230,-232, gamma scan, Strontium-90, Technetium-99, gross alpha, and gross beta.), volatiles (VOCs), semivolatiles (BNAs), PCBs, pesticides, and miscellaneous parameters identified in Table 7.6. Field measurements will include water depth, temperature, pH, specific conductance, and dissolved oxygen.

7.4.4.3 Sanitary Landfill biota

Biota sampling for the OLF, BYBY, and SL 1 will be carried out as one single coordinated sampling event. This biota sampling plan is described in section 7.4.5.3.

7.4.5 Boneyard/Burnyard and HCDA

Operational information from this site is limited, and soils at the BYBY and HCDA may be contaminated with metals, radionuclides, PCBs, dioxin, furan, and VOCs. The sources of contamination at the BYBY are the waste materials in trenches in both the Boneyard and the Burnyard, residues from burning wastes in those trenches, and contaminated soils resulting from operations at the HCDA. The site action plan describes activities at the BYBY to be carried out to complete Stage 1 of the RI/FS process and the first phase of sampling to be carried out in Stage 2. Stage 2 activities will be conducted in two phases of soil sampling because of the large size of this site and the uncertainty concerning contaminant locations.

A walk-over observational and radiological survey will be performed, followed by a walk-over soil electromagnetic induction (EM) survey and a magnetometry survey. The purpose of the observational survey is to identify unearthed items of waste that will need to be removed prior to RI work at the site. The radiological and geophysical surveys will be carried out to identify hot spots to target subsequent boring activities.

Surface soil sampling, soil boring sampling, and groundwater sampling will be conducted at the BYBY. Locations of surface soil samples will be based on a grid with 76.2-m (250-ft) sample spacing (Fig. 7.4). This grid will be refined, based on data from the geophysical surveys, to concentrate on areas with high radiologic emissions and/or with high conductivity. Locations of soil borings will be based on geophysical anomalies and are designed to sample soil in or adjacent to former waste trenches.

7.4.5.1 Screening surveys

Surveys will be carried out over a grid designed to cover the known boundaries of the BYBY and to incorporate possible sites of waste materials that are located outside the current boundaries. The UltraSonic Ranging and Data System (USRADS), developed by

ORNL, will be used to obtain quadrature equipment readings simultaneously for each survey. The radiological survey will be carried out for beta and gamma radiation using a Geiger-Mueller type counter. The EM survey will use an EM31 (Nyquist and Blair 1991); the magnetometry survey will use a Proton-precession type magnetometer. Geophysical surveys will be carried out on lines in a north-south direction. The whole area will then be repeated in an east-west direction. Rapid evaluation of high conductivity areas will allow closer spacing surveys of selected sites to be carried out in order to accurately locate soil borings. The number of lines and the line spacing for the surveys will be determined after an initial walk-over by the project geologist and geophysicist. Line spacing will be close enough to delineate variations in soil conductivity and magnetism.

7.4.5.2 Boneyard/Burnyard soils

Surface soil samples will be taken from the top 12 in. of soil at 26 sites over BYBY, located on a grid with 76.2-m (250-ft) spacing. Surface soil samples will be taken using a hand auger. Surface soil samples will be analyzed using rapid screening level analysis for metals and PCBs. Metals will be analyzed using ICP with QC equivalent to Level II, and PCBs will be analyzed colorimetrically after solvent extraction. Five surface soil samples will be analyzed at an off-site laboratory for metals, radiologicals (Uranium-234,-235,-238, Thorium-228,-230,-232, gamma scan, gross alpha, and gross beta), volatiles (VOCs), semivolatiles (BNAs), PCBs, and pesticides.

Soil borings will be located using the following two methods:

- borings located downgradient of the HCDA cap, and
- borings located at geophysical anomalies or radiological "hot spots" (i.e., regions of the survey grid with abnormally high radiological emissions).

Three soil borings will be located downgradient of the HCDA, and five located at geophysical anomalies. Soil boring samples will be composited over 4-ft intervals to refusal at or near the base of the residuum layer. Soil boring samples will be analyzed at an off-site laboratory for metals, radiologicals (Uranium-234,-235,-238, Thorium-228,-230,-232, gamma scan, gross alpha, and gross beta), volatiles (VOCs), semivolatiles (BNAs), PCBs, and pesticides. The three borings down gradient of the HCDA will also have soil samples analyzed for Dioxin 2,3,7,8-TCDD.

Soil borings downgradient from the HCDA cap will be used to determine the nature and extent of contamination in soils that may have resulted from liquid waste handling operations at this site. The five borings at geophysical anomalies will be used to address the nature of contamination associated with former trenches. These borings will be drilled as close as possible to the location of the trenches as indicated by the geophysical anomalies while avoiding drilling through metal objects. All sampling locations, with the exception of those that will be based on geophysical anomalies, are shown on Fig. 7.4.

Second-phase soil sampling may be made over a smaller grid spacing over "hot spot" areas. The location and spacing of borings on finer grids will be determined after the screening surveys. Soil samples will be selected at random from nodes on these finer grids. The number of samples to be taken from each grid will be such that the concentrations of COCs are defined within acceptable bounds at the 95% confidence level. Soils sampled in this

phase will be analyzed for COCs only. COCs at each "hot spot" will be determined from phase 1 sampling.

In phase 2, priority will be given to areas with identified soil contamination. For areas at which soil contamination was not indicated—though geophysical and radiological surveys suggested the presence of waste materials—the prioritization of "hot spots" will be as follows:

- areas with both high conductivity and high radiological emissions.
- areas with high radiological emissions without identifiable high conductivity readings.
- areas with high conductivity without significant elevated radiological emissions.

Soil borings will be sampled to refusal at or near the residuum/bedrock interface. On average this is anticipated to be 6.1 m (20 ft) below grade. The borings will be sampled at 0.6-m (4-ft) intervals using split spoons. All samples will be surveyed in the field with an OVA meter and scanned for beta and gamma radiation with a GM pancake probe. The OVA reading will be taken immediately after opening the split spoon followed by the GM pancake. The readings from these meters will be recorded for every 0.6-m (2-ft) interval on the soil lithology log. Table 7.7 presents the sample apportionment, and Table 7.8 presents a detailed summary of the number of samples and quality control samples to be collected.

Two shelly tube samples will be collected from soils at the BYBY. The samples will be analyzed for the geotechnical and hydrological parameters described in Table 6.2. The exact locations and sample intervals of the shelly tubes will be determined in the field by the project geologist after inspection of soil lithology logs prepared for the first soil boring drilled.

7.4.5.3 Boneyard/Burnyard groundwater

Groundwater will be sampled from each soil boring by one of three methods:

- using direct push methods at the time of sampling the soil,
- using a temporary monitoring well that will be removed after the sample is taken, or
- sampling twice, once at high flow and once at low flow, in existing monitoring wells GW-005, GW-067, and GW-087.

The first two methods will be one time grab samples at each of the eight soil borings. Where the recharge to the temporary screen or well is low, samples will be taken over an extended period of time using very low flow rates. Groundwater will be analyzed for metals, radiologicals (Uranium-234,-235,-238, Thorium-228,-230,-232, gamma scan, Strontium-90, Technetium-99, gross alpha, and gross beta.), volatiles (VOCs), semivolatiles (BNAs), PCBs, pesticides, and miscellaneous parameters identified in Table 7.8. Field measurements will include water depth, temperature, pH, specific conductance, and dissolved oxygen.

7.4.5.4 Boneyard/Burnyard biota

Up to five sampling locations will be chosen at random from contaminated areas identified by the initial radiological survey. Sample locations will be chosen from a maximum of 50 areas showing the highest levels of contamination. Grasses, browse, and earthworms will

Table 7.7. Sample apportionment by medium for the Boneyard/Burnyard

Medium	Source	Location	No. of samples ^a	Frequency	Analysis/procedure ^b	Rationale
Surface soils	Survey	Complete site	NA	Once	Scan for total activity, soil conductivity, and soil magnetics	Identify radiologically contaminated soils, locations of former trenches, and buried metal objects
Surface soils	Hand auger	See map	26 ^c	Once	Metals and PCBs by field screening.	Define the nature and extent of contamination in surface soils
Surface Soils	Hand Auger	TBD	5	Once	Metals, VOCs, SVOCs, radiological parameters ^d , PCBs, and pesticides	Confirm results of field analysis of soils. Provide high quality data for the risk assessment
Soils	Soil borings	BYBY 1 through 3. Five additional boring locations (BYBY 4-8) TBD	40 ^e	Once	Metals, VOCs, SVOCs, radiological parameters, ^d PCBs, and pesticides. Dioxin 2,3,7,8-TCDD at three borings adjacent to the HCDA	Define the nature and extent of contamination in soils
Soils	Soil borings	To be determined	2	Once	Physical, hydrological, and geochemical properties	Obtain information necessary to evaluate the potential for contaminant migration through soils and for screening of remedial technologies
Groundwater	Monitoring wells and direct push grab samples	GW-005, GW-67, and GW-087. BYBY 1, 2 and 3. Five additional boring locations TBD	11	Eight soil borings once and three monitoring wells twice	Metals, VOCs, radiological parameters ^f , PCBs and pesticides, major ions, and field parameters. Eight soil boring samples by direct push.	Assess the contribution to the integrator unit from contaminated soils
Habitat	Survey	Total area	NA	Once	Habitat survey	Evaluate vegetation types and distribution of habitats prior to biota sampling in phase 2

^aQuality assurance samples are not included.

^bAnalytical methods are discussed in the Quality Assurance Project Plan (Chap. 8 of this RI work plan).

^cThis number assumes one surface sample per site at 26 locations with a 250-ft sample separation. No QA samples included in this estimate.

^dRadiological parameters include gross α , gross β , gamma spectroscopy, ⁹⁰Sr, ⁹⁹Tc, ²²⁸Th, ²³⁰Th, ²³²Th, ²³⁴U, ²³⁵U, and ²³⁸U.

^eThis number assumes eight sample borings will be sampled continuously with 4-ft intervals over 20 ft.

^fRadiological parameters include gross α , gross β , gamma spectroscopy, ⁹⁰Sr, ⁹⁹Tc, ²²⁸Th, ²³⁰Th, ²³²Th, ²³⁴U, ²³⁵U, ²³⁸U, ⁹⁰Sr, and ⁹⁹Tc

Table 7.8. Number of samples and quality control samples to be collected for the Boneyard/Burnyard

Parameter	Number of sites	Sampling frequency	Number of samples	Field duplicate ^a	Trip blanks ^b	Field blanks ^c	Equipment rinsates ^d	Total samples
<i>Surface soils</i>								
Metals ^e	26	1	26	2		4	2	34
PCBs ^e	26	1	26	2		4	2	34
Metals	5	1	5	1		2	1	9
Radiological parameters ^f	5	1	5	1		2	1	9
Volatile organics	5	1	5	1	2	2	1	11
Semivolatiles organics	5	1	5	1		2	1	9
PCBs and Pesticides	5	1	5	1		2	1	9
<i>Soil borings</i>								
Metals	8	1	40	2		4	2	48
Radiological parameters ^f	8	1	40	2		4	2	48
Volatile organics	8	1	40	2	8	4	2	56
Semivolatiles organics	8	1	40	2		4	2	48
PCBs and Pesticides	8	1	40	2		4	2	48
2,3,7,8-TCDD	3	1	15	1		2	1	19
Physical and hydrological parameters ^g	2	1	2	1				3
<i>Groundwater</i>								
Metals ^h	11	1	14	1		2	1	18
Anions ⁱ	11	1	14	1		2	1	18
Radiological parameters ^j	11	1	14	1		2	1	18

Table 7.8 (continued)

Parameter	Number of sites	Sampling frequency	Number of samples	Field duplicate ^a	Trip blanks ^b	Field blanks ^c	Equipment rinsates ^d	Total samples
Volatile organics	11	1	14	1	9	2	1	27
PCBs and pesticides	11	1	14	1		2	1	18
Miscellaneous ^k	11	1	14	1		2	1	18

^aField duplicates will be collected at a frequency of 5% of the samples collected per matrix per event (see Sect. 8.7.1.4).

^bOne trip blank is to accompany each cooler containing VOC samples (see Sect. 8.7.1.1).

^cAt a minimum, one sample per event for decontamination source water, and one per lot for ASTM water. At least one set of field blanks will be collected for every three borings completed (see Sect. 8.7.1.3).

^dEquipment rinsates will be collected at a rate of 5% per event (see Sect. 8.7.1.2).

^eBy field screening methods

^fRadiological parameters include gross α , gross β , gamma spectroscopy, ⁹⁰Sr, ⁹⁹Tc, ²²⁸Th, ²³⁰Th, ²³²Th, ²³⁴U, ²³⁵U, and ²³⁸U

^gSee Table 6.2 for parameters under physical, hydrological, and geochemical.

^hUnfiltered samples only will be collected.

ⁱAnion analyses include sulfate, fluoride, nitrate, chloride, and cyanide.

^jRadiological parameters include gross α , gross β , gamma spectroscopy, ⁹⁰Sr, ⁹⁹Tc, ²²⁸Th, ²³⁰Th, ²³²Th, ²³⁴U, ²³⁵U, ²³⁸U, ⁹⁹Tc, and ⁹⁰Sr

^kMiscellaneous analyses include alkalinity, total suspended solids, total dissolved solids, pH, specific conductance, and turbidity.

be sampled at these locations; however, if their numbers are reduced, soil will be sampled for toxicity testing as described in Sect. 6.1.1.4. One field duplicate sample each of vegetation and earthworms will be submitted for analysis. One equipment rinsate will be submitted for each matrix (clippers and bowl for grasses, clippers and bowl for browse, and depuration chambers for earthworms).

Approximately 150 g of vegetation and 400 g of earthworms will be required per sample, except that one sample will be twice as large to allow for preparation of a duplicate sample. Earthworm samples will be divided and 235 g will be depurated before being submitted for analysis. Samples of each matrix will be held at 4° C and will be combined in sample groups with like samples from the OLF and the Burial Grounds. Vegetation and earthworm samples will be analyzed for metals and radiological parameters, and earthworms will also be analyzed for PAHs and PCBs (Table 7.8).

7.4.6 Burial Grounds

Operational information and previous sampling at the Burial Grounds indicate that soil and groundwater are contaminated with metals, radiological parameters, PCBs, and VOCs. The sources of contamination at the Burial Grounds are solid wastes in burial trenches, the residues of liquid wastes that remain in trenches, and soils surrounding those ponds that were contaminated by reprecipitation or sorption of contaminants migrating away from the trenches. Seeps identified in the Burial Grounds will be screened for ionizing radiation. A walk-over radiological survey will be conducted over parts of the Burial Grounds site, excluding the waste areas, to identify areas of soil contamination outside the known waste trenches.

Soil boring, groundwater, stream water, and sediment sampling will be conducted at the Burial Grounds. No direct sampling of the waste materials is proposed, and the nature and extent of waste will be determined using a statistical sampling of Y-12 Plant records (see Sect. 6.1.2.5). No borings through RCRA caps will be proposed. Soil sampling will be carried out in two phases. For phase 1, stratified random soil sampling will be performed outside the waste trenches. Stratification of sample locations will be based on knowledge of the location and the nature of the wastes. For the second phase of sampling, sample grids will be refined, based on data from the first phase, and areas will be selected for closer study. The purpose of phase 2 sampling is to determine the extent of secondary soil contamination within the Burial Grounds site boundaries. Table 7.9 presents the sample apportionment, and Table 7.10 presents a detailed summary of the number of samples and quality control samples to be collected.

For the purpose of selecting sample locations for the Burial Grounds, and in anticipation of a second phase of sampling, the Burial Grounds area has been divided into four sample areas (Fig. 7.5):

- Area 1: BG-A South, BG-A North (except trenches A-16 and A-17), ORP 1, ORP 2, NT-7, and soils to the south of BG-A as far as Bear Creek
- Area 2: Trenches A-16, A-17, and A-18
- Area 3: BG-B, BG-C East and -D East, and Walk-in Pits 1 and 2
- Area 4: BG-C West, BG-D West, -E, and -J and soils to the south of BG-C West as far as the access road to the Burial Grounds

Fig. 7.5. Proposed soil samples and

LEGEND:

- STORM FLOW TUBES
- ▽ SURFACE WATER STORM FLOW GRAB SAMPLE
- ⊕ TEMPORARY PIEZOMETERS
- ▽ SURFACE SOIL OR SEDIMENT SAMPLE
- SOIL BORING LOCATION AND TEMPORARY PIEZOMETER

RADIOLOGICAL WALKOVER ONLY

EXISTING GROUNDWATER WELLS

- SHALLOW - ACTIVE
- SHALLOW - ABANDONED
- INTERMEDIATE - ACTIVE
- INTERMEDIATE - ABANDONED
- ▲ DEEP - ACTIVE
- * RCRA COMPLIANCE MONITORING WELLS

SAIC

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BEAR CREEK VALLEY OPERABLE UNIT 1 BURIAL GROUNDS

REVISION	DRAWN BY:	CHKD. BY:	DATE:
0	P. HOLM	D. MOSS	2/25/94
1	P. HOLM	D. MOSS	3/7/94

XREFERENCES	PLOT FILES
/93044/XREF/BASEMAP	/PLOT/593SMPC.PLT
/93044/XREF/MANGAREA	

93044	/DWG/593SMPC	N/A
DRAWING #	CAD FILE #	ROT. ANGLE

Monitoring well locations for the Boneyard/Burnyard.

Table 7.9. Sample apportionment by medium for the Burial Grounds

Medium	Source	Location	No. of samples ^a	Frequency	Analysis/procedure ^b	Rationale
Surface soils	Survey	Sample areas 3 and 4. See map	NA	Once	Scan for total activity	Identify radiologically contaminated soils
Surface soils	Hand auger	See map	33 ^c	Once	Metals and PCBs by field screening	Define the nature and extent of contamination in surface soils
Surface soils	Hand auger	TBD	5	Once	Metals, VOCs, radiological parameters ^d , PCBs, and pesticides	Confirm results of field analysis of soils. Provide high quality data for the risk assessment
Soils	Soil borings	BG 1 through 15	75 ^e	Once	Metals, VOCs, radiological parameters ^d , PCBs, and pesticides. Only BG-3, BG-4, BG-6, BG-11, and BG-17 for dioxin 2,3,7,8-TCDD	Define the nature and extent of contamination in soils
Soils	Soil borings	BG-2, BG-5, BG-11, BG-15	4	Once	Physical, hydrological, and geochemical properties	Obtain information necessary to evaluate the potential for contaminant migration through soils and for screening of remedial technologies
Surface water	Grab samples	NT-8	4	Twice	Metals, VOCs, radiological parameters ^d , PCBs, pesticides, major ions, and field parameters.	Samples used to determine if surface water draining the upper section of area 3 is contaminated
Groundwater	Soil Borings by direct push grab sample	BG-1 to BG-15	15	Once	Metals, VOCs, radiological parameters ^d , PCBs and pesticides, major ions, and field parameters. Samples taken by direct push	Assess the contribution to the integrator unit from contaminated soils
Stormflow	Stormflow tubes	Ten locations in areas 3 and 4	10	NA	Level measured after storm events	Used to determine if stormflow is active at this site
Habitat	Survey	Total area	NA	Once	Habitat survey	Evaluate vegetation types and distribution of habitats prior to biota sampling in phase

^aQuality assurance samples are not included.

^bAnalytical methods are discussed in the Quality Assurance Project Plan (Chap. 8 of this RI work plan).

^cThis number assumes one surface sample per site at 33 locations with a 250-ft sample separation.

^dRadiological parameters include gross α , gross β , gamma spectroscopy, ⁹⁰Sr, ⁹⁹Tc, ²²⁸Th, ²³⁰Th, ²³²Th, ²³⁴U, ²³⁵U, and ²³⁸U.

^eThis number assumes all 17 sample borings will be sampled continuously with 4-ft intervals to 20 ft.

^fRadiological parameters include gross α , gross β , gamma spectroscopy, ⁹⁰Sr, ⁹⁹Tc, ²²⁸Th, ²³⁰Th, ²³²Th, ²³⁴U, ²³⁵U, ²³⁸U, ⁹⁰Sr, and ⁹⁹Tc

Table 7.10. Number of samples and quality control samples to be collected for the Burial Grounds

Parameter	Number of sites	Sampling frequency	Number of samples	Field duplicate ^a	Trip blanks ^b	Field blanks ^c	Equipment rinsates ^d	Total samples
<i>Surface soils</i>								
Metals ^e	33	1	33	2		4	2	41
PCBs ^e	33	1	33	2		4	2	41
Metals	5	1	5	1		2	1	9
Radiological parameters ^f	5	1	5	1		2	1	9
Volatile organics	5	1	5	1	2	2	1	10
PCBs and Pesticides	5	1	5	1		2	1	9
<i>Soil borings</i>								
Metals	15	1	75	4		6	6	92
Radiological parameters ^f	15	1	75	4		6	6	92
Volatile organics	15	1	75	4	15	6	6	107
PCBs and Pesticides	15	1	75	4		6	6	92
2,3,7,8-TCDD	3	1	15	2		4	2	23
Physical and hydrological parameters ^g	3	1	3	1				4
<i>Groundwater</i>								
Metals ^h	15	1	15	1		2	1	19
Anions ⁱ	15	1	15	1		2	1	19
Radiological parameters ^f	15	1	15	1		2	1	19
Volatile organics	15	1	15	1	7	2	1	26
PCBs and pesticides	15	1	15	1		2	1	19

Table 7.10 (continued)

Parameter	Number of sites	Sampling frequency	Number of samples	Field duplicate ^a	Trip blanks ^b	Field blanks ^c	Equipment rinsates ^d	Total samples
Miscellaneous ^j	15	1	15	1		2	1	19
Surface Water								
Metals ^h	2	2	4	2		4	2	12
Anions ⁱ	2	2	4	2		4	2	12
Radiological parameters ^j	2	2	4	2		4	2	12
Volatile organics	2	2	4	2	2	4	2	14
PCBs and pesticides	2	2	4	2		4	2	12
Miscellaneous ^k	2	2	4	2		4	2	12

^aField duplicates will be collected at a frequency of 5% of the samples collected per matrix per event (see Sect. 8.7.1.4).

^bOne trip blank is to accompany each cooler containing VOC samples (see Sect. 8.7.1.1).

^cAt a minimum, one sample per event for decontamination source water, and one per lot for ASTM water. At least one set of field blanks will be collected for every three borings completed (see Sect. 8.7.1.3).

^dEquipment rinsates will be collected at a rate of 5% per event (see Sect. 8.7.1.2).

^eBy field screening methods

^fRadiological parameters include gross α , gross β , gamma spectroscopy, ^{90}Sr , ^{99}Tc , ^{228}Th , ^{230}Th , ^{232}Th , ^{234}U , ^{235}U , and ^{238}U

^gSee Table 6.2 for parameters under physical, hydrological, and geochemical.

^hFiltered and unfiltered samples will be collected.

ⁱAnion analyses include sulfate, fluoride, nitrate, chloride, and cyanide.

^jRadiological parameters include gross α , gross β , gamma spectroscopy, ^{90}Sr , ^{99}Tc , ^{228}Th , ^{230}Th , ^{232}Th , ^{234}U , ^{235}U , ^{238}U , ^{90}Sr , and ^{99}Tc

^kMiscellaneous analyses include alkalinity, total suspended solids, total dissolved solids, pH, specific conductance, and turbidity.

7.4.6.1 Screening surveys

A walk-over radiological survey will be carried out over a grid designed to cover sample areas 2, 3, and 4 using the USRADS system and a Geiger-Mueller type counter. Rapid evaluation of high radiological areas will allow closer spacing surveys of selected sites to be carried out to accurately locate soil borings for phase 2 soil sampling. USRADS will be used to collect grid quadrature and in-phase readings simultaneously. The number of lines and the line spacing for the surveys will be determined after an initial walk-over by the project geologist and geophysicist. Line spacing will be close enough to delineate variations in soil conductivity and magnetism.

7.4.6.2 Burial Grounds soils

Surface soil samples will be taken from the top 12 in. of soil at 33 sites in samples areas 3 and 4, located on a grid with 76.2-m (250-ft) spacing. Surface soil samples will be taken using a hand auger. Surface soil samples will be analyzed using rapid screening level analysis for metals and PCBs. Metals will be analyzed using ICP with QC equivalent to Level II, and PCBs will be analyzed colorimetrically after solvent extraction. Five surface soil samples will be analyzed at an off-site laboratory for metals, radiologicals (Uranium-234,-235,-238, Thorium-228,-230,-232, gamma scan, gross alpha, and gross beta), volatiles (VOCs), semivolatiles (BNAs), PCBs, and pesticides.

Phase 1 borings in each area are located downgradient of waste trenches; the locations of the borings are shown on Fig. 7.5. Fifteen soil borings were proposed for phase 1 sampling. The purpose of phase 1 sampling is to determine the nature of secondary soil contamination that has resulted from migration of contaminants downgradient from waste trenches. Soil samples will be analyzed at an off-site laboratory for metals, radiologicals (Uranium-234,-235,-238, Thorium-228,-230,-232, gamma scan, gross alpha, and gross beta), volatiles (VOCs), PCBs, and pesticides. Borings BG-3, 4, 6, 11, and 17 that are down gradient from suspected burn areas also have soil samples analyzed for Dioxin 2,3,7,8-TCDD.

Second-phase soil borings will be made within a smaller grid spacing over "hot spot" areas identified by the radiological survey and over regions where downgradient soil contamination was indicated by phase 1 sampling. The location, number, and spacing of borings on finer grids will be determined after the screening surveys. Soil will be sampled at nodes from these finer grids selected at random. The number of samples to be taken from each grid will be such that the concentrations of COCs are defined within acceptable bounds at the 95% confidence level. COCs at each "hot spot" will be determined from phase 1 sampling. In phase 2, priority will be given to areas with identified soil contamination.

Soil borings will be sampled to refusal at or near the residuum/bedrock interface. On average this is anticipated to be 6.1 m (20 ft) below grade. The borings will be sampled at 1.2-m (4-ft) intervals using 2-ft split spoons. All samples will be surveyed in the field with an OVA and scanned for beta and gamma radiation with a GM pancake probe. The OVA reading will be taken immediately after opening the split spoon followed by the GM pancake. The readings from these meters will be recorded for every 0.6-m (2-ft) interval on the soil lithology log.

Three shelly tube samples will be collected from soils at the Burial Grounds. The samples will be analyzed for geotechnical and hydrological parameters described in Table 6.2.

The exact location and sample interval of the shelly tubes will be determined in the field by the project geologist after inspection of soil lithology logs prepared for the first soil boring drilled. When a shelly tube sample is taken, chemical and radiological parameters will be analyzed for the first 0.6 m (2 ft) of a 1.2-m (4-ft) sample interval. The shelly tube sample will be the second 0.6 m (2 ft) of the 1.2-m (4-ft) sample interval.

7.4.6.3 Burial Grounds groundwater

Groundwater will be sampled from each soil boring by one of two methods:

- using direct push methods at the time of sampling the soil, or
- using a temporary monitoring well that will be removed after the sample is taken.

These samples will be a one time grab sample for groundwater from each of the fifteen soil borings. Where the recharge to the temporary screen is low, samples may be taken over an extended period of time using very low flow rates. Groundwater will be analyzed for metals, radiologicals (Uranium-234,-235,-238, Thorium-228,-230,-232, gamma scan, Strontium-90, Technetium-99, gross alpha, and gross beta.), volatiles (VOCs), semivolatiles (BNAs), PCBs, pesticides, and miscellaneous parameters identified in Table 7.8. Field measurements will include water depth, temperature, pH, specific conductance, and dissolved oxygen.

7.4.6.4 Burial Grounds stormflow

It is possible that some of the upgradient trenches in the Burial Grounds will be affected by stormflow. Pine Ridge may act as a recharge area for the stormflow system, and there are no apparent barriers between the recharge areas and the up-slope trenches except some of the grading that has been carried out. Wastes that are likely to be affected are in BG-D, BG-E, and BG-J. Based on the conceptual model for stormflow (Solomon et al. 1992), these trenches probably intercept most of the stormflow from Pine Ridge.

Stormflow tubes will be installed at ten locations (Fig. 7.6) upgradient and downgradient of BG-D, E, and J. Construction and installation of the stormflow tubes will follow procedures outlined by Moore (1992). An example of a stormflow tube construction is shown in Fig. 7.2. Water levels at each location will be measured after each storm event.

