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RESULTS FROM LONG-TERM DISSOLUTION TESTS USING OXIDIZED SPENT FUEL

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ABSTRACT

Two semi-static dissolution tests using oxidized PWR spent fuel specimens are being conducted under ambient hot cell conditions in Nevada Test Site J-13 well water and unsealed fused silica vessels. The test specimens were oxidized at 250° C in air to bulk oxygen-to-metal (O/M) values of 2.21 and 2.33. Following an initial 191-day test cycle, the specimens were restarted in fresh J-13 water for a second long-term test cycle. Results through the first 40 months of Cycle 2 are compared with results from similar tests at 25° C and 85° C using unoxidized spent fuel specimens.

Increased concentrations of U, Am, Cm and Np were measured in 0.4- μ m filtered samples from the oxidized fuel tests compared to the unoxidized fuel tested at 25°C; Pu concentrations were not affected by the fuel oxidation state. Most of the Am and Cm, and a portion of the Pu, measured in 0.4- μ m filtered samples was removed by 2-nm filtration. Fission product release results were normalized to specimen inventories and reported as fractional release. No attempt was made to normalize the data to surface area. Initial ⁹⁹Tc release was greatly increased, and prolonged increases in the fractional release rates of ⁹⁹Tc and ¹²⁹I occurred as a result of fuel oxidation. Fractional release to levels similar to those observed with unoxidized fuel after equivalent testing times, suggesting that matrix dissolution rates normalized to fuel mass were not increased as a result of oxidation.

INTRODUCTION

Spent fuel disposed of in a geological site located above the water table in the unsaturated zone, such as the Yucca Mountain site under investigation in the state of Nevada, may be exposed to oxidizing conditions (air) for extended periods prior to contact by water. Under such conditions spent fuel may potentially be degraded or altered as a result of oxidation of the UO₂ fuel matrix. With the exception of ¹⁴C, which may be transported in the vapor phase, dissolution and transport as a result of groundwater flow is generally accepted as the primary mechanism by which radionuclides could be released from a repository to the biosphere[1]. Laboratory studies of spent fuel dissolution and radionuclide release behavior have typically been conducted using unoxidized specimens prepared from freshly sectioned fuel rod segments. Two scoping tests using air oxidized spent fuel specimens were started in August 1986 in support of the Waste Package Task of the U.S. Department of Energy Nevada Nuclear Waste Storage Investigations (NNWSI)^(b). The purpose of these tests was to provide preliminary data on the dissolution behavior of oxidized fuel prior to testing with fuel specimens from long-term low-temperature spent fuel oxidation tests [2], which were in progress at that time and are continuing as of this reporting.

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(b) Renamed the Yucca Mountain Project, the Waste Package Task is managed by Lawrence Livermore National Laboratory.

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TEST DESCRIPTION

The oxidized fuel scoping tests were similar to the NNWSI Series 2 Bare Fuel Tests [3], which were conducted at ambient hot cell temperature in 250 mL of Nevada Test Site J-13 well water. Ambient hot cell temperature varied over a range of about 22°C to 28°C depending on time of day and time of year and is estimated to have averaged 25°C. Turkey Point fuel particles were oxidized in air at 250°C for time periods of 45 hours and 116 hours producing test specimens having bulk oxygen-to-metal (0/M) stoichiometries of 2.21 and 2.33 based on weight gain. Burnup was approximately 27,000 MWd/MTU and fission gas release was reported to be 0.3% for this fuel. The fuel particles were screened to +10 mesh prior to test specimen preparation, and geometric surface area is estimated at about 2 cm²/g before oxidation.

