**Environmental Consequences of
Postulated Plutonium Releases from
Atomics International's Nuclear
Materials Development Facility
(NMDF), Santa Susana, California, as
a Result of Severe Natural Phenomena**

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ENVIRONMENTAL CONSEQUENCES OF POSTULATED
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SUMMARY

Potential environmental consequences in terms of radiation dose to people are presented for postulated plutonium releases caused by severe natural phenomena at the Atomics International's Nuclear Materials Development Facility (NMDF), in the Santa Susana site, California. The severe natural phenomena considered are earthquakes, tornadoes, and high straight-line winds. Plutonium deposition values are given for significant locations around the site. All important potential exposure pathways are examined. The most likely 50-year committed dose equivalents are given in Table 1 for the maximum-exposed individual and the population within a 50-mile radius of the plant. The maximum plutonium deposition values likely to occur offsite are also given in Table 1.

The most likely calculated 50-year collective committed dose equivalents are all much lower than the collective dose equivalent expected from 50 years of exposure to natural background radiation and medical x-rays. The most likely maximum residual plutonium contamination estimated to be deposited offsite following the earthquake, and the 150-mph and 170-mph tornadoes are above the Environmental Protection Agency's (EPA) proposed guideline for plutonium in the general environment of $0.2 \mu\text{Ci}/\text{m}^2$. The deposition values following the 110-mph and the 130-mph tornadoes are below the EPA proposed guideline.

**TABLE 1. Most Likely 50-Year Committed Dose Equivalents
and Plutonium Deposition Values**

Event	Organs of Reference	50-Year Committed Dose Equivalent ^(a)		Maximum Plutonium Deposition Offsite ($\mu\text{Ci}/\text{m}^2$)
		Population (person-rem)	Nearest Residence (rem)	
Earthquake	Lungs	2.0×10^5	3.0	30
	Bone	3.8×10^5	5.6	
110-mph Tornado	Lungs	0 ^(b)	0	0
	Bone	0	0	
130-mph Tornado	Lungs	7.5×10^3	1.2×10^{-2}	5.7×10^{-3}
	Bone	1.4×10^4	2.2×10^{-2}	
150-mph Tornado	Lungs	3.3×10^5	0.44	0.21
	Bone	6.0×10^5	0.80	
170-mph Tornado	Lungs	1.4×10^5	0.10	0.73
	Bone	2.6×10^5	0.19	

(a) Translocation Class Y has been assumed.

(b) The most likely release from the 110-mph tornado is insignificant ($<10^{-5}$ g).

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INTRODUCTION

This study estimates the potential environmental consequences in terms of radiation dose to people resulting from postulated plutonium releases caused by severe weather or other natural phenomena. The accident scenarios considered include earthquakes, tornadoes, and high winds. This is one in a series of reports on commercial plutonium fabrication facilities sponsored by the U.S. Nuclear Regulatory Commission and coordinated by Argonne National Laboratory.

Figure 1 illustrates the information requirements for such a study and how the data are utilized to estimate dose. The amount and form of plutonium released into the atmosphere was estimated by Mishima and Ayer (1981). The atmospheric transport and dispersal of released plutonium was estimated by Pepper (1980) for tornadoes, and by the NRC for earthquakes.^(a) The site characteristics and demography around Vallecitos, California were provided by the NRC.^(b) The population distribution given in Table 2 was used to calculate the population doses.

-
- (a) Annual average atmospheric dispersion values for the Site transmitted by a letter from J. E. Carson of ANL to J. D. Jamison of PNL, June 3, 1980.
(b) "Description of the Site Environment," transmitted by letter from Leland C. Rouse of NRC to Rockwell International, Attn: Dr. M. E. Remley, May 7, 1980.

FIGURE 1. Accidental Environmental Consequences Evaluation

TABLE 2. 1980 Population Distribution Around The Atomics International NMDF(a)

Distance (mi)	0-1	1-2	2-3	3-4	4-5	5-10	10-20	20-30	30-40	40-50
<u>Direction</u>										
N	0	4	690	3,642	8,694	210	1,150	535	781	304
NNE	0	0	86	4,002	1,153	131	2,241	870	825	822
NE	0	0	115	4,318	3,310	7,045	11,572	21,662	5,220	41,587
ENE	0	0	0	676	820	4,277	47,051	4,338	4,000	18,925
E	0	0	0	892	5,613	22,408	341,737	188,251	147,971	89,996
ESE	0	0	0	72	9,644	65,555	272,816	811,904	1,075,509	834,921
SE	0	0	58	303	7,774	32,012	50,693	585,058	1,042,501	811,545
SSE	0	20	14	0	360	2,384	10,165	0	38,612	12,422
S	0	0	0	29	130	1,465	1,912	0	0	0
SSW	0	0	0	0	2,995	3,604	2,928	0	0	0
SW	0	0	0	0	892	6,871	3,194	0	0	0
WSW	0	0	0	0	0	35,903	19,615	37,296	0	0
W	0	0	0	0	0	5,809	33,993	158,872	26,144	0
WNW	0	0	0	2,692	4,644	1,778	11,548	22,460	24,688	13,578
NW	0	43	4,894	9,961	9,356	1,740	5,731	5,856	793	117
NNW	0	0	4,606	12,666	3,527	13	7,116	54	69	1,018
Total	0	67	10,463	39,253	58,932	191,205	824,182	1,837,156	2,367,113	1,825,235
										Total 7,153,606

(a) The population distribution around Santa Susana, California provided by the NRC was based on 1970 census data. Using county growth rate information from the Statistical Abstracts of the United States and other sources, estimates of the 1980 sector populations were made to give a truer picture of dose consequences.

ENVIRONMENTAL EXPOSURE PATHWAYS FOR PLUTONIUM

The potential environmental exposure pathways for radionuclides released to the atmosphere and water are shown in Figure 2. Our experience has shown that the more important pathways for exposure to atmospheric releases of plutonium and daughter products are inhalation, cloud submersion, ingestion, and direct ground irradiation. For chronic atmospheric releases of plutonium, the most important of these pathways is inhalation (Selby 1975, Friedman 1976, Anspaugh et al. 1975, EPA 1977). It can also be shown that inhalation is the only important pathway for acute atmospheric releases of plutonium (see Appendix A). Therefore, only the radiation doses from inhalation during initial cloud passage and from inhalation of resuspended environmental residual contamination are calculated.

FIGURE 2. Potential Exposure Pathways for Radionuclides in the Biosphere

For liquid releases during a flood, the important exposure pathways are aquatic food ingestion, water consumption, irrigation with contaminated water and subsequent food ingestion, and shoreline exposure. The significant potential exposure pathways that have been discussed are shown in Figure 3.

FIGURE 3. Significant Potential Exposure Pathways Through Which People May Be Exposed from an Accidental Release of Plutonium

RADIATION DOSE MODELS FOR AN ATMOSPHERIC RELEASE

The equation for calculating committed radiation dose equivalents from acute inhalation is:

$$DC_{ir} = Q_i(E/Q)(BR)(DCF)_{ir} \quad (1)$$

where

- DC_{ir} • the committed dose equivalent to organ r from acute inhalation of radionuclide i , rem
- Q_i • the quantity of radionuclide i released to the atmosphere, μg
- E/Q • the accident atmospheric exposure coefficient, $\mu\text{g} \cdot \text{sec}/\text{m}^3$ per μg released
- BR • the ventilation rate of the human receptor during the exposure period, m^3/sec
- $(DCF)_{ir}$ • the acute committed dose equivalent factor, rem per μg inhaled; a number specific to a given nuclide i and organ r which can be used to calculate radiation dose from a given radionuclide intake.

Human ventilation rates for three time periods were derived from ICRP recommendations (ICRP 1975): $3.3 \times 10^{-4} \text{ m}^3/\text{sec}$ for the period 0-8 hours; $2.3 \times 10^{-4} \text{ m}^3/\text{sec}$ for 8-24 hours; and $2.7 \times 10^{-4} \text{ m}^3/\text{sec}$ for greater than 24 hours.

Fifty-year committed dose equivalent factors were calculated using the computer code DACRIN (Houston, Streng and Watson 1975). This code incorporates the ICRP Task Group Lung Model (TGLM) to calculate the dose commitment to the lung and other organs of interest (ICRP 1966). The organ masses used in the code have been modified to reflect the changes reported in ICRP-23 (1975). The translocation of americium from the blood to the organs of interest has been changed to the values suggested in ICRP-19 (1972).

Fifty-year committed dose equivalents per unit isotopic mass inhaled for particles with an AMAD^(a) of one micrometer are listed in Appendix B, Tables B.1 and B.2, for each plutonium isotope and ²⁴¹Am. The organs of interest in plutonium dosimetry are the total body, kidneys, liver, bone, and lungs.

The plutonium postulated to be released to the atmosphere is assumed to be in the form of plutonium oxides (Mishima and Ayer 1981). Lung retention, the TGLM, depends upon the chemical nature of the compound inhaled. Compounds of plutonium largely fall into Class Y (retained for years) or Class W (retained for weeks). There is no evidence of plutonium existing in the environment as Class D (retained for days). Actinides in the oxide form are currently classified as Class Y (ICRP 1972), which is assumed in this study. Doses for plutonium as Class W material, however, are included in Appendix B.