7.4.6.5 Burial Grounds surface water

Surface water will be sampled at two locations in the upper reaches of NT-8. One location is in the tributary directly north of the C-west cap, the second is located in the drainage channel between C-west and C-east areas. Each site will be sampled twice, once at high flow and once at low flow. Samples will be grab samples and will be analyzed for the same analyte list as groundwater samples discussed in Sect. 7.4.6.3.

7.4.6.6 Burial Grounds biota

Up to five sampling locations will be chosen at random from contaminated areas identified by the initial radiological survey. Sample locations will be chosen from a maximum of 50 areas showing the highest levels of contamination. Grasses, browse, and earthworms will be sampled at these locations; however, if their numbers are reduced, soil will be sampled for

Fig. 7.6. Proposed soil samples for soil sample area 1.

toxicity testing as described in Sect. 6.1.1.4. One field duplicate sample each of vegetation and earthworms will be submitted for analysis. One equipment rinsate will be submitted for each matrix (clippers and bowl for grasses, clippers and bowl for browse, and depuration chambers for earthworms).

Approximately 150 g of vegetation and 400 g of earthworms will be required per sample, except that one sample of earthworms will be three times as large to allow for preparation of matrix spike and matrix spike duplicate samples. Earthworm samples will be divided, and approximately 235 g (or 700 oz) will be depurated before being submitted for analysis. Samples of each matrix will be held at 4 degrees C and will be combined in sample groups with like samples from the OLF and the BYBY. Vegetation and earthworm samples will be analyzed for metals and radiological parameters, and earthworms will also be analyzed for PAHs and PCBs (Table 7.10).

7.5 SAMPLE TRACKING AND RECORDS MANAGEMENT

7.5.1 Objectives

The primary objectives of an effective sample tracking and records management program are to provide timely feedback regarding laboratory performance and project status and to ensure that data are complete and auditable. Meeting this objective is critical to the overall success of this RI because (1) the data are important in determining potential environmental and human health effects, (2) the data must be technically defensible and legally admissible, and (3) data collection and analysis costs are high. Existing records management plans, information management procedures, and database management systems will be used to track, document, and manage all field and analytical data.

7.5.2 Roles and Responsibilities

Data Coordinator. Coordinates data support for the technical team. Prepares work agreements as necessary, reviews sample-receipt acknowledgments, and receives laboratory data packages. Participates in QA/QC activities including nonconformance investigations and reporting and QA reviews.

Data Entry Clerk. Enters new or existing data generated by field activities or laboratory analyses. Performs data-entry verification.

Data Manager. Identifies and acquires all relevant data, including historical data. Ensures data completeness and consistency and resolves problems associated with data.

Field Team Leader. Helps prepare the sampling and analysis plan (SAP) and implements it in the field; assigns staff members to sampling teams and assigns sampling team members. Prepares for and coordinates sampling activities. Oversees field data collection, recording, and documentation activities. Supervises packing and shipping of samples.

Records Manager. Controls the recorded information about project activities that is generated and/or received by project organizations.

7.5.3 Presampling Database

The SAP specifies all sampling and analysis data to be collected. A temporary presampling database with structures identical to the central OREIS data structures will be created by the Data Manager and populated with the project SAP data (i.e., sample locations, sample identification codes, sample media, sample depths, types of analyses, field QC data, and chain-of-custody details). Before mobilizing the field team, the Field Team Leader and Data Manager will verify data entered into the presampling database. The verified presampling database will be used to print sample labels, sample tags, and field logbooks for each sampling event.

7.5.3.1 Sample labels and tags

Preprinted sample labels and tags will contain the following information:

- sample ID number;
- project number;
- sampling location, station code, and description;
- sample media;
- sample type;
- analysis;
- preservative;
- volume;
- destination laboratory.

Field staff members will complete the remaining information during sample collection:

- radiological screening information,
- date and time of sample collection,
- comments,
- name of collector.

If preprinted blank sample labels are not used, all information will be recorded manually during field sample collection.

7.5.3.2 Field logbooks

Each sampling team will have a field logbook that will contain, at a minimum, a table of contents, task team activity log sheets, and sample log sheets. The unbound logbooks will be reviewed for accuracy and completeness by the Field Team Leader. After making any needed corrections, the Data Coordinator will have the logbooks bound and numbered. All bound logbooks will be kept in secure storage and access will be controlled at all times.

Logbooks will be completed by the field sampling team during sampling according to relevant procedures. Entries will include sampling date and time, details on field conditions, field instrument measurements, deviations from the sampling plan, QA concerns, and any noteworthy observations. The field team will also complete chain-of-custody information for each sample collected in accordance with Energy Systems chain-of-custody procedure ESP-500 (Kimbrough et al. 1991).

7.5.4 Site Logbook

The site logbook is the master field investigation document. It is a bound book with a hard cover and sequentially numbered pages. Its primary purpose is to document each day's field activities; the personnel on each sampling team; and any administrative occurrences, conditions, or activities that may have affected the fieldwork or data quality of any environmental samples for any given day. The site logbook will be kept by the Field Team Leader. The front of the logbook will be used to list the location, the names of the companies performing the field sampling, the client, and the contract number of the work being performed. The site logbook will be updated daily. Following his/her review, the Field Team Leader will sign the bottom of the last page of the day.

The site logbook will, at a minimum, contain the following information (Fig. 7.6):

- day, date, time arrived on site, temperature, weather conditions, and names and titles of personnel present on site;
- names, titles, and organizations of any visitors who entered the site during the day;
- arrival time of any subcontractors on site, number of duplicate samples, number of split samples, laboratory to which split samples were sent, references to field logbook(s) or field forms that contain more specific field information for any of the above tasks;
- decontamination iterations, equipment decontaminated, and procedures used;
- specific comments relative to any problem areas that occurred during the day's activities, the final resolution, and the anticipated impact on the outcome of the field investigation or on data quality;
- a record of telephone calls (incoming or outgoing) pertaining directly to the decision-making process of the field investigation, along with the outcome of each conversation, and a reference to the telephone log to obtain more specific information on the call.

7.5.5 Corrections to Documents

All corrections to field logbooks and chain-of-custody forms will be made by drawing a single line through the error so as not to obliterate the correction, entering the correct information, and initialing and dating the change. If possible, the need for administering the change will be documented.

7.5.6 Sample Identification Numbers

To support the data management requirements specified in the CLP SOWs, sample identification will be limited to a length of six characters and/or numbers. While this may appear as a limitation, the data management system will provide additional data fields to

describe specific sample identification information. Adhering to the CLP SOW requirements will facilitate the overall data reporting and ensure the integrity of project-derived data. Sample identification numbers will be coded to enable identification of the location, sample pattern, media, and depth.

Position 1: Sample Location. A single character or digit will be used to identify the sampling location within the specific study area. This position of the sample identification will be keyed to descriptive text specifically identifying the location.

Positions 2–4: Sample Station/Borehole/Well identification. Three digits will be used to identify the specific sampling station within the above sampling location (i.e., borehole number, well identification number). This position of the sample identification will be keyed to descriptive text or ER-assigned station identification numbers. This field will identify the station.

Position 5: Sample Type and Media. A single digit will be used to identify the sample type and medium being sampled. This position will account for duplicate samples and will be keyed to descriptive text identifying the sample medium.

Position 6: Sample Depth. A single digit will be used to designate sample depth. This position of the sample identification will be keyed to descriptive text specifically identifying the depth in feet.

Trip Blanks, Equipment Rinsates, and Field Blanks. The first three positions of the sample identification will be used to identify trip blanks (TBL), rinsates (EQR), and field blanks (FBL). A sequential number in positions 4 through 6 will be assigned to each QC sample. Trip blank and field blank designations will be referenced to specific sampling dates and other pertinent sample information. Equipment rinsate designations will likewise be referenced to specific sampling equipment types, dates, and other pertinent information.

7.5.7 Data Entry of Field Information

The Field Team Leader will submit the field sampling documentation to the Data Coordinator on the schedule specified in the RI work plan. After verifying the accuracy and completeness of this information, the Data Coordinator will ensure that the database is updated accordingly. Data entry will be performed following established procedures. The field data captured include field sample measures, well construction logs, lithologic descriptions, and field screening results.

After the data are acquired and verified, the Data Manager and Coordinator will deliver the complete hard copy set of field sampling documentation to the Records Management Team, which will index, process, and place the documents in the project file. Upon completion of the field effort for each media, the Field Team Leader will turn over field documentation to the Document Coordinator. The Document Coordinator will lock the documentation in a fireproof safe, and copies of the material will be supplied to the Data Coordinator for entry into the project database. Upon conclusion of the field effort, all field documentation will be clearly labeled and placed in the project file.

7.5.8 Processing of Information on Samples Sent to and Received from Analytical Laboratories

The Field Team Leader will make available to the Data Coordinator copies of the chain-of-custody forms for samples sent to the laboratory. The laboratory also will return a Letter of Receipt indicating the sample identification, laboratory identification, and analyses scheduled for each sample along with the original or copy of the chain-of-custody form. The letter will be reviewed by the Data Coordinator and compared to the contents of the database to ensure that scheduled analyses concur with the requirements of the SAP and that expected turnaround will meet holding-time requirements. Any deviations from established procedures will be documented as Nonconformance Reports. The Data Coordinator will update the database with Letter of Receipt information and generate a status report. Hard copies of the chain-of-custody forms, Letter of Receipts, and periodic sample status reports will be forwarded to the Records Management Team for inclusion in the project files.

7.5.9 Receipt and Preliminary Processing of Analytical Data Packages and Diskettes

The Data Coordinator will receive and process the data package by date-stamping it, assigning it an accession number, photocopying it, and submitting the original to the project file. A copy will be retained for use in verification, validation, and data entry. The Data Coordinator will initiate tracking of the analytical data by logging the sample delivery group information into the database, linking sample identification numbers with the sample delivery group number.

The unvalidated data will then be transferred to the database for preliminary analysis. If an electronic deliverable exists, data will be stripped and added to the database using verified programs. If an error report is generated after loading, the Data Manager will resolve any discrepancies. If appropriate, Nonconformance Reports will be generated. When subsequent deliveries of corrected data are submitted electronically, printouts of the changes will be reviewed to ensure that any Noncompliance Reports are addressed. When Noncompliance Report resolution is completed, the data will be corrected in the database.

If results were not submitted on diskette, analytical results and laboratory information will be added to the database by the Data Entry Clerk. Data entry will be verified by entering data twice and resolving any discrepancies between resultant files or by performing a complete review.

Because data will be added to the unvalidated database, a series of checks and summaries will be performed to locate and correct any inconsistencies. This will include the use of data verification programs designed for sorting and comparing fields to determine whether data are missing, mislabeled, or duplicated. All such errors will be reconciled and documented.

After data validation, the Data Coordinator will oversee manual entry of flags and qualifiers into the database by the Data Entry Clerk. This data entry will be verified either by double entry and a compare program or by total review. The Data Coordinator then will migrate the completed records into a validated database and submit verification and validation documentation to the project file. All subsequent summaries, statistical analyses, and reports will be generated from the validated database using standardized and quality-controlled retrieval programs and validated application programs.

7.5.10 Data Consolidation and Storage

Once the data for a given sample or group of samples are complete and have been entered into the database, the Data Coordinator will ensure that logbooks and other field records are complete and properly stored. This will include both the electronic form and all the associated data packages. Each piece of information will be documented as to its sources. Hard-copy information will be appropriately indexed and filed. Metadata will be added to the validated database and linked to corresponding records.

Strict adherence to procedure-based routines established for data security, backup, and archival and the maintenance of proper database change procedures ensure database integrity. The database, once loaded, will be safe from physical corruption (i.e., hardware or software failure) or from unauthorized access and illegal updating. Corruption of the data causes loss of integrity. Physical security requires recovery procedures, time-stamping, and other related standard operating processes and controls. Any changes to the database will be documented using the Database Change Request Form.

7.5.11 Data and Document Archival

The Project Manager will instruct the Database Administrator to archive project data. The Database Administrator will archive project data to the appropriate electronic media. The Database Administrator also will prepare a data archive information package that describes the data system, the file format, and the method of archival. A copy of the checklist will be submitted to the Project Files.

Sufficient documentation will accompany the archived data to fully describe the source, contents, and structure of the data to ensure future data usability. Computer programs used to manipulate or report the archived data will be included in the data archive information package to further enhance the data's future usability.

Hard copies of all original field and analytical results, data reduction and summary programs, data packages, and associated QA/QC forms will be classified, indexed, and stored in appropriate file groups and record series within project files. The Records Manager will ensure that all required Administrative Record File documents are acquired and indexed.

• 7.6 DATA ASSESSMENT AND INTERPRETATION

Data assessment procedures are discussed in Sect. 8.10 of this RI work plan (i.e., in the QAPjP). Data interpretation is discussed in Sect. 6.3 of this work plan.

APPROVALS

QUALITY ASSURANCE PROJECT PLAN
FOR THE REMEDIAL INVESTIGATION OF THE
BEAR CREEK OPERABLE UNIT 1
AT THE OAK RIDGE Y-12 PLANT

DOE/OR/01-1161/V1&D2

March 1994

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Technical Support Contractor Program Manager Date

James C. Brannan 3/21/94
Technical Support Contractor QA/QC Officer Date

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S. P. Riddle (concur w/ comments of 3/28/94) 3/28/94
DOE-ER Division QA Program Manager Date

Tricia Wick 5/24/94
Energy Systems ER Site Program Manager Date

Garry A. Hodgins 3/23/94
Energy Systems ER Site Project Manager Date

James N. Dumont 3/28/94
Energy Systems ER Site QA Specialist Date

8. QUALITY ASSURANCE PROJECT PLAN

8.1 DATA QUALITY OBJECTIVES

QA objectives for data collection are developed from DQOs for this project. QA objectives are specified for each data collection activity and includes QC Levels I, II, III, and IV, which correlate with QC Levels 1, 2, 3, and 4, as described by the EPA document *Data Quality Objectives for Remedial Response Activities Development Process* (EPA 1987b). Data derived from sampling and analysis with EPA Level III and Level IV laboratory deliverables are of high quality and are both scientifically and legally defensible. For this RI, at least Level II and III laboratory deliverables will be used when development of a remedial strategy is required.

The BCV OU 1 location poses two problems that the DQOs will address: (1) whether the wastes and soil contamination pose a risk to human health or the environment now or in the future and (2) if the wastes and contamination do pose a risk, what feasible alternatives exist for reducing the risk to acceptable levels.

A QC program for analytical data develops information that can be used to accomplish the following:

- Evaluate the accuracy and precision of analytical data to establish their quality.
- Indicate the need for corrective actions, when a comparison with existing regulatory or program criteria or data trends shows that activities must be changed or monitored to a different degree.
- Determine the effect of corrective actions.

Because the laboratory for this project has not yet been specified, the QC criteria included herein should be considered minimum requirements. When the laboratory or laboratories have been specified, the QC requirements defined herein will be communicated to those facilities in analytical services statements of work. Any minor changes to these requirements will be documented and approved through the variance system. Requirements for analytical documentation for the selected laboratory will be QC Level II and III.

The QA objectives for all analytical data are to obtain reproducible, precise, and accurate measurements consistent with the intended use of the data and the limitations of the sampling and analytical procedures. The reasons for QC are (1) to screen out data of unacceptable degrees of precision or accuracy and (2) to obtain data that will accomplish the objectives of this project. QC will be accomplished through the assignment of measurement tasks to the appropriate analytical level (I-IV) as defined in *Data Quality Objectives for Remedial Response Activities Development Process* (EPA 1987b). This QAPjP specifies the documentation and QA requirements applicable to these data and are specified in Sect. 8.9.

8.1.1 QC Level of Field Laboratory Analysis and Data Quality Parameters

The DQOs for the plan determine the level(s) of field laboratory analysis and data quality parameters that will be required. Guides to measurements typical of the QC levels follow.

The collection and field screening/analysis of environmental samples are used for surface soil samples analysis of metals and PCB. These methods will provide QC Level II data. Additionally, Level I data will be generated by other field screening methods described below.

8.1.1.1 QC Level I analytical support

QC Level I analytical support is used to generate field screening data to determine the extent of contamination at a site, to provide real-time health and safety monitoring, to allow screening/optimization of sampling, and to acquire qualitative measurements of ambient conditions. General data quality criteria for QC Level I field screening support are specified below. Analytical parameters typically specified for QC Level I data collection include the following:

- measurement of ambient temperature (soil/water/air),
- measurement of electrical conductivity (soil/water),
- measurement of pH (soil/water),
- measurement of salinity (water),
- measurement of dissolved oxygen (water),
- total organic vapor concentration [photoionization detector/flame ionization detector (PID/FID)] (air/soil/gas),
- radiological monitoring (alpha/beta/gamma) (soil/air/equipment/personnel)

Calibration of QC Level I instrumentation, in accordance with the manufacturer's instructions as appropriate for the media and analytes of concern, is required before and upon completion of daily sampling. Positive response verification checks will be performed by the Health and Safety Officer, the Field Team Leader, or designee on these instruments for every 20 samples or every 4 h, whichever is more frequent. These checks consist of exposing the instrument to a source analyte (e.g., exposing a PID to a volatile organic source) and verifying a response. If no response is obtained, corrective action will be taken and documented and the instrument recalibrated. Results of calibration and response checks will be recorded in the instrument logbook.

Personnel responsible for using and/or maintaining particular instruments must maintain a log of calibration/response check procedures and performance. The Field Team Leader will inspect the logs, which will become part of the project records. Each operator must be trained in the proper use of the instruments, be familiar with their operation, be able to interpret data from them, and be able to produce documentation certifying these skills. This documentation will facilitate comparability between samples.

QC Level I DQOs for sampling conducted during the investigation are met by the following:

- sampling procedures,
- site surveillance,
- use of standard sampling and monitoring devices,

- personnel training,
- documentation of sampling location, method, and media,
- reporting of results in appropriate units,
- consistent use of types of screening/monitoring instruments,
- maintenance of measuring and test equipment (M&TE) and M&TE records.

8.1.1.2 QC Level II analytical support

QC Level II analytical support is used to provide real-time screening data for ongoing field activities, to identify samples for laboratory analysis, and to allow for a phased approach to sampling. Data quality criteria for QC Level II field analysis are given below. Analytical parameters, instrumentation, and media typically specified for QC Level II data collection may include the following:

- VOC analysis (soil/water);
- inorganic compound analysis (mobile lab) (soil/water/air);
- pesticide/PCB and dioxin/furan analysis (soil);
- special applications (e.g., down-hole monitoring, geotechnical applications).

Calibration of QC Level II instrumentation, in accordance with the manufacturer's instructions as appropriate for the media and analytes of concern, is required before and upon completion of the daily analyses or as required by the appropriate EPA-recommended methodology. Documentation and QC procedures for QC Level II field analyses are instrument- and analyte-specific. However, QC Level II data will be confirmed by submitting a minimum of 5% duplicate samples, with at least one sample per matrix, to a QC Level III or IV laboratory facility for analysis.

QC Level II DQOs for sampling conducted during the investigation are met by the following:

- sampling procedures;
- site surveillance;
- use of standard sampling and monitoring devices;
- personnel training;
- documentation of sampling location, method, and media;
- reporting of results in appropriate units;
- consistent use of types of screening/monitoring instruments;
- stringent control limits for the daily QC checks.

8.1.2 QC Level of Analytical Laboratory and Data Quality Parameters

Analytical laboratory data will be generated for environmental samples to provide confirmed identification and quantification of organic and inorganic compounds and radionuclides in each sampled media. Analytical methods that may be used include EPA CLP, SW-846, and others as long as the QC information that is appropriate to DQOs is being reported and/or delivered.

8.1.2.1 QC Level III analytical support

QC Level III analytical support is used to provide confirmed identification and quantification of contaminants to determine the extent of contamination, to determine the presence or absence of contaminants with required detection limits, and to confirm the results of lower QC Level (I & II) screening and analyses. QC Level III may provide data of comparable quality to QC Level IV but with less QC documentation, at a lower cost, and a shorter turnaround time. Analytical parameters and media typically specified for QC Level III data collection include these:

- organic and inorganic compound analysis (soil/water/air)
- TCLP

QC Level III laboratory QC procedures will follow requirements specified in the appropriate EPA-recommended methodologies. QC Level III protocols have built-in QA/QC, including calibration runs, surrogate standards, etc. External QA is employed in the form of trip blanks, replicate and duplicate samples, and blind spikes, which may be submitted with the samples. Documentation requirements depend on the laboratory and type of analysis performed.

Accuracy, precision, and method detection limit (MDL) information representative of QC Level III analytical support is provided by adherence to SW-846 and/or EPA CLP.

Laboratories performing QC Level III analyses must maintain a record of instrument calibration and QA/QC results. Such records will become part of the project records. Each operator must be trained in the proper use of the instrument, be familiar with the instrument's operation, be able to interpret data from the instrument, and have documentation certifying these skills.

8.1.2.2 QC Level IV analytical support

QC Level IV analytical support conforms to the EPA CLP for performing routine analytical services for analyses of all types of media. QC Level IV is used to provide confirmed identification and quantification of contaminants, to determine the extent of contamination, and to determine the presence or absence of contaminants with low required detection limits. QC Level IV provides documented data for risk assessment and remedial investigations. Data quality criteria for QC Level IV laboratory analyses are given below. Analytical specifications and media that may require QC Level IV data analysis include the following:

- Hazardous Substance List organic compounds (soil/water/air),
- priority pollutant inorganic compounds (soil/water/air),
- TCLP.

QC Level IV laboratory QC procedures will follow requirements specified in the current EPA SOW revision for CLP methodologies. Methods not included in the CLP will be elevated to QC Level IV by meeting CLP-type requirements and submitting all raw data. QC Level IV CLP routine analytical services protocols are very specific concerning QA/QC requirements and data package documentation. Routine analytical services deliverable packages contain information on initial and continuing calibration, gas chromatograph/mass spectrophotometer tuning, surrogate percent recovery, matrix spike duplicates (MSDs), and spectra for every sample and blank. External QA is employed in the form of trip blanks, replicate and duplicate samples, and blind spikes submitted with the samples.

QC Level IV precision, accuracy, representativeness, completeness, and comparability (PARCC) parameters are similar to those described for Level III. Routine analytical services documentation is sufficient to allow qualified personnel to review, verify, and validate analytical results and to evaluate data quality.

CLP routine analytical services documentation will be maintained as part of the project records. Chemists must be trained in the proper use of the instrument, be familiar with the instrument's operation, be able to interpret data from the instrument, and be able to produce documentation certifying these skills.

8.2 SAMPLE COLLECTION PROCEDURES

Subsurface and surface soil, surface water, groundwater, and sediment sampling are proposed for BCV OU 1. Characterization efforts will focus on several potential contamination sources at the Burial Grounds, the Oil Landfarm, the Boneyard/Burnyard, and the S-3 Site. Prior to commencing field activities, a readiness review will be conducted by SAIC following SAIC QAAP 2.2. Energy Systems will conduct a readiness review following ER/C-P1610 (see Appendix E).

A field procedures manual will be prepared and controlled by SAIC and approved by Energy Systems Y-12 ER Project Manager, Y-12 ER Quality Assurance Specialist, SAIC Project Manager, and SAIC QA/QC Officer. The procedures manual will include procedures from Energy Systems *Environmental Surveillance Quality Control Program*, ES/ESH/INT-14, Rev. 4 (Kimbrough et al 1994); Energy Systems *ERWM Programs Intersite Procedures Manual*, ES/ER/INT-26/R1, and the SAIC Field Technical Procedures Manual. The Energy Systems procedures will be modified, as necessary, and submitted for Y-12 ER approval for ESP procedures with modifications and BCV OU 1 site/project specifications on the cover sheet in front of each procedure.

8.2.1 Soil Sampling

Soil sampling will be conducted in accordance with the ESP-303 series and SAIC procedures (Table 8.1).

Table 8.1 Sample collection procedures

Procedure number	Subject activity	Issue date	Revision
Activity series 301: surface water sampling			
ESP-310-1	Water sampling using a dipper	8/30/88	0
ESP-310-4	Grab sampling with Kemmerer bottles	8/30/88	0
ESP-310-5	Stream flow measurements	8/30/99	0
Activity series 302: groundwater sampling			
ESP-302-1	Water level measurements using a water level indicator	8/29/88	0
ESP-302-5	Using a bladder pump	8/29/88	0
FTP-352	Groundwater sampling procedures: using a hydropunch	6/30/93	0
FTP-357	Groundwater sampling using micropurge and low flow sampling	TBA	
FTP-616	Installation, development and abandonment of temporary piezometers	TBA	
Activity series 303: soil sampling			
ESP-303-1	Soil sampling with a spade and scoop	8/30/88	0
ESP-303-2	Soil sampling with an auger	9/01/88	0
ESP-303-3	Soil sampling using a trier	8/31/88	0
ESP-303-4	Penetration test and split-barrel sampling	8/24/88	0
ESP-303-5	Subsurface soil sampling with shelby tubes	8/30/88	0
FTP-551	Soil sampling by direct push	TBA	
Activity series 304: sediment sampling			
ESP-304-1	Streambeds	8/29/88	0
Activity series 307: field measurements			
FTP-880	Field measurement procedures: pH, temperature, and conductance	8/29/88	0
ESP-307-5	Oxidation/reduction potential of water	8/29/88	0
FTP-750	Field measurement procedures: organic vapor detection	8/29/88	0
FTP-451	Field measurement procedures: operation of radiation survey instruments	1/31/90	0
FTP-701	Field measurement procedures: using a proton magnetometer	TBA	
FTP-710	Field measurement procedures: using the EM-31	TBA	

8.2.2 Sediment Sampling

Sediment sampling will be conducted in accordance with the ESP-304 series (Table 8.1).

8.2.3 Surface Water Sampling

Surface water sampling will be conducted in accordance with the ESP-301 series (Table 8.1).

8.2.4 Groundwater Sampling

Groundwater sampling will be conducted in accordance with the ESP-302 series and SAIC procedures (Table 8.1).

8.2.5 Field Measurement of Physical Parameters

Field measurement of physical parameters will be conducted in accordance with the ESP-307 series and/or SAIC procedures (Table 8.1). Due to the wide variety and case-specific application of these types of instrumentation, each procedure will include manufacturer's instructions for specific instrument operation and maintenance procedures.

8.2.6 Sample Compositing and Special Sample Collection Procedures

Sample compositing and specialized sample collection procedures will be conducted in accordance with ESP-308 series or SAIC procedures.

8.3 SAMPLE PREPARATION AND ANALYSIS PROCEDURES

8.3.1 Sample Containers, Sample Preservatives, and Holding Times

The selection criteria for appropriate sample containers, sample preservatives, and holding times shall be in accordance with ESP-701. Types of sample containers and sample preservation methods used will be documented in the sampling logbook. Field and laboratory records will indicate the sample holding time before analysis (Table 8.2).

8.3.1.1 Sample packaging and transportation

Handling, shipping, and storage of samples and data resulting from field activities will adhere to custody (Sect. 8.4) and will ensure the integrity for analytical purposes is maintained. The procedures required to properly package, ship, handle, and store containers of environmental samples are described in ER/C-P2302.

8.3.1.2 Decontamination of equipment and devices

Decontamination of sample containers and sampling devices will be performed according to ESP-801, Section VI, C, page 19-20. Sampling equipment will be decontaminated before use and between sampling locations/intervals. Each decontamination activity will be recorded in the field logbook. Equipment used during field activities, including drilling

Table 8.2. Sample holding times, sample containers, sample preservation, and minimum sample size

Parameter	Matrix	Holding time ^a (from time of collection)	Container	Preservative ^b	Sample size ^c
Volatile organics	Water	14 d	Two 40 mL vials with Teflon-lined caps	HCl to pH < 2 Cool to 4°C	80 mL
	Soil	14 d	4 oz glass jar with Teflon-lined cap	Cool to 4°C	50 g
Extractable organics (semivolatile organics, or pesticides/PCBs)	Water	7 d, extraction 40 d, analysis	1-L amber glass with Teflon-lined cap	Cool to 4°C	1000 mL
	Soil	14 d, extraction 40 d, analysis	8 oz glass jar with Teflon-lined cap	Cool to 4°C	100 g
Metals (other than mercury)	Water Soil	180 d 180 d	1-L poly bottle 8 oz plastic jar	HNO ₃ to pH <2	900 mL 80 g
Mercury	Water Soil	28 d 28 d	1-L poly bottle 8 oz plastic jar	HNO ₃ to pH <2 Cool to 4°C	100 mL 10 g
Cyanide	Water Soil	14 d 14 d	500 mL poly 1-L poly bottle 8 oz plastic jar	Water NaOH to pH > 12, Cool to 4°C	1000 mL 10 g
Radionuclides	Water Soil	180 d 180 d	4-L cubitainer 8 oz poly jar	HNO ₃ to pH < 2	4 L 100 g
Anions ¹	Water	28 d	250 mL poly	Cool to 4°C	250 mL
Specific conductance	Water	28 d	Polyethylene	4°C	100 mL
Miscellaneous ²	Water	24 hr	250 mL poly	Cool to 4°C	250 mL
Nitrate/Nitrite	Water	28 d	250 mL poly	H ₂ SO ₄ to pH < 2 Cool to 4°C	250 mL
pH	Water Soil	Analyze immediately 24 h	In situ, beaker or bucket Polyethylene	None None	100 mL 10 g

Table 8.2 (continued)

Parameter	Matrix	Holding time ^a (from time of collection)	Container	Preservative ^b	Sample size ^c
Temperature	Water	Determine on-site	In situ, beaker or bucket	None	100 mL
Soil moisture content	Soil	14 d	Polyethylene, glass	4°C	10 g
Percent organic matter, particle size, cation exchange capacity, ignitability, corrosivity, reactivity	Soil	28 d	Polyethylene, glass	4°C	100 g

^aHolding times are consistent with CLP validation guidelines.