Microstructures of fuel particles removed during Cycle 2 of each test are shown in Figure 1. The particle from the 0/M = 2.33 test appears to be relatively uniformly oxidized. The surface of this ceramographic section was double impregnated to reduce grain pull out during preparation. Only a few of the larger grains in the higher magnification photographs of this particle show unoxidized UO, visible as a darker gray phase at the center of the grains. The bright areas in the lower magnification photograph of the particle from the 0/M = 2.21 test are areas where the UO_2 fuel matrix did not exidize. The higher magnification composite strip shows fracturing and pull out of the more oxidized grains near the particle edge. A larger average grain size in this particle may indicate that the particle was from the center of a fuel pellet where a higher irradiation temperature induced grain growth, although other examinations of fuel from this fuel assembly have suggested variations in the as-fabricated fuel microstructure. The highest magnification photograph of this particle shows oxidation occurring initially along the grain boundaries forming oxidized shells surrounding unoxidized UO, grain cores. Oxidation appears to have caused grain cracking and the loss of several loose grains during sample preparation.

Solution samples (20 mL) were periodically removed from the vessels for analysis and the solution sample volume replaced with fresh J-13 well water to maintain a constant water volume during the tests. The first test cycle was terminated after 191 days, and the fuel specimens were restarted in new vessels and fresh J-13 well water for a second test cycle. Cycle 2 was not terminated and is still in progress as of this reporting. Samples were divided to provide unfiltered, 0.4- μ m filtered and 2-nm filtered aliquots for analysis. The unfiltered aliquots were typically analyzed for pH, U, 239+240Pu, 244Cm and 129I. The 0.4- μ m filtered aliquots were typically analyzed for U, 239+240Fu, 241Am, 244Cm, 237Np, 137Cs, 90Sr and 99Tc. The 2-nm filtered aliquots of Cycle 2 samples were typically analyzed for 239+240Pu, 244Am and 244Cm. Nonradiochemical inorganic analyses were also performed on 0.4- μ m filtered aliquots by inductively coupled plasma (ICP) emission spectrometry, ion chromatography, and for HCO₃. J-13 well water is a relatively pure low-ionic-strength groundwater containing about 170 μ g/mL dissolved NaHCO₃, 12 μ g/mL Ca and 30 μ g/mL Si as its major dissolved components. Very little change in the J-13 water nonradiochemical inorganic analyses occurred during the 25°C tests and pH remained in a range of 7.9 to 8.5. Partial precipitation of Ca and Si and reduction in bicarbonate concentration was noted in the 85°C test.

Results from the two oxidized fuel dissolution tests are compared to results from the 25°C Series 2 and 85°C Series 3 tests using unoxidized (0/M = 2.00) Turkey Point fuel specimens from the same fuel assembly [3,4]. The two oxidized fuel specimens were tested in unsealed fused silica vessels identical to those used in the Series 2 tests, whereas the 85°C Series 3 test was conducted in a sealed stainless steel vessel. The oxidized fuel specimens weighed approximately 10 g each. The Series 2 specimen weighed about 27 g and the Series 3 specimen weighed about 85 g. The Series 2 and 3 specimens were unscreened and included the empty cladding hulls; the oxidized fuel tests did not include cladding. The Series 2 specimen ran for

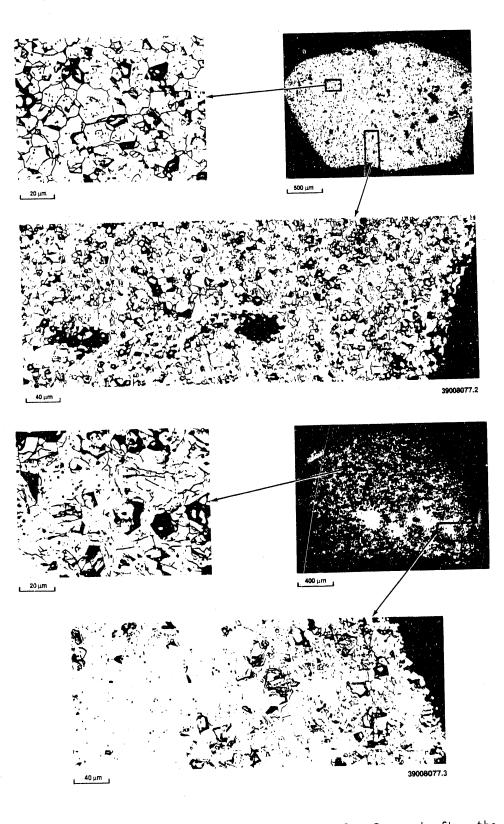


Figure 1. Ceramographic Sections of Fuel Particles Removed after the 1196day Cycle 2 Solution Samples. Particle from the O/M = 2.33 test shown above, and particle from the O/M = 2.21 test shown below. Etched 85% $H_2O_2 - 15\% H_2SO_4$.