Plutonium particulates that deposit onto the ground surface from a plume can be resuspended to the atmosphere by natural processes, and subsequently inhaled by people. Therefore, ground contamination is an important factor when calculating doses via inhalation. Where deposition values were not provided (distances less than 5000 meters for the 0-2 hour period following an earthquake), the deposition velocity concept was used to estimate the plutonium deposition (Equation 2).

$$W_i = Q_i(E/Q)V_d \quad (2)$$

where

- W_i • the concentration of radionuclide i on the ground surface, $\mu\text{g}/\text{m}^2$
- Q_i • the quantity of radionuclide i released to the atmosphere, μg
- E/Q • the accident atmospheric exposure coefficient, $\mu\text{g} \cdot \text{sec}/\text{m}^3$ per μg released
- V_d • particle deposition velocity, m/sec

(a)Activity median aerodynamic diameter

The deposition velocity of plutonium particles cannot be specified exactly because it will vary depending on the size distribution of the particles, the nature of the surface on which deposition occurs, the wind speed, and other meteorological variables. The deposition velocity for plutonium has been reported to range from 1×10^{-4} to 3×10^{-2} m/sec (Selby et al. 1975, Cohen 1977, Baker 1977, Gudiksen et al. 1976). A value of 1×10^{-3} m/sec is used in this report (Baker 1977). Deposition values for tornadoes were reported by Pepper (1980). The NRC estimated deposition values during earthquakes and annual average conditions.

Resuspension rates for material deposited on the ground are time dependent and tend to decrease with time after initial deposition. Local conditions can be expected to strongly affect the rate, with rainfall, winds, and surface characteristics predominant. The exact relationships are not well-enough understood to account for these effects (Selby et al. 1975). However, the airborne concentration from resuspended material can be estimated using a resuspension factor, K. The resuspension factor is defined as the resuspended air concentration divided by the surface deposition. Values for K in the environment between 10^{-4} and 10^{-13} m⁻¹ have been measured and reported (Selby et al. 1975, Friedman 1976, Anspaugh et al. 1975, EPA 1977, Cohen 1977, FES 1974, Bennett 1975, Hanson 1975, Martin and Bloom 1975, Sehmel 1977, Healy 1977, Anspaugh 1976). Until a more general model is available, which considers all the important variables affecting the resuspension process, Anspaugh (1975) recommends using a simple time-dependent model to predict the average airborne concentration of a resuspended contaminant:

$$K(t) = 10^{-4} \exp(-0.15 t^{\frac{1}{2}}) + 10^{-9} \quad (3)$$

where

- t • time since the material was deposited on the ground, days
- 10^{-4} • resuspension factor at time $t = 0$, m^{-1}
- 10^{-9} • resuspension factor after 20 years, m^{-1}

The second term in Equation 3, $10^{-9} m^{-1}$, was added based on the assumption that there may be no further measurable decrease in the resuspension factor after 20 years. This assumption was deemed appropriate since the model was empirically derived to simulate experimental measurements out to 17 years, and contains no fundamental understanding of the resuspension process (Anspaugh et al. 1975). Figure 4 illustrates the time dependence of the resuspension factor.

FIGURE 4. Time Dependence of the Environmental Surface Resuspension Factor

Equation 3 was integrated over each year post-deposition and divided by the integrated time period to determine the average resuspension factor for each year considered. Ninety-nine percent of the total 50-year exposure from resuspension occurs in the first 5 years. The chronic 50-year committed dose equivalent factor for inhalation remains relatively constant over this time period. Therefore, the 50-year committed dose equivalent from 50 years of exposure to resuspended plutonium can be estimated using chronic 50-year committed dose equivalent factors, and only the first 5 years of exposure to the resuspended material needs to be included. The committed dose equivalent from inhalation of resuspended material was calculated by:

$$DC_{ir} = W_i \bar{K}(BR)(DCF)_{ir} (3.16 \times 10^7) \quad (4)$$

where

- DC_{ir} • the 50-year committed dose equivalent to organ r from one year of inhalation of radionuclide i, rem/yr of inhalation
- W_i • the concentration of radionuclide i on the ground surface for the year of consideration, $\mu\text{g}/\text{m}^2$
- $\bar{K}$ • the average resuspension factor for the year of consideration, m^{-1}
- BR • the ventilation rate of the human receptor (for a duration of greater than 24 hours), m^3/sec
- $(DCF)_{ir}$ • chronic committed dose equivalent factor, rem/ μg inhaled
- 3.16×10^7 • conversion factor, sec/yr

Radiological decay of the deposited radionuclides and the buildup of ^{241}Am from the decay of ^{241}Pu were accounted for. Chronic 50-year committed dose equivalent factors for a one-year intake were calculated using DACRIN and are listed in Appendix B, Tables B.3 and B.4.

RESULTS

EARTHQUAKES

Committed radiation dose equivalents to several organs of the human body were calculated for three earthquake events using the source terms given in Table 3.

TABLE 3. Estimated Quantity of Plutonium Released to the Atmosphere Following an Earthquake Exceeding 0.55 g^(a)

Time Period	Quantity Released (grams)		
	Upper Bound	Most Likely	Lower Bound
0-2 hr	4	4	4
2-8 hr	6×10^{-2}	1×10^{-4}	1×10^{-4}
8-24 hr	0.2	3×10^{-4}	3×10^{-4}
1-4 days	0.7	2×10^{-3}	1×10^{-3}
Total	5.0	4.0	4.0

(a) Taken from Mishima and Ayer (1981). Only the quantity released in the respirable particle size range (less than 10 μ m) was used to calculate doses.

A peak ground acceleration in excess of 0.55 g was found to be necessary for substantial damage to the NMDF. At that acceleration, the exterior walls are postulated to fail causing the roof to collapse. Significant damage was not postulated for ground acceleration less than 0.55 g (Mishima and Ayer 1981), and no radiation doses were calculated.

The isotopic composition assumed for the plutonium mixture, given in Table 4, is from the Atomics International Environmental Impact Assessment (Rockwell International 1976).

For the 0-2 hour time period, accident atmospheric dispersion values for a 5% and 50% condition, calculated by the NRC for the Santa Susana site, were used to estimate potential committed dose equivalents to the population and a maximum individual. Annual average atmospheric dispersion and deposition values

TABLE 4. Isotopic Composition of the Plutonium Mixture

<u>Isotope</u>	<u>Weight Percent</u>
^{238}Pu	1.1
^{239}Pu	61.6
^{240}Pu	20.9
^{241}Pu	12.6
^{242}Pu	3.8
^{241}Am	<u>-(a)</u>
	100

(a) ^{241}Am was not considered in the release. However, the buildup of ^{241}Am from ^{241}Pu in the environment is accounted for.

also calculated by the NRC were used for all other time periods. For the 5% condition (conservative) at times greater than 2 hours, the annual average dispersion and deposition values were multiplied by a factor of 4, as recommended by Carson.^(a) Four combinations of release and dispersion are considered: most likely release with most likely dispersion; most likely release with conservative dispersion; conservative release with most likely dispersion; and conservative release with conservative dispersion. These combinations are referred to as Case I, Case II, Case III, and Case IV respectively, and are used for all of the natural phenomena scenarios. The calculated committed dose equivalents are listed in Table 5 for the Earthquake. The estimated maximum plutonium ground depositions at the site boundary, nearest residence, and farm are listed in Table 6. All the directions and distances given in the report are referenced to the NMDF (055 Building), the primary facility at the Santa Susana site for plutonium processing.

(a) Letter transmitted from J. E. Carson of ANL/EIS to R. B. McPherson of PNL/ESD, October 24, 1978.

TABLE 5. Fifty-Year Committed Dose Equivalents from Inhalation Following an Earthquake (Class Y)

Organ of Reference	Committed Dose Equivalents for:							
	Population (person-rem) ^(a)				Nearest Residence (rem) ^(b,c)			
	Case I ^(d)	Case II	Case III	Case IV	Case I	Case II	Case III	Case IV
Total Body	1.7E+4	2.0E+5	1.9E+4	2.1E+5	2.5E-1	1.4E+1	2.7E-1	1.4E+1
Kidneys	7.3E+4	8.8E+5	8.3E+4	9.2E+5	1.0E+0	6.0E+1	1.2E+0	6.1E+1
Liver	2.3E+5	2.7E+6	2.6E+5	2.9E+6	3.4E+0	1.9E+2	3.7E+0	1.9E+2
Bone	3.8E+5	4.5E+6	4.3E+5	4.7E+6	5.6E+0	3.1E+2	6.1E+0	3.1E+2
Lungs	2.0E+5	2.4E+6	2.3E+5	2.6E+6	3.0E+0	1.7E+2	3.3E+0	1.7E+2

(a) Population within a 50-mile radius of the NMDF.

(b) Located 2000 meters SSE of the NMDF.

(c) Note that in the conservative dispersion cases (II and IV), the dose at the nearest residence is more than 50 times greater than in the most likely dispersion cases with the same release (I and III respectively). This is the result of a comparable difference between the 5% (conservative) and 50% (most likely) dispersion factors calculated by the NRC for the 0- to 2-hour time period, during which most of the dose commitment at the nearest residence is incurred.

(d) Case I - most likely release with most likely dispersion; Case II - most likely release with conservative dispersion; Case III - conservative release with most likely dispersion; Case IV - conservative release with conservative dispersion.

TABLE 6. Estimated Maximum Plutonium Deposition at Significant Locations Following an Earthquake (all particle sizes)

Location	Pu Deposition ($\mu\text{Ci}/\text{m}^2$)			
	Case I	Case II	Case III	Case IV
Site Boundary ^(a)	3.0E+1	3.7E+2	3.2E+1	3.7E+2
Residence ^(b)	5.2E-1	3.0E+1	5.8E-1	3.1E+1
Farm ^(c)	2.3E-1	2.8E+0	2.4E-1	2.8E+0

(a) Located 410 meters NW of the NMDF.

(b) Located 2000 meters SSE of the NMDF.

(c) Located 3200 meters ENE of the NMDF.