^bDissolved metals require filtration before pH adjustment.

^cAdditional sample must be collected for matrix spike/matrix spike duplicate samples or matrix spike/duplicate.

^dAnions include chloride, sulfate, and fluoride.

^eMiscellaneous includes TSS, TDS, alkalinity, and turbidity.

equipment, soil sampling equipment, and field test equipment will be decontaminated in accordance with ESP-802.

8.3.1.3 Sample identification and traceability

Each environmental sample collected during these investigations will be assigned a unique sample identifier, which will be permanently affixed to the sample container and recorded in the field logbook and on the chain-of-custody form. The identifiers used for samples will be established and maintained in accordance with ESP-501.

Identification systems will ensure traceability of samples to the appropriate source. Sample identification shall be transferred to each subdivision when the sample is split.

8.3.1.4 Site surveying

Site surveying will be conducted to establish location and elevation of sampling points and monitoring well measuring points in relation to the appropriate plant or site grid. A licensed surveyor will be contracted to perform this effort.

8.3.1.5 Field variance system

Procedures cannot fully encompass all conditions encountered during a field investigation. Variances from operating procedures, the work plan, the sampling and analysis plan, and/or health and safety plan will, therefore, likely occur and must be documented on a field change order (FCO) form following ER/C-P1719 and noted in the appropriate logbooks. The Y-12 ER Project Manager may give verbal approval for the field change, if the risk issue is minor, until the FCO is reviewed and categorized. If a variance is anticipated (e.g., due to a change in field instrumentation), the applicable procedure should be modified and the change noted in the field logbooks.

Any variances from the health and safety plan must be approved by the Health and Safety Officer. Copies of the field change order forms and a field change order log will be maintained by the sampling teams until the fieldwork is complete. Field Change order forms and log will be included in the project file and the subcontractor's Central Records Facility.

Routine (minor) FCO will be completed in the field, noted on the FCO log, then routed through the FCO Coordinator for delivery to the Energy Systems Y-12 ER Project Manager and QAS for signature approval. A verbal approval from the Y-12 ER Project Manager or designer will be noted in the field logbook.

Emergency or urgent (major) FCOs will be delivered to the Y-12 ER Project Manager and QAS for immediate signature approval, then forwarded to the FCO Coordinator.

8.3.2 Analytical Procedures

Environmental samples collected during this projects will be analyzed for potential contaminants of concern using EPA-approved methods where applicable (Table 8.3).

The approved methods and protocols that may be used by the analytical laboratory are provided in the following documents:

Table 8.3. Analytical methods, parameters, and quantitation limit goals for surface water/groundwater and soil/sediment samples

Parameters	Analytical methods		Quantitation limit goals ^a	
	Water	Soil/sediment	Water (µg/L)	Soil/sediment (µg/kg)
Volatile organic compounds (VOCs)	CLP SOW ^b	CLP SOW ^b		Low soil/sediment ^c
Chloromethane			10	10
Bromomethane			10	10
Vinyl chloride			10	10
Chloroethane			10	10
Methylene chloride			5	5
Acetone			10	10
Carbon disulfide			5	5
1,1-Dichloroethene			5	5
1,1-Dichloroethane			5	5
1,2-Dichloroethene (total)			5	5
Chloroform			5	5
1,2-Dichloroethane			5	5
2-Butanone			10	10
1,1,1-Trichloroethane			5	5
Carbon tetrachloride			5	5
Bromodichloromethane			5	5
1,2-Dichloropropane			5	5
cis-1,3-dichloropropene			5	5
Trichloroethene			5	5
Dibromochloromethane			5	5
1,1,2-Trichloroethane			5	5
Benzene			5	5
trans-1,3-dichloropropene			5	5
Tribromomethane			5	5
4-Methyl-2-pentanone			10	10
2-Hexanone			10	10
Tetrachloroethene			5	5
Toluene			5	5
1,1,2,2-Tetrachloroethane			5	5
Chlorobenzene			5	5
Ethylbenzene			5	5
Styrene			5	5
Xylenes (total)			5	5

Table 8.3 (continued)

Parameters	Analytical methods		Quantitation limit goals ^a	
	Water	Soil/sediment	Water (µg/L)	Soil/sediment (µg/kg)
Additional compounds:				
Trichlorotrifluoroethane			10	10
Semivolatile organic compounds (SVOCs)	CLP SOW ^b	CLP SOW ^b		Low soil/sediment ^c
Phenol			10	330
bis(2-Chloroethyl) ether			10	330
2-Chlorophenol			10	330
1,3-Dichlorobenzene			10	330
1,4-Dichlorobenzene			10	330
1,2-Dichlorobenzene			10	330
2-Methylphenol			10	330
2,2'-Oxybis(1-chloropropane)			10	330
4-Methylphenol			10	330
N-nitroso-di-n-dipropylamine			10	330
Hexachloroethane			10	330
Nitrobenzene			10	330
Isophorone			10	330
2-Nitrophenol			10	330
2,4-Dimethylphenol			10	330
bis(2-Chloroethoxy) methane			10	330
2,4-Dichlorophenol			10	330
1,2,4-Trichlorobenzene			10	330
Naphthalene			10	330
4-Chloroaniline			10	330
Hexachlorobutadiene			10	330
4-chloro-3-methylphenol			10	330
2-Methylnaphthalene			10	330
Hexachlorocyclopentadiene			10	330
2,4,6-Trichlorophenol			10	330
2,4,5-Trichlorophenol			25	800
2-Chloronaphthalene			10	330
2-Nitroaniline			25	800
Dimethylphthalate			10	330
Acenaphthylene			10	330
2,6-Dinitrotoluene			10	330

Table 8.3 (continued)

Parameters	Analytical methods		Quantitation limit goals*	
	Water	Soil/sediment	Water ($\mu\text{g/L}$)	Soil/sediment ($\mu\text{g/kg}$)
3-Nitroaniline			25	800
Acenaphthene			10	330
2,4-Dinitrophenol			25	800
4-Nitrophenol			25	800
Dibenzofuran			10	330
2,4-Dinitrotoluene			10	330
Diethylphthalate			10	330
4-Chlorophenyl-phenyl ether			10	330
Fluorene			10	330
4-Nitroaniline			25	800
4,6-Dinitro-2-methylphenol			25	800
N-nitrosodiphenylamine			10	330
4-bromophenyl-phenylether			10	330
Hexachlorobenzene			10	330
Pentachlorophenol			25	800
Phenanthrene			10	330
Anthracene			10	330
Carbazole			10	330
Di-n-butylphthalate			10	330
Fluoranthene			10	330
Pyrene			10	330
Butylbenzylphthalate			10	330
3,3'-Dichlorobenzidine			10	330
Benzo(a)anthracene			10	330
Chrysene			10	330
bis(2-Ethylhexyl)phthalate			10	330
Di-n-octylphthalate			10	330
Benzo(b)fluoranthene			10	330
Benzo(k)fluoranthene			10	330
Benzo(a)pyrene			10	330
Indeno(1,2,3-cd)pyrene			10	330
Dibenzo(a,h)anthracene			10	330
Benzo(g,h,i)perylene			10	330

Table 8.3 (continued)

Parameters	Analytical methods		Quantitation limit goals ^a	
	Water	Soil/sediment	Water (µg/L)	Soil/sediment (µg/kg)
Pesticides	CLP SOW ^b	CLP SOW ^b		Low soil/sediment
alpha-BHC			0.05	1.7
beta-BHC			0.05	1.7
delta-BHC			0.05	1.7
gamma-BHC (Lindane)			0.05	1.7
Heptachlor			0.05	1.7
Aldrin			0.05	1.7
Heptachlor epoxide			0.05	1.7
Endosulfan I			0.05	1.7
Dieldrin			0.10	3.3
4,4'-DDE			0.10	3.3
Endrin			0.10	3.3
Endosulfan II			0.10	3.3
4,4'-DDD			0.10	3.3
Endosulfan sulfate			0.10	3.3
4,4'-DDT			0.10	3.3
Methoxychlor			0.5	17.0
Endrin ketone			0.10	3.3
Endrin aldehyde			0.10	3.3
alpha-Chlordane			0.05	1.7
gamma-Chlordane			0.05	1.7
Toxaphene			5.0	170.0
Dioxin/Furon				
2,3,7,8-TCDD	SW 8290	SW 8290	.01 ng/L	.001-0.2
PCBs	CLP SOW ^b	CLP SOW ^b		Low soil/sediment ^c
Aroclor-1016			1.0	33.0
Aroclor-1221			2.0	67.0
Aroclor-1232			1.0	33.0
Aroclor-1242			1.0	33.0
Aroclor-1248			1.0	33.0
Aroclor-1254			1.0	33.0
Aroclor-1260			1.0	33.0

Table 8.3 (continued)

Parameters	Analytical methods		Quantitation limit goals ^a	
	Water	Soil/sediment	Water (µg/L)	Soil/sediment (µg/kg)
Metals (target analyte list)	CLP SOW ^d	CLP SOW ^d		(mg/kg)
Aluminum			200	20
Antimony GFAA			3	0.3
Arsenic GFAA			5	0.5
Barium			200	20
Beryllium			5	0.5
Cadmium			5	0.5
Calcium			5000	500
Chromium			10	1
Cobalt			50	5
Copper			25	2.5
Iron			100	10
Lead GFAA			3	0.3
Magnesium			5000	500
Manganese			15	1.5
Mercury CVAA			0.2	0.02
Nickel			40	4
Potassium			5000	500
Selenium GFAA			5	0.5
Silver			10	1
Sodium			5000	500
Thallium GFAA			5	0.5
Vanadium			50	5
Zinc			20	2
Additional elements:				
Molybdenum			40	2
Tin			50	5
Common anions			(mg/L)	(mg/kg)
Nitrate-nitrite	E353.2 ^e	NA	0.2	NA
Chloride	E325.1 ^e	NA	1.0	NA
Sulfate	E375.4 ^e	NA	5	NA
Fluoride	E340.2 ^e	NA	0.1	NA
Miscellaneous				
TSS	E160.2 ^e	NA	10	NA
TDS	E160.1 ^e	NA	10	NA

Table 8.3 (continued)

Parameters	Analytical methods		Quantitation limit goals ^a	
	Water	Soil/sediment	Water (µg/L)	Soil/sediment (µg/kg)
ALK	E310.0 ^d	NA	NA	NA
Turb	E180.1 ^d	NA	0.1 NTUL	NA
Indicator parameters				
Total organic carbon (TOC)	E415.1/E415.2 ^e	NA	1	NA
Radionuclides			(pCi/L)	(pCi/g dry)
Gamma spectrometry	EPA 901.1 ^f	EPA 901.1 ^f		
Gamma scan ^g			g	g
Alpha spectrometry	EPA 907.0 ^f	EPA 907.0 ^f		
Americium-241			0.1	0.1
Plutonium-238,339,240			0.1	0.1
Thorium-228,230,232			1	1
Uranium-234,235,238			1	1
Nepunium-237			0.1	0.1
Strontium-90	EPA 905.1 ^f	EPA 905.1 ^f	1	1
Technetium-99	E-Tc-01		30	30
Gross alpha	EPA 900.0 ^f	EPA 900.0 ^f	10	20
Gross beta	EPA 900.0 ^f	EPA 900.0 ^{if}	5	20

NA = not applicable.

^aThese are expected quantitation limits based on reagent grade water or a purified solid matrix. Actual quantitation limits may be higher depending on the nature of the sample matrix. The limit reported on final laboratory reports will take into account the actual sample volume or weight, percent moisture (where applicable), and the dilution factor, if any. The quantitation limits for additional analytes attached to the Target Analyte List of CLP SOWs may vary, depending on the results of laboratory studies.

^bContract Laboratory Program, Statement of Work for Organics Analysis, Multi-Media, Multi-Concentration, EPA Document OLM01.8. Quantitation limits are contract-required quantitation limits with the exception of additional organic compounds. Minimum quantitation limits stated in the table will be reported by the laboratory.

^cMedium soil/sediment contract-required quantitation limits are 120 times the low soil/sediment contract-required quantitation limits for volatile organic compounds and 30.3 times the low soil/sediment contract-required quantitation limits for semivolatile organic compounds. Estimated detection limits for metals in soil are based on a 1-g sample diluted to 100 mL.

^dContract Laboratory Program Statement of Work for Inorganics Analysis, Multi-Media, Multi-Concentration, EPA Document ILM01.1. Determination limits are contract-required detection limits except for tin and molybdenum. Minimum detection limits stated in the table will be reported by the laboratory.

^eMethods for Chemical Analysis of Water and Wastes, U.S. EPA. EPA-600/4-79-020 revision March 1983.

^fPrescribed Procedures for Measurement of Radioactivity in Drinking Water, U.S. EPA, EPA-600/4-80-032, latest version.

^gThe subcontracted laboratory will resolve and quantify nonspecified gamma-emitting radionuclides. These nonspecified gamma-emitting radionuclides should be detected and semiquantitatively determined based on photon yield and energy relative to the ¹³⁷Cs MDL and standard.

^hNephelometric Turbidity Units.

- *Test Methods for Evaluating Solid Wastes* (EPA 1986)
- *Methods for Chemical Analysis of Water and Wastes* (EPA 1983)
- *Prescribed Procedures for Measurement of Radioactivity in Drinking Water* (EPA 1980a)
- *Eastern Environmental Radiation Facility Radiochemistry Procedure Manual* (EPA 1984)
- *CLP SOWs for Organics and Inorganics* (EPA 1990b,c)
- 40 CFR 136, Appendix A
- 40 CFR 141.30
- 40 CFR 136, Appendix C
- APHA, *Standard Methods for the Analysis of Water and Wastewater*
- *EML Procedures Manual*, HASL-300 (Krey and Beck 1992)

The MDL is defined as the minimum concentration of a substance that can be measured and reported with 99% confidence that the value is greater than zero. The MDL actually achieved in a given analysis varies depending on instrument sensitivity and interferences. Contracts will be established with analytical laboratories to analyze environmental samples collected during these investigations. Each contract laboratory that analyzes samples will provide quantification limits for each constituent analyzed.

Laboratory equipment QA/QC will be drawn from the supporting laboratory's (or laboratories') QA plan(s).

8.4 SAMPLE CUSTODY

8.4.1 Field Documentation

An integral part of the QA/QC plan for the field activities will be to maintain current, accurate, and complete field records including logbooks, chain-of-custody forms, and airbills. Field logbooks (e.g., Field Health and Safety, Site Activity, M&TE, Decontamination, Sampling Management logbook for each activity) shall be of hardcover construction. All information pertinent to field activities will be recorded. Each page must be signed and dated. Entries in the logbooks or on the data forms will be made in water resistant black or blue ink and will include the information specified in ESP-500, Sect. VII, Part D methodologies. Corrections must be marked out with a single line, dated, and initialed. Field logbooks will be copied once a week with the copies remaining in the SAIC offices. When the field work is complete copies will be made of the completed logbooks, entered into the SAIC Control Records Facility, the Project File, and a copy forwarded to the Y-12 ER Project Manager.

Appropriate field data forms will be prepared based on the requirements of the current revision of the *Environmental Restoration Quality Program Plan*, ES/ER/TM-4/R3 (Energy Systems 1993c). No blank spaces should appear on completed forms. If information requested is not applicable, the space shall be marked "N/A."

Chain-of-custody procedures require documentation of sample possession from the time of collection to time of disposal. These procedures allow the possession and handling of samples from the time of collection through analysis and final disposal to be traced.

Samples will be accompanied by the original completed Chain-of-Custody Form (see Appendix F). Should a sample be split, another copy will remain with the sample. As the samples are transferred, the present custodian and the new custodian will complete the required sections of the record as well as note any discrepancies. This Chain-of-Custody Form will remain with the sample from the field while it is being transported to the laboratory, and it will be checked upon receipt at the laboratory. The laboratory will retain a copy of the form, and the original will be sent to the designated subcontractor personnel. This original will be routed to the Central Records Facility and used in the final analytical report.

8.4.1.1 Field custody procedures

Sample custody will be initiated at the time of sample collection. Field samples will be identified by sample tags and appropriate labels. Descriptions of sampling activities and sample identification data will also be recorded in a field logbooks. Field chain-of-custody forms containing the same information will be completed for each set of samples. A line item on the sample chain-of-custody form will be completed for each sample, and the sampling technician will confirm by signature the completeness of the information on the form. Each individual who assumes responsibility for the samples will sign and date the chain-of-custody form.

Field custody procedures include the following steps:

- Before sampling begins, the QA/QC Officer, or designee, will instruct sampling personnel on the chain-of-custody and sample labeling procedures, as necessary.
- A Chain-of-Custody Form, which corresponds to the sample identification label, will be initiated in the field for each sample.
- Each time sample custody is transferred, the person relinquishing the sample and the new custodian will sign the Chain-of-Custody Form and note the date and time.
- The analyses to be performed for each sample will be recorded on the Chain-of-Custody Form accompanying the samples to the laboratory.
- The Field Team Leader will confirm that proper custody procedures are followed during the fieldwork and that results were documented in the appropriate field logbooks.

8.4.2 Shipping Custody Procedures

Sample custody will be maintained by personnel until custody is transferred to a common carrier/air freight company, if applicable. Shipments sent by common carrier/air freight will have a bill of lading/air bill accompanying them with the original Chain-of-Custody Form packed in the containers. These bills will be maintained as permanent records, and the air bill number will be noted on the Chain-of-Custody Form. The common carrier/air freight company delivers samples and transfers custody to particular analytical laboratories, where their intralaboratory chain-of-custody procedures will be in effect. If the Y-12 Plant

Laboratory is used to analyze the samples, the samples will be delivered by a sample team member to the laboratory.

8.4.3 Laboratory Custody Procedures

Upon receipt at the laboratory, each sample identification will be compared to the information contained on the Chain-of-Custody Form. If discrepancies exist, appropriate notes (signed and dated) will be made on the Chain-of-Custody Form and the Project Manager or designated person will be notified.

At receipt and initial inspection of samples and accompanying forms, the following items will be checked:

- Seals and tapes on the transportation container to verify that they are unbroken.
- Sample containers in the transportation container to ensure that they are intact and at correct temperature.
- External activity with radiation survey instruments and smear surfaces for removal of contaminants.
- The pH of preserved samples (except VOAs).
- The identification on the sample bottles to verify that it corresponds to the entire description on the Chain-of-Custody Form.
- The number of sample containers received to verify that it is equal to the number of samples listed on the Chain-of-Custody Form(s).

If there are discrepancies, the Project Manager or designee shall be called immediately to rectify them. Discrepancies will be recorded on appropriate forms and incorrect information marked out with a single line, initialed, and correct information added.

Once samples are in the possession of the laboratory, their internal chain-of-custody and sample handling procedures will be followed. If samples are to be shipped from one laboratory to another, proper chain-of-custody and packaging procedures will be maintained. Anticipation of reanalysis requires proper preservation of samples following analysis. Samples requiring refrigeration will remain refrigerated.

8.4.3.1 Sample labeling

Sample labels and tags will contain sufficient information to identify the sample in the absence of other documentation. The label and tag will be directly affixed to the sample container, will be completed with black water resistant ink, and will follow ESP-501.

Refer to Sect. 7.5 for details concerning the sample numbering system and preparation of pre-printed sample labels. In the event samples arrive damaged or custody seals are broken, a nonconformance report (Sect. 8.11.2) will be initiated. The Project Manager will be advised and will make a decision as to the fate of the nonconforming sample(s) and initiate corrective actions. The laboratories will retain tags in their project files.

8.5 CALIBRATION PROCEDURES AND FREQUENCY

All M&TE will be calibrated against certified equipment and/or standards having known valid traceability to nationally recognized standards. M&TE shall be calibrated, adjusted, and maintained at prescribed intervals and/or before use (see Table 8.4). If no nationally recognized standards exist, the basis for calibration will be documented.

8.5.1 Field Instrument Calibration Procedures and Frequency

A list of all M&TE to be used, along with a schedule for calibration, will be prepared before initiating fieldwork. Field instrumentation will be calibrated according to the procedures specified in the manufacturer's operating manual or more frequently should the conditions dictate it for the particular instrument. Instrument logbooks or notebooks will be established and maintained by the cognizant field team members, Field Team Leader, or the Health and Safety Officer, as appropriate.

8.5.2 Laboratory Instrument Calibration Procedures and Frequency

Laboratory equipment will be calibrated according to the procedures specified in the analytical methods and in the operating manual for the particular instrument. Calibration frequency will be based on the analytical methods employed, type of equipment, inherent stability, manufacturer's recommendations, values given in national standards, intended use, and experience.

For those cases for which analyses are governed by EPA CLP protocols, the calibration frequencies and procedures as designated in the CLP routine analytical services SOW will be followed. In this scenario, the calibration requirements in the CLP routine analytical services SOW take precedence.

For laboratory equipment, Class A volumetric glassware will be used to prepare calibration standards, bench standards, samples for analysis, etc. Class A glassware may be purchased with known accuracy per American Society of Testing and Materials specifications.

It should be noted that other nonanalytical instrumentation (such as thermometers) must be properly maintained and calibrated. The temperatures of ovens and refrigerators used in sample handling will be recorded and the control limits defined. If these limits are not met, corrective action will be required. All calibration and maintenance records will be kept with the equipment if practical. Otherwise, they will be maintained by the laboratory QA personnel.

8.5.3 Equipment Categories

Calibrated equipment will be uniquely identified by using either the manufacturer's serial number or other means. All equipment will be categorized as one of the following:

- Category A—Casual devices and systems (e.g., rulers, tape measures, graduated cylinders) that are not to be calibrated in service (i.e., not calibrated other than by the manufacturer).

Table 8.4. Level I instrument uses, detection limits, and calibration^a

Instrument	Uses	Detection limits	Calibration	Comments
Total organic vapor meters	Sample screening for VOCs	PID—0.2 ppm benzene	1 point—PID benzene daily	Action level must be stated in work plan
	Soil gas screening	FID—1.0 ppm methane	1 point—FID methane daily	Instrument cannot differentiate naturally occurring compounds from contaminants
	Health and safety screening	Daily calibration	Verification check every 20 samples	PID cannot detect compounds with ionization potentials >11 eV
Radiological monitoring	Monitoring of beta-gamma surface, gross gamma, alpha surface contamination levels	Daily calibration check varies by equipment	Daily source check per manufacturer	Validation labels include minimum and maximum acceptable levels
Indicator kits (colormetric)	Field screening for a variety of analytes	Varies kit by kit	Check color response w/level of analyte	Need to check expiration dates
	Gross quantitation	One-time check		One analyst should conduct the test to ensure consistency in interpretation Interferences from other compounds are possible
pH kits (litmus paper)	Field screening—water only	N/A	Check color response w/known pH standards of 4.0, 7.0, 10.0	Instrument is not to be used on samples to be submitted for analysis One-time check for each lot of kits is necessary
pH meters	Field screening—water and soil	N/A	2 point with standards at pH 7.0 and 4.0 or pH 7.0 and 10.0 daily	Accuracy is to ± 0.5 pH units

Table 8.4 (continued)

Instrument	Uses	Detection limits	Calibration	Comments
Conductivity meter	Determining conductivity of water	N/A	1 point in 0.01 in KCl solution	Calculations and acceptance criteria must be available in the field
Dissolved oxygen meter	Determining amount of dissolved oxygen in water	N/A	To manufacturer instructions	Calculations must be available in the field

^aPID = photoionization detector, FID = flame ionization detector, N/A = not applicable, VOCs = volatile organic compounds.

- Category B—Routine devices and systems (e.g., balances, spectrophotometers) that are to be included in a regular calibration recall program.
- Category C—Field experiment devices and systems (e.g., pH meters, turbidimeters) that are to be calibrated before, during, and after use.

The appropriate category decal with the identification number and the due date of the next calibration will be attached to the equipment. If this identification is not possible, records traceable to the equipment will be readily available for reference.

8.5.4 Calibration Failures

Scheduled periodic calibration of equipment will not relieve personnel of the responsibility of employing properly functioning equipment. If an individual suspects an equipment malfunction, he/she should remove the device from service, initiate a nonconformance report, tag it so it is not inadvertently used, and notify project management. If equipment is found to be out of calibration, the appropriate project management personnel will evaluate and document (in the instrument logbook) the validity of previous inspection or test results and the acceptability of similar equipment previously inspected or tested. The responsible supervisor will ensure the devices that are out of calibration are (1) tagged or segregated from other equipment and (2) disposed of or not used until they are calibrated. Any equipment that is consistently found to be out of calibration will be repaired or replaced. Any such action should be recorded in the instrument logbook or notebook.

All standards used for equipment calibration will be traceable to the EPA, National Institute of Standards and Technology, or a commercially available certified standard. The source of the standard used must be documented in a calibration logbook.

8.5.5 Calibration Records

Calibration data will be recorded in the instrument logbook or notebook. The information will include the date, operator, signature, and standard that was used. Records will be prepared and maintained for each piece of calibrated equipment to indicate that established calibration procedures have been followed. The appropriate project management personnel will ensure that records of calibration data are kept current. Records for field equipment used will be maintained by the Field Team Leader and the Health and Safety Officer and kept in the project files. Records for laboratory equipment used will be maintained by the laboratory supervisor and kept in the project files.

8.6 PREVENTIVE MAINTENANCE PROCEDURES/SCHEDULES

Any equipment (an inclusive term for tools, gauges, instruments, and other items that require specific preventive maintenance) will be serviced and documented as specified by the manufacturer's recommended schedule.

All service will be performed by qualified and trained individuals. The operators are responsible for seeing that the equipment is scheduled for service, serviced, and properly maintained. Properly maintained equipment helps reduce unnecessary "downtime."

A complete list of equipment will be developed by the operator, and the parts or replacement equipment will be immediately available (either from the supplier/manufacturer or on site). Having replacement equipment or critical spare parts available minimizes "downtime."

The implementation of a preventive maintenance program depends on the specific instruments and equipment used for the field. SAIC will insure a preventive maintenance program that includes the following:

- a list of the instruments and equipment that are included in the program, including backup alternatives;
- the frequency of maintenance, taking into consideration the manufacturer's recommendations and/or previous experience with the equipment. The frequency should be stated in terms of monthly, quarterly, etc.;
- for each instrument in the program provide the following:
 - external service contracts
 - items to be checked and/or serviced during maintenance
 - directions for performing maintenance (if external service is not provided, or if directions are not stated in the manufacturer's instrument manuals)

Preventive maintenance will be documented. A file will be maintained for each instrument at the SAIC office. The instrument file should include, at a minimum:

- external service contracts
- a checklist of items to be serviced and directions for maintenance or manufacturer's instrument manuals
- a record of periodic maintenance
- comments noting any replacement of parts, observed deterioration, etc.

Equipment used throughout the laboratory must be controlled through a calibration program. Calibration of analytical balances must be performed by a trained and qualified instrument technician using weights traceable to National Institute of Standards and Technology specifications. The gas chromatograph/mass spectrophotometer instruments must be calibrated, tuned, and operated in accordance with EPA CLP protocol OLM01.8 (EPA 1990c), which also meets the criteria specified in other EPA documents and the *Federal Register*. The ICP, atomic adsorption spectroscopy, gas chromatograph, and similar instruments must be calibrated and operated in accordance with requirements of the procedure being employed. The calibration of all equipment must be performed or verified with each analysis run, and the record of calibration for each analysis must be maintained as a permanent record. Any deviation from the calibrations required by the appropriate EPA method must be approved Using an FCO that is authorized by the SAIC Project Chemist; the Y-12 ER QA Specialist, and the Y-12 ER Project Manager. Project Chemist and the Y-12 ER Project Manager.