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five test cycles of 181, 195, 224, 240, and 132 days. The Series 3 specimen ran for three test cycles of 174, 181, and 97 days.

RESULTS AND DISCUSSION

The actinides (U, Np, Pu, Am and Cm) have low solubilities in neutral aqueous solutions and reached solubility-limited steady-state concentrations during the tests. The solution concentrations of U and activities of $^{239+240}$ Pu and 241 Am measured in 0.4- μ m filtered samples are plotted in Figure 2. Maximum U concentrations in the oxidized fuel tests reached about 20 μ g/mL during Cycles 1 and 2, decreasing to about 10 μ g/mL later in Cycle 2. With unoxidized fuel at 25°C, U concentrations decreased from about 4 μ g/mL during Cycle 1 to about 1 μ g/mL at the end of Cycle 5, with lower values measured in the initial samples from each of the later test cycles. Lower U concentrations in the range of 0.1 to 0.2 μ g/mL in the 85°C test are attributed to precipitation secondary uranium phases. Uranophane, a calcium-uranium-silicate phase, was identified in the 85°C test [4]. Fuel oxidation did not affect the 100-200 pCi/mL $^{239+240}$ Pu activities measured in 0.4- μ m filtered samples at 25°C. However, $^{239+240}$ Pu activities decreased to less than 1 pCi/mL in the 85°C test. A similar effect of temperature was observed on 241 Am activities. Activities measured for 244 Cm were very similar to those measured for 241 Am in all samples.

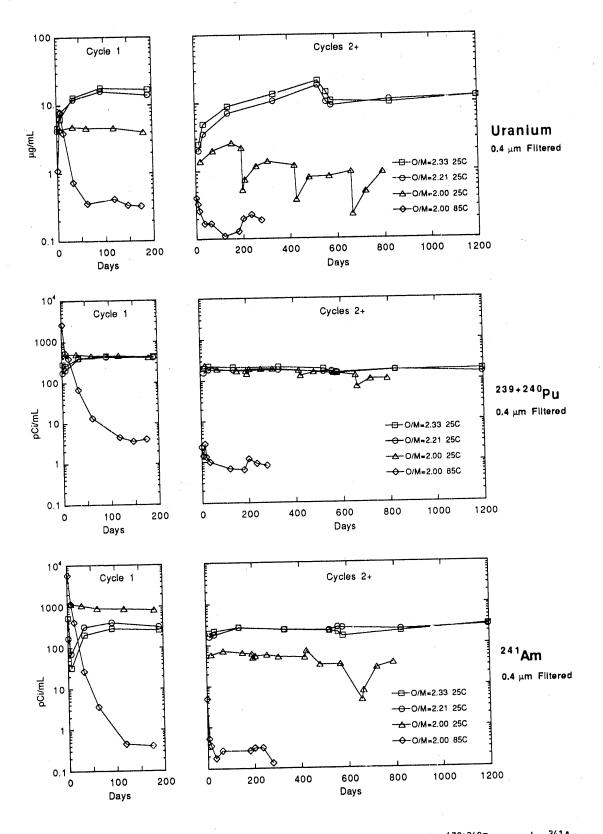
Approximate concentrations at which the actinide elements appeared to stabilize during later test cycles are compared in Table I for the 0.4- μ m and 2-nm filtrations. These data suggest that about 40-80% of the Pu and more than 90% of the Am measured in C.4- μ m filtered samples from the 25°C tests were associated with colloids. Lower concentrations and reduced effects of filtration in the 85°C test suggest that formation of non-suspended stable secondary phases was favored over colloid formation at the higher temperature.