HIGH WINDS

Fujita (1977) reported that the probability of a 90-mph straight wind occurring at the NMDF site was approximately the same as for a 90-mph tornadic wind. Below 90 mph, the probability of a straight-line wind gust is higher than for a tornado of the same speed. Above 90 mph, the probability of tornadic winds was judged to exceed that of a comparable straight wind.

At 110 mph, the lowest wind speed for which a significant release is reported (Mishima and Ayer, 1981), the probability of a tornado is far greater than that of a straight wind. High straight winds are therefore not considered to be a likely cause of major structural damage at the NMDF site and are not discussed further.

TORNADOES

Plutonium releases following four tornadoes with maximum total wind speeds of 110 mph, 130 mph, 150 mph, and 170 mph were estimated by Mishima and Ayer (1981). Releases for four time periods are presented in Table 7.

TABLE 7. Estimated Quantity of Plutonium Released to the Atmosphere Following a Tornado

Time Period	Airborne Release of Pu (g) for the Following Maximum Total Wind Speeds:							
	110 mph		130 mph		150 mph		170 mph	
	Conservative	Most Likely	Conservative	Most Likely	Conservative	Most Likely	Conservative	Most Likely
0-2 hr	3E-3	-	1.3E-2	5E-3	3.3E-2	2E-2	4	4
2-8 hr	9E-3	-	9E-3	3E-5	9E-3	1E-3	6E-2	1E-2
8-24 hr	-(a)	-	8E-5	8E-5	1E-3	1E-3	2E-2	3E-2
1-4 days	-	-	4E-4	4E-4	2E-2	2E-2	7E-1	2E-1
Total	1.2E-2	-	2.2E-2	5.5E-3	6.3E-2	4.2E-2	4.8	4.2

(a) - denotes $<10^{-7}$ g released.

Atmospheric dispersion and deposition values most likely to occur during a tornado were calculated by Pepper (1980). These values were assumed to apply during the first two hours after the event. During this time period, the tornadoes were assumed to move in a northeasterly direction. Annual average atmospheric dispersion and deposition values were used for all other time

periods. As recommended by Pepper^(a) and Carson, the tornado dispersion values were multiplied by a factor of 10 and the annual average atmospheric dispersion and deposition values were multiplied by a factor of 4 to represent the conservative case. Committed radiation dose equivalents are given in Tables 8 through 11 for Class Y plutonium. The estimated maximum ground contamination levels from plutonium deposition at the significant locations are listed in Tables 12 through 15.

TABLE 8. Fifty-Year Committed Dose Equivalents from Inhalation Following a 110-mph Tornado (Class Y)

Organ of Reference	Committed Dose Equivalents for:							
	Population (person-rem)				Nearest Residence (rem) ^(a)			
	Case I ^(b)	Case II	Case III	Case IV	Case I	Case II	Case III	Case IV
Total Body	-(c)	-	3.4E+2	3.2E+3	-	-	5.8E-4	5.8E-3
Kidneys	-	-	1.5E+3	1.4E+4	-	-	2.5E-3	2.5E-2
Liver	-	-	4.6E+3	4.4E+4	-	-	7.9E-3	7.9E-2
Bone	-	-	7.6E+3	7.2E+4	-	-	1.3E-2	1.3E-1
Lungs	-	-	4.1E+3	3.9E+4	-	-	7.1E-3	7.1E-2

(a) Located 16,000 to 32,000 meters from the NMDF in the direction the tornado travels.

(b) Case I - most likely release with most likely dispersion; Case II - most likely release with conservative dispersion; Case III - conservative release with most likely dispersion; Case IV - conservative release with conservative dispersion.

(c) No doses calculated for Cases I and II, as most likely release is insignificant ($<10^{-7}$ g Pu).

(a) Letter transmitted from D. W. Pepper of SRL/ETD to R. B. McPherson of PNL/ESD, February 21, 1979.

TABLE 9. Fifty-Year Committed Dose Equivalents from Inhalation Following a 130-mph Tornado (Class Y)

Organ of Reference	Committed Dose Equivalents for:							
	Population (person-rem)				Nearest Residence (rem)(a)			
	Case I	Case II	Case III	Case IV	Case I	Case II	Case III	Case IV
Total Body	6.1E+2	6.1E+3	1.6E+3	1.6E+4	9.6E-4	9.6E-3	2.5E-3	2.5E-2
Kidneys	2.7E+3	2.7E+4	7.1E+3	7.0E+4	4.2E-3	4.2E-2	1.1E-2	1.1E-1
Liver	8.4E+3	8.4E+4	2.2E+4	2.2E+5	1.3E-2	1.3E-1	3.4E-2	3.4E-1
Bone	1.4E+4	1.4E+5	3.6E+4	3.6E+5	2.2E-2	2.2E-1	5.6E-2	5.6E-1
Lungs	7.5E+3	7.5E+4	2.0E+4	2.0E+5	1.2E-2	1.2E-1	3.1E-2	3.1E-1

(a) Located 16,000 to 32,000 meters from the NMDF in the direction the tornado travels.

TABLE 10. Fifty-Year Committed Dose Equivalents from Inhalation Following a 150-mph Tornado (Class Y)

Organ of Reference	Committed Dose Equivalents for:							
	Population (person-rem)				Nearest Residence (rem)(a)			
	Case I	Case II	Case III	Case IV	Case I	Case II	Case III	Case IV
Total Body	2.7E+4	2.7E+5	4.0E+4	4.0E+5	3.6E-2	3.6E-1	5.3E-2	5.3E-1
Kidneys	1.2E+5	1.2E+6	1.8E+5	1.8E+6	1.6E-1	1.6E+0	2.3E-1	2.3E+0
Liver	3.7E+5	3.7E+6	5.5E+5	5.5E+6	4.9E-1	4.9E+0	7.3E-1	7.3E+0
Bone	6.0E+5	6.0E+6	9.0E+5	9.0E+6	8.0E-1	8.0E+0	1.2E+0	1.2E+1
Lungs	3.3E+5	3.3E+6	4.9E+5	4.9E+6	4.4E-1	4.4E+0	6.5E-1	6.5E+0

(a) Located 16,000 to 32,000 meters from the NMDF in the direction the tornado travels.

TABLE 11. Fifty-Year Committed Dose Equivalents from Inhalation Following a 170-mph Tornado (Class Y)

Organ of Reference	Committed Dose Equivalents for:							
	Population (person-rem)				Nearest Residence (rem) ^(a)			
	Case I	Case II	Case III	Case IV	Case I	Case II	Case III	Case IV
Total Body	1.2E+4	1.1E+5	1.3E+4	1.2E+5	8.3E-3	8.3E-2	2.2E-2	8.9E-2
Kidneys	5.1E+4	5.0E+5	5.8E+4	5.3E+5	3.6E-2	3.6E-1	9.6E-2	3.9E-1
Liver	1.6E+5	1.6E+6	1.8E+5	1.6E+6	1.1E-1	1.1E+0	3.0E-1	1.2E+0
Bone	2.6E+5	2.6E+6	3.0E+5	2.7E+6	1.9E-1	1.9E+0	5.0E-1	2.0E+0
Lungs	1.4E+5	1.4E+6	1.6E+5	1.5E+6	1.0E-1	1.0E+0	2.7E-1	1.1E+0

(a) Located 32,000 to 48,000 meters from the NMDF in the direction the tornado travels for Cases I and II; located 2000 meters SSE of the NMDF for Cases III and IV.

TABLE 12. Estimated Maximum Plutonium Deposition at Significant Locations Following a 110-mph Tornado (all particle sizes)

Location	Pu Deposition ($\mu\text{Ci}/\text{m}^2$)			
	Case I	Case II	Case III	Case IV
Site Boundary ^(a)	-(b)	-	2.7E-2	6.3E-2
Residence ^(c)	-	-	3.4E-3	3.4E-2
Farm ^(c)	-	-	3.4E-3	3.4E-2

(a) Located 410 meters NW of the NMDF.

(b) No deposition; Cases I and II had no significant release.

(c) Located 16,000 to 32,000 meters from the NMDF in the direction the tornado travels.

TABLE 13. Estimated Maximum Plutonium Deposition at Significant Locations Following a 130-mph Tornado (all particle sizes)

Location	Pu Deposition ($\mu\text{Ci}/\text{m}^2$)			
	Case I	Case II	Case III	Case IV
Site Boundary ^(a)	1.6E-3	3.6E-3	2.9E-2	6.6E-2
Residence ^(b)	5.7E-3	5.7E-2	1.5E-2	1.5E-1
Farm ^(b)	5.7E-3	5.7E-2	1.5E-2	1.5E-1

(a) Located 410 meters NW of the NMDF.

(b) Located 16,000 to 32,000 meters from the NMDF in the direction the tornado travels.

TABLE 14. Estimated Maximum Plutonium Deposition at Significant Locations Following a 150-mph Tornado (all particle sizes)

Location	Pu Deposition ($\mu\text{Ci}/\text{m}^2$)			
	Case I	Case II	Case III	Case IV
Site Boundary ^(a)	6.7E-2	1.5E-1	9.1E-1	2.1E-1
Residence ^(b)	2.1E-1	2.1E+0	3.1E-1	3.1E+0
Farm ^(b)	2.1E-1	2.1E+0	3.1E-1	3.1E+0

(a) Located 410 meters NW of the NMDF.

(b) Located 16,000 to 32,000 meters from the NMDF in the direction the tornado travels.

TABLE 15. Estimated Maximum Plutonium Deposition at Significant Locations Following a 170-mph Tornado (all particle sizes)

Location	Pu Deposition ($\mu\text{Ci}/\text{m}^2$)			
	Case I	Case II	Case III	Case IV
Site Boundary ^(a)	7.3E-1	2.9E+0	2.9E+0	1.2E+1
Residence ^(b)	4.5E-2	4.5E-1	4.5E-2	4.5E-1
Farm ^(b)	4.5E-2	4.5E-1	4.5E-2	4.5E-1

(a) Located 410 meters NW of the NMDF.