Volumetric glassware must be used to prepare calibration standards, bench standards, and samples for analysis; therefore, the glassware used for these preparations must be of known

accuracy. The glassware must be purchased in accordance with federal and American Society for Testing and Materials specifications.

The standards used for instrument calibration, internal controls, bench standards, spike solution, and traceability of sample analyses to these specific standards must be documented.

8.7 DATA REDUCTION AND REPORTING

Data reduction and reporting will be in accordance with the *Environmental Restoration Quality Program Plan*, ES/ER/TM-4/R3 (Energy Systems 1993c). The data reduction will follow the guidelines of the EPA CLP SOWs (EPA 1990b, 1990c) and the *Test Methods for Evaluating Solid Waste, Physical/Chemical Methods* (EPA 1990a). Generally, results must be expressed to two significant figures. The results for aqueous samples must be expressed in milligrams per liter or micrograms per liter, and the results for nonaqueous samples must be expressed on a dry weight basis in milligrams per kilogram, micrograms per kilogram, or micrograms per gram. The moisture content of samples must also be reported.

Table 8.5 depicts data set deliverables for Level III and Level IV QA. In addition to following field and laboratory documentation and QA/QC procedures, data may be verified using a variety of computerized checks for reasonableness. These procedures will ensure that data are entered, encoded, manipulated in a consistent way, and available in a usable format.

8.7.1 Field Data Reduction and Reporting

Data collected during field activities will be evaluated by checking the procedures used and comparing the data to previous measurements. The QA/QC Officer or designee and appropriate field personnel will be responsible for checking field QC sample results to ensure that field measurement and sampling protocols have been observed. These reviews will check date and time sampled, preservation, standard operating procedures, calibration method and frequency, and chain-of-custody documentation.

Reviewers are responsible for ensuring that data reduction calculations are documented and checked by qualified personnel. Written reports including reduced and summarized data may include the raw data in appendixes. Specific calculations used for data reduction may also be included.

8.7.2 Laboratory Data Reduction

In general, the analyst will process the data, either manually or by inputting the data into a computer. All computerized programs used to reduce and analyze data should be in compliance with software QC [Energy Systems 1990 (Sect. 2.19)] and appropriate Energy Systems procedures. For manually processed data, all the steps in the computation must be provided, including equations used and the source of input parameters such as response factors, dilution factors, and calibration constants. If calculations are not performed directly on the data sheet or chromatogram, the calculations must be provided on company letterhead paper and attached to the data sheets. All pages of the calculations must be signed and dated by the analyst performing the calculations.

Table 8.5. Data set deliverables for EPA Levels III and IV quality assurance^a

Method requirements	Deliverables
<i>Requirements for all methods</i>	
Holding time information and methods requested	Signed chain-of-custody forms
Discussion of laboratory problems	Case narratives
LCS with results on control charts (run with each batch of samples processed)	Control charts
<i>Organics</i>	
Sample results	CLP Form 1 or equivalent
Surrogate recoveries. Surrogates to be used in volatiles, semivolatiles, pesticides/PCBs. For volatiles by GC, surrogate names should reflect the appropriate surrogate used	CLP Form 2 or equivalent
Matrix spike/spike duplicate (will be 1 spike and spike duplicate per 20 samples of similar matrix)	CLP Form 3 or equivalent
Method blank data	CLP Form 4 or equivalent
GC/MS tuning for volatiles/semivolatiles	CLP Form 5 or equivalent
GC/MS initial calibration data for volatiles/semivolatiles	CLP Form 6 or equivalent
Pesticide/PCB calibration data if calibration factors are used	CLP Form 9 or equivalent
For volatiles by GC, initial calibration data: If calibration factors are used If a calibration curve is used	CLP Form 8D or equivalent, with five columns for multilevel calibration factors A plot of the calibration curve is required and a linear regression determination, with flagged correlation coefficient, if it is less than 0.995
For volatiles by GC, continuing calibration data; if calibration factors are used, calibration factors and their percent differences from the initial calibration must be reported; retention time (RT) windows and analyte RTs for the analytes must be included on this form	CLP Form 9 or equivalent
GC/MS continuing calibration data: No chromatograms or mass spectra are presented for calibration; these data should be filed in the laboratory and available if problems arise in reviewing/validating the data	CLP Form 7 or equivalent
GC/MS internal standard area data	CLP Form 8 or equivalent
Second column confirmation for all GC work when compounds are detected above reporting limits	CLP Form 10 or equivalent for all positive hits
<i>Metals</i>	
Sample results	CLP Form 1 or equivalent
Initial and continuing calibration	CLP Form 2 or equivalent

Table 8.5 (continued)

Method requirements	Deliverables
Method blank taken through digestion (1 per 20 samples of same matrix)	CLP Form 3 or equivalent
ICP interference check sample	CLP Form 4 or equivalent
Spike sample recovery (1 per 20 samples of similar matrix)	CLP Form 5A or equivalent
Postdigestion spike sample recovery for ICP metals (only done if predigest spike recovery exceeds CLP limits)	CLP Form 5B or equivalent
Postdigestion spike for graphite furnace atomic absorption	Recovery noted on raw data
Duplicates (one per 20 samples will be split and digested as separate samples)	CLP Form 6 or equivalent
LCS	CLP Form 7 or equivalent
Standard addition. The decision process outlined in CLP (p. E-3) will be used to determine when standard additions are required	CLP Form 8 or equivalent
Holding times	CLP Form 10 or equivalent
<i>Wet chemistry</i>	
LCS (one/batch)	Control charts
Method blank (one/batch)	Report result
Sample results	Report result
Spike/spike duplicate	Percent recovery and percent RPD
Continue calibration check	Recovery and percent difference
<i>Radiochemical analysis</i>	
Sample results	Report results
Initial calibration	Efficiency determination
Efficiency check	Percent difference from calibration
Background determinations	Report results
Minimum detectable activity	Report results
Method blank	Report results
Spike recovery results	Spike added and percent recovery
Internal standard results	Standard added and percent recovery
Duplicate results	Report results and percent RPD
Self-absorption factors (α , β)	Report factors
Run log	Copy of run log (Level IV only)

^aLCS = laboratory control standard, CLP = contract laboratory program, PCB = polychlorinated biphenyl, GC = gas chromatograph, MS = mass spectrometry, ICP = inductively coupled plasma.

For data input by an analyst and processed using a computer, a copy of the input must be kept and uniquely identified with the project number and other pertinent information as necessary. The samples to which the data processing refers must be clearly stated, and the input must be signed and dated by the analyst performing the input.

When processing data acquired through the use of instrumentation, the analyst must verify that the correct project, sample numbers, calibration constants, response factors, units, and numerical values used for detection limits are present.

After required chemical analyses have been completed, the samples will be maintained at a temperature of 4°C for at least 2 weeks after expiration of the appropriate holding time. The laboratory will be responsible for disposal of all laboratory-generated wastes. Upon the receipt of written notification from the subcontractor's Project Manager, the laboratory will return unused sample portions to the project.

8.8 QC CHECKS

8.8.1 Field QC Checks

The number of required QC samples per site is based on a set minimum percentage. The number of QC samples to be collected may change from site to site; however, the types of field QC samples remain constant. The types of QC samples are trip blanks, equipment rinsates, field blanks, and field duplicates.

In defining the number of field blanks required, it is important to note that a sampling event is from the time sampling personnel arrive at the site until personnel leave for more than 24 h. For example, sampling personnel go to a site for 10 d, conduct soil borings, and install groundwater wells. During this visit, soil and water samples are collected. The crew leaves the site for 2 months and later returns to collect another set of groundwater samples over a 3-d period. Two sampling events have occurred: a 10-d event and a 3-d event.

8.8.1.1 Trip blanks

Trip blanks are used to detect contamination by VOCs during sample shipping and handling. Trip blanks are 40-mL VOA vials of American Society of Testing and Materials Type I water that are generally filled in the laboratory, transported to the sampling site, and returned to the laboratory with VOA samples. Trip blanks are not to be opened in the field. One trip blank is to accompany each cooler containing VOA samples. Each trip blank is to be stored at the laboratory with associated samples and analyzed with those samples. Trip blanks are only analyzed for VOCs. Trip blanks are also sent with soil samples; however, because the blanks cannot be analyzed through the use of soil methods, they are analyzed and reported as water samples.

8.8.1.2 Field blanks

Field blanks are samples of the source water, and American Society of Testing and Materials Type I, that was used in the decontamination and cleaning of equipment used in sample collection. Type I water will be sampled per lot number. The decontamination source water will be sampled once in each event.

8.8.1.3 Equipment rinsates

Equipment rinsates are samples of American Society of Testing and Materials Type I water passed through decontaminated sampling equipment. They are used as a measure of decontamination process effectiveness. Equipment rinsates will be collected at a rate of 5% of the samples collected per matrix per event (i.e., 1 to 20 samples collected equals 1 equipment rinsate; 21 to 40 samples collected equals 2 equipment rinsates). Matrices for this purpose are considered to be soils, sediments, groundwater, and surface water. Equipment rinsates are analyzed for the same analytes as the associated samples. Equipment rinsates are reported as water samples.

Representative numbers of equipment rinsate samples will be sent to the mobile laboratory and the fixed analytical laboratory. Each laboratory will receive equipment rinsates in proportion to the number of routine samples received.

8.8.1.4 Field duplicates

Additional samples may be taken near the field sample co-located to determine total measurement error variance. Field duplicates for water samples will be collected simultaneously. Samples submitted for VOA will not be homogenized. Field duplicates will be taken at the rate of one for every 20 samples.

8.8.2 Laboratory QC Procedures

To ensure data quality, laboratory quality control samples, data review, and data validation are used to measure the quality of the sampling and analysis process. A quality control operation or component is only useful if it can be measured or documented. The components of analytical QC that are defined in this section are related to the analytical batch or to those samples that are analyzed together with the same method sequence, the same lots of reagents, and the manipulations common to each sample within the same time period or in continuous sequential time periods. Samples in each batch should be of similar composition. All QC data and records specified in this section will be reported as part of the data deliverables package. These procedures will be performed at least once with each analytical batch with a minimum of 1 per 20 samples.

The types of laboratory QC samples are method blanks, LCSs (blank spikes), MSs, MSDs, and duplicates. These samples are defined in the following sections. More detailed ER Division requirements are found in *Requirements for Quality Control of Analytical Data* (DOE 1990a).

8.8.2.1 Method blanks

Method blanks are used to determine the existence and magnitude of possible contamination encountered during sample preparation and analysis. They consist of American Society of Testing and Materials Type II water for water samples and clean sand or soil for soil samples if available. They are carried through the entire analytical procedure with the samples. At least one method blank must be prepared with each batch of samples. A batch cannot exceed 20 samples and is determined by the number of samples of similar matrix that can be processed simultaneously through the entire preparation process. Analysis is performed for the same analytes as those associated with the respective environmental samples.

8.8.2.2 Matrix spikes/matrix spike duplicates

An MS is an aliquot of a sample spiked with known quantities of analytes and subjected to the entire analytical procedure. It is used to indicate the appropriateness of the method for the matrix by measuring recovery. An MSD is a second aliquot of the same sample as the MS with known quantities of compounds added. The purpose of the MSD, when compared to the MS, is to determine method precision. Precision is the measure of the reproducibility of a set of replicate results among themselves or the agreement among repeat observations made under the same conditions. MSs and MSDs will be performed per 20 samples of similar matrix.

8.8.2.3 Duplicate samples

Duplicate samples are identical splits of individual samples that are analyzed by the laboratory to test for method reproducibility or precision. These samples are split in the laboratory during sample preparation. Duplicates are used mainly for inorganic analysis, while MSDs are used for organics. Duplicates will be performed per 20 samples of similar matrix.

8.8.2.4 Surrogate spikes

A surrogate spike is prepared by adding a pure compound to a sample before extraction. The compound in the surrogate spike should be of a similar type to that being assayed in the sample. The purpose of a surrogate spike is to determine the efficiency of recovery of analytes in the sample preparation and analysis. The percent of recovery of the surrogate spike is then used to gauge the total accuracy of the analytical method for that sample.

8.8.2.5 Laboratory control sample

If specified in the SOW, the laboratory will use an LCS system to monitor sample preparation and analysis.

The laboratory will employ an LCS program that monitors sample preparation and analysis. The LCS program will be monitored through the use of control charts. The purpose of this program is to demonstrate that the laboratory process for sample preparation and analysis is in control. This type of program is required for all ER Waste Management projects.

An LCS is a laboratory blank with a known amount of analyte(s) added or a commercially available standard consisting of known analyte concentrations representative of the contaminants to be determined. Analytes selected for spiking should be representative of the compound class. It is suggested that these criteria be followed.

- For VOCs and semivolatile organic compounds the method blank surrogate standards can be used as the LCS. All surrogates in the method blank must be monitored on control charts.
- For pesticides, at least two target compounds are added to a blank other than the method blank. These compounds are monitored on control charts.
- For PCBs, at least one target Aroclor is added to a blank other than the method blank. The compound is monitored on a control chart. If pesticides and PCBs are both being

determined, it is recommended that a single LCS with both pesticide and PCB spikes be used.

- For metals determined by ICP, an LCS must be monitored by means of control charts for at least three metals.
- For metals determined by atomic absorption (graphite furnace atomic absorption), furnace atomic absorption, and cold vapor atomic absorption), an LCS must be monitored by means of a control chart for each element.
- For wet chemical and radiochemical methods, an appropriate LCS for each method must be used (e.g., for cyanide, a control standard of sodium cyanide from a source other than that used for calibration may be spiked into water and analyzed with the water samples). The value for the control is monitored on the control chart.

The LCS matrix should be comparable to the sample matrix (i.e., analyze water control samples when water samples are analyzed).

In the LCS, problems encountered with matrix effects and sample nonhomogeneity will be minimized through the use of blank water and well characterized soil. This information, used in conjunction with sample MS recoveries, can aid in determining whether an out-of-control condition is due to laboratory problems or matrix problems.

Five percent of the samples will be LCSs, which may vary depending upon analytical method.

8.9 DATA VALIDATION PROCEDURES

8.9.1 Analytical Laboratory Data Verification and Validation

Data verification/validation is a systematic process for evaluating a body of data against a defined set of criteria to provide assurance that the data are adequate for their intended use. The evaluation process consists of data screening, range checking, auditing, verification, flagging, and quality evaluation. Specifically, data will be evaluated per derived DQO requirements. This will include review of sample custody forms, photocopied pages of laboratory notebooks, and data forms completed by laboratory staff, including: sample weights, dilutions, laboratory data reduction, instrument logs, and raw data, as necessary. The final task report will certify that the data have been verified/validated and qualified in accordance with the defined process.

Analytical data generated during and following field activities will be reviewed by the laboratory(s) and reported per the systematic requirements identified and defined in Sect. 8.3. The analytical laboratory(s) will not perform data verification/validation. Verification/validation is independent of the analytical laboratory data review.

Data validation will be consistent with the specifications as outlined in *Requirements for Quality Control of Analytical Data for the Environmental Restoration Program* (ES/ER/TM-16). All project data will be evaluated to ensure a complete, consistent, and usable project data set.

8.9.1.1 Data Validation Approach

Data validation will be performed to ensure that the precision and accuracy of the analytical data are adequate for their intended use. Because the greatest uncertainty in a measurement is often a result of the sampling process and inherent variability in the environmental media rather than the analytical measurement, validation will be performed only to the level necessary to minimize the potential of using false positive or false negative concentrations in the decision-making process (i.e., to assure accurate identification of detected versus non-detected compounds). This approach is consistent with the DQOs for the project and with the methods, listed in Sect. 1.8.7.1, for determining COCs and calculating risk.

A minimum of 10% of project data will be evaluated by data validation criteria consistent with the guidelines in TM-16 Level D validation. The 10% subset of the project data will be selected with respect to sample medium, analytical method, analytical laboratory, and project schedule (i.e., sample data packages will be selected throughout the project). The remaining 90% of the project data will be validated based on a reduced set of TM-16 Level C validation criteria as described in Sect. 8.9.1.2.

8.9.1.2 Data Validation Rationale

The data validation criteria listed below have been determined as critical in the evaluation of analytical data usability. These are a subset of the comprehensive TM-16 Level C data validation criteria that, based on analytical process knowledge, contribute significantly to the qualification and associated uncertainty of the reported results. Evaluation of this subset will allow identification of potential false positive or negative results. Because these criteria are associated with random rather than systematic error, they require evaluation throughout the analytical measurement process. They cannot be comprehensively determined by reviewing only a portion of the data. Consistent with the data quality requirements as defined in the DQOs and based on the above rationale, all project data and associated QC must be evaluated on these criteria and qualified as per the outcome of the review. The Level C criteria by which the data will be evaluated are discussed in Sect. 8.9.1.3.

The reduced Level C data validation criteria are as follows.

<i>Organics</i>	<i>Inorganics</i>	<i>Radiochemistry</i>
Holding times	Holding times	Holding times
Blanks	Blanks	Blanks
Surrogate recovery	Laboratory control sample	Laboratory control sample
Internal standards	Furnace atomic absorption QC	Sample specific chemical or tracer recovery
Calibration	Calibration	Calibration
Sample reanalysis		
Secondary dilutions		
Case narrative	Case narrative	Case narrative

While the criteria listed above are those that most impact the determination of data usability, it is also necessary to evaluate whether any systematic errors are occurring during the analytical process that could adversely affect the sample results. Therefore, some portion of the data must be evaluated against the additional criteria. Because the laboratory(s) will be reviewed, audited, and approved based on extensive selection criteria prior to providing analytical support to the project and because they will be subject to periodic audits during the project, selection of a minimum of 10% of the data packages for full TM-16 Level D validation will adequately identify any systematic problems. The 10% subset of data packages will be selected across media, methods, laboratories, and time to ensure an adequate sampling of the process. If a systematic problem is discovered in the 10% subset, additional data packages will be reviewed based on the problem criterion to determine the extent of and resolve the problem. This approach to validation will result in a minimum of 10% of the data packages being validated to TM-16 Level D criteria and the remaining 90% of the data packages being reviewed based on reduced TM-16 Level C criteria.

8.9.1.3 Reduced Level C Data Validation Criteria

Holding Times. Evaluation of holding times ascertains the validity of results based on the length of time from sample collection to sample preparation or sample analysis. Verification of sample preservation must be confirmed and accounted for in the evaluation of sample holding times. The evaluation of holding times is essential to establishing sample integrity and representativeness. Concerns regarding physical, chemical, or biochemical alteration of analyte concentrations can be eliminated or qualified through this evaluation.

Blanks. The assessment of blank analyses is performed to determine the existence and magnitude of contamination problems. The criteria for evaluation of blanks applies to any blank associated with the samples, including field, trip, equipment, and method blanks. Contamination during sampling or analysis, if not discovered, results in false positive data.

Surrogate Recovery. System monitoring compounds are added to every sample, blank, matrix spike, matrix spike duplicate (MS/MSD), and standard. They are used to evaluate extraction, cleanup, and analytical efficiency by measuring recovery on a sample-specific basis. Poor system performance as indicated by low surrogate recoveries is one of the most common reasons for data qualification. Evaluation of surrogate recovery is critical to the provision of reliable sample-specific analytical results.

Calibration. The purpose of initial and continuing calibration verification analyses is to verify the linear dynamic range and stability of instrument response. Relative instrument response is used to quantitate the analyte results. If the relative response factor is outside acceptable limits, the data quantification is uncertain and requires appropriate qualification.

Internal Standards. Internal standards are utilized to evaluate and compensate for sample-specific influences on the analyte quantification. They are evaluated to determine whether data require qualification due to excessive variation in acceptable internal standard quantitative or qualitative performance measures. For example, a decrease or increase in internal standard area counts for organics may reflect a change in sensitivity which can be attributed to the sample matrix. Because quantitative determination of analytes is based on the use of internal standards, evaluation is critical to the provision of reliable analytical results.

Sample Reanalysis. When instrument performance-monitoring standards indicate an analysis out-of-control, the laboratory is required to reanalyze the sample. If the reanalysis does not solve the problem (i.e., surrogate compound recoveries are outside the limits for both analyses), the laboratory is required to submit data from both analyses. An independent review is required to determine which is the appropriate sample result.

Secondary Dilutions. When the concentration of any analyte in any sample exceeds the initial calibration range, a new aliquot of that sample must be diluted and reanalyzed. The laboratory is required to report data from both analyses. When this occurs, an independent review of the data is required to determine the appropriate results to be used for that sample. An evaluation of each analyte exceeding the calibration range must be made, including a review of the dilution analysis performed. Results chosen in this situation may be a combination of both the original results (i.e., analytes within initial calibration range) and the secondary dilution results.

Laboratory Control Samples. The laboratory control sample (LCS) serves as a monitor of the overall performance of the analytical process, including sample preparation, for a given set of samples. Evaluation of this standard provides confidence in or allows qualification of results based on a measurement of process control during each sample analysis.

Furnace Atomic Absorption QC. Duplicate injections and furnace post-digestion spikes are evaluated to establish precision and accuracy of individual analytical determinations. Because of the nature of the furnace atomic absorption technique and because of the detailed decision tree and analysis scheme required for quantitation of the elements, evaluation of the QC is critical to assuring reliable analytical results.

Sample Specific Chemical or Tracer Recovery. Laboratory performance on individual samples subject to chemical process and separation is established by means of spiking with tracer quantities of other radioisotopes of the same element or carrier quantities of an inactive isotope of the same or a chemically similar element. This process is analogous to surrogate or internal standard recovery, dependent on the analyte and method being evaluated, and is a common reason for data qualification.

8.9.2 Project Data Quality Assessment

The data quality indicator parameters precision, accuracy, representativeness, completeness, and comparability (PARCC) will be used to evaluate data quality and quantity. Reviewers are responsible for ensuring that data reduction calculations follow appropriate data calculation procedures, are documented, and are checked by qualified personnel. Written reports, including reduced and summarized data, will include the raw data in appendixes. Specific calculations used for data reduction will also be included.

Both qualitative and quantitative criteria are used as indicators of the quality of the data. In determining data usability, especially in the decision-making process, the integrity and authenticity of the data must be evaluated and the analytical uncertainty must be determined. Parameters used to assess data quality for this investigation are precision, accuracy, representativeness, comparability, completeness, and sensitivity. These parameters will be evaluated for the entire project data set.

The following procedures will be used to assess data precision, accuracy, and completeness. These equations apply to both field and laboratory measurements.

8.9.2.1 Precision

Precision is defined as the reproducibility or degree of agreement between duplicate measurements under a given set of conditions. The closer the measurements approach each other, the more precise the measurement. The level of precision is determined by calculating the relative percent difference (RPD) between the two measurements using the following equation:

$$RPD = \frac{|S - D|}{(S + D)/2} \times 100\%$$

where:

S = analyte concentration of the original sample,
D = analyte concentration of the duplicate sample.

Precision is determined using MS/MSD analyses. The laboratory will select one sample in 20 to split into three aliquots. The first aliquot will be analyzed routinely for the parameters of interest; the other two aliquots will be spiked with known quantities of the parameters of interest before analysis. The RPD between the two spike results will be calculated and used as an indication of the precision of the analyses performed.

When the analyte of interest provides a measurable quantity without spiking, precision determinations can be obtained through duplicate analysis comparisons. Field duplicate samples are evaluated as an indication of overall precision. This evaluation measures variability introduced during both field sampling and laboratory analysis.

Analytical precision goals for the proposed radiochemical and inorganic parameters will be ≈ 20 RPD. Goals for analytical precision for organic parameters will be those identified in the analytical methods for specific analytes. In general, these range up to 25 RPD (volatile organics) and 50 RPD (semivolatile organics) in soils. If analytical precision goals are exceeded, a determination will be made through the data validation process relative to the useability of that information.

8.9.2.2 Accuracy

Accuracy is defined as the degree of difference between measured values and the true values. Sampling accuracy will be maximized by adhering to the QAPjP. All procedures used will be documented as SOPs or Analytical Operating Procedures (AOPs). Analytical measurement accuracy is evaluated through implementation of analyte matrix spike or standard reference material determinations. The following equation will be used to calculate percent recovery (%R):

$$\%R = \frac{A_r - A_o}{A_r} \times 100\%$$

where:

A_t = total compound or element concentration detected in the spiked sample,
 A_o = concentration of the compound or element detected in the unspiked sample, and
 A_f = concentration of the compound or element added to the sample.

For situations in which a standard reference material (SRM) is used rather than or in addition to MS analyses, the following equation will be used:

$$\%R = \frac{C_m}{C_{srn}} \times 100\%$$

where:

C_m = measured concentrate of SRM and
 C_{srn} = actual concentration of SRM.

A recovery of 100% is equivalent to 100% accuracy. Values varying from 100% may indicate a sample matrix effect and a false reading. A periodic program of sample spiking is required (e.g., 1 MS and 1 MSD per 20 samples).

The proposed radiochemical and inorganic accuracy goals for this investigation will be 80 to 120% as expressed in recovery of known analytical spikes into the sample media. Analytical accuracy goals for organic parameters will be those identified in the CLP analytical methods for specific analytes.

In general, volatile organic compound spike recoveries range between 60% and 150% in soils, while semivolatile organic compound spike recoveries range between 20% and 140%. In the event analytical accuracy goals are exceeded, a determination will be made through the data validation process relative to the usability of those data.

8.9.2.3 Representativeness

Representativeness expresses the relative degree to which the data depict the characteristics of a population, parameter, sampling point, process condition, or environmental condition. The objective of this study is to accurately represent the chemical concentrations of target analytes.

Representative samples for this investigation will be acquired through implementation of approved sampling and analytical procedures that will generate data representative of the sampling point location. Sampling procedures are designed to minimally impact the sample obtained, so that conditions representative of the sampling location will be maintained. Analytical methods will be selected that most accurately represent the true concentration of the parameter of interest. The accumulation of QC procedures and information (i.e., RPD values, blank QC concentrations, MS percent recoveries, etc.) employed for a given analysis combine to exhibit the representativeness of the data generated.

The goal for representative sample data will therefore be met through the proper documentation of field and analytical protocols. If these procedures and methods cannot be implemented, the appropriate corrective action documentation will encompass the impact on

the representativeness of the information. When review of the data and documentation determines that the data are nonrepresentative, the information will be qualified in its use or will not be used by the project. Statistical formulas may be implemented to further define sample population, parameter, or process characteristics relative to representativeness.

8.9.2.4 Completeness

Completeness is the measure of the amount of data obtained from a measurement process that achieves the project goals compared to the amount of data planned to be obtained by the project.

The following are three measures of completeness determined for this investigation.

Sample Collection:

$$\text{Completeness} = \frac{\text{Number of Sample Points Sampled}}{\text{Number of Sample Points Planned}} \times 100\%$$

Field Measurements:

$$\text{Completeness} = \frac{\text{Number of Valid Field Measurements Made}}{\text{Number of Field Measurements Planned}} \times 100\%$$

Laboratory Analysis:

$$\text{Completeness} = \frac{\text{Number of Valid Laboratory Measurements Made}}{\text{Number of Laboratory Measurements Planned}} \times 100\%$$

Completeness goals established for this project for sample collection and field measurements are 95%. If these goals are not met, the necessity for resampling will be determined on a case-by-case basis. The completeness goal for laboratory analysis for this project is 90% overall. For critical data points, which consist of background sample locations, the completeness goal for sample collection, field measurements, and laboratory analysis is 100%. If review of the data and documentation determines that the data are incomplete, the impact relative to the project objectives will be assessed and documented.

To complete the intent of this project, all data will be validated against the reduced Level C criteria, with a minimum of 10% of the data validation being completed at Level D, to determine its useability. For determination of completeness in this project, all data not flagged as rejected by the validation process will be considered valid.

Additional formulas and statistically generated data may be utilized to further define data set completeness.

8.9.2.5 Comparability

Comparability expresses the confidence with which one data set can be compared with another. Comparability of the data generated in this investigation will be obtained through the implementation of the identified protocols for sampling and analysis of samples. Use of traceable reference materials as laboratory standards, expression of results in standard

concentration units, and successful participation by the laboratories in external performance evaluation programs will enable the data produced through this investigation to be compared with future data sets, if required.