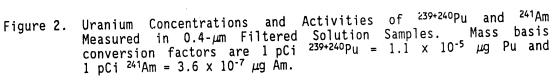
Fission product release data were normalized to specimen inventory to obtain fractional release results. Results for 137 Cs, 90 Sr, and 99 Tc are plotted in Figure 3. Each plotted data point includes the inventory fraction in the 250 mL of water in the test on the sampling day plus the inventory fractions calculated to have been removed in prior samples from the test cycle. Cumulative release is plotted for Cycles 2-5 of the 25°C and Cycles 2 and 3 of the 85°C unoxidized fuel tests. Fractional releases measured during Cycle 1 are given in Table II. The effect of temperature on initial releases of 137 Cs appeared to be greater than the effect of oxidation. However, initial 137 Cs release may have been reduced in the oxidized

| Table I. | Actinide | Concentrations, | $Log(\mu g/mL)$ | |
|----------|----------|-----------------|-----------------|--|
|----------|----------|-----------------|-----------------|--|

| <u>Actinide</u> | Filter | 0/M=2.33 | 0/M=2.21 | 0/M=2.00 | 0/M=2.00 85°C |
|-----------------|------------------------|--------------|--------------|--------------|------------------|
| U | 0.4 <i>µ</i> m 2 nm | 1.1 | 1.1 | 0.0 | -0.8 -0.8 |
| Pu | 0.4 µm 2 nm | -2.7 -3.3 | -2.7 -3.5 | -2.7 -2.9 | -5.0 -5.0 |
| Am | 0.4 μm 2 nm | -4.1 -5.4 | -4.0 -5.4 | -4.7 -6.1 | -7.0 -7.3 |
| Np | 0.4 μm 2 nm | -2.6 | -2.6 | -3.2 | -3.8 <-3.6 |

Estimated steady-state concentrations based on Cycle 2 data for the oxidized fuel tests, Cycles 2-5 of the $0/M = 2.00 \ 25^{\circ}C$ test and Cycles 2 and 3 of the $0/M = 2.00 \ 85^{\circ}C$ test.





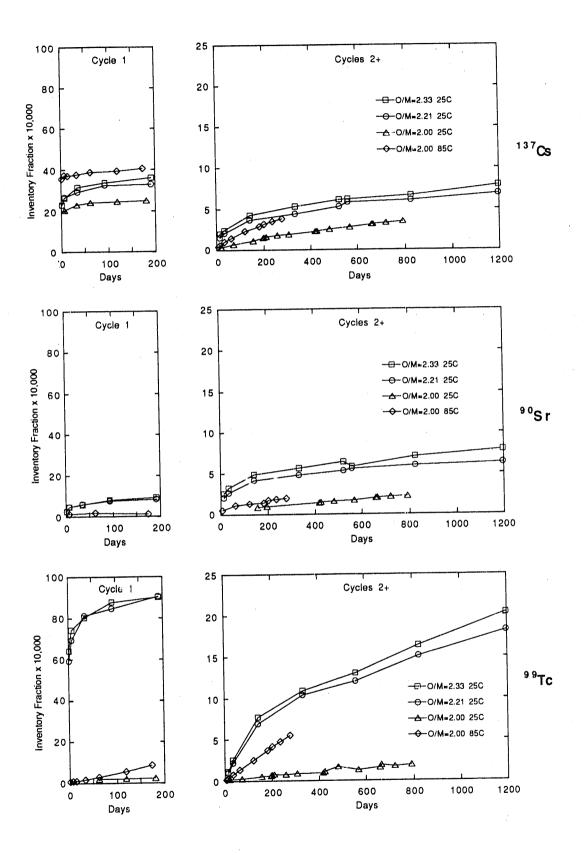


Figure 3. Fractional Releases of Soluble Fission Product Radionuclides ¹³⁷Cs, ⁹⁰Sr, and ⁹⁹Tc

fuel tests due to the lack of cladding in these tests. Initial $^{129}\mathrm{I}$ release was increased both as a result of oxidation and increased temperature. $^{90}\mathrm{Sr}$ was not measured during Cycle 1 of the 25°C unoxidized fuel test and precipitated during the 85°C tests.