(b) Located 32,000 to 48,000 meters from the NMDF in the direction the tornado travels.

DISCUSSION

The calculated committed dose equivalents are based on the ICRP Publication 2 Metabolic Model, the ICRP Task Group Lung Model and standard man parameter values. To the best of our knowledge, there are no reported assessments of the accuracy of dose calculations using these models and parameter values. Dose results are usually presented with no indication of the error associated with their use. Present insights into the degree of uncertainty involved are very limited and qualitative (Hoffman 1978). Dose results presented in this paper are probably within a factor of ± 10 . However, studies should be conducted to determine the uncertainties associated with these kinds of calculations.

The estimated average annual whole-body radiation dose from natural background radiation in California is reported to be 115 mrem/yr (Klement 1972). Therefore, an individual receives a total-body dose of about 5.8 rem from exposure to natural background radiation during a 50-year period. The collective dose equivalent from 50 years of exposure to natural background radiation to the total body of the population within a 50-mile radius of the NMDF is 4.1×10^7 person-rem. The average annual dose to the total body of an individual from medical x-ray examination is about 20 mrem (United Nations 1977). This average dose corresponds to a 50-year collective dose equivalent of 7.2×10^6 person-rem. The dose contribution from fallout is negligible when compared to natural background radiation and medical x-ray exposure. If a radiation worker was involved in an occupational accident and received a maximum permissible bone burden of ^{239}Pu , the 50-year committed dose equivalent to the bone would be greater than 1000 rem. As can be seen, in all cases, the calculated 50-year committed dose equivalents to the population for the severe natural phenomena scenarios considered in this report are lower than the collective dose equivalent from 50 years of exposure to natural background radiation and medical x-rays.

Existing guidelines on acceptable levels of soil contamination from Pu can be found to range from $0.01 \mu\text{Ci}/\text{m}^2$ to $270 \mu\text{Ci}/\text{m}^2$ (Selby et al. 1975; EPA 1977; Martin and Bloom 1975; Healy 1977; U.S. Code 1976; Healy 1974; Guthrie and Nichols 1964; Hazle and Crist 1975; Kathren 1968; Dunster 1962). The EPA has proposed a guideline of $0.2 \mu\text{Ci}/\text{m}^2$ for plutonium in the general environment (EPA 1977). This guideline is based on annual doses of one mrad to the lung from inhalation and three mrad to the bone from ingestion. If the broad range of current guidelines are normalized to these lung and bone doses and the same resuspension factor is used, the guidelines are all in reasonable agreement with $0.2 \mu\text{Ci}/\text{m}^2$.

The predicted maximum residual plutonium contamination levels on the ground following the earthquake, the 150-mph and 170-mph tornadoes are above the EPA proposed guideline at some or all of the significant locations. The estimated (most likely) contamination levels at these locations range from about 0.2 to $30 \mu\text{Ci}/\text{m}^2$. The predicted ground contamination levels for the 110-mph and the 130-mph tornadoes are below the EPA proposed guideline at all significant locations.

APPENDIX A

EVALUATION OF ENVIRONMENTAL PATHWAYS
BY WHICH PLUTONIUM MAY REACH PEOPLE
FROM AN ACCIDENTAL AIRBORNE RELEASE

APPENDIX A

EVALUATION OF ENVIRONMENTAL PATHWAYS BY WHICH PLUTONIUM MAY REACH PEOPLE FROM AN ACCIDENTAL AIRBORNE RELEASE

Twelve environmental exposure modes for an accidental airborne release are considered for evaluation. Three are from exposure to the radioactive cloud, and four result from radioactive material deposited on the ground. The remaining five are via the waterborne pathway, assuming radioactive material was deposited onto a nearby surface body of water. For the Babcock & Wilcox study, it is assumed that irrigation does not occur.

The following exposure modes are included in the study:

1. inhalation during the initial cloud passage,
2. inhalation of resuspended radioactive material,
3. direct exposure from cloud submersion,
4. direct exposure to radioactive material deposited on the ground,
5. ingestion of leafy vegetables, beef, milk, water and fish,
6. direct exposure from swimming and boating in contaminated water, and
7. external exposure to radioactive material concentrated in the shoreline sediment.

One isotope of plutonium is considered, ^{239}Pu , but the conclusions from this evaluation apply to most of the other isotopes, as well as a typical mixture of plutonium isotopes. The dose models used to evaluate the exposure modes are taken from Regulatory Guide 1.109, Revision 1 (1977) and are modified accordingly for an accidental release. In the evaluation, doses are calculated to a hypothetical maximum adult individual for 50 years of exposure.

AIRBORNE PATHWAYS

Airborne Release Assumptions and Dispersion

A normalized plutonium release of $1 \mu\text{g } ^{239}\text{Pu}$ is assumed, and an arbitrary average accident exposure coefficient (E/Q) of $1 \times 10^{-3} \mu\text{g} \cdot \text{sec}/\text{m}^3$ per μg at the location of the maximum exposed individual is chosen. It is assumed that the particle size of the plutonium released is in the respirable range, less than $10 \mu\text{m}$ in diameter.

Inhalation

The committed dose equivalent from radioactive material inhaled during passage of the initial cloud is calculated by:

$$DC_i = Q(E/Q)(DCF)_i \quad (A-1)$$

where

- DC_i • committed dose equivalent to organ i (rem)
- Q • total quantity of radioactive material released during the accident (μg)
- (E/Q) • accident atmospheric exposure coefficient ($\mu\text{g} \cdot \text{sec}/\text{m}^3$ per μg)
- $(DCF)_i$ • committed dose equivalent factor for organ i ($\text{rem per } \mu\text{g} \cdot \text{sec}/\text{m}^3$)

The committed dose equivalent factors for acute inhalation of ^{239}Pu were calculated using the computer code DACRIN (Houston et al. 1975). This code incorporates the ICRP Task Group Lung Model to calculate the dose to the lungs and other organs of interest. The organ masses were modified to reflect the changes reported in ICRP-23 (1975). The 50-year committed dose equivalent factors per unit mass exposure for ^{239}Pu particles, with an activity median aerodynamic diameter (AMAD) of one micrometer, are listed in Table A.1 for the total body, lungs, bone, and the lower large intestine (GI-LLI), along with the calculated committed dose equivalents. The translocation classes which minimize the contribution from the inhalation pathway are used. Cloud depletion is not considered. The location of the maximum exposed individual would

TABLE A.1. Fifty-Year Committed Dose Equivalents from
Inhalation of 1 μm AMAD ^{239}Pu Particles

Organ	Translocation Class	Committed Dose Equivalent Factors (rem per $\mu\text{g} \cdot \text{sec}/\text{m}^3$)	Committed Dose Equivalent ^(a) (rem)
Total Body	Y	5.6E-04 ^(b)	5.6E-07
Lungs	W	9.9E-04	9.9E-07
Bone	Y	1.2E-02	1.2E-05
GI-LLI	W	7.3E-07	7.3E-10

(a) At location where $E/Q = 1 \times 10^{-3} \mu\text{g} \cdot \text{sec}/\text{m}^3$ per μg .

(b) 5.6E-04 is identical to 5.6×10^{-4} .

probably be a few hundred meters from the point of release, and the inclusion of cloud depletion would only lower the doses by a few percent (USNRC Guide 1.111 1977, Gudiksen 1976).

Resuspension is an important aspect to be considered when calculating the dose from inhalation of plutonium. The airborne concentration from resuspended material can be predicted using a resuspension factor, k . The resuspension factor is defined as the resuspended air concentration divided by the surface deposition. Reasonable values for k between 10^{-4} m^{-1} and 10^{-13} m^{-1} have been measured and reported (Anspaugh 1975, Cohen 1977, Hanson 1975, FES 1974, Selby 1975, Bennett 1975, EPA 1977, Martin and Bloom 1975, Sehmel 1977, Healy 1977, and Anspaugh 1976). Until a more general model is available, which considers all the important variables affecting the resuspension process, Anspaugh (1975) recommends using a simple time-dependent model to predict the average airborne concentration of a resuspended contaminant:

$$K(t) = 10^{-4} \exp(-0.15t^{1/2}) + 10^{-9} \quad (\text{A-2})$$

where

K • resuspension factor (m^{-1})

t • time since the material was deposited on the ground (days)

The second term was added based upon the assumption that there may be no further measurable decrease in the resuspension process after 17 years, which is the longest period post deposition that measurements have been reported. This time-dependent model accounts for the observed decrease in air concentrations which has been noted to occur in the absence of a significant net loss of the deposited contaminant.

The resuspension factor was integrated over 50 years to include the total potential intake from resuspended plutonium. It was determined that 78% of the total 50-year exposure from resuspension occurs in the first year and 99% occurs in the first five years. To simplify the calculation, it is assumed that the total 50 years of exposure to resuspended plutonium is received during the first year, and the particle size is in the respirable range. The committed dose environment factor remains relatively constant over the first five years. Thus, bringing this term out of the integral does not affect the results. When Equation A-2 is integrated over 50 years, the total exposure to resuspended material is calculated to be 8.9×10^{-3} days/meter.

The committed dose equivalent from inhalation of resuspended material is calculated by:

$$DC_i = W(DCF)_i (8.64 \times 10^4) \int_0^{T_d} K(t) dt \quad (A-3)$$

where

W • surface concentration of radioactive material initially deposited from the cloud onto the ground ($\mu g/m^2$)

T_d • dose commitment time (days);

$T_d = 1.83 \times 10^4$ days (50 years in this study)

$(DCF)_i$ • 50-year committed dose equivalent factor for organ i
(rem per $\mu\text{g} \cdot \text{sec}/\text{m}^3$)

8.64×10^4 • constant which converts days to seconds (sec/day)

The terms K and DC_i have already been defined.