8.10 AUDITS AND SURVEILLANCES

Audits are performed to review and evaluate the adequacy of field and laboratory performance and to ascertain whether the QA/QC plan is being completely and uniformly implemented. Planned and scheduled audit will be performed to verify compliance with all aspects of the QA program and to determine the program's effectiveness. Energy Systems audits will be conducted in accordance with ESS-QA-18.0 and ESP-QA-18.1. An SAIC audit will be conducted in accordance with SAIC QAAP 18.1 and will be performed by personnel who do not have direct responsibility for performing the activities being audited.

Surveillance activities include monitoring and observing documents and work activities to provide an effective real-time means of evaluating the adequacy and effectiveness of methods for achieving quality and for assessing the quality of final results. Energy Systems surveillances will follow ER/Y-I1600.

At least one surveillance will be scheduled to include observation of all procedures implemented.

Field surveillances will be conducted to assess technical and quality activities including sampling, instrumentation use, equipment decontamination, sample shipment, and field record maintenance.

8.11 CORRECTIVE ACTION

Energy Systems corrective actions to audit/surveillance findings and nonconformances will be managed in accordance with ESS-QA-16.0, 16.1, 16.2, 16.3, 16.4 and ESS-QA-15.0 and 15.1. SAIC corrective actions to audit/surveillance findings and nonconformances will be managed in accordance with SAIC QAAP 15.1 and 16.1. The Y-12 ER Project Manager will be notified when a nonconformance is documented and furnished a copy as soon as possible. Copies of nonconformances and their dispositions will be forwarded to the Y-12 ER Project Manager for placement in the Y-12 ER Records Center (DMC). Nonconformance Reports issued as a result of an audit or surveillance will identify the root cause of the problem.

8.12 QA REPORTS TO MANAGEMENT

All QA records concerning the project (e.g., internal and external correspondence, sampling and analysis plan, QAPjP, field logbooks and forms, chain-of-custody forms, data packages, audit reports, surveillance reports, nonconformance reports, corrective action reports, etc.) will be submitted to the Central Records Facility for dual storage and retrieval.

Table 8.6 is the ER Records Checklist to indicate the BCU OUX project records to be submitted to the ER DMC. Records concerning the project will be forwarded to the Y-12 ER Project Manager for placement in the Y-12 ER DMC during the project. A complete copy of the project file will be submitted to the Y-12 ER Project Manager at the end of the period of performance.

Table 8.6. Bear Creek OU 1 field sampling environmental restoration checklist

**Bear Creek OU 1 Field Sampling
Environmental Restoration Records Checklist**

3/23/94

Item*	Reqd	QA
Reviews/Audits/Verifications		
Operational Readiness Review	X	X
Audits/Surveillances/Corrective Actions	X	X
Verification Matrix Review		X
Davis-Bacon Review	X	X
Legal Review		X
Inspection Reports		X
Classification Review		X
DOT/Shipping Documents		X

Design

Constructibility Assessment		X
Constructibility Review		X
Design to Cost Review		X
Review		X
Professional Engineer Stamping of Design		X
Design Criteria/SOW		X
Design Reviews & Design Review Record		X
Analyses and Calculations		X
Shop (Vendor) Drawings		X
Design Specifications		X
Design/Installation Drawings		X
Design Change Requests		X
Design Deviation Reports		X
As-Built (As-Constructed) Drawings		X
Test Plans		X

Procurement

Lease/Buy Decision	X	X
Proposal Evaluations	X	X
Procurement Documents	X	X
Technical Reviews and Documentation	X	X
Receiving Inspections of Procured Items	X	X
Vendor Certifications	X	X

Construction

Permits		X
Construction Job Instructions		X
Cost Estimates and Control		X
Schedules and Control		X
Training		X
Certified for Construction Package		X
Construction Management Plan		X
Health and Safety Review		X
Testing and Test Reports		X
Acceptance Reports		X
Nonconformance Report Status		X

Item*	Reqd	QA
Surveys		
Site Survey		X
Industrial Hygiene Survey		X
Health Physics Survey (Assessment)		X
Topographical Survey		X
Fire Protection Survey		X
Geotechnical Survey		X

Reports and Documentation

System Requirements Document		X
Field Change Orders	X	X
Justification for Continued Ops. Approval		X
Closure Summary Report		X
Supporting Technical Documents (Eg., maps)		X
Data Summary Sheets		X
Field Data Sheets	X	X
Field Logbooks	X	X
Field Equip. Calibration/Mntnce Records		X
Hydrogeologic & Geophysical Test Data		X
Boring and Well Construction Logs		X
Analytical Data Packages	X	X
Computer Aided Mfg. (CAM) Data Packages		X
Site Background Documents		X
Topo., Aerial, and Hydrogeologic Maps		X
Modeling/Software Control Documentation		X
Correspondence addressing Performance	X	X
Comments and Responses	X	X
Photographs		
Operable Unit Bibliography		
Videos		

Training Records

Technical Staff Resumes	X	X
Training Requirements		X
Individual Training Summaries		
Training Records	X	X
Record of Qualification		X

Design Const. Reports & Docs.

Design Basis Report		X
Project Validation		X
Periodic Proj Reviews & Status Reports		X
Construction Completion Report		X
Construction Proj. Requirements Summaries		X
Constr. Eng. (Title III) Test/Insp Reports		X
As-constructed Verification (Title III Insp)		X

*Shaded Items indicate ADS/program level documents

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Table 8.6 (continued)
Bear Creek OU 1 Field Sampling
Environmental Restoration Records Checklist

3/23/94

Item*	Reqd	QA	Item*	Reqd	QA
Financial/Scheduling			Plans and Planning Documents		
Financial Plan			Federal Facility Agreement		X
Engineering Service Order		X	Primary Document (RFI/RIWP -ROD)	X	X
Activity Data Sheets (ADS)			FFA Secondary Docs. (RFI/RIWP-ROD)		X
Current Year Work Plans		X	Public Information Fact Sheets		X
Cost Estimating Backup			Public Legal Notices/Newspaper Articles		
Baseline		X	Public Comments		X
Baseline Change Control		X	Procedures (Modified)		X
Project Schedules		X	Best Management Practices Plan		X
Progress Tracking System Report			Closure Plan		X
Cost Performance Reports			Community Relations Plan		X
Monthly Budget Review			Configuration Management Plan		X
Monthly/Quarterly/Annual Progress Reports			Contingency Plan		X
DOE Directive		X	Data Management Plans (Program)		X
Subcontractor Progress Reports			Data Management Plans (Project)		X
NEPA Documentation			Field Sampling Plans	X	X
Archeological Survey		X	Health and Safety Plan	X	X
Critical Habitat Survey		X	Mobilization Plan		X
Wetlands Survey		X	Program QA Plan		X
Threatened/Endangered Animal Species Survey		X	Project Management Plan		X
Site Historical Preservation Office Survey		X	Project QA Plan	X	X
Environmental Checklist	X	X	Records Management Plan		X
Request for Categorical Exclusion (CX)	X	X	Safety Action Plan		X
Response (eg., CX Approval)	X		Sampling and Analysis Plan	X	X
Environmental Assessment (EA)		X	Security Plan		X
Action Description Memorandum (ADM)		X	Systems Engineering Plan		X
Analyses/Assessment/Evaluation			Waste Management Plan	X	X
Risk/Problem Analysis		X	Permits		
Preliminary & Final Safety Analysis		X	Air Permits (PSD Permit/Clean Air Act)		X
Safeguards & Security Evaluation		X	Aquatic Resource Alternations Permit		X
Life Cycle Cost Analysis			Asbestos Removal Permit		X
Maintenance Assessment		X	Burning Permit		
Self-Assessments		X	Const. Safety-Work Permit EP-D-08		X
Value Engineering Study		X	Corp of Engrs 404 Permit (Corp.-Major)		X
Service Subcon. S & H Proc (PC 164)	X	X	Excavation/Penetration Permit EP D-06	X	X
Site Location Selection		X	NPDES Permit Sec 402 (Outfall)		X
Contracts			RCRA Permit (Permit-By-Rule)		X
Statement of Work (SOW)	X	X	TSCA Permit (PCBs, Asbestos, Freon)		X
SOW Amendments	X	X	Waste Disposal Permit 2109	X	X
Cost Estimate/Proposal Info	X	X			
Technical Cost/Business Proposal Info	X				
Monthly Cost Information	X				
Invoices	X	X			

*Shaded Items indicate ADS/program level documents

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APPROVALS

HEALTH AND SAFETY PLAN
FOR THE REMEDIAL INVESTIGATION OF THE
BEAR CREEK OPERABLE UNIT 1
AT THE OAK RIDGE Y-12 PLANT

DOE/OR/01-1161/V1&D2

March 1994

3/21/94

Technical Support Contractor Program Manager Date

3/21/94

Technical Support Contractor Field Team Leader Date

3/21/94

Site Health and Safety Officer Date

3/23/94

Energy Systems ER Program Manager Date

3/24/94

Energy Systems ER Site Program Manager Date

3/28/94

DOE ER Site Program Manager Date

9. HEALTH AND SAFETY PLAN

It is Energy Systems policy to provide all Energy Systems employees and subcontractor personnel with the information and procedures necessary to conduct investigative environmental field activities in a safe and healthful manner. This plan applies to all sampling being conducted at the Y-12 Plant BCV OU 1 sites that could expose subcontractor and Energy Systems personnel to hazardous substances. This health and safety plan does not attempt to provide specific requirements for all conceivable situations that might be encountered. It does provide general requirements that are applicable to all sites. Specific health and safety requirements are identified in the Site-Specific Health and Safety Checklists (Appendixes G and H). DOE Orders that are pertinent to health and safety include DOE Order 5480.3 Safety Regulations for the Packaging and Transportation of Hazardous Materials, Hazardous Substances, and Hazardous Wastes; DOE Order 5482.1B Environment Safety and Health Appraisal Program; DOE Order 5480.1A Contractor Occupational Medical Program.

Energy Systems emphasizes the importance of conducting all activities in a safe and healthful manner and requires all personnel involved with sampling activities to receive safety and health training. Safety and health training, including site-specific training (as outlined in Sect. 9.3, Field Personnel Training Requirements), is required under OSHA standards to prepare site personnel to recognize and respond to site hazards.

It is impossible to anticipate all specific safety and health hazards; therefore, all personnel are expected to exercise common sense and good judgment in their approach to specific situations. All personnel are encouraged to bring to the attention of the Site Health and Safety Officer (SHSO) or the Project Health and Safety Officer (PHSO) any unsafe or hazardous conditions observed at their respective sites. Unsafe conditions are to be corrected by the subcontractor or Energy Systems as soon as possible.

9.1 SITE HEALTH AND SAFETY PERSONNEL

The PHSO will coordinate all health and safety corrective actions reported to the ER Project Manager. The PHSO will advise Energy Systems personnel and the SHSO of anticipated project activities or conditions that could adversely affect the health or safety of project personnel. The PHSO is authorized to take whatever actions are determined necessary to protect all site personnel and the environment during site investigations. The PHSO will perform the following duties:

- Determine the initial level of protective clothing needed at each site.
- Review the Health and Safety Plan.
- Provide guidance to the Field Operations Manager and SHSO, as needed.

The SHSO will coordinate health and safety activities in the field. The SHSO will perform the following duties:

- Verify that protective measures in use (including protective clothing) are adequate for current conditions.

- Notify the Field Operations Manager of any unexpected conditions or hazards.
- Notify the Field Operations Manager of all instances of noncompliance with this health and safety plan or with Y-12 Plant safety procedures.
- Provide site personnel with information on site hazards.
- Maintain a list of Y-12 Plant and subcontractor phone numbers.
- Ensure that all site personnel know how to obtain emergency aid from the Y-12 Plant Shift Superintendent's office.
- Perform on-site air monitoring.
- Stop work if an unsafe situation occurs.
- Conduct site-specific health and safety briefings.
- Maintain copies of Material Safety Data Sheets (MSDSs) for hazardous materials brought on site.

The Field Operations Manager will be responsible for implementing this health and safety plan. Specific health- and safety-related responsibilities include the following:

- Coordinating on-site operations
- Minimizing the number of personnel in the exclusion zone
- Stopping work if any unsafe condition exists
- Coordinating with SHSO and PHSO to ensure safe operations
- Obtaining excavation/penetration permits

9.2 SAFETY RISK ANALYSIS

9.2.1 BCV OU 1 Hazards

BCV OU 1 was used for the treatment, storage, or disposal of hazardous wastes and metals. All of the sites are considered hazardous, though the Boneyard/Burnyard and Burial Grounds pose the greatest threats because of the properties of the chemicals and contaminants that were known to have been placed or that may have been placed in these areas.

Hazards directly related to the Boneyard/Burnyard include potential exposure to high levels of radiation, uranium, organic compounds, acids, and metals in soil and groundwater in the area of trenches that were filled and covered with soil.

Hazards associated with the Burial Grounds include all the hazards listed for the Boneyard/Burnyard. BG-B and the Walk-in Pits contain picric acid, uranium, and thorium saw-fines, which possess pyrophoric qualities. No foot or vehicle traffic will be permitted in these areas.

The S-3 Site, Oil Landfarm, and ORPs all present lesser hazards. Oils, PCBs, heavy metals, depleted uranium, acids, and other wastes were disposed of at these sites.

9.2.2 Physical Hazards

9.2.2.1 Utilities

Before the start of sampling, drilling, or excavation, all underground utility lines, other underground structures, and above-ground utility lines must be clearly delineated in the selected work area. The SHSO will ensure that the proper excavation/penetration permit, including utility review and survey, has been acquired before sampling, drilling, or excavation activities begin.

9.2.2.2 Heavy machinery

During activities involving heavy equipment (drill rigs), care will be taken to avoid overhead equipment, falling or flying objects, or being caught between moving and idle pieces of machinery. At a minimum, the general protective clothing (Sect. 9.4) will be worn within the restricted access area around the sampling site as defined in Sect. 9.9. Lifting will comply with the DOE Hoisting and Rigging Manual.

9.2.2.3 Traffic

Traffic cones, flagging, or barriers appropriate for the traffic hazard will be used to cordon off the work area, if the operations are to be conducted in areas of vehicular traffic.

9.2.2.4 Weather

Serious hazards can result from adverse weather; therefore, the SHSO will be authorized to discontinue sampling or other field activity because of severe and threatening weather conditions (e.g., lightning, strong wind, heavy rain, temperature extremes). In hot weather, cool fluids and frequent breaks will be provided to help prevent heat exhaustion/stroke.

9.2.3 Radiation Hazards

Alpha, beta, and gamma radiation each pose different hazards. Alpha particles do not penetrate the skin but can pose serious internal exposure problems if the emitters are inhaled or ingested. Beta particles can penetrate the human body and are, therefore, internal radiation hazards. Beta particles can also cause skin burns at high doses. Gamma rays are a highly penetrating form of radiation and present an external radiation exposure hazard. Protective clothing, good personal hygiene, and decontamination afford good protection against alpha and beta radiation but will not provide protection against gamma radiation.

Radiation detection instruments are essential for detecting radiological contamination and determining external radiation exposure levels. Exposure measurements will be used to determine the time, distance, and shielding alternatives that will be considered to protect personnel from external radiation exposure.

All personnel will be issued a thermoluminescent dosimeter (TLD) upon entering the Y-12 Plant site. During sampling activities, downhole equipment and material will be surveyed when removed from the sample point.

The Y-12 Plant Health Physics Department will determine radioactive dose limits and surveying requirements at sites within controlled areas. Their determination will be placed on the field checklist to be used by field personnel. The Health Physics Department must be notified before work commences in an area of potential radiation exposure.

9.2.4 Chemical Hazards

Because of the nature of the sampling activities, workers may encounter some of the following chemicals:

- Metals, including combustible metals
- Inorganic and organic compounds (acids, salts, polar and nonpolar solvents, halogenated hydrocarbons)
- Oils (hydraulic fluids, PCB-contaminated electrical equipment)
- Uranium and uranium-contaminated materials
- Thorium-contaminated materials
- Beryllium and beryllium-contaminated materials
- Laboratory wastes
- Plutonium wastes

Workers may encounter hazardous contaminants during sampling activities in a matrix of groundwater, soil, or rock cuttings, or as airborne contaminants. The most likely avenues of chemical exposure are direct contact with the waste or contaminated material (allowing possible dermal adsorption or ingestion) and inhalation of gases, vapors, or contaminated dust. The possibility of chemical exposure will be mitigated through the use of appropriate protective clothing and respirators. The level of protection necessary will be determined by on-site monitoring.

9.3 FIELD PERSONNEL TRAINING REQUIREMENTS

Before being permitted access to the site, each worker will be required to provide documentation of successful completion of the Energy Systems General Employee Training, Energy Systems Radiation Worker II training, 40-h Hazardous Waste Operations training, and an annual 8-h refresher training. Each worker must obtain medical approval to perform the intended work. Respirator users must provide documentation of a respirator fit test as required by 29 *CFR* 1910.134. OSHA 8-h supervisor training is required for the site supervisors. Additionally, before activities begin at each site, workers will be required to attend a site-specific health and safety briefing conducted by the SHSO. This briefing will fulfill the site-specific requirements of 29 *CFR* 1910.1200, the Hazard Communication Standard. Certain sites may require health and safety briefings by Y-12 Plant health physics personnel or industrial hygiene personnel. Site-specific health and safety briefings will be coordinated through the SHSO or Y-12 Plant ER Project Manager.

9.4 PERSONAL PROTECTIVE APPAREL AND EQUIPMENT

The SHSO will determine and specify the types of personal protective equipment (PPE) to be worn during site investigations based on specific hazards identified during preliminary site hazard evaluation and site characterization. All workers must be familiar with the use of protective clothing and its limitations. This includes disposable coveralls, gloves/inner gloves, goggles/safety glasses, hard hats, and safety shoes.

9.4.1 Levels of Personal Protection

9.4.1.1 Level A

Level A protection should be provided when either (or both) of the following conditions exist:

- A chemical has been identified on site that requires the highest level of respiratory and dermal protection because of its toxicity and high concentration.
- Visible emissions or other conditions indicate a high potential for exposure to respiratory and dermal hazards. Level A protection will be worn until the absence of substances requiring such protection is determined by air monitoring or other site characterization methods.

Level A protection includes a supplied-air respirator (approved by the Mine Safety and Health Administration and the National Institute for Occupational Safety and Health) worn with a fully encapsulating, chemical-resistant suit. The respirator may be a positive pressure-demand self-contained breathing apparatus or a positive pressure-demand airline respirator [with a 5-min escape bottle for Immediately Dangerous to Life or Health (IDLH) or potentially IDLH atmospheres].

9.4.1.2 Level B

Level B protection should be provided when any of the following conditions exist.

- A type and atmospheric concentration of contaminants has been identified that requires a high level of respiratory protection but not complete dermal protection.
- Airborne contaminant concentrations are high or suspected to be high, precluding use of air-purifying respirators.
- The atmosphere contains less than 19.5% oxygen.

Level B protection includes a Mine Safety and Health Administration- or National Institute for Occupational Safety and Health-approved supplied-air respirator. The respirator may be a positive pressure-demand self-contained breathing apparatus or a positive pressure-demand airline respirator (with an escape bottle for IDLH or potentially IDLH atmospheres).

9.4.1.3 Level C

Level C protection should be provided when all of the following conditions exist.

- Oxygen concentrations are not less than 19.5% by volume.
- All potential atmospheric contaminants and concentration ranges have been identified through monitoring.
- Air concentrations of identified substances will be reduced by the respirator to below the substance's Threshold Limit Value (TLV), and the contaminant concentrations in air do not exceed the maximum protection level provided by the canister or cartridge.
- Atmospheric contaminant concentrations do not exceed IDLH levels.
- The airborne contaminants known to exist at the site would give sufficient warning, through odor, taste, or irritation, to alert the user that the canister/cartridge sorbent is saturated and/or is not functioning properly.
- The potential for encountering unknown airborne contaminants or high concentrations of known site contaminants is low.

Level C protection includes a full-face, canister- or cartridge-equipped air-purifying respirator (Mine Safety and Health Administration- or National Institute for Occupational Safety and Health-approved). Half-face respirators may be substituted for full-face respirators if conditions warrant and if approved by the SHSO.

The PHSO will determine the proper cartridge for use with Level C respiratory protection based on the contaminants known or suspected to occur at the site.

9.4.1.4 Level D

Level D personal protection provides no respiratory protection; therefore, it should be used only when no potential for overexposure to airborne contaminants exists.

9.4.2 Protective Clothing

The PHSO will be responsible for selecting initial personal protective clothing. The selection will be based on a comparison of the specific materials and clothing ensemble limitations with the potential hazards expected to occur at the site. The PHSO will select the protective clothing that will be most impermeable to the mixture of the specific chemicals known or suspected to exist at the site.

In situations where maximum skin protection is not required, the protective clothing requirements will be reduced to levels appropriate for the contaminant exposure hazards during specific activities. The general protective clothing requirements include steel-toed boots, work clothing, gloves, and safety glasses or goggles. Protective clothing requirements will be upgraded or downgraded according to the actual known or suspected exposure hazards.

The PHSO and the SHSO will identify specific site conditions or project activities that influence the selection of protective clothing and the development of special use procedures.

These factors may include heat stress, environments with unknown exposure hazards, low-temperature environments, and differences in work activities within the same contaminated work zone. These factors will always be taken into account in selecting protective clothing. For example, protective clothing for site operations in high-temperature environments should have the lowest practical thermal insulation properties to allow body cooling. Long cotton underwear should be worn under protective clothing in warm environments to absorb moisture and to protect the skin from contact with heat-absorbing protective clothing. Insulating layers of clothing should be worn under protective clothing to protect against low-temperature environments.

Varying levels of protective clothing may be selected for workers performing different activities within the same work zone, if their activities involve different exposure hazards. For example, drillers who are frequently in contact with contaminated soil may be required to wear both inner and outer chemical-resistant gloves and polyvinyl chloride (PVC) chemical-resistant protective suits; however, soil samplers and geologists who have limited contact with contaminated soils may be required to wear only inner chemical-resistant gloves and disposable lightweight chemical-resistant suits. Protective clothing requirements for unknown environments will be made by the PHSO at the site on a case-by-case basis with the assistance of the Y-12 Plant ER Project Manager and the SHSO. Air monitoring will be performed in unknown environments to ensure that the protective ensembles are appropriate for the type and concentration of contaminants.

After daily fieldwork is completed, outer protective clothing will be removed. Disposable clothing will be collected and packaged in garbage bags on site and disposed of according to the waste management plan (Chap. 10). Reusable boots and gloves will be decontaminated and left on site for future use.

The SHSO will ensure that all field workers have the following items available for their use:

- Gloves: inner (surgical type, disposable) with outer (chemical-protective, neoprene or nitrile) *or* inner (chemical-protective, neoprene or nitrile) with outer (heavy-duty work gloves)
- Protective clothing: Tyvek™ (regular or coated, full length with zipper)
- Footwear: steel-toed boots
- Ear protection: disposable foam earplugs *or* industrial earmuffs
- Eye protection: safety glasses with side shields *or* safety goggles to fit over prescription glasses (to be worn at all times when sampling)
- TLDs: TLDs are to be worn at all times, as required by Energy Systems
- Head/face protection: hard hats (when power equipment is in use) *or* hard hats with face shield attached (used when decontaminating equipment)
- Respirator: full-face mask air-purifying respirator equipped with cartridges for protection against organic vapors and acid gases, dusts and mists, and radionuclides
- First-aid kits
- Portable emergency eyewash

- Appropriate fire extinguisher
- Spill/cleanup materials

9.5 PERSONAL HYGIENE

A personal hygiene program helps to ensure that no workers are exposed to harmful levels of hazardous materials. The following rules will be enforced:

- Workers will consume food only after washing hands and only in a safe area removed from the work zone.
- No eating, drinking, using tobacco, chewing gum, or applying cosmetics will be permitted in the cordoned working area around the Contamination Reduction Zone or Exclusion Zone.

9.6 ADDITIONAL EQUIPMENT

In addition to PPE, other equipment will be maintained on site. This equipment will include the following:

- An appropriate fire fighting device, such as a 20- or 30-lb ABC fire extinguisher
- Plastic bags for trash, contaminated equipment, and clothing disposal
- A supply of clean water that is both potable and suitable for equipment decontamination

9.7 MEDICAL SURVEILLANCE

All hazardous waste and hazardous materials (HAZMAT) operations personnel will participate in a medical surveillance program.

The medical surveillance program will consist of the following required physical examinations:

- Preassignment
- Periodic [at least once every 12 months for each employee covered unless the attending physician believes a longer interval (not to exceed 2 years) is appropriate]
- Postemployment
- Postaccident

Medical examinations for hazardous waste and HAZMAT operations personnel will include the following:

- Medical and work history
- Physical examination

- Audiogram
- Blood chemistry (SMAC 24)
- Complete blood count
- Chest X-ray (posterior, anterior, and lateral), initially and every 5 years thereafter
- Electrocardiogram (optional)
- Pulmonary function test (spirometry)
- Urinalysis
- Visual acuity
- Other tests as directed by the examining physician

If the SHSO suspects that an individual has been exposed to a significant level of hazardous materials, he/she will notify the PHSO and/or the Y-12 Plant ER Project Manager and request that either Y-12 Plant medical personnel or the subcontractor's physician examine the individual to determine the impact of the exposure.

9.8 MONITORING EQUIPMENT

The SHSO will ensure that all necessary monitoring equipment is provided, functioning properly, and currently calibrated before starting the investigation. Instrumentation will be used only by the SHSO or qualified personnel who have had prior experience or training in instrument calibration, operation, and care.

On-site monitoring of the immediate area (breathing zone) during all sampling activities will be essentially the same. Site personnel and sampling locations will be monitored with an HNu or OVA, Jerome mercury vapor analyzer, and radiation meters.

To assist in evaluating potential site hazards, the SHSO will use one or more of the following (or an equivalent):

- Foxboro-Century OVA Model 128 and/or HNu Model 33 Photoionization Detector (PID) or equivalent
- Ludlum Model 3 portable radiation survey meter with a Ludlum Model 44-9 pancake probe to detect beta and gamma radiation
- Ludlum Model 12 portable radiation survey meter with a scintillation counter probe to detect alpha radiation
- Portable explosimeter or Foxboro-Century OVA Model 108 to determine flammable/explosive gases or vapors
- Portable pH/conductivity meter to monitor drilling fluids and groundwater that may contact the skin
- Jerome mercury vapor analyzer

It should be noted that all direct-reading instruments are limited in their ability to detect hazards, and no one instrument will detect all possible hazardous compounds.

TLD badges, which must be worn by all personnel working at the Y-12 Plant site, provide Health Physics Department personnel with radiological exposure data. Abnormally high exposure information will be reported to the SHSO, and appropriate investigations will be conducted by Y-12 Plant personnel as necessary.

9.9 SITE CONTROL MEASURES

Site access is tailored according to whether sampling activities will be conducted in noncontaminated sites, radiologically contaminated sites, or sites where hazardous substances are likely to be encountered (Fig. 9.1). Access requirements may be upgraded or downgraded as conditions warrant.

In noncontaminated areas, a working area large enough to accommodate all direct sampling activities (e.g., sampling, containerizing, logbook keeping) will be cordoned.

If drilling activities are to be conducted, the cordoned zone will have a radius equivalent to the rig mast height.

At hazardous work sites, only those persons directly involved in the site activities will be allowed in the exclusion zone. Visitors who must observe work being conducted in the exclusion zone must wear all appropriate PPE and must be properly trained before entering the area. Should respiratory protection be necessary, all visitors must also produce evidence that they have had a complete physical, respirator training, and a respirator fit test within the last 12 months. Visitor entry into the work zones will be left to the discretion of the SHSO. The following areas will be clearly identified at each site where required:

- **Exclusion Zone**—This is the area where physical, chemical, and/or radiological hazards exist or could occur. The purpose of this zone is to limit the spread of contaminants to clean areas and to provide for the safety of those persons not authorized to enter the zone. The boundaries will be clearly marked, and the flow of personnel and equipment will be restricted to the contamination reduction zone. All personnel in the exclusion zone will abide by a buddy system.
- **Contamination Reduction Zone**—This is the area located between the exclusion zone and the support zone. This zone serves as a buffer to further reduce the possibility of the spread of contamination to clean areas. Decontamination equipment will be available at the boundary between the exclusion zone and the contamination reduction zone.
- **Support Zone**—This is the area located outside the zones of potential contamination. The support zone serves as an entry area for personnel, material, and equipment to the area of site operations. It also serves as an exit area for decontaminated personnel, materials, and equipment. Uncontaminated safety and work equipment will be stored in the support zone. Eating, drinking, and all other activities supporting the sampling teams are permitted in the support zone.