The substantial increase in ⁹⁹Tc release as a result of oxidation is thought primarily to be the result of increased release from grain boundaries. Gray and Strachan [5] have also observed increased ⁹⁹Tc release from separated fuel grains as a result of fuel oxidation, indicating that the increased release is not entirely due to increased water penetration of grain boundaries as a result of oxidation. One hypothesis is that Tc is concentrated in finely dispersed metallic phases present at the grain boundaries that are insoluble in their reduced state, and that the solubility of these particles is also increased when the fuel is oxidized.

Table III gives annual fractional release rates for the soluble fission product radionuclides during later time periods of the tests after much of the initial enhanced release from exposed gap and grain boundary inventories had been completed. An unexpected result is that the ¹³⁷Cs and ⁹⁰Sr fractional release rates in the oxidized fuel tests eventually decreased to levels similar to those observed with unoxidized fuel after equivalent testing times, even though grain boundary opening may have increased wettable fuel surface area as a result of oxidation. If ¹³⁷Cs and ⁹⁰Sr release rates are indicative of fuel matrix dissolution rates at later time periods of the tests, these data suggest that inventory normalized matrix dissolution rates were not increased by oxidation, and that matrix dissolution rates normalized to surface area were not increased as a result of oxidation. However, matrix dissolution rates measured in water types such as J-13 well water, which contains dissolved Ca, Si, NaHCO₃ and O₂, may be ambiguous. Flow-through tests with unirradiated UO₂ in such water types have suggested that the dissolution rate of UO₂ may decrease with time as result of Ca and/or Si incorporation with the UO₂ surface [6].

| Nuclide | 0/M=2.33 | 0/M=2.21 25°C | 0/M=2.00 | 0/M=2.00 85°C |
|-------------------|--------------|------------------|--------------|------------------|
| ¹³⁷ Cs | 362 | 330 | 251 | 406 |
| 90Sr | 92 | 84 | · | |
| %Tc | 903 | 906 | 25 | 88 |
| 129 I | 159 | 129 | 29 | 104 |

Table II. Cycle 1 Fractional Releases x 100,000

Table III. Annual Fractional Release Rates x 100,000

| Nuclide | 0/M=2.33 | 0/M=2.21 25°C | 0/M=2.00 | 0/M=2.00 85°C |
|------------------|--------------|------------------|--------------|------------------|
| 137Cs | 12 | 11 | 12 | 46 |
| ⁹⁰ Sr | 10 | 8 | 8 | · |
| ⁹⁹ Tc | 44 | 39 | 7 | 73 |
| 129 I | 62* | 58* | 8 | 43 |

Average rates for days 140-1196 of cycle 2 for oxidized fuel tests, cycles 3 through 5 for the $0/M = 2.00 25^{\circ}C$ test, and cycles 2 and 3 for the $0/M = 2.00 85^{\circ}C$ test.

* Average for days 0-1196 of Cycle 2.

CONCLUSIONS

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Fuel oxidation appeared to cause moderate increases in the eventual steady-state concentrations measured for the actinide elements U, Am, Cm and Np at 25°C. Pu concentration was not affected by fuel oxidation. Am and Cm, and to a lesser extent Pu, appeared to associate with colloids in the 25°C tests. Lower actinide concentrations were generally measured in the 85°C test where kinetic factors appeared to favor formation of nonsuspended secondary phases. Fuel oxidation resulted in a substantial increase in the initial release of 99 Tc, and in greater fractional release rates for 99 Tc and 129 I over the duration of the tests. Fractional release rates measured for 137 Cs and 90 Sr in the oxidized fuel tests eventually decreased to levels similar to those measured with unoxidized fuel after equivalent testing times, suggesting that inventory normalized fuel matrix dissolution rates in J-13 well water were not increased as a result of oxidation. Fractional release rates for the soluble fission products 137 Cs, 99 Tc and 129 I from unoxidized fuel increased by factors ranging from about 4 to 10 when temperature was increased from 25°C to 85°C.

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