The initial ground deposition concentration for an accidental release of plutonium is calculated using the following equation:

$$W = E V_d \quad (\text{A-4})$$

where

E • exposure ($\mu\text{g} \cdot \text{sec}/\text{m}^3$), the product of Q and E/Q as defined earlier
 V_d • deposition velocity (m/sec)

A deposition velocity of 1×10^{-3} m/sec was chosen as is used in the computer code FOOD (Baker 1977).

Chronic committed dose equivalent factors were calculated for one year of inhalation of ^{239}Pu using DACRIN (Houston et al. 1975). The organ masses were modified to reflect the changes reported in ICRP-23 (1975) as before. The 50-year committed dose equivalent factors per unit mass inhaled during the first year for ^{239}Pu particles with an AMAD of one micrometer are listed in Table A.2 for the total body, lungs, bone, and GI-LLI, along with the calculated committed dose equivalents. The translocation classes which minimize the contribution from inhalation are used to be consistent with Table A.1.

TABLE A.2. Fifty-Year Committed Dose Equivalents from 50 Years' Inhalation of 1 μm AMAD Resuspended ^{239}Pu Particles

<u>Organ</u>	<u>Translocation Class</u>	<u>Committed Dose Equivalent Factors (rem per $\mu\text{g} \cdot \text{sec}/\text{m}^3$)</u>	<u>Committed Dose Equivalent^(a) (rem)</u>
Total Body	Y	4.6E-04	3.5E-07
Lungs	W	8.1E-04	6.2E-07
Bone	Y	9.7E-03	7.5E-06
GI-LLI	W	5.9E-07	4.5E-10

(a)At the same location where initial inhalation was calculated.

Cloud Submersion

A semi-infinite cloud model is used for calculating the external doses from cloud submersion during cloud passage. The doses are calculated with the following equation:

$$D_i = E_T(DF)_i S_F \quad (\text{A-5})$$

where

- E_T • sum of the initial exposure from initial cloud passage and resuspended radioactive material ($\mu\text{g} \cdot \text{sec}/\text{m}^3$)
- $(DF)_i$ • dose factor for cloud submersion for organ i
(rem per $\mu\text{g} \cdot \text{sec}/\text{m}^3$)
- S_F • attenuation factor which accounts for shielding provided by residential structures (dimensionless)

The total exposure is calculated by:

$$E_T = E + W(8.64 \times 10^4) \int_0^{T_d} K(t) dt \quad (\text{A-6})$$

where all terms have already been defined.

The total exposure to the airborne particulate, E_T , is calculated to be $1.8 \times 10^{-3} \mu\text{g}\cdot\text{sec}/\text{m}^3$. A value of 0.70 is used for the attenuation factor, S_F (USNRC Guide 1.109 1977).

Doses for submersion in a semi-infinite cloud of ^{239}Pu were calculated using dose factors taken from Soldat (1974) and converted to dose per unit mass. The calculated doses and dose factors are given in Table A.3.

TABLE A.3. Air Submersion Doses from Exposure to ^{239}Pu

<u>Organ</u>	<u>Dose Factor (rem per $\mu\text{g} \cdot \text{sec}/\text{m}^3$)</u>	<u>Dose (rem)</u>
Total Body	9.6E-13	1.2E-15
Skin	1.3E-11	1.6E-14

Ground Exposure

Dose from external exposure to radioactive material deposited on the ground is calculated by:

$$D_i = W(DF)_i S_F T \quad (\text{A-7})$$

where

- $(DF)_i$ • dose rate factor for organ i (rem/hr per $\mu\text{g}/\text{m}^2$)
- S_F • attenuation factor defined in Equation A-5 (dimensionless)
- T • time of exposure (hours)

W has been defined previously ($\mu\text{g}/\text{m}^2$).

The assumption is made that the ground concentration of plutonium is constant (radiological decay is insignificant and was ignored), and the exposure time, T , is 50 years or 4.38×10^5 hours. A value of 0.70 is used for S_F (USNRC Guide 1.109 1977).

The dose rate factors were again taken from Soldat et al. (1974) and converted to dose per unit mass. The calculated dose factors and doses are listed in Table A.4.

TABLE A.4. Fifty Years of External Exposure to ^{239}Pu Deposited on the Ground

<u>Organ</u>	<u>Dose Rate Factor</u> (rem/hour per $\mu\text{g}/\text{m}^2$)	<u>Dose (rem)</u>
Total Body	4.9E-11	1.5E-11
Skin	4.8E-10	1.5E-10

Crop Ingestion

The internal committed dose equivalent received from ingestion of contaminated vegetation is calculated by Equation A-8.

$$\text{DC}_i = C_p U_p (\text{DCF})_i \quad (\text{A-8})$$

where

- U_p • consumption rate for vegetation (kg/yr)
- $(\text{DCF})_i$ • 50-year committed dose equivalent factor for organ i from chronic ingestion of ^{239}Pu (rem per μg ingested per year)
- C_p • radionuclide concentration in the edible portion of the vegetation ($\mu\text{g}/\text{kg}$):

$$C_p = \frac{W V_d r T_r (8.64 \times 10^4) \exp(-\lambda_e t_2)}{Y} \left[10^{-4} \int_{T_1}^{T_2} \exp(\lambda_e s - 0.15\sqrt{s}) ds \right. \\ \left. + \frac{10^{-9}}{\lambda_e} \exp(\lambda_e t_2) - \exp(\lambda_e t_1) \right] + \frac{C_o r T_r}{Y} \exp[-\lambda_e (t_2 - t_1)] \quad (\text{A-9})$$

where previously undefined symbols are defined by:

- r • fraction of deposited radionuclide retained by the vegetation (dimensionless)
- T_r • factor for translocation of externally deposited radionuclides to the edible parts of the vegetation (dimensionless)

- Y • vegetation yield (kg/m^2)
- 8.64×10^4 • constant which converts days to seconds (sec/days)
- λ_e • effective decay constant for removal of radionuclides on leaf or produce surfaces by weathering and radiological decay (days^{-1})
- t_1 • time from the accident to the appearance of the vegetation (days)
- t_2 • time from the accident to harvest of the vegetation (days)
- C_0 • the concentration of radioactive material initially deposited on the vegetation ($\mu\text{g/m}^2$); equals zero if the vegetation was not present at the time of the accident.

Equation A-9 accounts for the radioactive material initially deposited on the vegetation following the accident as well as the contribution from the deposit of resuspended environmental residual radioactive material.

The contribution from root uptake of plutonium is negligible compared to the plutonium deposited directly onto the vegetation (less than 1%), and is ignored. It is assumed that the accident occurred a few days before harvest during the beginning of the growing season. A five-month growing season and a 90-day growing period for vegetation are used.

The integral was evaluated numerically. It was determined that almost all of the plutonium deposited on the vegetation occurs during the first year. Therefore, the assumption was made that the total intake occurs during the first year. Fifty-year committed dose equivalent factors were used to calculate the resulting committed dose equivalents.

Values for r , Y and λ_e are taken from Regulatory Guide 1.109 as 0.2 for particulates, 2.0 kg/m^2 for leafy vegetables and produce, and 0.0495 days^{-1} , respectively. Values of l , for leafy vegetables, and 0.1, for produce, are used for T_r (Baker et al. 1966). Consumption rates for the maximum individual are taken from Regulatory Guide 1.109 and adjusted for a five-month growing season. Consumption rates of 27 kg/yr , for leafy vegetables, and 217 kg/yr , for fruits, vegetables, and grain, were calculated. The average plutonium concentrations in the edible portion of leafy vegetables and produce are presented in Table A.5 for a five-year period.

TABLE A.5. Average ^{239}Pu Concentration Estimated in Leafy Vegetables and Produce for a Five-Year Period

<u>Year (a)</u>	<u>Plutonium Concentration ($\mu\text{g/kg}$)</u>	
	<u>Leafy Vegetables</u>	<u>Produce</u>
1	4.7E-08	4.7E-09
2	9.4E-10	9.4E-11
3	2.9E-10	2.9E-11
4	1.2E-10	1.2E-11
5	5.5E-11	5.5E-12

(a) Accident occurred during first year, a few days before the first harvest.

Fifty-year committed dose equivalent factors for a one-year chronic ingestion of ^{239}Pu were taken from NUREG-0172 (Hoenes and Soldat 1977), modified using updated organ masses from ICRP-23 (1975) and biological half-lives from ICRP-19 (1972), and converted to a unit mass intake. Committed dose equivalent factors and committed dose equivalents were calculated and are included in Table A.6.

TABLE A.6. Fifty-Year Committed Dose Equivalents from 50 Years' Ingestion of Leafy Vegetables and Produce Contaminated with ^{239}Pu

<u>Organ</u>	<u>Committed Dose Equivalent Factor (rem/50-year per μg ingested per year)</u>	<u>Committed Dose Equivalent (rem)</u>
Total Body	1.2E-03	2.8E-09
Bone	4.5E-02	1.1E-07
GI-LLI	4.6E-03	1.1E-08

Ingestion of Milk and Beef

The internal committed dose equivalents from the ingestion of animal products (milk and beef) are calculated by:

$$DC_i = \left[\sum (CQ) \right] S U(DCF)_i \quad (A-10)$$

where the summation of the product of the two terms, C and Q , represents the total radionuclide intake by the animal from consumption of contaminated forage, feed, and water, and

- C • radionuclide concentration in the animal's food, C_f (μg per kg forage or feed), or drinking water, C_w (μg per liter of water). Equation A-9 is used to determine values for C_f .
- Q • animals' consumption rate, Q_f (kg feed or forage/day), or Q_w (ℓ water/day)
- S • animal product transfer coefficient that relates the daily intake rate of an animal to the radionuclide concentration in milk, S_d (days/ ℓ) and beef, S_b (days/kg)
- U • human consumption rate for milk, U_d (ℓ /year), or beef, U_b (kg/year)
- $(DCF)_i$ • chronic ingestion committed dose equivalent factor for organ i given in Table A.6 (rem/50-year per μg ingested per year)

The dose contribution from the ingestion of milk or beef contaminated by the animal's drinking water is addressed in the next section.