Fig. 9.1. Site control zones.

9.10 DECONTAMINATION AND DISPOSAL

Each site in hazardous-substance or radiologically contaminated areas where reusable equipment is being worn will have one protective clothing decontamination location. The location will contain two large washpans and one large wash tub for equipment, rinse water, soap, brushes, and towels.

Workers will be responsible for decontaminating their boots, gloves, and equipment. Workers will also be required to check their clothing for radioactive contamination before leaving the site.

All noncontaminated disposable clothing will be packaged daily in a marked plastic garbage bag on site and disposed of according to the project waste management plan. Contaminated disposable clothing will be packaged in a marked plastic bag and disposed of according to the project waste management plan. The SHSO, at the direction of the Y-12 Plant ER Project Manager and PHSO, will ensure that waste minimization efforts are employed to keep contaminated and noncontaminated wastes segregated.

Large equipment will be decontaminated by pressure washer. Sampling equipment will be decontaminated with detergent and water and alcohol rinses.

The Y-12 Plant Health Physics Department will survey all subcontractor field equipment before it leaves the Y-12 Plant area. All radioactive contamination will be less than the Green Tag limits as determined by Y-12 Health Physics when equipment is scheduled to be removed from the site. A property removal request will be initiated through the Field Operations Manager so that Y-12 Plant Health Physics personnel can conduct required radiation surveys.

9.11 CONTINGENCY PLAN

Emergency actions are tailored according to whether investigative activities are being conducted on sites in noncontaminated areas, on sites that have been designated as radiologically contaminated, or on sites that have been surveyed and may contain hazardous substances.

Emergency actions for noncontaminated areas are discussed in Sect. 9.11.1. Hazardous-waste-site emergency actions require additional precautions and are discussed separately in Sect. 9.11.2.

9.11.1 Emergency Actions for Noncontaminated Areas

In the event of an accident involving actual or suspected personal injury, the Field Operations Manager will perform the following steps in the order shown here:

- Initiate emergency first aid and/or decontamination, if necessary.
- Contact the Y-12 Plant Shift Superintendent.
- Obtain transportation for injured personnel (e.g., driller's service truck, SHSO vehicle, escort vehicle, ambulance), if required. A medical exam will be required after each accident that potentially exposes personnel to hazardous substances above the allowable

limits or that results in injuries, other than minor first aid injuries, that might result in lost work days or reassignment of the injured worker.

- Initiate necessary corrective measures to abate any unsafe conditions.
- Withdraw uninjured site personnel to a safe distance from the accident scene until the SHSO determines that it is safe for work to resume.
- Contact the Y-12 Plant ER Project Manager and/or the PHSO for immediate advice. Emergency telephone numbers are listed in Appendix G. Contact the appropriate person, department, and/or organization listed in Appendix G of this document, as directed by the Y-12 Plant ER Project Manager or PHSO.

Personnel should not enter an area to attempt a rescue if their own lives would also be threatened because of inadequate personal protection (e.g., clothing, self-contained breathing apparatus). The SHSO will ensure that all employees can be rescued from any area of the site that they are allowed to enter and is responsible for developing any special decontamination procedures for injured personnel if necessary. Site personnel will be instructed to evacuate the work area if PPE fails.

Emergency planning will include establishing a procedure for evacuation from the site in case of a natural or man-made catastrophe. For work sites inside the Y-12 Plant, contractor and subcontractor personnel will assemble at the designated assembly point or where instructed by the Plant Shift Superintendent in the event of a plant emergency.

9.11.2 Emergency Actions Within Contaminated Areas

If personal injury or exposure of personnel to radiological contamination, suspected hazardous chemicals, vapors, or particulates occurs within a contaminated site, the following emergency procedures will be applied.

9.11.2.1 Personnel injury in the exclusion zone

All site personnel will assemble at the decontamination line. The rescue team will enter the exclusion zone (if required) to remove the injured person to the boundary (hotline) (Fig. 9.1).

- The SHSO or other qualified person will evaluate the nature of the injury. If the injury is determined not to be an immediate threat to life, the injured person will be decontaminated to the extent possible before being moved to the support zone. If the injury is determined to be an immediate threat to life, decontamination procedures will be postponed until the injured person stabilizes.
- The Y-12 Plant Shift Superintendent will be contacted for an ambulance, and the designated medical facility will be notified (if required). No person will re-enter the exclusion zone until the cause of the injury or symptoms is determined.

9.11.2.2 PPE failure

If any site worker experiences a failure or alteration of PPE that affects the protection factor, that person and his/her buddy will immediately leave the work area for a safe location. Re-entry will not be permitted until the equipment has been repaired or replaced. The SHSO

will have two air-purifying, full-face respirators on site for escape purposes in an emergency. Rescue operations from toxic atmospheres will be conducted by persons wearing either self-contained breathing apparatus or suitable air-line respirators with escape bottles.

9.11.3 First Aid

Any person designated as an SHSO will be certified in first aid and cardiopulmonary resuscitation (CPR). Workers trained and certified in first aid and CPR will render first aid and will attempt to stabilize any ill or injured worker needing assistance. Life support techniques such as CPR and treatment of life-threatening problems such as bleeding, airway obstruction, and shock will be given top priority. All on-site emergency assistance, including medical assistance, will be obtained as rapidly as possible by calling 911 or the Plant Shift Superintendent at 574-7172.

9.11.4 Spill Response

Site personnel will respond and contain any nonemergency spill such as hydraulic fluid or small quantities of decontamination fluid. Site personnel will not perform emergency response to significant spills such as large quantities of fuels. The PSS will be contacted for emergency response and site personnel will withdraw from the area.

9.12 ON-SITE COMMUNICATIONS

On-site radio communication is required. For emergency communication, the SHSO will use one of the following communication methods:

- mobile radio,
- mobile telephone,
- nearest plant or guard house telephone,
- fire pull-station,
- escort notification (in areas in which a security escort is required).

The Y-12 Plant ER Project Manager and the PHSO will be informed daily by the SHSO of proposed site investigative activities.

9.13 SITE HEALTH AND SAFETY CHECK LIST

9.13.1 Site Description

The Bear Creek Valley (BCV) Operable Unit 1 (BC-OU1) is located at the west end of the Y-12 Plant. This plan addresses the Bear Creek Burial Grounds, Oil Retention Ponds (ORPs) I and II, Oil Landfarm, Boneyard/Burnyard, Sanitary Landfill (SL I) and the S-3 ponds.

9.13.2 Site History

The units in Bear Creek Valley were used for the disposal of waste oils, mop water, uranium, thorium, acids, salts, solvents, asbestos, and other contaminated material. In most cases open earthen trenches were filled and either burned or buried. The S-3 ponds were treated with acid and lime for bioremediation. The ponds were drained and filled with soil, then paved to create a parking lot.

9.13.3 Health and Safety Hazard Evaluation

9.13.3.1 Physical hazards

☒ Heat Stress	☐ Cold Stress	☒ Noise
☒ Confined Space	☐ Enclosed Space	☒ Heavy Lifting
☒ Tripping/Falling	☐ High Voltage	☐ High Pressure Water
☐ Oxygen Deficient	☒ Explosive/Flammable	☐ Vibration

9.13.3.2 Construction hazards

☒ Trenching	☒ Excavating	☒ Heavy Equipment Operation
☐ Demolition	☐ High Work	☒ Welding/Cutting
☐ Ladders	☐ _____	☐ _____

9.13.3.3 Chemical hazards

☒ Organic Chemical	☒ Inorganic Chemical	☒ Carcinogen
☒ Corrosive	☒ Reactive	☐ OSHA Specific
☐ Mutagen	☐ Teratogen Substances	

9.13.3.4 Ionizing radiological hazards

☒ Internal Exposure ☒ External Exposure

9.13.3.5 Nonionizing radiological hazards

☐ UV	☐ RF	☐ Microwave
☐ Laser		

9.13.3.6 Biological/vector hazards

☒ Wildlife	☒ Plants	☐ Medical Waste
☐ Bacterial	☒ Parasites	

9.13.4 Previous Sampling Results

An extensive inventory of wastes present in Operable Unit 1 indicate that PCBs and uranium compounds are present in elevated quantities in water, soil, and sediment in and around the sites in BCV. For additional information see the RIWP.

9.13.5 Contaminants

Chemical: Total Uranium

PEL: 0.05 mg/m³

IDLH: 0.6 mg/m³

STEL: Carcinogen

FP: NA

Route of Exposure: absorption, ingestion, inhalation

Monitoring Equipment: alpha and beta/gamma meters (see monitoring section)

Symptoms/Effects of Exposure: carcinogen

Special Medical Monitoring: urinalysis

Chemical: Acetic Acid

PEL: 10 ppm

IDLH: 1000 ppm

STEL: 15 ppm

FP: 102°F

Route of Exposure: inhalation, contact

Monitoring Equipment: FID/PID

Symptoms/Effects of Exposure: irritation of sensitive areas

Special Medical Monitoring: none

Chemical: Beryllium

TLV: 0.002 mg/m³

IDLH: 10 mg/m³

STEL: Ceiling 0.005 mg/m³

FP: NA

Route of Exposure: inhalation

Monitoring Equipment: filtration and laboratory analysis

Symptoms/Effects of Exposure: respiratory distress, weakness

Special Medical Monitoring: NA

Chemical: PCBs

TLV: 0.5 mg/m³

IDLH: 5 mg/m³

STEL: NA

FP: NA

Route of Exposure: inhalation, ingestion, absorption

Monitoring Equipment: filtration and laboratory analysis

Symptoms/Effects of Exposure: acne-form dermatitis, liver damage

Special Medical Monitoring: NA

Chemical: Heavy Metals

PEL: 0.005 mg/m³ (lead)

IDLH: 700 mg/m³

STEL: NA

FP: NA

Route of Exposure: inhalation, ingestion, contact

Monitoring Equipment: filtration and laboratory analysis

Symptoms/Effects of Exposure: weakness, insomnia, abdominal pain

Special Medical Monitoring: urinalysis

Chemical: Organic Solvents

PEL: 1 ppm (benzene)

IDLH: 3000 ppm

STEL: NA

FP: <100°F

Route of Exposure: inhalation, ingestion, absorption, contact

Monitoring Equipment: FID/PID

Symptoms/Effects of Exposure: carcinogen, irritation, narcosis

Special Medical Monitoring: NA

Chemical: Nitric Acid

TLV: 2 ppm

IDLH: 100 ppm

STEL: NA

FP: NA

Route of Exposure: inhalation, ingestion, contact

Monitoring Equipment: FID

Symptoms/Effects of Exposure: irritation, pulmonary edema

Special Medical Monitoring: NA

Chemical: Isopropyl Alcohol (decontamination fluid)

TLV: 400 ppm

IDLH: 12000 ppm

STEL: 500 ppm

FP: 53°F

Route of Exposure: inhalation, ingestion, contact

Monitoring Equipment: FID/PID

Symptoms/Effects of Exposure: irritation, narcosis

Special Medical Monitoring: NA

9.13.6 Task Breakdown

9.13.6.1 Description: Trenching and associated soil sampling

Level of Personal

Protective Equipment: Primary: C Contingency: B (at SHSO's discretion), D (concurrence of Energy Systems H&S and SAIC H&S manager), withdraw (at SHSO or field manager's discretion)

Type of Work: Intrusive (X) Nonintrusive ()

Engineering Controls: none feasible

Administrative Controls: trenches deeper than 4 feet will comply with 29 CFR 1910.146 (Confined Space) and 29 CFR 1926.650 (Excavation) and relevant Energy Systems procedures.

Personal Protective Equipment:

(Specify Type)

a) Respiratory Protection: P: full or half face APR
C: SCBA

b) Protective Clothing: P: hooded Polyethylene or Saranax coated Tyvek
C: hooded Saranax or PVC

c) Head Protection: P: hard hats around overhead hazards
C: hard hats

d) Eye Protection: P: safety glasses or goggles unless using full face respirator
C: faceshield

- e) Foot Protection: P: PVC safety boots or PVC covers over leather safety boots
C: same
- f) Hand Protection: P: inner (natural rubber) and outer (PVC or nitrile) gloves
C: same
- g) Hearing Protection: P: foam plugs if in high noise areas
C: same
- h) Tape-up Required: P: no
C: yes

9.13.6.2 Description: Drilling

- Level of Personal Protective Equipment: Primary: C Contingency: B (at SHSO's discretion), D (concurrence of Energy Systems H&S and SAIC H&S manager), withdraw (at SHSO or field manager's discretion)
- Type of Work: Intrusive (X) Nonintrusive ()
- Engineering Controls: none feasible
- Administrative Controls: exclusion zone around drill rig
- Personal Protective Equipment:
(Specify Type)
- a) Respiratory Protection: P: full or half face APR
C: SCBA
- b) Protective Clothing: P: hooded Polyethylene or Saranax coated Tyvek
C: hooded Saranax or PVC
- c) Head Protection: P: hard hats around overhead hazards
C: hard hats
- d) Eye Protection: P: safety glasses or goggles unless using full face respirator
C: faceshield
- e) Foot Protection: P: PVC safety boots or PVC covers over leather safety boots
C: same
- f) Hand Protection: P: inner (natural rubber) and outer (PVC or nitrile) gloves
C: same

- g) Hearing Protection: P: foam plugs if in high noise areas
C: same
- h) Tape-up Required: P: no
C: yes

9.13.6.3 Description: Surface soil sampling (hand auger, backsaver)

- Level of Personal Protective Equipment: Primary: C Contingency: B (SHSO's discretion), D (concurrence of Energy Systems H&S and SAIC H&S manager), withdraw (at SHSO or field manager's discretion)
- Type of Work: Intrusive (X) Nonintrusive ()
- Engineering Controls: none feasible
- Administrative Controls: general project requirements
- Personal Protective Equipment:
(Specify Type)
- a) Respiratory Protection: P: full or half face APR
C: SCBA
- b) Protective Clothing: P: hooded Polyethylene or Saranax coated Tyvek
C: hooded Saranax or PVC
- c) Head Protection: P: hard hats around overhead hazards
C: hard hats
- d) Eye Protection: P: safety glasses or goggles unless using full face respirator
C: faceshield
- e) Foot Protection: P: PVC safety boots or PVC covers over leather safety boots
C: same
- f) Hand Protection: P: inner (natural rubber) and outer (PVC or nitrile) gloves
C: same
- g) Hearing Protection: P: foam plugs if in high noise areas
C: same
- h) Tape-up Required: P: no
C: yes

9.13.6.4 Description: Groundwater sampling

Level of Personal Protective Equipment: Primary: D Contingency: C (SHSO's discretion), withdraw (at SHSO or field manager's discretion)

Type of Work: Intrusive (X) Nonintrusive ()

Engineering Controls: none feasible

Administrative Controls: general project requirements

Personal Protective Equipment:
(Specify Type)

a) Respiratory Protection: P: none
C: APR

b) Protective Clothing: P: coveralls (Tyvek or company)
C: hooded Polyethylene or Saranax Tyvek

c) Head Protection: P: hard hats around overhead hazards
C: hard hats

d) Eye Protection: P: safety glasses or goggles unless using full face respirator
C: faceshield

e) Foot Protection: P: safety boots
C: PVC safety boots or PVC covers over leather safety boots

f) Hand Protection: P: inner (natural rubber) and outer (PVC or nitrile) gloves
C: same

g) Hearing Protection: P: foam plugs if in high noise areas
C: same

h) Tape-up Required: P: no

9.13.6.5 Description: Storm water sampling

Level of Personal Protective Equipment: Primary: D Contingency: C (SHSO's discretion), withdraw (at SHSO or field manager's discretion)

Type of Work: Intrusive () Nonintrusive (X)

Engineering Controls: none feasible

Administrative Controls: general project requirements

Personal Protective Equipment:

(Specify Type)

- a) Respiratory Protection: P: none
C: APR
- b) Protective Clothing: P: coveralls (Tyvek or company)
C: hooded Polyethylene or Saranax Tyvek
- c) Head Protection: P: hard hats around overhead hazards
C: hard hats
- d) Eye Protection: P: safety glasses or goggles unless using full face respirator
C: faceshield
- e) Foot Protection: P: safety boots
C: PVC safety boots or PVC covers over leather safety boots
- f) Hand Protection: P: inner (natural rubber) and outer (PVC or nitrile) gloves
C: same
- g) Hearing Protection: P: foam plugs if in high noise areas
C: same
- h) Tape-up Required: P: no
C: yes

9.13.6.6 Description: Decontamination of equipment

Level of Personal Protective Equipment: Primary: D Contingency: C (SHSO's discretion), withdraw (at SHSO or field manager's discretion)

Type of Work: Intrusive () Nonintrusive (X)

Engineering Controls: decontamination area will be constructed so that rinsate will be contained and collected

Administrative Controls: decontamination area will be treated as an exclusion zone

Personal Protective Equipment:

(Specify Type)

- a) Respiratory Protection: P: none
C: APR
- b) Protective Clothing: P: coveralls (Tyvek or company)
C: hooded Polyethylene or Saranax Tyvek

- c) Head Protection: P: hard hats around overhead hazards
C: hard hats
- d) Eye Protection: P: safety glasses or goggles unless using full face respirator
C: faceshield for spray cleaning
- e) Foot Protection: P: safety boots
C: PVC safety boots or PVC covers over leather safety boots
- f) Hand Protection: P: inner (natural rubber) and outer (PVC or nitrile) gloves
C: same
- g) Hearing Protection: P: foam plugs if in high noise areas
C: same
- h) Tape-up Required: P: no
C: yes

9.13.7 Monitoring Requirements

<u>Instrument</u>	<u>Task(s)</u>	<u>Monitoring Frequency</u>	<u>Action Guidelines</u>
LEL Meter	<u>6.1, 6.2</u>	<u>≥ once / 10 min.</u>	<u>≥ 10% LEL - stop</u>
O ₂ Meter	<u>6.1</u>	<u>≥ once / 10 min.</u>	<u>≤ 19.5% O₂ - stop</u>
Colorimetric Indicator Tubes	<u>all</u>	<u>if PID > 1 ppm</u>	<u>≥ TLV - reassess</u>
Photoionization Detector (PID)	<u>6.1 - 6.5</u>	<u>≥ once / 10 min.</u>	<u>≥ 1 ppm test benz</u> <u>≥ 5 ppm level C</u> <u>≥ 20 ppm level B</u>
Flame Ionization Detector (FID)	<u>6.1 - 6.5</u>	<u>if PID fails</u>	<u>same</u>
Alpha Meter	<u>6.1 - 6.2</u>	<u>sample removal</u>	<u>≥ 500cpm/100cm^{2a}</u>
Beta/Gamma Meter	<u>6.1 - 6.5</u>	<u>sample removal</u>	<u>≥ 100 cpm/100cm²</u>
Area Radiation Monitors	<u>none</u>		
Perimeter Radiation Monitors	<u>none</u>		
Noise Meter	<u>6.1 - 6.2</u>	<u>initially</u>	<u>85 dBA</u>
Area radiation dose survey performed by Y-12 HP	<u>all</u>	<u>prior to work</u>	<u>Y-12 limits</u>

^a Stop work and notify HP.

9.13.7.2 Personnel monitoring

Whole-body Dosimetry	<u>all</u>	<u>continuously</u>	<u>Y-12 limits</u>
Extremity Dosimetry	<u>none</u>		
Whole-body Count	<u>none</u>		
Urinalysis/Bioassay	<u>6.1 -6.5</u>	<u>pre and post job</u>	<u>Y-12 limits</u>
Chemical Air Sampling	<u>all</u>	<u>if PID in excess</u>	<u>TLV/PEL</u>
Radiation Air Sampling	<u>none</u>		

9.13.8 Hazard Analysis

9.13.8.1 Chemicals

Tasks: 6.6, isopropyl alcohol

Specific labeling requirements of site-generated waste: per EPA hazardous waste regulations

Chemical-specific disposal requirements: to be disposed by Y-12 waste management

9.13.8.2 Fire/explosion

Tasks: 6.6, isopropyl alcohol

Are flammable liquids present? Yes

Description isopropyl alcohol

Location in decontamination area

Quantity less than 10 gallons

Containment/Storage method original shipping packaging

For welding, cutting, or brazing is a Welding Permit required? Yes

9.13.8.3 Confined/enclosed spaces

Tasks: 6.1, It may be necessary to enter excavations.

Confined/enclosed space entry required (Low Risk)? Potentially

Operations Safety Work Permit required? Yes (if entry occurs)

Standard Operating Procedure required? Yes

9.13.8.4 Ionizing radiation

Tasks: all

Applicable detailed checklist? NA

Primary contaminating isotope(s) uranium

Location on site unknown

Containment/storage method	<u>contaminant</u>
Radiation type	Alpha/Gamma
Dose rate (maximum)	mR/h @ _____ meter(s)
(average)	mR/h
Note: to be determined for each work area prior to work in that area.	
Contamination level (fixed)	dpm/100 cm ²
(removable)	dpm/100 cm ²
Note: to be determined for each work area prior to work in that area.	
Water contamination potential?	Yes
Unrestricted airborne contamination release potential?	No
Radiation work permit required?	Yes/No
Note: to be determined for each work area prior to work in that area.	
Health Physics coverage	Conditional
Special task operation requirements	<u>no</u> Welding/cutting/brazing <u>no</u> Grinding/chipping <u>yes</u> Hydraulic/air hammer operation <u>no</u> Dusty conditions (sweeping, vacuuming, etc.) <u>yes</u> Equipment decontamination/free release

9.13.8.5 Electrical hazards

Tasks:	<u>potentially all</u>
Electrical shock hazard?	Yes
	<u>unknown</u> Voltage
	<u>unknown</u> Current

Location of hazard: potential for underground and overhead utilities. Each work area will be surveyed for visible hazards and these will be avoided. Excavation/penetration permits will be obtained.

9.13.8.6 Temperature extremes

Tasks:	<u>potentially all</u>
Temperature extremes (Hot/Cold)?	Yes
Average daily high temperature (during work shift)	°F/°C
Average daily low temperature (during work shift)	°F/°C
Average wind speed MPH	cloudy/sunny
Temperature WBGT °C	

Work load:

Light
☒ Moderate
 Heavy

Work/Rest regimen:

% work
 % rest

Threshold Limit Value

Precautions (*specify*): Temperatures will be determined prior to starting work and appropriate precautions will be taken.

Cooling/heating equipment needed: This will be determined prior to starting work. The need for this equipment varies by season.

9.13.8.7 Noise

Tasks:

6.1 - 6.2

Noise extremes?

Yes

Sound level

85 dB(A)

Noise source(s):

excavation equipment, drill
 rigs

Noise above 85 dB(A) (hearing protection required) Yes

Precautions: Work areas will be surveyed and high noise areas will be designated. Hearing protection will be required in these areas.

9.13.8.8 Sanitation

Tasks:

all

Potable water required?

Yes

Non-potable water used?

No

Eating, drinking, and smoking permitted?
 Where?

Yes
in break area

Toilet facilities required?

Yes/No

Location and number:

will determine prior to
 starting work.

Washing facilities required?

Yes/No

Location:

will be provided as
 necessary

Change rooms required?

No

Specify:

if personnel become
contaminated, existing
shower facilities will be
used

9.13.8.9 Safety hazards

Site posted information/notification required?	Yes
Site posting required?	Yes
Site guard required?	No
Access control required?	Yes
Entry/exit logs required?	Yes
Escape routing/posting required (include site map)?	Yes
Compressed gas cylinders? Location:	Yes <u>oxygen and acetylene cylinders on drilling support truck</u>

9.13.9 Emergency Preparedness

Evacuation route: see attached map

EMERGENCY CONTACTS: All emergency services at the OU 1 Site are to be contacted by first calling the Plant Shift Superintendent (PSS) at 911. In the event that the PSS is not available, emergency services may be reached at the telephone numbers shown below.

<u>Facility Personnel</u>	<u>Phone</u>
PSS	911
Industrial Hygiene	576-7172
Health Physics	574-3547
Industrial Safety	574-1562
SAIC Project Manager (Duncan Moss)	481-4600
SAIC Site Health and Safety Officer (Gary Seaver)	481-4600

<u>Emergency Assistance Service</u>	<u>Phone</u>
Police	574-7272
Ambulance	574-1577
Hospital	574-1577

9.13.10 Record Keeping Requirements

	<u>Required</u>
Hazardous Chemicals	
Area monitoring	<u>X</u>
Personnel monitoring	<u>X</u>
Oxygen Level Measurements	<u>X</u>
Flammability Measurements	<u>X</u>
Ionizing Radiation	
Worker dose	<u>X</u>
Contamination levels	<u>X</u>
Airborne contamination level	
Radiation work permit	<u>X</u>
Non-Ionizing Radiation	
UV level measurements	
Microwave level measurements	
Laser power level measurements	
Biological	
Personnel exposure monitoring	
Electrical	
Tag-out records	
Temperature extremes	
WBGT measurements	<u>X</u>
Daily temperature/wind speed	<u>X</u>
Noise	
Area monitoring	<u>X</u>
Personnel monitoring	
Personnel Medical Monitoring	<u>X</u>
Safety/Emergency Response	
OSHA accident records	<u>X</u>
Accident/incident reports	<u>X</u>
Personal Protective Equipment	
Inspection of: Clothing	
Respirators	
Gloves	
Boots	
Waste Disposal Manifests	
Spill Incident Reports	
Training of Employees	

APPROVALS

WASTE MANAGEMENT PLAN FOR THE REMEDIAL INVESTIGATION OF THE BEAR CREEK OPERABLE UNIT 1 AT THE OAK RIDGE Y-12 PLANT

DOE/OR/01-1161/V1&D2

March 1994

J. E. Myril 3/21/94
Technical Support Contractor Program Manager Date

James C. Bonnan 3/22/94
Technical Support Contractor QA/QC Officer Date

Sherry Sanford 3/20/94
DOE-ER Division Program Manager Date

Wanda J. Loe 3/24/94
Energy Systems ER Site Program Manager Date

Judy A. Hodgins 3/23/94
Energy Systems ER Site Project Manager Date

Chris Smith 3/24/94
Plant ER Waste Management Coordinator Date

10. WASTE MANAGEMENT PLAN

10.1 SCOPE

The objectives of this Waste Management Plan for the BCV OU 1 RI are to identify analytical requirements, personnel responsibilities and types, anticipated characterizations, and volumes of waste that will be generated during RI activities conducted by the Y-12 Plant ER Division. Another goal of the Waste Management Plan is to establish TSD objectives for project-derived wastes prior to final remedial activities.

The BCV OU 1 Waste Management Plan will include management of soil, protective clothing, well development purge water, and decontamination water wastes. Table 10.1 outlines the ER Waste Management Plan Checklist for BCV OU 1 to be conducted by the Y-12 Plant ER Division. Investigation-derived wastes (IDW), waste types, waste categories, estimated generation volumes, and expected contaminants for each waste stream are provided. This Waste Management Plan defines the procedures for segregating, characterizing, and controlling all IDW from the point of generation through TSD including (1) the processes and methods for sampling, analyzing, characterizing, and certifying IDW; (2) waste staging areas, satellite and 90-d accumulation areas, and permitted or interim status storage areas; (3) waste packaging and labeling procedures; (4) transportation routes and U.S. Department of Transportation (DOT) reporting requirements; (5) proposed treatment and/or disposal methods and facilities; and (6) waste minimization and reduction techniques to be used. Waste management plans in support of these RI activities are summarized below.

Decontamination procedures will be established to support equipment decontamination and reduce the potential for cross contamination during sampling activities. All equipment and personnel will be scanned for radioactive contamination before leaving a borehole location. Equipment or PPE exhibiting radioactive contamination will be decontaminated at the borehole location where it was used before being moved to another borehole location or returned to the temporary decontamination area. All equipment at each borehole location will be subjected to preliminary cleaning to remove surface dirt or mud that otherwise might be tracked to other areas of the Y-12 Plant.

Soil cuttings will be monitored for radioactivity and volatile organic contaminants in the field to screen the waste types (hazardous or nonhazardous). Nonhazardous soil will be returned to the borehole in accordance with CERCLA guidance (*Guide to Management of Investigation-Derived Wastes*, EPA Publication 9345.3-03FS, April 1992.) The top 4 ft of the borehole will be grouted for containment. Hazardous samples will be packaged as indicated in Sect. 10.10.2.2 and sent for analytical analysis and characterization. Hazardous solid wastes will be placed in appropriately labeled drums and coordinated through Y-12 Waste Management Organization for disposal at a hazardous waste storage facility with available capacity.

Table 10.1. Environmental Restoration Waste Management Project Plan

Environmental Restoration Waste Management Project Plan Revision No. Draft 1						
1. Project Name: Bear Creek OU 1 RI Project Plant: Y-12 Organization: ER Division/RI Subcontractor						
2. Responsible Project Manager: Y-12 ER						
3. Expected Start/Completion Dates: 5-4-94/3-29-96						
4. Project Description (Brief): Remedial Investigations to collect subsurface soil samples						
5. Project Participants (for waste management interface only): Waste Generator: RI Subcontractor Waste Handler: Y-12 ER Division Waste Transporter: Energy Systems Waste Treatment, Storage, and Disposal (WTSD) Interim Waste Storage: RI Subcontractor Permanent Waste Storage: WTSD Waste Treatment: Y-12 Waste Treatment Organization (WTO) Waste Disposal: Y-12 Sanitary Landfill and/or WTSD						
6. Waste Generation:						
Waste Stream	Type	Category	Volume (ft ³ or gal)	Contaminants	Characterization Based on:	
					Process Knowledge	Sampling and Analysis
Waste Stream 1 Unused Soil will be returned to borehole it came from. The upper 4 ft of each borehole will be grouted. Any remaining soil will be kept on site.	Soil	LLW	1.8 m ³ (60 ft ³)	Depleted and enriched uranium, nitrates, metals, acids, bases	✓	
Waste Stream 2	Decontamination water	LLW	418 L (110 gal)	Depleted and enriched uranium, metals, acids, bases	✓	✓
Waste Stream 3	Solids (clothing)	Sanitary (Dumpster)	0.45 m ³ (15 ft ³)	Same as Waste Stream 1	✓	
Waste Stream 4	Purge Water	LLW	(2000 gal)	Same as Waste Stream 2		✓
Waste Stream 5	Unused Samples Returned from Lab	LLW	(85ft ³)	Same as Waste Stream 1	✓	
7. Waste Analysis and Characterization: None to date						

Table 10.1 (continued)

Environmental Restoration Waste Management Project Plan Revision No. Draft 1	
8. Waste Staging Area: Y/N YES Location: Bear Creek Valley OU 1 Site Special Requirements: Flagging around area to segregate drums	
9. Transport Across Public Access Roads: N Energy Systems will be responsible for transporting drums from accumulation area Road Involved: DOT Regulations to be applied: N Reportable quantities of anticipated DOT-regulated hazardous materials (if required): N/A	
10. Waste Storage Requirements: Field Staging Area: Y/N/NA YES - CERCLA staging area Location: Parking lot in Y-12 exclusion area Responsible Organization: SAIC 90-Day Storage Area: Y/N/NA NO Location: Capacity: Waste Acceptance Criteria Requirements: Responsible Organization: Energy Systems Permitted (or Interim Status) Storage: Y/N/NA Location: Capacity: Waste Acceptance Criteria Requirements: Responsible Organization:	
11. Identification of Potential Treatment Options: SOIL - NO; CLOTHES - NO Location: West End Water Treatment Plant - Water Capacity: 418 L (110 gal) Responsible Organization: Energy Systems, Y-12 WTO Special Waste Acceptance Criteria Requirements: Analysis for process use	
12. Identification of Potential Disposal Options: Location: Capacity: Responsible Organization: Energy Systems Special Waste Acceptance Criteria Requirements: Pass through trash monitoring station before final disposal	

Table 10.1 (continued)

Environmental Restoration Waste Management Project Plan Revision No. Draft 1	
13. Waste Minimization and Reduction Techniques to be Implemented: Segregation: ✓ In-Field Pretreatment: Decontamination: Waste Handling (Spill Control): ✓ Compaction: ✓ Material Recycle (Solvents, Containers): Solvent Substitution: Materials Reuse (Solvents, Wash Waters): Sludge Dewatering: Cutting Fluids Recovery: Selection of PPE: ✓ Selection of Equipment: Other:	
Description of Special Techniques and Expected Effectiveness: Effort will be made to minimize all waste generation. An attempt will be made to compact protective clothing that is put into drums. Soil from borings will be returned to the borehole it came from.	
Prepared by:	
Review: Plant Environmental Compliance Organization: Central ER Waste Management: ER Waste Reduction Manager: Central Waste Management Organization: _____ Date: _____	
Approvals: Plant ER Waste Management Organization: _____ Date: _____ Plant Waste Management Organization: _____ Date: _____	

Protective clothing will be scanned for radioactive contamination; if not contaminated, it will be treated as nonhazardous solid waste and will be disposed of in a dumpster provided by Y-12 Plant ER Division. The solid waste from the dumpster will be disposed of at the Y-12 Plant landfill after passing through the trash monitoring station. Any protective clothing that is contaminated will be placed in 209-L (55-gal) drums and disposed of as hazardous waste by Y-12. This waste management plan will assume all soiled protective clothing to be hazardous waste that will be segregated according to level of contamination (radioactive and nonradioactive or sanitary), containerized (placed in 55-gal drum), properly labeled, and coordinated through Y-12 Waste Management Organization for disposal at an appropriate hazardous waste storage facility with available capacity.

All decontamination water will be contained and pumped into 209-L (55-gal) drums with bung hole tops or into poly tanks provided by the Y-12 Plant Water Treatment Plant for final

disposal. Analysis for process use will be run on the water by the Y-12 Plant Water Treatment Plant. The isopropyl alcohol used as a rinse during decontamination will be contained separately from the decontamination water and allowed to evaporate.

10.2 GENERAL INFORMATION

Remedial investigation of BCV OU 1 is expected to begin in May 4, 1994, and end March 29, 1996.

10.3 INVESTIGATION-DERIVED WASTE POLICY

The IDW Policy [as stated in the *Guide to Management of Investigation-Derived Wastes* (EPA 1992a) and the draft Central ER Program procedure on management of IDW] will be implemented to the extent practicable for BCV OU 1 RI indigenous and nonindigenous wastes generated during sampling activities. Indigenous IDWs expected to be generated during the RI include soil cuttings derived from installing soil borings and groundwater wells, and groundwater derived from well development, purging, and sampling. Nonindigenous IDW expected to be generated during the RI includes expendable sampling materials, disposable PPE, and decontamination fluids. The quantity of wastes generated will be minimized to the extent practicable. Considerations for protection of the public, workers, and the environment were reviewed, and a determination has been made that use of the IDW Policy option for indigenous waste would be protective of human health and the environment.

The indigenous IDW generated during the BCV OU 1 RI will be regulated under the IDW policy. This policy will allow waste generated during sampling to be returned to the area of contamination (AOC). Also, under IDW policy, past sampling data can be used to characterize waste as nonhazardous until the waste is characterized on the basis of analytical data received from environmental sampling. Indigenous IDW will be characterized primarily by using results from environmental sampling, analytical data, and best professional judgment. During project activities, records of environmental samples collected during the RI will be such that each sample can be directly tracked to indigenous and nonindigenous IDW.

Because surface and subsurface contamination may exist in the BCV OU 1, the preferred method of handling the indigenous IDW that will be generated during the RI will be to segregate and consolidate the potentially contaminated excess soil cuttings, groundwater, and well cuttings within the AOC. Soil cuttings from each borehole will be characterized on the basis of analytical results from environmental samples collected from the same borehole. Based on this characterization, contaminated soils will be consolidated, containerized, and stored in a waste stockpile within the AOC until they can be remediated in conjunction with the site remediation. All unused soil obtained during the drilling of boreholes will be returned to the borehole from which it came, in accordance with EPA guidance for management of IDW. The upper 4 ft of each borehole will be grouted, and any excess soil will be spread evenly around the borehole. Groundwater will be containerized within the AOC and characterized on the basis of analytical results from environmental sampling and waste characterization sampling for on-site or off-site wastewater treatment according to TSD facility waste acceptance criteria. The nonindigenous IDW (expendable sampling materials, disposable PPE, and decontamination solutions) will be containerized and analyzed for on-site or off-site wastewater treatment according to TSD facility waste acceptance criteria.

Nonindigenous wastes such as liquids (decontamination water/soap solutions, and decontamination water/isopropyl alcohol solutions) and compactible wastes (expendable sampling materials and disposable PPE) generated during the BCV OU 1 RI will be containerized in DOT specification drums and will be stored in field staging areas within the AOC for future TSD. Until analytical data are received characterizing the IDW as RCRA hazardous waste, the waste will be managed in a protective manner in field staging areas (FSAs). The FSAs will be clearly marked by signs and roping. Storage of suspected RCRA hazardous IDW within the BCV OU 1 AOCs will comply with the requirements of RCRA set forth in 40 *CFR* Part 262, with respect to labeling, inspecting, containerizing, and management of RCRA hazardous wastes. Storage of nonhazardous wastes within the BCV OU 1 AOCs will be managed in a manner that is protective of human health and the environment. Figure 10.1 indicates the AOC designations (encircled by heavy lines) for the BCV OU 1 RI.

10.4 HISTORY AND EXPECTED CONTAMINANTS

Facility history and expected contaminants are presented in Chap. 2 of this RI work plan. An all-inclusive list of contaminants expected to be encountered by the RI subcontractor personnel during field activities would be unrealistic. Typical COCs include radionuclides (uranium, thorium, radium, and ^{99}Tc), metals (mercury, lead, chromium, and beryllium), organics (PCBs, trichloroethene, tetrachloroethene, trichloroethene, petroleum products), and nitrates. Table 5.1, presented earlier, contains a complete list of preliminary COCs for sites in BCV OU 1.

10.5 APPLICABLE AND REFERENCE DOCUMENTS

The applicable laws governing waste management and related activities at the Y-12 Plant include the following:

- Atomic Energy Act of 1954, Public Law (P.L.) 83-703, as amended
- NEPA of 1969, P.L. 91-190, as amended
- RCRA of 1976, P.L. 94-580, as amended
- CERCLA of 1980, P.L. 96-510, as amended
- TSCA of 1976, P.L. 94-469, as amended
- Water Pollution Control Act (Clean Water Act) of 1972, P.L. 92-500, as amended
- Safe Drinking Water Act of 1974, P.L. 93-523, as amended
- Air Pollution Control Act of 1955, P.L. 84-159; the Clean Air Act of 1963, P.L. 88-206; the Air Quality Act of 1967, P.L. 90-148; Clean Air Act Amendments of 1970, P.L. 91-604; Clean Air Act of 1977, P.L. 95-190; and the Clean Air Act Amendments of 1990, P.L. 101-549; as amended
- OSHA of 1970, P.L. 91-596, as amended
- Transportation Safety Act of 1974, P.L. 93-633, as amended

Fig. 10.1. Designated AOC at the BCV OU 1.

- State of Tennessee Water Quality Control Act, Tennessee Code, Title 69—Waters, Waterways, Drains and Levees, Chapter 3: Water Pollution Control; enacted by Tennessee Public Act of 1971, Chapter 164, as amended
- Tennessee Solid Waste Disposal Act, Tennessee Code, Title 69—Waters, Waterways, Drains and Levees, Chapter 3; Water Pollution Control; enacted by Tennessee Public Act of 1971, Chapter 164, as amended
- State of Tennessee Hazardous Waste Management Act, Tennessee Code, Title 68—Safety and Health, Chapter 46: Hazardous Waste Management; enacted by Tennessee Acts of 1977, Chapter 175, as amended
- State of Tennessee Petroleum Underground Storage Tank Act, Tennessee Code, Title 68—Safety and Health, Chapter 53: Petroleum Underground Storage; enacted by Tennessee Acts of 1988, Chapter 984, as amended

10.5.1 Regulations

Numerous federal and state regulations, including DOE orders, derive their authority from, and provide standards for, regulation of provisions contained in the foregoing laws. Regulations relevant to various aspects of waste management at the Y-12 Plant include the following:

- DOE Order 5440.1C, “National Environmental Policy Act”
- DOE Order 5480.1A, “Environmental Protection, Safety, and Health Protection Programs for DOE Operations”
- DOE Order 5440.1B, “Implementation of the National Environmental Policy Act” canceled and replaced by: DOE Order 5440.1C, “National Environmental Policy Act,” April 9, 1985
- DOE Order 5480.1A, “Environmental Protection, Safety, and 6/22/82 Health Protection Programs for DOE Operations”
- DOE Order 5480.10, “Contractor Industrial Hygiene,” February 25, 1986
- DOE Order 5482.1B, “Environmental Safety and Health Appraisal”
- DOE Order 5483.1A, “Occupational Safety and Health Program for DOE” canceled and replaced by: DOE Order 5483.1A: “Occupational Safety and Health Program for DOE,” June 24, 1983
- DOE Order 5484.1, “Environmental Protection, Safety, and Health Protection Information Reporting Requirements,” May 5, 1982
- DOE Order 5500.2A, “Emergency Planning, Preparedness, and Responses for Operations” canceled and replaced by DOE Order 5500.2A “Emergency Notification,” April 13, 1988
- DOE-ORO Notice 5500.2A, “Procedures for the Classification, Notification, and Reporting of Emergency and Nonroutine Events”
- DOE Order 5820.2A, “Radioactive Waste Management”
- NRC Regulation 10 *CFR* 61, “Licensing Requirements for Land Disposal of Radioactive Waste”

- DOE Regulation 10 *CFR* 1021, "Compliance with the National Environmental Policy Act"
- EPA Regulation 40 *CFR* 61, "National Emission Standards for Hazardous Air Pollutants"
- EPA Regulation 40 *CFR* 122, "National Pollutant Discharge Elimination System"
- EPA Regulation 40 *CFR* 136, "Regulations on Test Procedures for the Analysis of Pollutants"
- EPA Regulation 40 *CFR* 141, "National Primary Drinking Water Regulations"
- EPA Regulation 40 *CFR* 142, "National Primary Drinking Water Regulations Implementation"
- EPA Regulation 40 *CFR* 143, "National Secondary Drinking Water Regulations"
- EPA Regulation 40 *CFR* 190, "Environmental Radiation Protection Standards for Nuclear Power Operation"
- EPA Regulation 40 *CFR* 191, "Environmental Radiation Protection Standards for Management, and Disposal of Spent Nuclear Fuel, High-Level, and Transuranic Radioactive Wastes"
- EPA Regulations, Title 40, "Protection of Environment," Subchapter I, "Solid Waste," Parts 240 through 299
- EPA Regulations, Title 40, "Protection of Environment," Subchapter I, "Superfund Programs" Parts 300 through 399
- EPA Regulations, Title 40, "Protection of Environment," Subchapter R, "Toxic Substances Control Act," Parts 700 through 799
- Council on Environmental Quality Regulations, Title 40, "Protection of Environment," Parts 1500 through 1508, "Regulations on Implementing National Environmental Policy Act Procedures"
- Department of Labor Regulation 29 *CFR* 1910, "Occupational Safety and Health Standards"
- DOT Regulations, Title 49, "Transportation," Subchapter C, "Hazardous Materials Regulations," Parts 171 through 179
- TDEC Regulations Chapter 1200-1-11, "Hazardous Waste Management Rules"
- TDEC Regulations Chapter 1200-1-7, "Solid Waste Regulations"
- TDEC Regulations Chapter 1200-4-1, "General Regulations"
- TDEC Regulations Chapter 1200-4-3, "Water Quality Criteria"
- TDEC Regulations Chapter 1200-4-5, "Effluent Limitations and Standards"
- TDEC Regulations Chapter 1200-1-15, "Underground Storage Tank Program"

10.5.2 Energy Systems Policies, Standards, and Procedures

- ESS-EP-125—"Management of Polychlorinated Biphenyls (PCBs) Immediate Action Directive" (IAD)
- ESS-ENV.1—"Environmental, Safety, Health, and Waste Management Quality Assurance"
- ESS-TR-101—"Transportation and Packaging Safety"
- ESP-IH-161—"Worker Protection in Hazardous Waste Operations and Energy Response"

10.5.3 Energy Systems Environmental Surveillance Procedures (Kimbrough et al. 1990)

- ESP-307—"Field Measurement Procedures"
- ESP-500—"Manual Chain of Custody Procedures"
- ESP-80—"Packaging Environmental Samples for Transportation"
- ESP-900—"Cleaning and Decontaminating Sample Containers and Sampling Devices"
- ESP-901—"Equipment Decontamination"
- ESP-1000—"Waste Management"

10.5.4 ER Division Procedures

- ER/C-P1300—"Reporting of Occurrences and Noncompliances for the Environmental Restoration Division"
- ER/C-P1301—"Implementation of Training Program Requirements for Environmental Restoration Division Employee"
- ER/C-P1719—"Controlling and Documenting Field Changes to Append Field Sampling Plans"
- ER/C-P2101—"Preparation of ER Task/Project Specific Waste Management Plans"
- ER/YP2102—"Waste Removal and Disposition Through the Y-12 ER Program" (IAD)

10.5.5 Waste Management Division Operating Procedures

- WMD-HS-5402—"Intra-Plant Movement of Hazardous Material" (old 1430.31 "Intra-Plant Movement of Hazardous Materials")
- SMD-HS-5501—"Off-Site Shipment of Hazardous Waste-Administration"
- WMD-OP-2025—"Temporary Outside Storage of Waste Containers"
- WSO-OP-2001—"Waste Storage Inspection Requirements"
- WSO-OP-2004—"Fissile Material Storage Unit Operations"
- WSO-OP-2008—"Waste Container Handling and Stacking Operations" (old "Handling and Storage of Hazardous Waste")
- WSO-OP-2014—"Corrosion Inspection of Storage Containers"

- WTDO-HS-5403—"Asbestos Waste Shipment Tracking and Reporting"
- WTDO-HS-5404—"K-25 Site Fill Area I Operations"
- WTDO-OP-2006—"Hazardous Waste Manifest and Tracking"
- WTO-HS-5502—"Packaging and Verification of Lab. Pack Quantity Waste for Off-Site" (old "Waste Verification and Waste Packaging")

10.5.6 Y-12 Plant Procedures and Standard Practices

- 70-303—"Control of Industrial Wastes"
- 70-903—"Transfer, Storage, or Disposal of Waste"
- 50-41-WT8-2.64—"Hazardous Waste Collection, Packaging, Storage and Disposal"

10.6 PROJECT CONTACTS

Names and affiliations of key personnel involved in the generation, management, and disposal of wastes in association with the BCV OU 1 RI project are identified in Table 10.2. Active waste management facilities at the Y-12 Plant are identified in Table 10.3.

10.7 PROJECT WASTE MANAGEMENT/WASTE HANDLING

The types of waste most likely to be generated and managed during field investigations on the Y-12 Plant include decontamination fluids, well development/purge water, debris, sludge, soil, solvents/oils, PPE, trash, PCB equipment, equipment, metal, miscellaneous solid waste, and miscellaneous liquid waste. The RI Subcontractor will characterize all IDW according to the eight categories identified: (1) RCRA waste, (2) TSCA waste, (3) low-level radioactive waste (LLW), (4) LLW/RCRA waste, (5) LLW/TSCA waste, (6) LLW/RCRA/TSCA waste, (7) RCRA/TSCA waste, or (8) sanitary/industrial waste.

10.7.1 Anticipated Decontamination Activities

Field activities conducted by the RI subcontractor on the Y-12 Plant will require decontamination of equipment, sampling devices, the outside portion of sample containers, and personnel. The subcontractor's project Health and Safety Officer will be responsible for personnel decontamination, which will not be permitted in the same wash tubs used for equipment decontamination. To decontaminate sample containers, the RI subcontractor will follow the procedures outlined in ESP-900 Part VII D (Kimbrough et al. 1990). Decontamination of equipment and sampling devices will be accomplished in accordance with ESP-900 and ESP-901.

10.7.2 Investigation Derived Waste

Liquid, soil, and solid waste resulting from decontamination activities will be managed as CERCLA hazardous substances and placed in appropriately labeled drums. The RI subcontractor will obtain empty drums from the ER Division Project Manager. In the event

Table 10.2. Key project personnel, affiliations, and telephone numbers

Affiliation	Name	Telephone
Project Manager, Science Applications International Corporation (SAIC)	Duncan Moss	481-4107
Field Operations Manager, SAIC	Greg Schank	481-8760
Site Health and Safety Officer, SAIC	Eric Corbin	481-8726
Y-12 Site Health Physics Department	Rick Howard	574-4085
Y-12 ER Field Operations Manager	Steve Walker	576-5364
Y-12 ER Field Coordinator	Ricky Walker	576-2151
Y-12 Site Waste Transportation and Disposal Department	Ken Delius	576-4965
Y-12 Site Waste Disposal Coordinators	Chris Smith	576-6526
Y-12 ER Project Manager	Judy Hodgins	576-2368
Engineering Division Principal Engineer	Bill Barton	576-0519
	John Baker	574-1064
Y-12 Site Industrial Hygiene	Rudy Weigel	241-2487
Y-12 Site Environmental Management	Wayne McMahon	574-7535
Y-12 Site Industrial Safety	John Peters	574-2178

that Energy Systems is unable to supply drums, they will be supplied by the drilling subcontractor. The subcontractor project manager will contact the Y-12 Plant Waste Disposal Coordinator (WDC) to arrange for treatment or disposal of the filled drums at the close of the field activity.

10.7.3 Liquid IDW

Most liquid IDW resulting from decontamination activities will be generated by steam cleaning of equipment such as hollow-stem augers and drill rods. Smaller quantities of liquid IDW will be generated as a result of washing and rinsing sampling devices such as hand augers and bailers. All decontamination water will be placed in 209-L (55-gal) drums with bung openings and analyzed for using pH and conductivity measurements.

10.7.4 Soil IDW

Soil IDW resulting from decontamination activities will include large soil clumps removed from equipment and soil cuttings from each borehole. Soil wastes derived from each borehole will be placed in open-top, 209-L (55-gal) drums, managed and stored separately within the AOC. Soil cuttings from each borehole will be characterized on the basis of analytical results from environmental samples collected from the same borehole. Based on this characterization, contaminated soils will be consolidated, containerized, and stored in a waste stockpile within the AOC until they can be remediated in conjunction with the site remediation. All unused

Table 10.3. Active waste management facilities at the Y-12 Plant

	Facility	Waste managed	Contact
1	Y-12 Centralized Sanitary Landfill	Sanitary Waste	J.R. McDeerman 576-8461
2	New Salvage Yard	Clean scrap metal and unserviceable equipment	B.R. Patton 574-0446
3	Industrial Waste Landfill IV	Classified, nonhazardous, nonradioactive, solid waste	J.A. Coffey 576-7898
4	9720-12 Non-SNM Warehouse	Low-level radioactive waste	V.W. Tassey 576-8462
5	9720-25	Classified, radioactive, classified hazardous, and mixed storage wastes	M.C. Wiginton 576-7768
6	Spoil Area I	Clean demolition spoil	J.A. Coffey 576-7898
7	Interim Reactive Treatment (NAK)	Water reactive waste	S.R. Fain 576-5816
8	OD-8	Clean, PCB, radioactive, hazardous and mixed oils and solvents	C.F. Wisener 574-6847
9	OD-9	PCB, radioactive, hazardous and mixed oils and solvents	C.F. Wisener 574-6847
10	9407-7 PCB Storage	PCB and uranium solids	L.C. Duncan 574-0882
11	9720-31 (RCRA Motel)	RCRA waste, nonradioactive	S.A. Bowen 574-1707
12	PCB Shed	PCB waste	L.C. Duncan 574-0882
13	DARA Solids Storage Facility	PCB, uranium 574-0468	R.W. Sellers
14	DARA Oil Land Soils Storage	PCB, uranium	L.C. Duncan 574-0882
15	Uranium Oxide Storage Vaults	Radioactive solid waste	V.W. Tassey 576-8462
16	Waste Feed Preparation Facility	Radioactive solid waste	V.W. Tassey 576-8462
17	Old Salvage Yard	Radioactive scrap metal	B.R. Patton 574-0466
18	Interim Drum Yard	Hazardous and mixed waste	S.R. Fain 576-5816
19	East Chestnut Ridge Waste Pile	Mercury-contaminated soil	J.A. Coffey 576-7898
20	Contaminated Soils Storage Area	Hazardous and mixed soils and spoil material	S.R. Fain 576-5816
21	Interim Reactive Waste Area	Water reactive waste	S.R. Fain 576-5816

soil obtained during the drilling of boreholes will be returned to the borehole from which it came, in accordance with EPA guidance for management of IDW. Any excess soil will be spread evenly around the borehole.

10.7.5 Solid IDW

Solid IDW resulting from decontamination activities will include used disposable PPE and expendable cleaning items, such as paper towels, brushes, and rags used in decontamination activities. Solid waste will be placed in open-top, 209-L (55-gal) drums. Solid and soil wastes will not be stored in the same container.

10.8 CONTAINERIZING AND LABELING REQUIREMENTS

Wastes generated by the RI subcontractor during field investigations and sampling activities may be collected, disposed of immediately, or provided some level of interim storage management. Until the IDW is positively identified as a RCRA hazardous waste, the subcontractor Field Operations Manager must manage IDW in a protective manner but not necessarily in accordance with RCRA Subtitle C requirements. Under CERCLA, IDW may be stored in a container within the AOC and returned to its source without meeting the specified LDR treatment standards. When IDW is disposed of off site or is transported on public roads, DOT requirements for manifesting, containerizing, labeling, and transporting waste may apply. Samples collected and shipped for analysis are exempt from the RCRA packaging, labeling, manifesting, and transportation requirements. These samples will be managed in accordance with approved procedures.

10.8.1 Containerization

Containers used to store RCRA waste, TSCA waste, and LLW will comply with the DOT regulations specified in 49 *CFR* 171-179 for hazardous and radioactive waste as appropriate. LLW will be packaged according to DOE Order 5820.2A and 49 *CFR* 273. Each container type will be reviewed for compatibility with wastes before use. IDW will be placed only in containers approved by the Y-12 Plant Waste Management Organization.

Oak Ridge Y-12 PP 70-903, "Transfer, Storage, or Disposal of Waste," contains specific instructions for selecting the type of container for the waste type being containerized. The RI Subcontractor Project Manager will coordinate with the Y-12 Plant WDC to ensure appropriate containers for the waste generated are used. Subcontractor drilling teams normally will provide containers for their field activities; however, the subcontractor Project Manager will obtain containers for supporting subcontractor field activities and the decontamination pad in the central staging area.

Empty containers that are stored in the central staging area while awaiting use must be kept on clean pallets in good condition.

Accessory equipment required includes pallets (including clean pallets for new containers), vent plugs, bung wrenches, plastic liners for open-top-drums, labels (see Sect. 10.8.2), PPE, and handling equipment (such as forklifts and/or a pickup or flatbed truck with a hydraulic tailgate and tie-down straps).

10.8.1.1 RCRA hazardous waste container requirements

RCRA hazardous waste will be placed in containers or containers with liners that are compatible with the hazardous wastes being stored. Tanks containing RCRA hazardous waste will have a secondary containment system that includes a liner, a vault, or an equivalent device.

10.8.1.2 TSCA waste container requirements

PCB or TSCA waste containers will comply with the specifications of 49 *CFR* 178. Nonliquid PCB waste will be placed in containers that comply with 49 *CFR* 178.

10.8.1.3 Low-level waste

LLW will be placed in containers that are approved in writing by the Waste Management Officer.

10.8.1.4 Mixed waste—RCRA/LLW, LLW/TSCA, and LLW/RCRA/TSCA

Mixed waste will be placed in containers compatible with the waste contents and will meet the regulatory requirements for each component of the waste characterization.

10.8.1.5 RCRA/TSCA

RCRA/TSCA waste will be contained in a DOT-specification container compatible with the RCRA waste and in compliance with the DOT waste container specification for PCB waste.

10.8.1.6 Sanitary waste

Sanitary waste will be stored and handled in accordance with the generator's agreements with site-specific utilities and maintenance personnel.

10.8.1.7 Empty containers

Empty containers that are stored in the central staging area while awaiting use must be kept on clean pallets in good condition and labeled with an "Empty" label. Labels are never to be placed on the top of the container lid. Labels are to be applied adjacent to but not covering the container's seam which is to be visible for inspection purposes.

Accessory equipment required includes 120- by 120-cm (48- by 48-in.) pallets (including clean pallets for new containers), vent plugs, bung wrenches, plastic liners for open-top drums, labels (see Sect. 10.8.2), PPE, and handling equipment (such as forklifts and/or a pickup or flatbed truck with a hydraulic tailgate and tie-down straps).

10.8.2 Labeling

Each container of waste generated during field investigations at the Y-12 Plant will have a label identifying its contents. Information on the container contents label will include contents, location generated, hazard class (if known), name of contact, and the date that the

waste was first placed in the container. Specific and any additional requirements for labeling are provided in the appropriate plant-specific waste operations procedures and/or ER site procedures.