An eight-month grazing season is assumed for both beef cattle and milk cows. For the remaining four months, the animals are fed stored feed and grain, which were grown during the previous five-month growing season. The radionuclide concentration in the grain is calculated using Equation A-9. The same parameters used to calculate the concentration in produce are used for grain. It is assumed that the accident occurred a few days before the harvest, during the beginning of the growing season. A 90-day growing period is used.

To estimate the average radionuclide concentration deposited on the pasture from the initial accident and resuspended radioactive material, Equation A-9 was integrated with respect to t , evaluated over the eight-month grazing season using a 30-day buildup period (USNRC Guide 1.109 1977), and divided by the integration time period of 30 days. All other parameters remained unchanged, except that values of 1.0 (Baker et al. 1966) and 0.70 kg/m^2 (USNRC Guide 1.109 1977) were used for T_p and Y , respectively.

It was determined that almost all of the plutonium deposited on the grain and pasture occurs during the first year. Regulatory Guide 1.109 uses a value of 50 kg/day for feed or forage consumption for both beef cattle and milk cows. Values for S_d of 2.5×10^{-8} days/l for milk and for S_b of 5.0×10^{-3} days/kg for beef are reported by Baker (1966). Values for the human consumption rates are taken from Regulatory Guide 1.109 as 310 l of milk/year and 110 kg beef/year.

The average plutonium concentrations in the edible portion of the grain and in the fresh forage are presented in Table A.7 for a five-year period.

TABLE A.7. Average ^{239}Pu Concentration Estimated in Grain and Forage for a Five-Year Period

Year ^(a)	Plutonium Concentration ($\mu\text{g/kg}$)	
	Grain	Forage
1	4.7E-09	2.7E-08
2	9.5E-11	1.4E-09
3	2.9E-11	4.8E-10
4	1.2E-11	2.0E-10
5	5.5E-12	9.9E-11

(a) Accident occurred at beginning of first year, 90 days into the growing and grazing season.

Using the 50-year ingestion committed dose equivalent factors given in Table A.6, committed dose equivalents were calculated for the consumption of contaminated animal products and are presented in Table A.8.

TABLE A.8. Fifty-Year Committed Dose Equivalents from 50 Years' Ingestion of Milk and Beef Contaminated with ^{239}Pu

Organ	Committed Dose Equivalent (rem)	
	Milk	Beef
Total Body	9.8E-15	6.9E-10
Bone	3.7E-13	2.6E-08
GI-LLI	3.7E-14	2.7E-09

WATERBORNE PATHWAYS

To account for the contribution from radionuclides deposited onto a nearby surface body of water, a small lake one meter deep is assumed. Deposition over the lake is assumed to occur at the same rate as over land (USNRC Guide 1:111 1977). The plutonium is assumed to be soluble; however, it is most likely insoluble in water at normal pH levels. It is assumed that the radioactive particulates are uniformly mixed in the lake upon contact. A four-year removal half-life is used (Wahlgren and Marshall 1975).

A differential equation was set up with a dynamic source term (this includes the initial deposition onto the water and the contribution from deposited resuspended material) to reflect the changing water concentration. When this equation is solved for the water concentration, an expression is obtained similar to Equation A-9. Due to the slower removal process (four years versus 14 days for vegetation), and since the majority of the radioactive material deposited on the lake takes place during the first year, the water concentration curve approaches a simple exponential expression. The initial water concentration can be determined by extrapolating the curve back to time zero, which yields Equation A-11.

$$C_w = 1.86 \times 10^{-9} \exp(-\lambda t) \quad (A-11)$$

where

- C_w • radionuclide concentration in water ($\mu\text{g}/\ell$)
- λ • removal constant for plutonium in the lake (days^{-1});
 $\lambda = 0.693/1461 \text{ days}^{-1}$
- t • time since deposition onto the lake surface (days)

1.86×10^{-9} • extrapolated water concentration at $t = 0$ ($\mu\text{g}/\ell$)

Equation A-11 is much simpler than the complex solution obtained by solving the differential equation, and it only overestimates the water

concentration by a few percent over the 50-year period considered. Equation 11 was integrated to estimate the average radionuclide concentration in the water for each year.

Ingestion of Animal Products

If the animals' drinking water supply was to come from the lake, a small fraction of the plutonium would eventually be transferred to the animals' milk or meat and be subsequently ingested by the maximum exposed individual. Since the potential radionuclide intake from this pathway occurs for approximately 30 years, ingestion committed dose equivalent factors must be calculated to account for the decreasing commitment time as the exposure time approaches the end of the 50-year exposure period. Committed dose equivalent factors for chronic ingestion of radionuclides are calculated with Equation A-12.

$$(DCF)_i = (2.92 \times 10^{-7}) SA (F_w \epsilon/m)(\tau^2) [\lambda_e t_1 + \exp(-\lambda_e t_2) - \exp(-\lambda \Delta t)] \quad (A-12)$$

where

- $(DCF)_i$ • committed dose equivalent factor for organ i (rem per μg ingested per year)
- SA • specific activity of the radionuclide ($\text{pCi}/\mu\text{g}$)
- F_w • fraction of the ingested radionuclide reaching organ i (dimensionless)
- ϵ • effective energy of the radionuclide in organ i [$(\text{MeV/dis})(\text{rem/rad})$]
- m • mass of organ i (g)
- τ • effective half-life of the radionuclide in organ i (days)
- λ_e • effective decay constant in organ i (days^{-1}); $\lambda_e = \ln 2/\tau$
- t_1 • duration of intake (days)

t_2 • time over which the dose commitment is calculated, including the duration of intake (days)

Δt • $(t_2 - t_1) \geq 0$

One- through 50-year ingestion committed dose equivalent factors were calculated for ^{239}Pu using parameter values found in ICRP-2 (1959), ICRP-23 (1975), and ICRP-19 (1972). The committed dose equivalent factor to the GI-LLI from ingestion does not change.

Committed dose equivalents were calculated using Equation A-10 for each year during the 50-year exposure period, using average yearly water concentrations and the appropriate dose commitment factor for that year. Baker (1966) and Regulatory Guide 1.109 use animal water consumption values of 60 l/day for milk cows, and 50 l/day for beef cattle. The total committed dose equivalents for the 50-year period to the maximum individual from ingestion of animal products for this exposure pathway are presented in Table A.9. The sum of the contributions from animal product ingestion for the airborne pathway and the waterborne pathway is also shown in Table A.9.

TABLE A.9. Fifty-Year Committed Dose Equivalents from 50 Years' Ingestion of Animal Products Contaminated with ^{239}Pu (waterborne pathway and waterborne plus air)

Organ	Committed Dose Equivalents (rem) from Waterborne Pathway		Total Committed Dose Equivalents (rem) from Airborne and Waterborne Pathways	
	Milk	Beef	Milk	Beef
Total Body	5.2E-15	3.1E-10	1.5E-14	1.0E-09
Bone	2.0E-13	1.2E-08	5.7E-13	3.8E-08
GI-LLI	2.3E-14	1.4E-09	6.0E-14	4.1E-09

Drinking Water Ingestion

The committed dose equivalent from consumption of contaminated drinking water is calculated by:

$$DC_i = \bar{C}_w U_w (DCF)_i \quad (A-13)$$

where

- $\bar{C}_w$ • average radionuclide concentration in water during the year of interest ($\mu\text{g}/\ell$); calculated earlier
- U_w • consumption rate (ℓ/year)
- $(DCF)_i$ • chronic ingestion committed dose equivalent factor for organ i (rem per μg ingested)

A water consumption rate of 730 ℓ/year is used (USNRC Guide 1.109 1977). Committed dose equivalent factors were calculated using Equation A-12. The dose calculations were taken out to 28 years, at which time the contribution to the total dose commitment from drinking water is insignificant. The resulting committed dose equivalents are listed in Table A.10.

TABLE A.10. Fifty-Year Committed Dose Equivalents from 50 Years' Consumption of Water Contaminated with ^{239}Pu

<u>Organ</u>	<u>Committed Dose Equivalent (rem)</u>
Total Body	8.3E-09
Bone	3.2E-07
GI-LLI	3.6E-08

Fish Ingestion

The equation used to estimate the dose from consumption of fish, assuming immediate transfer and equilibrium after deposition of the radionuclide onto the lake, is:

$$DC_i = \bar{C}_w B U_f (DCF)_i \quad (\text{A-14})$$

where

- $\bar{C}_w$ • average radionuclide concentration in the lake during the year of interest ($\mu\text{g}/\ell$); calculated earlier

- B • equilibrium bioaccumulation factor expressed as the ratio of the concentration in fish to the radionuclide concentration in water (l/kg)
- U_f • fish consumption rate (kg/year)
- $(DCF)_i$ • same committed dose equivalent factor calculated for the ingestion of animal products and drinking water, for organ i (rem per μg ingested)

A fish consumption rate of 21 kg/year is used (USNRC Guide 1.109 1977) and the value for the bioaccumulation factor, B, is selected to be 3.5 l/kg (Soldat et al. 1974). The duration of fish consumption is assumed to be 50 years. However, the contribution from the last 22 years of fish consumption is negligible.