In addition to the container contents label, other labeling requirements are required for wastes known or suspected to be hazardous and for wastes determined to be radioactive, as specified in the waste management project plan. The RI Subcontractor Project Manager will coordinate with the Y-12 Plant WDC to obtain the appropriate labels for IDW containers.

Labeling requirements are further described in Y-12 PP 70-903. The RI Subcontractor Project Manager will ensure that the following tags/accessories are available:

- Forms UCN-2114A and UCN-2114B—Hazardous Waste Identification and Waste Identification Tags
- Form UCN-100732—Classified Information Tag
- Form UCN-14B—Health Physics Material Transfer Clearance Form
- Form UCN-15C—Contamination Radiation Form
- Form UCN-2491B—Health Physics Restriction Tag
- Form UCN-15265—Empty Label
- Form UCN-11904A—Beryllium Label
- “<50 ppm PCBs” and “PCB Label” labels
- Self-adhesive vinyl envelopes
- “CONTAINER” bar code labels (tankers, polytanks, or tuff tanks will have permanent bar code labels applied by the Y-12 Plant Waste Management Division)

10.9 WASTE ESTIMATES

Volume estimates for each anticipated waste stream are given in Table 10.1. These estimates were prepared based on anticipated sampling activities described in Chap. 7, Table 7.3. Table 10.4 provides an estimate of waste generation volumes per site during the BCV OU 1 RI.

10.10 SAMPLE MANAGEMENT—WASTE ANALYSIS AND CHARACTERIZATION PLAN FOR DISPOSAL

The subcontractor will manage samples collected during field investigations and sampling activities investigations at the Oak Ridge Y-12 Plant in accordance with relevant federal, state, and DOE guidelines and existing Energy Systems policies and procedures. Samples collected during these CERCLA driven activities shall be collected in accordance with Energy Systems Environmental Surveillance Procedures, ESP-301-1 through 301-5; ESP-302-1 through 302-5; ESP-303-1 through 303-7; ESP-304-1 and ESP-304-2; ESP-305-1 and 305-2; ESP-306; and Y-12 Plant Procedures, “Waste Transportation, Storage, and Disposal

Table 10.4. Waste generation per site for the BCV OU 1 RI

Site	Media	Source	Number of drill holes	Auger diameter (in.)	Depth (ft)	Minimum volume estimate (ft ³)
S-3	Soil	Soil borings	5	6	20	19.63
	Soil	Soil borings	2	6	28	11.00
Oil Landfarm	Soil	Shallow soil	6	6	0.5	0.59
	Soil	Soil borings	6	6	16	18.85
Boneyard/ Burnyard	Soil	Shallow soil	25	6	0.5	2.45
	Soil	Soil borings	25	6	16	78.54
Burial Grounds	Soil	Shallow soil	23	6	0.5	2.26
	Soil	Soil borings	23	6	20	90.32
Subtotal volumes						133.32
Volume increase factor				2	Estimated volume	266.64

Department Sampling Procedures.” Samples will be categorized as “Classified” or “Non-Classified” and will be packaged, labeled, marked, and shipped for analysis accordingly. “Non-Classified” samples will be managed in accordance with the procedure, “Sample Classifying, Packaging, Marking, Labeling, and Shipping for Analysis Through Y-12 and K-25 Environmental Restoration Programs.” QA Samples and sample residuals will be archived as required by the Project Manager. An overview of the applicable provisions of the referenced procedures and management strategies is provided in the sections that follow.

Sample management is regulated at the federal and state level. Hazardous waste samples as defined in 40 *CFR* 261.3 are exempt from RCRA regulation if the terms of 40 *CFR* 261.4(d) are met. This exemption does not apply if the laboratory determines that the waste is hazardous because the laboratory is no longer meeting any of the conditions stated in paragraph (d)(1) of that section. No such standards, exclusions, or limitations exist for hazardous substance samples. However, managing of hazardous substance samples as if they were hazardous waste may be appropriate based on Section 104(c)(3)(B) of CERCLA, which requires that any off-site treatment, storage, or disposal of hazardous substances be conducted in compliance with Subtitle C of the Solid Waste Disposal Act.

10.10.1 Classified Samples

Samples will be categorized as “classified” when the sample, due to its composition, structure, or function, either directly or through analysis, reveals restricted data or other

classified information. Y-12 Plant ER personnel must be contacted and will determine whether the samples will be categorized and managed as classified.

10.10.1.1 Collecting

Samples will be collected in accordance with the ESP procedures governing the specific sampling activity (e.g., groundwater sampling). Additionally, special precautionary actions must be taken to ensure that information recorded in the field logbook and the Request for Analysis form does not include classified information or include enough bits and pieces of information that will identify the sample as classified. If this is unavoidable, the field logbook and Request for Analysis form must be treated and managed as classified material. (A unique classified numbering system will be developed for classified samples so that no one other than the individual maintaining the numbering system list can identify the sample as classified by the sample number alone.)

10.10.1.2 Packaging

All classified samples must be packaged such that the classified component of the waste is not revealed during routine material handling and transport. Y-12 Plant ER and the Waste Treatment, Storage, and Disposal (WTSD) Department will be contacted for specific packaging requirements. Waste and containers must be compatible. Containers will be labeled with the properly completed Form UCN-2114A (tag).

10.10.1.3 Documentation

The following documentation must be generated and maintained as a minimum for sample management: quality control records, field logbook, field laboratory equipment calibration and maintenance records, chain-of-custody records, and shipping papers.

Chain-of-Custody Records. The chain-of-custody form will be completed by the sampling team in accordance with ESP-500, "Manual Chain of Custody Procedures," at the time of the sampling event. This record documents sample custody transfer from the sampler, often through another person, to the sample custodian in the laboratory. The form(s) will be signed as relinquished or received each time the sample changes possession, from collection to final deposition in the laboratory. The chain-of-custody record will include the following information:

- Header information (project number and name, Contract Laboratory case number or SAS number); for each station number, enter date, time, composite/grab, station location, number of containers, analytical parameters, and Sample Identification Tag number
- Signature, date, and the time next to "Relinquished by" entry
- Person receiving the sample signature next to the "Received by" entry, or enter the name of the carrier (e.g., United Parcel Service, Federal Express) under "Received by" (Receiving laboratory will sign "Received for Laboratory by" on the lower line and enter the date and time)
- The bill-of-lading or Federal Express air bill number under "Remarks or Reason for Change of Custody," if appropriate.

Place the original (top, signed copy) of the chain-of-custody record form in the appropriate sample shipping package. Retain a copy with project files.

Quality control records. Sample labels and sample seals will be used during all sampling events as quality control documents. Sample labels will be affixed to all sample containers prior to or at the time of sampling. To the extent practicable, sample bottles will be labeled prior to filling. Sample labels will be waterproof paper or plastic with gummed backs or waterproof tags, as appropriate. Sample labels will contain the following information:

- Unique sample number
- Sample location
- Date and time of sample collection
- Sample media (oil, water, soil, sludge, other)
- Analyses required
- Comments and special precautions as needed
- Sampler signature

Sample seals will be used to detect possible tampering with samples following sample collection and prior to the time of analysis. The seal will be attached in such a way that it is necessary to break the seal in order to open the sample container. Samples that are destined for off-site laboratories, for overnight storage, or to leave the sampler's site will have a sample seal affixed to the sample container or will be locked to prevent tampering with samples. Sample seals are not required for containers delivered directly from the field to the on-site laboratory. Sample seals will include the following information:

- Requisition number or chain-of-custody number
- Name(s) of collector(s)
- Date and time of sample collection

Field logbook. Field logbooks will be maintained during all sampling events. An entry will be made to the field logbook and completed at the time the sample is taken. This entry will include the following information at a minimum:

- Unique sample number
- UCN-2109 number
- Date and time of sample collection
- Sample media (oil, water, soil, sludge, etc.)
- Suspected composition, including concentrations, as appropriate (e.g., mercury or PCBs)
- Analytical parameters to be measured
- Preservatives used and number of sample bottles
- Location of sampling point (e.g., field or building coordinates of sample location, grid number, well number, etc.)

- Designation of QC samples (e.g., duplicates, rinsates, or blanks)
- Sampling methodology (equipment used and depth in container from which sample was taken)
- Observations made during sampling (e.g., odors, colors, and textures)
- Chain-of-custody control number
- Field observations and measurements
- Sample distribution (appropriate lab)
- Identification of sampling team
- Diagonal line drawn from the beginning of the last entry to the next to the bottom line (for any partially empty pages at the end of the day)

At the completion of each day, the sampling team will give the field logbook to the sampling team leader for review. The sampling team leader will review all log entries and will sign off daily on the last line on the bottom of the page.

Field equipment calibration and maintenance logbook. Calibration and maintenance activities must be entered into the respective equipment calibration or maintenance logbook. These entries should be logged at the time the activities are initiated and completed and the results of each documented.

10.10.2 Nonclassified Samples

Samples that are not “classified” will be termed “nonclassified.” Nonclassified samples will include environmental, hazardous, or radiological samples. Based on best judgement of regulatory knowledge, historical knowledge of the site, and field screening activities, the sampling team leader or designee will categorize the sample. A sample will be categorized as radiological based on historical knowledge or when readings three to five times above background are detected. A sample will be categorized as hazardous when its characteristics indicate that it has the potential to contain hazardous materials. These samples are to be shipped and packaged in accordance with DOT regulations. Environmental samples include samples of naturally occurring matrices such as soil, sediment, groundwater, surface water, and air. Environmental samples containing hazardous materials are exempt from the RCRA packaging, labeling, marking, transporting, and manifesting requirements when these samples are shipped for analysis.

10.10.2.1 Collecting

Samples will be collected in accordance with appropriate ESP procedure(s) for a specific sampling activity. The sampling team leader is responsible for ensuring the sample team members are knowledgeable of the proper collecting techniques.

10.10.2.2 Packaging

Nonclassified samples must be packaged in accordance with the procedure, “Sample Classifying, Packaging, Marking, Labeling, and Shipping for Analysis Through the K-25 and Y-12 Environmental Restoration Programs.” Y-12 ER and WTSD Department will be

contacted for specific packaging requirements. Waste and containers must be compatible. Containers will be labeled with the properly completed Form UCN-2114A (tag). A brief description of these packaging requirements are discussed below for each respective sample classification.

Radiological samples. Radiological samples will be packaged in accordance with EPA, Energy Systems guidance, and DOT requirements. Accordingly, the inner package of radiological samples are packaged by:

- Wiping down the sample container exterior, placing the container in at least one plastic bag, and securely taping the bag shut with duct tape (note: closure of Zip-Loc bags is not sufficient; these bags must be taped closed)
- Marking inner packages with appropriate radioactive materials tags (UCN-2785) based on the results of radiological survey(s) conducted by the survey technician
- Annotating the dose rate at the surface of the container provided by survey technician

These inner packages will be placed in either a strong tight container or a Type A package for use as required by the category of shipping. Samples will be packed to prevent shifting of the inner containers during shipment. *Note: Chempack™ or other K-25/Y-12 approved packing materials will be used to pack all water and soil samples.*

Samples requiring Type A packaging will be packaged in the same manner as LSA materials, with the additional requirement that the cooler or strong tight container must be placed in a steel drum that conforms to DOT specification 7A. Rigid polyethylene foam inserts will be placed around the container to keep it from shifting within the drum.

Hazardous Samples. *Samples will be packaged in a cooler unless directed otherwise by the Site Transportation Operations Group.* Hazardous materials will be packaged in accordance with DOT HM-181 requirements and procedure, "Sample Classifying, Packaging, Marking, Labeling, and Shipment for Analysis."

10.10.23 Marking and labeling of the shipping package

The sampling team leader or designee will mark and label each outer package with the proper shipping name, identification number (HM-181 Classification number), and the names, addresses, and telephone numbers of the consignee and consignor. *Note: Environmental samples do not need shipping name/number.* In addition, Type A packages will be clearly marked "DOT Type A" in letters at least 0.5 in. high. Packages containing liquids will have "This Side Up" or "This End Up" orientation markings and a double orientation arrow. Packages in excess of 49 kg (110 lb) will be plainly marked with the gross weight, and packages containing glass will be marked "Fragile." Packages sent by commercial courier will be marked in accordance with the courier's procedures.

Labels will be printed on the package or affixed to the package near the marked proper shipping name. Labels will be placed on a tag if the package has irregular surfaces. Labels will be placed on two sides or two ends of the package (excluding the bottom) and will not be obscured by markings or attachments. A "Cargo Aircraft Only" label will be placed on all Type A shipments. Place this label near the proper shipping name in such a manner that any

other labels or markings are not obscured. Packages sent by commercial courier will be labeled in accordance with the courier's procedures.

10.10.2.4 Documentation

The following documentation must be generated and maintained as a minimum for sample management: quality control records, field logbook, field laboratory equipment calibration and maintenance records, chain-of-custody records, and shipping papers. These documents will be prepared and maintained in accordance with guidance outlined in Sect. 10.10.1.3.

10.10.3 Shipment for Analysis

The sampling team leader will determine whether the sample(s) will be transported on or off site for analysis. Shipment will be conducted in accordance with ER/C-P2302, "Sample Classifying, Packaging, Marking, Labeling, and Shipping for Analysis Through the K-25 and Y-12 Environmental Restoration Programs." The sampling team leader will ensure that a copy of the contracted laboratory's radiation license is in his possession before any mixer waste or radiological samples are shipped for analysis. In addition, applicable packaging, handling, and shipping requirements identified by the contracted laboratory will be followed.

10.10.4 Analytical Waste

Conformance with 40 *CFR* 261.4(d) requires that hazardous samples be returned to their generator for proper management after the analysis. This return should be specified as an agreed-upon last task in analytical contracts for hazardous samples if the project manager wishes to avoid the effort entailed in treating the material as other than a sample. Without the RCRA sample exclusion, samples would require manifesting for shipment to the laboratory; the receiving facility would need to be a RCRA TSD facility; and off site ultimate disposal would require yet another manifest.

When samples received by laboratories are clearly not hazardous by RCRA or CERCLA definitions or are determined by analysis to be nonhazardous, they are not required to be managed in accordance with the RCRA exclusion paragraph. However, before any nonhazardous samples are disposed of as part of laboratory solid refuse or wastewater, the state and local solid waste codes and industrial wastewater discharge codes should be examined to ensure that their terms are being met. Even for these nonhazardous samples, it might be necessary to have contract conditions or additional fees to cover the disposal of samples.

Analytically derived waste may or may not be classified as hazardous wastes or hazardous substances. Nonhazardous analytical waste may be disposed of in accordance with state and local solid waste and industrial waste water discharge requirements. Typically, these wastes can be disposed of in the wastewater discharged from the laboratory to the sanitary sewer. However, analytical waste derived from listed hazardous wastes, or the chemicals used to obtain the derivative, could cause the waste to be classified as hazardous. RCRA regulations provide for such waste to be disposed of with laboratory wastewater if certain conditions are met. Wastes that are considered hazardous only because of a characteristic are no longer hazardous once they are mixed to eliminate the characteristic. Mixing small volumes of analytical waste with the sanitary sewer flow would cause the waste to become so diluted that it no longer exhibits hazardous characteristics.

CLP laboratories will assume responsibility for sample residuals at the laboratories, with the exception of mixed waste; all unused mixed waste will be returned. Non-CLP laboratories will not assume responsibility of sample residuals; accordingly, the project manager must make arrangements for the proper disposal of archived or unused samples.

10.10.5 Archiving of Samples

Unused samples returned from laboratory analysis and testing, will be characterized based on analytical analysis. Nonhazardous solid wastes will be returned to AOC; Hazardous solid wastes will be coordinated through Waste Management Organization for disposal at a hazardous waste treatment, storage, disposal facility with available capacity. Liquid hazardous wastes will be coordinated through Y-12 Plant WTO and treated at Y-12 WETF.

At EPA's direction and for quality assurance reasons, duplicate samples are often collected and stored or archived for later use. The project manager will identify samples to be archived, determine where and for how long samples will be archived, and make the necessary arrangements with the storage facility. It is important to manage these samples properly because reanalysis may be required. However, reanalyses may yield unreliable data if the samples are stored past their holding time.

Unused samples generated by the on-site analytical laboratories will be stored by the laboratories for a predetermined period before repackaging for management at the appropriate treatment, storage, or disposal facility. Unused samples generated at an off-site analytical laboratory will be returned to the site from which the sample originated. Unused samples accountability upon return from the off-site laboratory will be the responsibility of the project manager.

10.11 WASTE MINIMIZATION

As described in Sect. 10.3, efforts will be implemented to minimize the generation of IDW and to segregate waste generated during the field investigations.

10.12 PROJECT PARTICIPANTS

Table 10.1 provides names of project participants including waste generator, waste handler, waste transporter, waste treatment and disposal organizations, and responsible parties.

10.13 FIELD STAGING AREAS

In accordance with the work plan, the RI Subcontractor Project Manager will select an area within the AOC to serve as the FSA for the site. The Project Manager will base selection of the FSA on the preliminary assessment of the site and IDW volume estimates by type to be generated during RI. The FSA will be clearly posted and adequate aisle spacing will be provided. In accordance with this same procedure, a logbook will be maintained to record weekly inspections of the containers stored in this area. Record of any leaks or spills,

missing labels, tags, placards, or unusual conditions will be maintained. Inspections will be conducted weekly until completion of remediation activities.

Upon completion of RI activities, the RI Subcontractor Project Manager will decide if continued storage of the IDW is appropriate until the remedial action is implemented. Following characterization of wastes that will be transferred to a permitted TSD facility, the Project Manager will notify the Waste Management Organization of the quantities of IDW involved (i.e., those IDW that cannot be returned within the AOC) and will make arrangements for disposition of IDW and transfer of responsibility to the Waste Management Organization.

In addition to the FSA, the RI Subcontractor may need to establish project-specific equipment and material FSAs when an investigation site is located within a secured area of the Y-12 Plant that has restricted access. The ER Division Project Manager will be responsible for locating and securing access to an area that can be used for temporary staging and decontamination.

All FSAs will be clearly posted and will have barriers. FSA barriers will be constructed of fencing and will be marked with flags, identification markers, and tape, as deemed necessary. The type of barrier used will depend on the condition of the site, its accessibility/location, its security level (i.e., restricted/unrestricted), the equipment/materials stored in the FSA, and the activity or testing being performed.

Access to all FSAs will be properly maintained and controlled. Necessary precautions will be taken to prevent unauthorized access, damage, alteration, or tampering in FSAs that contain expensive or critical test equipment, irreplaceable test data, and records.

10.13.1 Locations

Waste FSAs will be designated within the confines of the BCV OU 1 RI where appropriate. Figure 10.2 denotes the general location of the FSAs. There will be three staging areas, one located in the vicinity of each waste area: the S-3 Site, the Oil Landfarm area (to serve the Oil Landfarm and the Boneyard/Burnyard, and the Burial Grounds. This will reduce the amount of sample transport on site and will reduce the possibility of cross-contamination between waste areas.

10.13.2 Special Requirements

Special requirements for the FSAs will involve the use of DOE-specification drums and poly tanks for solids and liquids as stated in Sect. 10.8. for low-level solids wastes, Waste Management Division-approved B-25 boxes will be used. Labeling requirements (Sect. 10.8.2) for these areas will apply for drums and boxes. Areas will be flagged or roped off and applicable warning signs will be posted at areas where needed indicating either a 90-d accumulation area or a radiological area. Indigenous waste (contaminated soil) will be placed in a central location in the AOC. The soil will be stockpiled in separate piles according to contamination (hazardous, low-level, or mixed waste) with adequate protection and labeling for each pile. Stockpiles will be flagged or roped off with applicable warning signs where needed, indicating characterization of waste. Poly tanks will be used to store decontamination fluids, groundwater, and well water. The poly tanks will be labeled according to the labeling requirements in Sect. 10.8.2. All wastes are to be inspected weekly and handled in a manner

Fig. 10.2. General location of FSAs.

that is protective of human health and the environment and in accordance with K-25 Site standard operating procedure (WSO-OP-2008) and in accordance with Energy Systems ER IDW policy ER/Y-P2103 (IAD).

10.13.3 Responsible Organizations

RI Subcontractor will be responsible for transporting wastes from the point of generation to designated FSAs, under the guidance and direction of the Y-12 Plant ER Division. The Y-12 Plant ER Division will be responsible for transporting the wastes from the field staging area to a permitted or interim status TSD facility.

10.14 REQUIRED TRANSPORT ACROSS PUBLIC ACCESS ROADS

IDWs generated during the RI investigations will remain on the Y-12 Plant site and are not expected to be transferred across public access roads. Before transporting waste from the field investigation site for treatment, storage, or disposal, the Project Manager, in coordination with the Y-12 Plant WDC identifies a TSD facility; completes a request for disposal package [using Form UCN-2109, Request for Transfer, Storage, or Disposal of Waste, or WSIN CALL-IN FORM or State Identification Form (see Y-12 PP 70-903, Appendix A)]; selects a route of transport; and performs a visual inspection of the waste containers. These activities are described below.

After verifying that all waste transfer requirements have been met, the Y-12 Plant WDC coordinates on-site and off-site waste transportation with Y-12 Plant waste management personnel and the Energy Systems Central Waste Management Division.

10.14.1 TSD Evaluation and RFD Documentation

The WDC prepares a request for disposal package that includes the original site-approved request for disposal, copies of analytical results, and supporting documentation [i.e., MSDSs]. The WDC reviews the request for disposal package for technical accuracy, and determines the appropriate TSD option.

10.14.2 Visual Inspection of Waste Containers

The WDC will conduct a visual inspection of waste containers before they are removed from the project site. This review shall determine if transfer standards are met and if wastes are ready for transport. Additionally, the WDC verifies, in accordance with DOT regulations and the EPA waste manifest requirements in 40 *CFR* 262, the following:

- The container count according to the manifest (RFD form for on-site transportation and the EPA uniform Hazardous Waste Manifest for off-site transportation)
- Each container is labeled properly, legibly, and in accordance with the instructions provided by the WDC on the manifest
- The contents of each container have been characterized
- The containers are sealed securely

- Each container has been surveyed by health physics personnel for external gross alpha, beta, and gamma radiation and approved for transport. If approval is granted, the WDC shall sign the RFD form and waste container tags

10.14.3 Transportation

Upon verification that all waste transfer requirements have been met, the WDC coordinates on-site and off-site waste transportation with the Site Waste Management personnel and Central Waste Management Division. Waste scheduled for removal from the project site shall be transported on a public thoroughfare to an on-site or off-site TSD facility in accordance with all DOT regulations.

10.15 WASTE STORAGE REQUIREMENTS

Waste stored for on-site or off-site disposal (outside an AOC) will be characterized, packaged, and labeled in accordance with the waste acceptance criteria associated with the designated waste storage facility.

Under certain circumstances the Project Manager may place IDW within the AOC by returning the waste to its source as described in the scenarios of the ER procedure for management of IDW. If IDW is determined to be a RCRA hazardous waste and subject to LDRs, "land disposal" of the IDW will be prohibited unless specified treatment standards are met. "Land disposal" applies to wastes from different AOCs that are consolidated into one AOC; wastes that are moved outside an AOC (for treatment or storage) and returned to the same or a different AOC; or wastes that are excavated, placed in a separate hazardous waste management unit, and then redeposited into the AOC. However, RCRA waste stored in a container (as defined in 40 *CFR* 260.10) within the AOC and then returned to its source is permissible without meeting the specified LDR treatment standards. This action does not constitute "land disposal" because a container does not constitute a hazardous waste management unit. Therefore, returning IDW that has been stored in containers (not tanks or other RCRA-regulated units) within the AOC will be an acceptable waste management option, as long as containers are not managed in such a manner as to constitute a RCRA storage unit as defined at 40 *CFR* 260.10.

10.16 RCRA SATELLITE AND 90-D ACCUMULATION AREA REGISTRATION NUMBERS AND LOCATIONS

RCRA hazardous constituents are not the primary contaminants of concern during the RI. However, if analytical data indicate that IDW is RCRA hazardous, these wastes will be stored in the satellite or 90-d accumulation areas. The satellite and 90-d accumulation areas will be registered with and approved by the Y-12 Site Environmental Management Department and Y-12 Plant Waste Management Division for waste characterized by analytical data as RCRA hazardous. A unique registration number will be assigned to each identified satellite and accumulation area.

10.16.1 Waste Storage Area Locations

Waste that is stored in 90-d accumulation or low-level waste storage areas after being characterized by analytical data will be transported to permanent storage by the Y-12 Plant WTSD.

10.16.2 Waste Acceptance Criteria Requirements

Waste acceptance criteria requirements for the satellite areas and the 90-d accumulation area will be in accordance with 40 *CFR* 262, "Standards Applicable to Generators of Hazardous Waste," with respect to containers, labeling, accumulation times, and inspections.

10.16.3 Responsible Organization

The Y-12 ER Division Project Manager, or designated representative, will be responsible for registering the satellite and accumulation areas with the Y-12 Plant ER Division RI Subcontractor, under the direction and guidance provided by the Y-12 Plant ER Division Project Manager, will be responsible for transferring wastes from the satellite area to the 90-d accumulation area, and ensuring that the standards set forth in 40 *CFR* 262 are followed.

10.17 IDENTIFICATION OF POTENTIAL TREATMENT OPTIONS AND FACILITIES

Potential treatment options identified by the Y-12 Plant ER Division for liquid wastes are the Y-12 Plant's West End Treatment Facility, the K-25 Site's Central Neutralization Facility (CNF), Central Pollution Control Facility, or the Groundwater Treatment Facility upon meeting the specific facility's waste acceptance criteria. An option identified for solid low-level wastes may be incineration and compaction by Scientific Ecology Group, Inc.. Another potential option could be the K-1435 TSCA Incinerator. Treatment options will be decided based on analytical results, waste acceptance criteria requirements for treatment facilities, and type of waste that is generated.

10.17.1 Special Waste Acceptance Criteria Requirements of Potential Treatment Facilities

Specific waste acceptance criteria for the K-25 Site CNF are specified in K/SS-538, The Oak Ridge Gaseous Diffusion Plant K-1047-H and K-1407-A Central Neutralization Facility Waste Acceptance Criteria; and the K-1435 TSCA Incinerator waste acceptance criteria are specified in K/HS-252/R2, Waste Acceptance Plan and Analytical Protocol for the K-1435 Toxic Substances Control Act Incinerator. The Y-12 Plant's West End Treatment Facility, Central Pollution Control Facility, and Groundwater Treatment Facility waste acceptance criteria are not identified in formal documents. The analytical data provided by the waste generator to the treatment facility operator are reviewed against permit limits to determine acceptance criteria are not identified in formal documents. The analytical data provided by the waste generator to the treatment facility. The Y-12 Plant WDC will review the analytical data provided by the waste acceptance criteria are not identified in formal documents. The analytical data provided by the waste generator to the treatment facility. The Y-12 Plant's West End Treatment Facility, Central Pollution Control Facility, and Groundwater Treatment Facility waste acceptance criteria are not identified in formal documents. The analytical data

provided by the waste generator to the treatment facility operator are reviewed against permit limits to determine acceptance of wastes at the treatment facility. The Y-12 Plant WDC will review the analytical data to be provided to determine if contaminant levels are below or exceed the criteria established for the treatment facilities. If the contaminant levels exceed the criteria, the wastes may require on-site storage.

10.18 IDENTIFICATION OF POTENTIAL DISPOSAL OPTIONS AND FACILITIES

The WDC prepares a request for disposal package that includes the original site-approved request for disposal, copies of analytical results, and supporting documentation (i.e., MSDSs). The WDC reviews the request for disposal package for technical accuracy, and determines the appropriate TSD option.

10.19 AMENDMENTS

Amendments and revisions to the Waste Management Plan will be documented and submitted to the ER Project Manager for approval. Amendments will be numbered and dated consecutively and submitted through the review and approval cycle.

10.20 WASTE MANAGEMENT PLAN REVIEW/APPROVAL SIGNATURES

An example of the Approval Signatures form is shown in Fig. 10.3.

*Approval signatures indicate concurrence with and approval of the contents
and intent of this WMP.*

APPROVAL SIGNATURES

Y-12 ER Waste Certification Officer

Date

Y-12 ER Project Manager

Date

SAIC Project Manager

Date

Y-12 Site Waste Management Division

Date

Prepared by: _____ SAIC

Y-12 ER Program

March 1993

Revision No. _____

Fig. 10.3. Sample Approval Signatures Form.

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