The committed dose equivalents from fish consumption were calculated and are given in Table A.11.

TABLE A.11. Fifty-Year Committed Dose Equivalents from 50 Years' Consumption of Fish Contaminated with ^{239}Pu

<u>Organ</u>	<u>Committed Dose Equivalent (rem)</u>
Total Body	3.8E-10
Bone	3.2E-08
GI-LLI	3.6E-09

Swimming and Boating

The following equation is used to calculate the dose from swimming:

$$D_i = \bar{C}_w (DF)_i U T \quad (\text{A-15})$$

where

- $\bar{C}_w$ • average radionuclide concentration in the lake during the period of exposure ($\mu\text{g/l}$); calculated earlier

- (DF)_i • water immersion dose rate factor for organ i (rem/hr per $\mu\text{g}/\ell$)
- U • exposure rate (hours/year)
- T • period of exposure (years)

The dose rate factors were taken from Soldat et al. (1975) and converted to dose per unit mass. Equation A-11 was integrated over a 50-year period and divided by 50 years to obtain an average radionuclide concentration in the lake during the period of exposure of $2.1 \times 10^{-10} \mu\text{g}/\ell$. Using a value of 100 hrs/year for the exposure rate (Soldat et al. 1974), and assuming a 50-year exposure time, the doses from swimming were calculated and are listed, along with the dose rate factors, in Table A.12.

TABLE A.12. Fifty Years of External Exposure to ^{239}Pu from Swimming

<u>Organ</u>	<u>Dose Rate Factor (rem/hour per $\mu\text{g}/\text{liter}$)</u>	<u>Dose (rem)</u>
Total Body	7.4E-09	7.8E-15
Skin	1.1E-07	1.2E-13

The doses received from boating are calculated using Equation 15 and by dividing the dose rate factors by 2, to correct for the geometry (Soldat et al. 1974). The same values assumed for U and T, in the dose calculations for swimming, are used (Soldat et al. 1974). The doses for boating are given in Table A.13.

TABLE A.13. Fifty Years of External Exposure to ^{239}Pu from Boating

<u>Organ</u>	<u>Dose (rem)</u>
Total Body	3.9E-15
Skin	5.8E-14

Shoreline Exposure

The doses received from exposure to shoreline deposits are calculated by:

$$D_i = U (DF)_i \bar{C}_s T \quad (A-16)$$

where

- U • exposure rate (hours/year)
- $(DF)_i$ • dose factor for organ i given in Table 4 (rem/hour per $\mu\text{g}/\text{m}^2$)
- $\bar{C}_s$ • average radionuclide surface concentration in the top 2.5 cm of shoreline sediments ($\mu\text{g}/\text{m}^2$)
- T • period of exposure (years)

A differential equation was set up to represent the buildup of radionuclide concentration in the shoreline sediments from the transport of the radionuclide particulates in the water adjacent to the sediment. The procedure is similar to that discussed in Regulatory Guide 1.109, Rev. 1, except that the water concentration is decreasing with time. Assuming a sediment surface density of $40 \text{ kg}/\text{m}^2$ (USNRC Guide 1.109 1977) and a water-to-sediment transfer coefficient of $7.2 \times 10^{-2} \text{ l}/\text{kg}$ per hour (USNRC Guide 1.109 1977), ignoring radiological decay, and using Equation A-11 to predict the water concentration, the following equation for the sediment surface concentration is obtained:

$$C_s = C_s^0 + (1.29 \times 10^{-7}) S_w [1 - \exp(-\lambda t)]/\lambda \quad (A-17)$$

where

- C_s^0 • the surface concentration of radioactive material initially deposited from the cloud onto the sediments ($\mu\text{g}/\text{m}^2$); same as W defined in Equation A-4

- 1.29×10^{-7} • the extrapolated water concentration at $t = 0$, $1.86 \times 10^{-9} \mu\text{g}/\ell$, times the water-to-sediment transfer coefficient, $7.2 \times 10^{-2} \ell/\text{kg} \cdot \text{hr}$, times the sediment surface density, $40 \text{ kg}/\text{m}^2$, times $24 \text{ hr}/\text{day}$ ($\mu\text{g}/\text{m}^2$ per day)
- S_w • shore-width factor that describes the geometry of the exposure (dimensionless)
- λ • removal constant for plutonium in a lake, defined in Equation A-11 (days^{-1})
- t • time since deposition onto the lake surface (days)

A shore-width factor of 0.3 is assumed for a lake shore (USNRC Guide 1.109 1977). Equation A-17 was integrated over a 50-year period and divided by 50 years to obtain an average radionuclide surface concentration in the sediments of $7.3 \times 10^{-5} \mu\text{g}/\text{m}^2$, during the period of exposure. Using an exposure rate of 50 hrs/year (Soldat et al. 1974), doses were calculated for a 50-year period and are listed in Table A.14.

TABLE A.14. Fifty Years of Shoreline Exposure to ^{239}Pu

<u>Organ</u>	<u>Dose (rem)</u>
Total Body	8.9E-11
Skin	8.8E-10

SUMMARY OF FIFTY-YEAR COMMITTED DOSE EQUIVALENTS
FROM ALL MODES OF EXPOSURE

A summary of the 50-year committed dose equivalents from all modes of exposure is given in Table A.15. The inhalation exposure mode contributes greater than 98% of the total dose to the total body, lungs, and bone. All other exposure modes contribute less than 1%, except water consumption, which contributes 1.6% to the total bone dose. The assumptions used for this exposure mode were very conservative, and, in reality, this mode of exposure is not expected to contribute significantly. Therefore, for accidental airborne releases of plutonium, only inhalation from initial cloud passage and resuspension requires consideration. Resuspension could contribute as much as 39% of the total dose to the total body, lungs, and bone from inhalation, if all plutonium particles released and available for resuspension are in the respirable range.

The dose equivalents to the skin and GI-LLI are insignificant when compared to the lungs and bone dose, and can be ignored. Although ^{239}Pu was the only isotope considered, these conclusions apply to most isotopes of plutonium and a typical mixture of plutonium isotopes.

TABLE A.15. Fifty-Year Committed Dose Equivalents from an Acute Release of ^{239}Pu to the Atmosphere

Exposure Mode	Fifty-Year Committed Dose Equivalents (rem) to the Following Organs:				
	Total Body	Skin	Lungs	Bone	GI-LLI
Initial Inhalation	5.6E-07 (61)		9.9E-07 (61)	1.2E-05 (60)	7.3E-10 (1.3)
Inhalation from Resuspension	3.5E-07 (38)		6.2E-07 (39)	7.5E-06 (38)	4.5E-10 (0.8)
Cloud Submersion	1.2E-15 (<0.1)	1.6E-14 (<0.1)	1.2E-15 (<0.1)	1.2E-15 (<0.1)	1.2E-15 (<0.1)
Ground Exposure	1.5E-11 (<0.1)	1.5E-10 (15)	1.5E-11 (<0.1)	1.5E-11 (<0.1)	1.5E-11 (<0.1)
Crop Consumption	2.8E-09 (0.3)			1.1E-07 (0.6)	1.1E-08 (20)
Milk Consumption	1.5E-14 (<0.1)			5.7E-13 (<0.1)	6.0E-14 (<0.1)
Beef Consumption	1.0E-09 (0.1)			3.8E-08 (0.2)	4.1E-09 (7.3)
Water Consumption	8.3E-09 (0.9)			3.2E-07 (1.6)	3.6E-08 (64)
Fish Consumption	8.3E-10 (<0.1)			3.2E-08 (0.2)	3.6E-09 (6.4)
Swimming	7.8E-15 (<0.1)	1.2E-13 (<0.1)	7.8E-15 (<0.1)	7.8E-15 (<0.1)	7.8E-15 (<0.1)
Boating	3.9E-15 (<0.1)	5.8E-14 (<0.1)	3.9E-15 (<0.1)	3.9E-15 (<0.1)	3.9E-15 (<0.1)
Shoreline Exposure	8.9E-11 (<0.1)	8.8E-10 (85)	8.9E-11 (<0.1)	8.9E-11 (<0.1)	8.9E-11 (0.2)
Totals	9.2E-07 (100)	1.0E-09 (100)	1.6E-06 (100)	2.0E-05 (100)	5.6E-08 (100)

() - percent contribution of pathway to total organ dose.

APPENDIX B

DOSE FACTORS FOR INHALATION, AND DOSE
CALCULATION RESULTS FOR CLASS W PLUTONIUM

APPENDIX B

TABLE B.1. Fifty-Year Committed Dose Equivalent Factors from Acute Inhalation for Class W Material^(a)

Isotope	(rem per μg inhaled)				
	Total Body	Kidneys	Liver	Bone	Lungs
^{238}Pu	$1.2\text{E}+3^{(b)}$	$4.8\text{E}+3$	$1.5\text{E}+4$	$2.4\text{E}+4$	$9.2\text{E}+2$
^{239}Pu	$4.6\text{E}+0$	$1.9\text{E}+1$	$5.9\text{E}+1$	$9.7\text{E}+1$	$3.0\text{E}+0$
^{240}Pu	$1.7\text{E}+1$	$6.9\text{E}+1$	$2.2\text{E}+2$	$3.6\text{E}+2$	$1.1\text{E}+1$
^{241}Pu	$1.3\text{E}+2$	$6.1\text{E}+2$	$1.8\text{E}+3$	$3.2\text{E}+3$	$1.8\text{E}+0$
^{242}Pu	$2.8\text{E}-1$	$1.1\text{E}+0$	$3.6\text{E}+0$	$5.7\text{E}+0$	$1.8\text{E}-1$
^{241}Am	$2.0\text{E}+2$	$1.5\text{E}+3$	$3.2\text{E}+3$	$5.2\text{E}+3$	$1.7\text{E}+2$

(a) Committed dose equivalent factors calculated using DACRIN for $1\text{ }\mu\text{m}$ AMAD size particles. Organ masses are those reported in ICRP-23.

(b) $1.2\text{E}+3$ is identical to 1.2×10^3 .

TABLE B.2. Fifty-Year Committed Dose Equivalent Factors from Acute Inhalation for Class Y Material

Isotope	(rem per μg inhaled)				
	Total Body	Kidneys	Liver	Bone	Lungs
^{238}Pu	$4.3\text{E}+2$	$1.8\text{E}+3$	$5.8\text{E}+3$	$8.9\text{E}+3$	$9.0\text{E}+3$
^{239}Pu	$1.7\text{E}+0$	$7.1\text{E}+0$	$2.3\text{E}+1$	$3.7\text{E}+1$	$3.0\text{E}+1$
^{240}Pu	$6.3\text{E}+0$	$2.6\text{E}+1$	$8.3\text{E}+1$	$1.3\text{E}+2$	$1.1\text{E}+2$
^{241}Pu	$4.3\text{E}+1$	$2.0\text{E}+2$	$6.0\text{E}+2$	$1.1\text{E}+3$	$9.6\text{E}+1$
^{242}Pu	$1.0\text{E}-1$	$4.3\text{E}-1$	$1.4\text{E}+0$	$2.2\text{E}+0$	$1.8\text{E}+0$
^{241}Am	$7.8\text{E}+1$	$5.6\text{E}+2$	$1.2\text{E}+3$	$1.9\text{E}+3$	$1.7\text{E}+3$

TABLE B.3. Fifty-Year Committed Dose Equivalent Factors from
One-Year Chronic Inhalation for Class W Material

<u>Isotope</u>	<u>(rem per μg inhaled in first year)</u>				
	<u>Total Body</u>	<u>Kidneys</u>	<u>Liver</u>	<u>Bone</u>	<u>Lungs</u>
^{238}Pu	1.2E+3	4.8E+3	1.5E+4	2.4E+4	9.2E+2
^{239}Pu	4.5E+0	1.9E+1	5.8E+1	9.7E+1	3.0E+0
^{240}Pu	1.7E+1	6.8E+1	2.2E+2	3.6E+2	1.1E+1
^{241}Pu	1.3E+2	6.1E+2	1.8E+3	3.2E+3	1.8E+0
^{242}Pu	2.8E-1	1.1E+0	3.6E+0	5.7E+0	1.8E-1
^{241}Am	2.0E+2	1.5E+3	3.2E+3	5.1E+3	1.7E+2

TABLE B.4. Fifty-Year Committed Dose Equivalent Factors from
One-Year Chronic Inhalation for Class Y Material

<u>Isotope</u>	<u>(rem per μg inhaled in first year)</u>				
	<u>Total Body</u>	<u>Kidneys</u>	<u>Liver</u>	<u>Bone</u>	<u>Lungs</u>
^{238}Pu	4.3E+2	1.8E+3	5.7E+3	8.8E+3	9.0E+3
^{239}Pu	1.7E+0	7.0E+0	2.2E+1	3.6E+1	3.0E+1
^{240}Pu	6.2E+0	2.6E+1	8.2E+1	1.3E+2	1.1E+2
^{241}Pu	4.3E+1	2.0E+2	6.0E+2	1.0E+3	9.6E+1
^{242}Pu	1.0E-1	4.3E-1	1.4E+0	2.1E+0	1.8E+0
^{241}Am	7.7E+1	5.6E+2	1.2E+3	1.9E+3	1.7E+3

TABLE B.5. Fifty-Year Committed Dose Equivalents from Inhalation Following an Earthquake (Class W)

Organ of Reference	Committed Dose Equivalents for:							
	Population (person-rem)				Nearest Residence ^(a) (rem)			
	Case I ^(b)	Case II	Case III	Case IV	Case I	Case II	Case III	Case IV
Total Body	4.8E+4	5.8E+5	5.5E+4	6.0E+5	7.1E-1	4.0E+1	7.8E-1	4.0E+1
Kidneys	2.1E+5	2.5E+6	2.4E+5	2.6E+6	3.1E+0	1.7E+2	3.4E+0	1.7E+2
Liver	6.5E+5	7.8E+6	7.4E+5	8.1E+6	9.6E+0	5.3E+2	1.1E+1	5.4E+2
Bone	1.1E+6	1.3E+7	1.2E+6	1.4E+7	1.6E+1	8.9E+2	1.8E+1	9.0E+2
Lungs	2.0E+4	2.3E+5	2.2E+4	2.5E+5	2.9E-1	1.6E+1	3.2E-1	1.6E+1

(a) Located 2000 meters SSE of the NMDF.

(b) Case I - most likely release with most likely dispersion; Case II - most likely release with conservative dispersion; Case III - conservative release with most likely dispersion; Case IV - conservative release with conservative dispersion.

TABLE B.6. Fifty-Year Committed Dose Equivalents from Inhalation Following a 110-mph Tornado (Class W)

Organ of Reference	Committed Dose Equivalents for:							
	Population (person-rem)				Nearest Residence ^(a) (rem)			
	Case I	Case II	Case III	Case IV	Case I	Case II	Case III	Case IV
Total Body	-(b)	-	9.7E+2	9.2E+3	-	-	1.7E-3	1.7E-2
Kidneys	-	-	4.2E+3	4.0E+4	-	-	7.2E-3	7.2E-2
Liver	-	-	1.3E+4	1.2E+5	-	-	2.2E-2	2.2E-1
Bone	-	-	2.2E+4	2.1E+5	-	-	3.7E-2	3.7E-1
Lungs	-	-	4.0E+2	3.8E+3	-	-	6.8E-4	6.8E-3

(a) Located 16,000 to 32,000 meters from the NMDF in the direction the tornado travels.

(b) No doses calculated for Cases I and II, as most likely release is insignificant ($<10^{-7}$ g Pu).

TABLE B.7. Fifty-Year Committed Dose Equivalents from Inhalation
Following a 130-mph Tornado (Class W)

Organ of Reference	Committed Dose Equivalents for:							
	Population (person-rem)				Nearest Residence ^(a) (rem)			
	Case I	Case II	Case III	Case IV	Case I	Case II	Case III	Case IV
Total Body	1.8E+3	1.8E+4	4.6E+3	4.6E+4	2.8E-3	2.8E-2	7.2E-3	7.2E-2
Kidneys	7.7E+3	7.7E+4	2.0E+4	2.0E+5	1.2E-2	1.2E-1	3.1E-2	3.1E-1
Liver	2.4E+4	2.8E+5	6.3E+4	6.2E+5	3.7E-2	3.7E-1	9.7E-2	9.7E-1
Bone	4.0E+4	3.9E+5	1.0E+5	1.0E+6	6.2E-2	6.2E-1	1.6E-1	1.6E+0
Lungs	7.2E+2	7.2E+3	1.9E+3	1.9E+4	1.1E-3	1.1E-2	2.9E-3	2.9E-2

(a) Located 16,000 to 32,000 meters from the NMDF in the direction the tornado travels.

TABLE B.8. Fifty-Year Committed Dose Equivalents from Inhalation
Following a 150-mph Tornado (Class W)

Organ of Reference	Committed Dose Equivalents for:							
	Population (person-rem)				Nearest Residence ^(a) (rem)			
	Case I	Case II	Case III	Case IV	Case I	Case II	Case III	Case IV
Total Body	7.7E+4	7.7E+5	1.2E+5	1.2E+6	1.0E-1	1.0E+0	1.5E-1	1.5E+0
Kidneys	3.4E+5	3.4E+6	5.0E+5	5.0E+6	4.4E-1	4.4E+0	6.7E-1	6.7E+0
Liver	1.0E+6	1.0E+7	1.6E+6	1.6E+7	1.4E+0	1.4E+1	2.1E+0	2.1E+1
Bone	1.7E+6	1.7E+7	2.6E+6	2.6E+7	2.3E+0	2.3E+1	3.4E+0	3.4E+1
Lungs	3.1E+4	3.1E+5	4.7E+4	4.7E+5	4.2E-2	4.2E-1	6.3E-2	6.3E-1

(a) Located 16,000 to 32,000 meters from the NMDF in the direction the tornado travels.

TABLE B.9. Fifty-Year Committed Dose Equivalents from Inhalation
Following a 170-mph Tornado (Class W)

Organ of Reference	Committed Dose Equivalents for:							
	Population (person-rem)				Nearest Residence ^(a) (rem)			
	Case I	Case II	Case III	Case IV	Case I	Case II	Case III	Case IV
Total Body	3.4E+4	3.3E+5	3.8E+4	3.5E+5	2.4E-2	2.4E-1	6.3E-2	2.5E-1
Kidneys	1.5E+5	1.4E+6	1.7E+5	1.5E+6	1.0E-1	1.0E+0	2.8E-1	1.1E+0
Liver	4.5E+5	4.4E+6	5.2E+5	4.7E+6	3.2E-1	3.2E+0	8.5E-1	3.4E+0
Bone	7.6E+5	7.3E+6	8.6E+5	7.8E+6	5.3E-1	5.3E+0	1.4E+0	5.7E+0
Lungs	1.4E+4	1.3E+5	1.6E+4	1.4E+5	9.7E-3	9.7E-2	2.6E-2	1.0E-1

(a) Located 32,000 to 48,000 meters from the NMDF in the direction the tornado travels for Cases I and II; located 2000 meters SSE of the NMDF for Cases III and IV.

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